



# Saint Peter and Saint Paul Archipelago barcoded: Fish diversity in the remoteness and DNA barcodes reference library for metabarcoding monitoring

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## Abstract

In order to monitor the effects of anthropogenic pressures in ecosystems, molecular techniques can be used to characterize species composition. Among molecular markers capable of identifying species, the cytochrome c oxidase I (*COI*) is the most used. However, new possibilities of biodiversity profiling have become possible, in which molecular fragments of medium and short-length can now be analyzed in metabarcoding studies. Here, a survey of fishes from the Saint Peter and Saint Paul Archipelago was barcoded using the *COI* marker, which allowed the identification of 21 species. This paved the way to further investigate the fish biodiversity of the archipelago, transitioning from barcoding to metabarcoding analysis. As preparatory steps for future metabarcoding studies, the first extensive *COI* library of fishes listed for these islands was constructed and includes new data generated in this survey as well as previously available data, resulting in a final database with 9,183 sequences from 169 species and 63 families of fish. A new primer specifically designed for those fishes was tested *in silico* to amplify a region of 262 bp. The new approach should guarantee a reliable surveillance of the archipelago and can be used to generate policies that will enhance the archipelago's protection.

**Keywords:** Biodiversity, conservation, DNA barcoding, island, primer.

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## Introduction

Impacts of human-induced climate change, habitat fragmentation, and over-exploitation of natural resources have depleted global biodiversity, in particular in the marine environment (Díaz *et al.*, 2006; Butchart *et al.*, 2010; Pinsky *et al.*, 2019). Conservation efforts based on robust biomonitoring programs are necessary to identify and mitigate ecological issues (Stat *et al.*, 2017; Berry *et al.*, 2019); therefore, preservation of diversity depends on species classification accuracy (Thomsen and Willerslev, 2015; Lin *et al.*, 2020). The species composition and distribution can act as an environmental indicator of human activity (DiBattista *et al.*, 2020).

Species are rapidly going extinct as a result of these anthropogenic activities, and it is impossible to describe the true magnitude of the loss with traditional monitoring approaches (Blaxter, 2003; Hubert and Hanner, 2015; Zamani *et al.*, 2022); hence, molecular techniques have been developed to characterize species diversity quickly and reliably (Krishnamurthy and Francis, 2012; Elbrecht *et al.*, 2019). Since the early 1990s, the mitochondrial gene cytochrome c oxidase I (*COI*) has been used as a tool to describe biodiversity (Folmer *et al.*, 1994). The field was revolutionized when Hebert *et al.* (2003) proposed that the “Folmer region” of *COI* could be used to identify and discriminate species as a molecular barcode (Hebert *et al.*, 2003; Hebert and Gregory, 2005). This

658 bp genetic fragment can be easily obtained from animal tissues, and once sequenced, it provides greater than 97% confidence for differentiating species by the divergence in their *COI* sequences (Hajibabaei *et al.*, 2005; Meusnier *et al.*, 2008). After nearly two decades, the method has been widely accepted as the standard procedure for surveying biodiversity (Hubert and Hanner, 2015; Delrieu-Trottin *et al.*, 2019).

However, for reliable species descriptions, DNA barcoding is not sufficient, and additional taxonomic approaches are necessary (Zamani *et al.*, 2022). In fact, one of the major limitations of the technique is the need to have a reference library of DNA sequences that is built from morphologically identified species (Christoffer and Endre, 2005). This need for reference specimens imposes further difficulties because some species are rare or difficult to sample (Ogwang *et al.*, 2020). This is exacerbated when sampling specimens from remote marine protected areas, which is the case of the Saint Peter and Saint Paul fishes.

The Saint Peter and Saint Paul Archipelago (SPSPA) is a small group of plutonic rocks uplifted from the upper mantle of the earth, located in the central equatorial Atlantic Ocean between Brazil and the African continent (Figure 1; Campos *et al.*, 2005). The archipelago is a rare non-volcanic formation resulting from the Mid-Atlantic Ridge's exhumed mantle rocks (Mohriak, 2020). As a consequence of unique geological traits, along with latitude, weather, marine currents, and biogeographic features, the biodiversity of the SPSPA is commensurately singular. The archipelago is an important migratory, breeding, and feeding site for fishes (Mendonça *et al.*, 2018). Also, its isolation spawned the evolution of a unique biodiversity of fishes, with a variety of color morphs and genetically divergent lineages (Pinheiro *et al.*, 2020).

Due to this, the fish biodiversity of SPSPA has been intensively studied since the time when Lubbock and Edwards (1981) listed 50 fish species. The authors surprisingly considered the species diversity the lowest of any tropical island studied to date. Following the inauguration of the archipelago's first scientific station in 1998, SCUBA (Self-Contained Underwater Breathing Apparatus) expeditions were made possible (Viana *et al.*, 2009), and gradually the number of identified species increased from 75 (Feitoza *et al.*, 2003) to 116 (Vaske Jr *et al.*, 2005); and, most recently, to 225 species (Pinheiro *et al.*, 2020). Contrary to Lubbock and Edwards's (1981) considerations, the last survey pointed to the archipelago as having the third-highest level of endemism in the Atlantic (10 endemic species; Pinheiro *et al.*, 2020).

Among the 225 listed species, 112 are pelagic, 86 are shallow, and 27 are deep reef shore fishes. The inventory classification consists of 202 *Teleostei* distributed in 16 orders and 23 *Elasmobranchii* in six orders (Pinheiro *et al.*, 2020). There are at least 29 endangered species inhabiting the SPSPA waters according to the IUCN and Brazilian Red lists (Pinheiro *et al.*, 2020). Naturally, the research collection of these species is limited by strict policies meant to protect the species; therefore, other sampling strategies are required to survey the genetic diversity of these fishes.

Fortunately, advanced molecular technologies including new DNA extraction protocols (Taberlet *et al.*, 2018) and high-throughput sequencing have made it possible to sequence DNA molecules expelled by organisms into the environment through urine, reproductive and digestive materials, hair, skin, tissues, and decaying carcasses (Thomsen and Willerslev, 2015; Wangenstein *et al.*, 2018). The genetic assessment of multiple taxa from bulk environmental samples is denominated "DNA metabarcoding" (Taberlet *et al.*, 2018). And now ecologists have the necessary tools to analyze the species composition of environmental samples (Taberlet *et al.*, 2012; Creer *et al.*, 2016).

However, the genetic material extracted from ecosystems is highly fragmented (Deagle *et al.*, 2006); to this extent, it may be challenging in practice to retrieve full-length *COI* barcode sequences (658 bp) from environmental samples (Meusnier *et al.*, 2008). Metabarcoding analyses are contingent on targeting shorter DNA regions (<350 bp) than the traditionally defined barcoding regions (Yu *et al.*, 2012; Clarke *et al.*, 2014; Thomsen and Willerslev, 2015). In this context, alternative target metabarcoding markers (metabarcodes) have been developed to obtain biodiversity information in short-length (150-250 bp) PCR products (Taberlet *et al.*, 2018).

One metabarcode option is the much shorter "*mini-COI*" barcode, a 130 bp fragment of the full ca. 658 bp *COI* barcode; Meusnier *et al.* (2008) developed a universal primer set for the amplification of *mini-COI* that provides sufficient taxonomic resolution to differentiate between 1,587 metazoan species. Their results suggested that the region provides efficient taxonomic identification success, and its use was proposed to analyze environmental mixtures (Meusnier *et al.*, 2008); however, the mini-barcode is not variable enough to differentiate between fish species. (Sultana *et al.*, 2018).

Medium-sized (~320 bp) barcodes that are capable of differentiating between fish species have been developed and used in marine metabarcoding studies, and to identify fish

species in processed forms. (Shokralla *et al.*, 2015; Collins *et al.*, 2019). Despite the successful use of these markers in fish biodiversity assessment via metabarcoding (Singer *et al.*, 2019; McClenaghan *et al.*, 2020; Russo *et al.*, 2021), biodiversity assessments could be maximized by the use of regional-specific reference barcode libraries (Lin *et al.*, 2020).

In order to better characterize the baselines of Saint Peter and Saint Paul's fish biodiversity, we collected fishes and generated full barcode sequences. For future metabarcoding monitoring of this region, we constructed a *COI* reference library of listed fish species from SPSPA, adding our sequences to those previously published. Using this library, we identified a primer pair that would be appropriate to meta-amplify fragmented *COI* barcodes of SPSPA fishes.

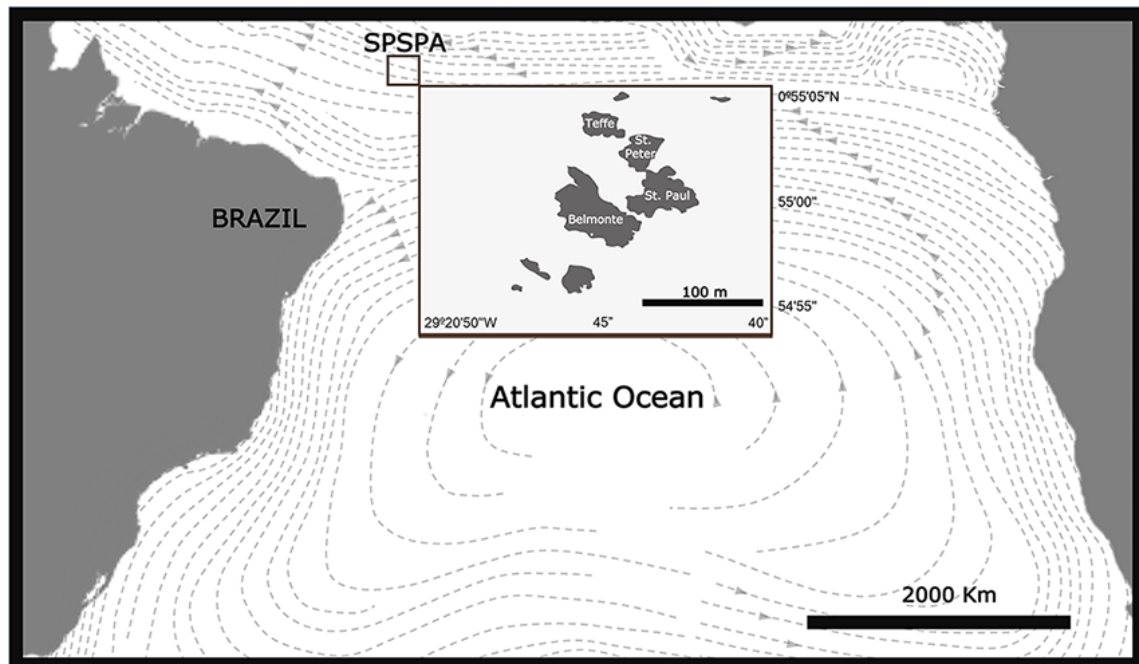
## Material and Methods

Five field expeditions were conducted between 2005 and 2015 in surroundings of the Saint Peter and Saint Paul Archipelago (000° 55' N and 029° 21' W; Fig 1). Fishes were opportunistically sampled from authorized longline catches targeting wahoos and tunas (license number SISBIO/ICMBio 014/2005). Muscle fragments were labeled (numbered) and preserved in 96% ethanol at -20°C until their extraction. Sampled fishes were identified following on-site taxonomic guides (Menezes *et al.*, 2003).

DNA was extracted using the PureLink™ Genomic DNA Mini Kit (Thermo Fisher Scientific, Massachusetts, United States) following the manufacturer's protocol. The forward FishF2 (5' TCG ACT AAT CAT AAA GAT ATC GGC AC 3') and reverse FishR2 (5' ACT TCA GGG TGA CCG AAG AAT CAG AA 3') primer pair (Ward *et al.*, 2005) was used to amplify the cytochrome c oxidase I (*COI*) gene by polymerase chain reaction (PCR). Each PCR reaction was conducted in a total volume of 25 µL, consisting of 0.2 mM of dNTPs, buffer 1× 1.5 mM of MgCl<sub>2</sub>, 0.2 µM of each primer, 1 U of AmpliTaq Gold DNA polymerase (Thermo Fisher Scientific, Massachusetts, United States), 50-100 ng of template DNA quantified using NanoDrop 2000 (Thermo Scientific, Massachusetts, United States), and ultrapure water to a final volume.

The thermal cycling condition began with an initial denaturing at 94 °C for 5 minutes, followed by 35 repeated cycles of denaturing (94 °C for 0.5 minutes), annealing (50 °C for 0.5 min) and extension (72 °C for 1 min), then concluded with a final extension at 72 °C for 7 min. The size and specificity of amplification products were confirmed in 1% agarose gel stained with GelRed (Biotium, Fremont, California). The successful products were purified using exonuclease I and Shrimp Alkaline Phosphatase enzymes (Amersham Biosciences, Little Chalfont, UK). Finally, they were sequenced by the Sanger method on an ABI3730XL DNA sequencer (Thermo Fischer Scientific, Massachusetts, United States) in Macrogen Inc. (Seoul, South Korea), with the forward primer used for amplification.

The sequences were quality checked, and low-quality regions were removed by using the software Geneious Pro version 9 (Biomatters Ltd, Auckland, New Zealand). The removal of chimeric sequences and alignment using ClustalW (Edgar, 2004) were also performed in Geneious software.



**Figure 1** – Saint Peter and Saint Paul Archipelago (SPSPA) in a map showing its geographical location (white square) in the Mid-Atlantic Ridge.

Species were identified using the “Identification Engine” of the Barcode of Life Data System (BOLD) by selecting ‘Animal Identification (*COI*)’ and the ‘Species Level Barcode Records’ (accessed 10 June 2021).

The taxonomic identity of each sequence was assigned to the deposited sequence with the highest similarity score. Also, a neighbor-joining tree was constructed based on the aligned dataset using the Kimura 2-Parameter (K2P) model (Kimura, 1980) with 1,000 bootstrap replicates and pairwise deletion in Geneious to cluster candidate species based on their sequences’ similarities.

As the sequenced samples represent only a small fraction of listed Saint Peter and Saint Paul fishes, the names listed in the Pinheiro *et al.* (2020) study were used to perform a mining within BOLD. Globally distributed *COI* sequences from the listed species were added to a new SPSPA *COI* reference database for further reference database expansion. The scientific fish names from the Pinheiro *et al.* (2020) checklist were searched on the BOLD “Taxonomy Browser” (accessed 15 June 2021). All available *COI* sequences were subsequently deposited in the SPSPA *COI* database. A detailed list of specimens and their BOLD IDs is given in Table 1. Then overall mean distance by (K2P) was computed using MEGA X software (Kumar *et al.*, 2018).

A new primer pair exclusively curated (based on the physical properties, penalties of hairpin formations and primer-dimers of the SPSPA sequences database) was designed in the Primer3 plugin featured in Geneious Software (Untergasser *et al.*, 2012). The performance of the newly designed primers was tested *in silico* against Saint Peter and Saint Paul fish sequences repository using the “Add Primers to Sequence” Geneious tool. Among the candidates’ primer pairs, the selected was the one with the highest “Pairwise Identity” targeting all the sequences of the database and with a product size appropriate for future metabarcoding studies.

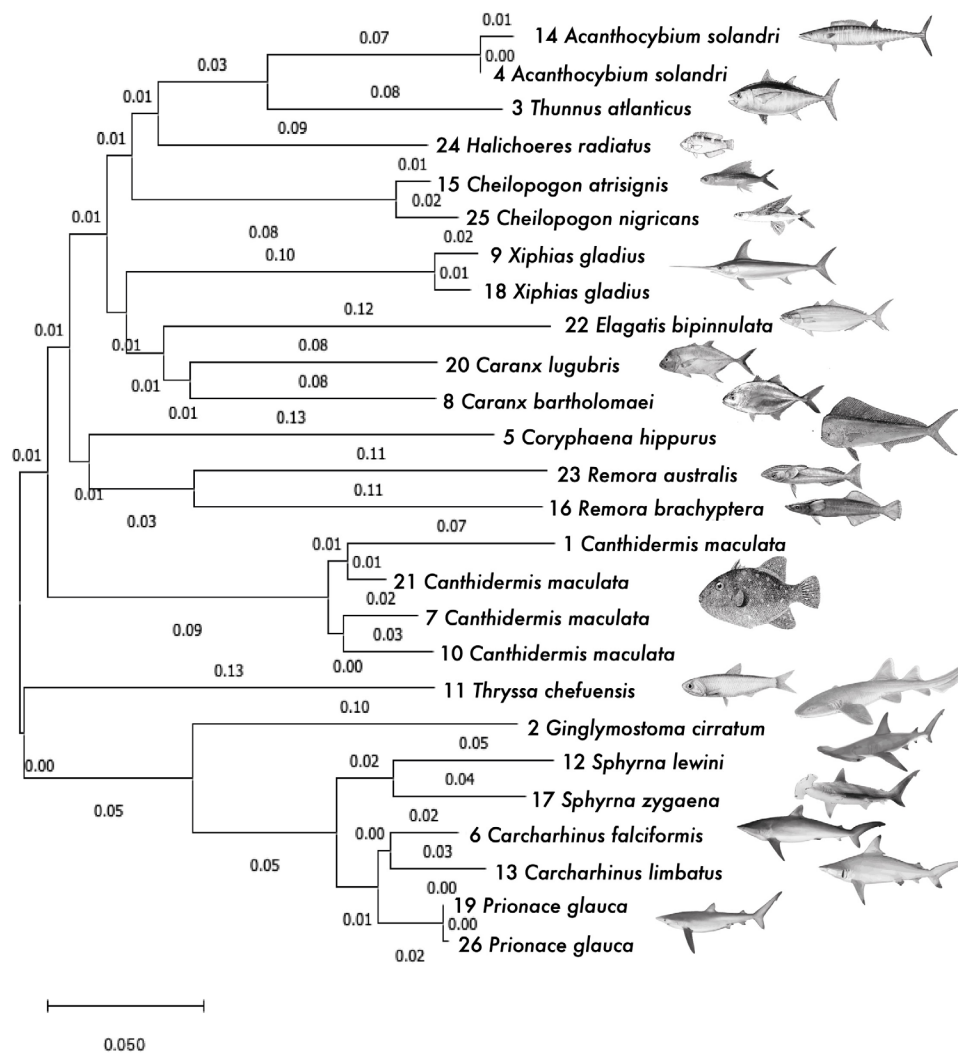
## Results

The first attempt to barcode fishes from SPSPA waters resulted in 28 captured samples, following strict collection rules as a maximum of six fishes could be caught per expedition. The extraction, amplification, and sequencing methods were successful for 26 out of 28 samples (representing 11.55% of the known SPSPA fishes). Among the 26 samples, the *COI* Barcode could be identified on BOLD with a high percentage of similarity (98.04%-100%; Table 1), revealing 21 species that are found in 11 families of fishes (graphically represented in Figure 2). The sequences were deposited in GenBank under accession numbers OK030800-OK030825. The neighbor-joining tree revealed expected patterns - closely related species in the same genus clustered together while dissimilar species appeared on different branches. Among the 21 species of fish, *Canthidermis maculata* was the most abundant (three of the samples), followed by *Acanthocybium solandri*, *Xiphias gladius*, and *Prionace glauca* (two samples each). Table 1 also indicates the closest match and where the matching sequence was collected.

Of the 21 newly identified fishes, four were not listed in Pinheiro *et al.* (2020). Those records were then added to a new database. While 165 of the 225 species listed in Pinheiro *et al.* (2020) have *COI* sequences deposited in the BOLD database from a fish caught somewhere else, these were also used to complete the database. Therefore, the new Saint Peter and Saint Paul sequence database has 9,183 sequences from 169 species and 63 families of fish. The full reference library can be found at <https://github.com/marcelomcruz4/SPSPAfishes>. From this species list, 84 are pelagic, 83 are reef-associated or deep-water residents, and two are endemic (*Emblemariopsis signifier* and *Stegastes sanctipauli*). The overall mean distance among all sequences was 0.4. Coherently, the AT content was higher than the GC content in the barcoded collected fishes (56.30%), and among the constructed database (AT content: 55.70%).

**Table 1** – Sample identification, identified species, their family, similarity to the BOLD database candidate species (%), location of the BOLD matching sequence, deposited sequence (GenBank accession number), and size of the fragment. Identified fishes of Saint Peter and Saint Paul Archipelago.

Sample identification	Candidate species name (BOLD accession number)	Family	Identity (%)	Sampling location of the matching sequence	Deposited sequence (GenBank accession number)	Size of the fragment
1	<i>Canthidermis maculata</i> (LIDB123-11)	Balistidae	98.04	Belize	OK030800	540 bp
2	<i>Ginglymostoma cirratum</i> (PHANT057-08)	Ginglymostomatidae	100	United States	OK030801	515 bp
3	<i>Thunnus atlanticus</i> (MFLE487-14)	Scombridae	99.84	Honduras	OK030802	625 bp
4	<i>Acanthocybium solandri</i> (MXIII111-07)	Scombridae	100	Mexico	OK030803	660 bp
5	<i>Coryphaena hippurus</i> (MXII093-07)	Coryphaenidae	100	Mexico	OK030804	606 bp
6	<i>Carcharhinus falciformis</i> (GBMND3415-21)	Carcharhinidae	100	Brazil	OK030805	629 bp
7	<i>Canthidermis maculata</i> (GBMND69325-21)	Balistidae	100	United States	OK030806	628 bp
8	<i>Caranx bartholomaei</i> (BZLWD025-07)	Carangidae	100	Belize	OK030810	625 bp
9	<i>Xiphias gladius</i> (ANGBF8490-12)	Xiphiidae	100	Not informed	OK030811	642 bp
10	<i>Canthidermis maculata</i> (FOAH793-08)	Balistidae	100	Indonesia	OK030807	630 bp
11	<i>Thryssa chefuensis</i> (ANGBF1012-12)	Coryphaenidae	100	South Korea	OK030812	625 bp
12	<i>Sphyrna lewini</i> (GBMND3593-21)	Sphyrnidae	100	Brazil	OK030813	650 bp
13	<i>Carcharhinus limbatus</i> (ANGBF48501-19)	Carcharhinidae	100	Brazil	OK030814	651 bp
14	<i>Acanthocybium solandri</i> (MXIII111-07)	Scombridae	100	Mexico	OK030809	612 bp
15	<i>Cheilopogon atrisignis</i> (ANGBF32051-19)	Exocoetidae	100	Taiwan	OK030815	635 bp
16	<i>Remora brachyptera</i> (MFC279-08)	Echeneidae	100	Panama	OK030816	652 bp
17	<i>Sphyrna zygaena</i> (GBMNC59337-20)	Sphyrnidae	100	United States	OK030817	655 bp
18	<i>Xiphias gladius</i> (ANGBF36944-19)	Xiphiidae	100	Belgium	OK030818	620 bp
19	<i>Prionace glauca</i> (GBGC9258-09)	Carcharhinidae	100	Italy	OK030819	598 bp
20	<i>Caranx lugubris</i> (SABA054-11)	Carangidae	100	Saba (Caribbean Netherlands)	OK030820	633 bp
21	<i>Canthidermis maculata</i> (MEFM383-06)	Balistidae	100	Mexico	OK030808	522 bp
22	<i>Elagatis bipinnulata</i> (MXIII391-09)	Xiphiidae	100	Mexico	OK030821	620 bp
23	<i>Remora australis</i> (TZSAL697-13)	Echeneidae	100	South Africa	OK030822	627 bp
24	<i>Halichoeres radiatus</i> (BZLWA436-06)	Labridae	100	Belize	OK030823	607 bp
25	<i>Cheilopogon nigricans</i> (ANGBF32059-19)	Exocoetidae	100	Atlantic Ocean	OK030824	648 bp
26	<i>Prionace glauca</i> (GBMND3512-21)	Carcharhinidae	100	Brazil	OK030825	598 bp



**Figure 2** – Neighbor-Joining Tree of the Saint Peter and Saint Paul Archipelago surveyed fish species labeled with substitutions *per site*.

From this database four new primer pairs were designed. The one with the highest “Pairwise Identity” rate (74.6%) and with the most adequate target size to be amplified is presented below:

SPSPAF-5’ GCTGGAGCATCTGTTGACCT3’,  
 SPSPAR-5’ CTCCTCCTGCAGGGTCAAAG3’.

This marker is suited to amplify a product size of 262 base pairs from the *COI* region and performs *in silico* capacity to amplify 73.6% of Saint Peter and Saint Paul’s sequences.

## Discussion

As expected from the revised theory of island biogeography for marine fishes, the SPSPA represents an important reservoir of biological diversity and a refuge for many endemic species that have diversified on these islands through time (Pinheiro *et al.*, 2017). Naturally, the isolation has played a crucial role in the genetic diversity and endemism of the smallest remote tropical island in the world (Luiz *et al.*, 2015). Aside from the distance, seamounts may also have played an essential function in the marine evolution of the SPSPA. The site (as a peak of the mountain range) acted as a “stepping stone” for fishes during successive periods of

sea-level changes (Ludt and Rocha, 2015; Dias *et al.*, 2019). Also, the topography and strategic location of the area make it an important feeding and reproduction ground for several migratory pelagic species, mostly with high commercial value (Viana *et al.*, 2015; Macena and Hazin, 2016; Pimentel *et al.*, 2020). Our results confirm the presence of some of these species, such as the blackfin tuna (*Thunnus atlanticus*), the wahoo (*Acanthocybium solandri*), the rainbow runner (*Elagatis bipinnulata*), the flying fishes (*Cheilopogon sp.*), the silky shark (*Carcharhinus falciformis*), and the blue shark (*Prionace glauca*). Due to the heterogeneity of migrants and residents of the region, molecular techniques are a useful tool to catalog and uncover the biodiversity of SPSPA.

## DNA Barcoding advantages and limitations

DNA barcoding technology provides an efficient molecular technique for species identification to elucidate global biodiversity (Hebert *et al.*, 2003; Krishnamurthy and Francis, 2012). The mitochondrial *COI* gene has been barcoding fish species with high efficiency (Ward *et al.*, 2009; Ward, 2012). The marine ichthyofauna was successfully characterized in Australia (Ward *et al.*, 2005), the Antarctic (Rock *et al.*, 2008; Mabrugaña *et al.*, 2016), Canada (Steinke



*et al.*, 2009), the Arctic (Mecklenburg *et al.*, 2010), Japan (Zhang and Hanner, 2011), India (Lakra *et al.*, 2011), Portugal (Costa *et al.*, 2012), Brazil (Ribeiro *et al.*, 2012), Germany (Kneibelsberger *et al.*, 2014), Taiwan (Bingpeng *et al.*, 2018), Indonesia (Limmon *et al.*, 2020), Pakistan (Ghouri *et al.*, 2020), and Bangladesh (Ahmed *et al.*, 2021).

In this unprecedented study, we successfully amplified the *COI* barcode sequences for Saint Peter and Saint Paul Archipelago fishes. The surveyed site is a remote and protected oceanic island (Soares and Lucas, 2018). This bio-blitz was the first effort to barcode representatives from the SPSPA. To this extent, the sample size is limited and for this reason, the samples of this study were opportunistically collected over different expeditions. Despite these sampling challenges, the *COI* barcoding genes of 26 fish specimens were successfully amplified and sequenced. The differentiation between species through individual *COI* barcodes validates the efficiency of *COI* barcodes for identifying marine fish species.

Even though a complete and robust identification process requires additional steps (such as diagnosable morphological characters and natural history/ecological studies), a DNA bio-scan is an extremely useful method for an initial sorting of new and known biodiversity (Zamani *et al.*, 2022). In this way, our survey opened up the possibility of uncovering the hidden biodiversity of the archipelago.

The feasibility of gathering new species' records for the region is sustained by the fact that the DNA barcoding revolution has hastened species discovery during the last 15 years (Cao *et al.*, 2016; DeSalle and Goldstein, 2019; Lopez-Vaamonde *et al.*, 2021). In turn, efforts to collect and barcode fish species from specific regions aided new fish records in other regions of the globe, such as Bangladesh, Sri Lanka, and the Bay of Bengal (Rathnasuriya *et al.*, 2019; Ahmed *et al.*, 2021; Sharifuzzaman *et al.*, 2021).

The methodology applied in this study revealed four new records to the Saint Peter and Saint Paul region: *Cheilopogon atrisignis*; *Cheilopogon nigricans*; *Remora australis*; and *Thryssa chefuensis*. Considering the natural history of these species, it is plausible that *Cheilopogon nigricans* and *Remora australis* inhabit the SPSPA, as their distribution is described to be in the neighboring waters of the Atlantic Ocean (Fishbase, 2021). In fact, *Remora australis* is already photo-documented at SPSPA waters (Hoffmann *et al.*, 2008; Wingert *et al.*, 2021); our survey corroborates the inclusion of this species in future checklists. Whereas *Cheilopogon atrisignis* and *Thryssa chefuensis* are related to the Indian and Pacific oceans respectively (Fishbase, 2021). Additional morphometric approaches must be applied in order to confirm the presence of these species in the SPSPA. In particular, the presence of *Thryssa chefuensis* must be investigated carefully, as there are no other members of the family Engraulidae reported to the archipelago (Pinheiro *et al.*, 2020) and DNA Barcoding has the capacity to detect alien species which invade different ecosystems (Nagarajan *et al.*, 2020).

The identification of two species from the genus *Cheilopogon* represents new records for the site and confirms the vast diversity of flying fishes in SPSPA. It is reported that at least five species of the genus inhabit the site (Pinheiro *et al.*, 2020); thus, the assignment of *Cheilopogon atrisignis* or

*Cheilopogon nigricans* could be a case of misidentification due to closely related species with low differentiation between *COI* sequences. This illustrates one of the limitations of *COI* barcoding methodologies; i.e., the *COI* gene is not sufficiently variable to distinguish between some closely related species (Moritz and Cicero, 2004). To overcome this limitation and confirm species identities, more data are needed from morphological characters and/or additional genetic markers.

### Future monitoring

DNA Barcoding technical limitations prompted additional research towards the technological transition to Metabarcoding. In other words, to transition from sampling individuals (DNA Barcoding) to whole communities (DNA metabarcoding; Porter and Hajibabaei, 2020). Metabarcoding is a capture-free and non-invasive tool useful for detecting rare, elusive, controlled, protected, or threatened species (Wilcox *et al.*, 2013; Schwentner *et al.*, 2021). With the impossibility to sample individuals from SPSPA, metabarcoding emerges as the solution to survey and monitor SPSPA fish diversity. This approach is becoming a well-established tool for monitoring fishes not only from water samples (Miya, 2022), but also from various types of samples such as air (Lynggaard *et al.*, 2022), sediment (Ip *et al.*, 2021), bottom trawl fishing vessels (Maiello *et al.*, 2022), and feces (Creer *et al.*, 2016; Jarman *et al.*, 2018).

Although the ability to identify and describe new species is limited using *COI* metabarcoding approaches, the amount of data generated is informative for biodiversity assessment (Taberlet *et al.*, 2018; Meierotto *et al.*, 2019). The collection impediment compromises the construction of a barcode reference database that optimally should be composed only of local specimens (Delrieu-Trottin *et al.*, 2019; Lin *et al.*, 2020). To overcome this limitation, we added to the SPSPA *COI* reference database *COI* sequences that were available on BOLD from the listed species but were collected elsewhere. As future metabarcoding steps, the constructed database, as well as the generated primer pair, must be tested in vitro, preferably with SPSPA samples and then directly with SPSPA environmental samples in a pilot study (Taberlet *et al.*, 2018). Another future perspective is the constant update of the SPSPA *COI* database, this would potentially increase the coverage of endemic species in the database, which currently only has two of the 11 listed endemic species. In this case, collected specimens in the archipelago vouchered in museums, especially the endemic ones, should be barcoded and added to the database (Ward *et al.*, 2009).

Rather than designing primers to target all fishes (Miya *et al.*, 2015; Collins *et al.*, 2019), here we designed primers capable of amplifying fishes found in the target geographical region. We did this by generating an alignment of *COI* sequences for fishes known to be present in the SPSPA. Fishes are the largest group of vertebrates, and the teleost and elasmobranch species are evolutionarily distant; therefore, their genetic fingerprints are dissimilar (Nelson *et al.*, 2016). We chose to focus on only the fishes of the SPSPA in order to increase the probability of amplification using environmental samples, thus ensuring accurate monitoring and protection.

A cocktail of primers targeting other metabarcodes such as the mitochondrial *12S* or *16S* rRNA genes (Epp *et al.*, 2012)

should be considered for a comprehensive metabarcoding study of the total fish biodiversity of the region (Collins *et al.*, 2019).

### Conservation Considerations

Due to the presence and connectivity of key species of corals, crustaceans, mollusks, fishes, marine birds, and cetaceans, SPSPA has been protected by the Ministry of the Environment of Brazil since 1986 (Francini-Filho *et al.*, 2018). Despite the protection, commercial fishing boats were allowed to operate in the SPSPA regularly (Viana *et al.*, 2015). In 2018, the environmental protection of the islands and surroundings was increased by the Brazilian government (Brasil, 2018). However, the vast majority of the new areas are classified as “Areas of Sustainable Use”, where “subsistence” fisheries are specifically allowed in the management plan. In practice, commercial fishing and industrial activities by regional fishing companies are also taking place in these areas, as reported by Giglio *et al.* (2018). Furthermore, the habitats considered more vulnerable to high environmental impact have not received integral protection. The areas of integral protection were designated in places where these activities are already unlikely or rare (Magris and Pressey, 2018).

Fine-scale geographical and temporal studies are crucial to define boundaries and to set goals for Marine Protected Areas. Therefore, systematic data collection along time and space is necessary to understand the protected ecosystem better and promote possible zoning changes. Considering the richness of SPSPA biodiversity and its lack of protection, advanced genetics tools for monitoring ecosystems are needed. In this case, DNA metabarcoding of marine water has the potential to effectively monitor and give solid periodic information to managers and policymakers (Gold *et al.*, 2021).

### Conclusion

The Saint Peter and Saint Paul Archipelago is a reservoir of biodiversity. The strategic location of the archipelago is an important feeding and reproductive ground for a variety of migratory fishes; likewise, it is a refuge to the third-highest fish endemism level in the Atlantic. The checklist of fishes that live in shallow and deep waters has already elucidated these outstanding patterns (Pinheiro *et al.*, 2020); as yet the genetic signatures of SPSPA fish species have remained unknown. Thereupon, this research endeavored to barcode surveyed species of the site and catalog all deposited sequences of listed fishes in the region. Challenges and limitations of the application of DNA Barcoding methodology on SPSPA fishes reveals there is yet more diversity to be discovered. Due to this, the protection of the archipelago should be enhanced and well monitored with more robust approaches. In this case, DNA metabarcoding is an emerging tool that could assist in safeguarding SPSPA fauna; therefore, the reference library and the primer pair specifically designed to study the fishes of these islands should be considered for future metabarcoding monitoring activities.

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### Conflict of Interest

The authors declare that there are no competing interests.

### Authors Contributions

MMC, LSH and TROF conceived and the study, LSH conducted the sampling, MMC conducted all other experiments, MMC analyzed the data, MMC wrote the manuscript, LSH and TROF reviewed the manuscript, all authors read and approved the final version.

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