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## **OPEN** Thermodynamic and hydrochemical controls on CH<sub>4</sub> in a coal seam gas and overlying alluvial aquifer: new insights into CH<sub>4</sub> origins

D. Des. R. Owen<sup>1</sup>, O. Shouakar-Stash<sup>2</sup>, U. Morgenstern<sup>3</sup> & R. Aravena<sup>4</sup>

Using a comprehensive data set (dissolved CH<sub>4</sub>,  $\delta^{13}$ C-CH<sub>4</sub>,  $\delta^{2}$ H-CH<sub>4</sub>,  $\delta^{13}$ C-DIC,  $\delta^{37}$ Cl,  $\delta^{2}$ H-H<sub>2</sub>O,  $\delta^{18}$ O-H<sub>2</sub>O, Na, K, Ca, Mg, HCO<sub>3</sub>, Cl, Br, SO<sub>4</sub>, NO<sub>3</sub> and DO), in combination with a novel application of isometric log ratios, this study describes hydrochemical and thermodynamic controls on dissolved CH4 from a coal seam gas reservoir and an alluvial aguifer in the Condamine catchment, eastern Surat/north-western Clarence-Moreton basins, Australia.  $\delta^{13}$ C-CH $_4$  data in the gas reservoir (-58% to -49%) and shallow coal measures underlying the alluvium (-80% to -65%) are distinct. CO<sub>2</sub> reduction is the dominant methanogenic pathway in all aquifers, and it is controlled by SO<sub>4</sub> concentrations and competition for reactants such as H<sub>2</sub>. At isolated, brackish sites in the shallow coal measures and alluvium, highly depleted  $\delta^2$ H-CH<sub>4</sub> (<310‰) indicate acetoclastic methanogenesis where SO<sub>4</sub> concentrations inhibit CO<sub>2</sub> reduction. Evidence of CH<sub>4</sub> migration from the deep gas reservoir (200–500 m) to the shallow coal measures (<200 m) or the alluvium was not observed. The study demonstrates the importance of understanding CH<sub>4</sub> at different depth profiles within and between aquifers. Further research, including culturing studies of microbial consortia, will improve our understanding of the occurrence of CH<sub>4</sub> within and between aquifers in these basins.

Methane ( $CH_4$ ) is a ubiquitous substance that occurs in adsorbed, dissolved and free gas forms in a range of aquifer, surface water, soil and atmospheric environments<sup>1,2</sup>. In surface waters and the shallow subsurface, CH<sub>4</sub> production and consumption are mediated by microbial processes that are stimulated by changes in redox conditions, availability of suitable fermentation substrates and electron acceptors<sup>2-4</sup>. The relative abundance of heavy and light stable isotopes of carbon  $({}^{12}C/{}^{13}C)$  and hydrogen  $({}^{2}H/{}^{1}H)$  that comprise CH<sub>4</sub> is influenced by certain processes including: the type of methanogenic (production) and consumption pathways; transport processes such as diffusion and desorption; competition for substrates with other reducing organisms (e.g. SO<sub>4</sub>-reducers); and thermodynamic conditions<sup>3,5-12</sup>. As a result, isotopic data of  $CH_4$  can be ambiguous, particularly when interpreting data from larger scales, where multiple sources of  $CH_4$  exist, where  $CH_4$  has potentially moved or where thermodynamic conditions change. There has been a range of research dedicated to understanding these complexities<sup>3-6,8,13-16</sup>. Yet, while information on the complex behaviour of CH<sub>4</sub> isotopes is readily available, it creates some uncertainty about the value of CH<sub>4</sub> isotopes as indicators of broader processes, such as fugitive emissions or aquifer connectivity (see Figure S1 for examples of  $\delta^{13}$ C-CH<sub>4</sub> values under a range of different pathways and conditions).

Recent rapid development of unconventional gas resources such as coal seam gas (coal bed methane) or shale gas (CH<sub>4</sub> sorbed under pressure in coal measures or shale deposits) has spurred interest in understanding the extent of unconventional gas resources and the potential for gas migration within and between aquifers. Dissolved gas can migrate within and between aquifers either via advection or by a diffusion process<sup>12,17</sup>. Some previous research in shale-gas-bearing basins has shown that  $CH_4$  can migrate with brines from underlying gas-bearing aquifers but, in the absence of hydrocarbon reservoirs, it can also be generated in situ<sup>18-20</sup>.

<sup>1</sup>School of Earth, Environmental and Biological Sciences, Queensland University of Technology, Brisbane, Queensland, 4000, Australia. <sup>2</sup>Isotope Tracer Technologies, Waterloo, ON N2V 1Z5, Canada. <sup>3</sup>GNS Science, Lower Hutt 5014, P.O. Box 30368, New Zealand. <sup>4</sup>Department of Earth and Environmental Sciences, University of Waterloo, Ontario N2L 3G1, Canada. Correspondence and requests for materials should be addressed to D.D.R.O. (email: dr. owen@qut.edu.au)

Approach	Application
$\delta^{13}\text{C-CH}_4$ and $\delta^2\text{H-CH}_4$	Collectively these two isotopes allow more informative assessments of potential CH <sub>4</sub> origins than $\delta^{13}$ C-CH <sub>4</sub> alone. E.g. Acetoclastic methanogenesis produces an enriched $\delta^{13}$ C-CH <sub>4</sub> value that is similar to both CSG CH <sub>4</sub> and other thermogenic CH <sub>4</sub> , but it can be distinguished from these by the highly depleted $\delta^{2}$ H-CH <sub>4</sub> <sup>3.26</sup> .
Isotope fractionation factors $\alpha_{\text{DIC-CH4}}$ and $\alpha_{\text{H2O-CH4}}$	The ratio of isotope values between source carbon/hydrogen and that of $CH_4$ provides insight into the production and consumption pathways <sup>3,5,26</sup> .
Thermodynamic data	Reducing organisms, such as $SO_4$ reducers, operate at thermodynamic thresholds that inhibit less- competitive methanogenic processes <sup>4–6</sup> . Comparisons of Gibbs free energy values for production and consumption pathways provide information on the extent to which certain reaction processes have proceeded in the subsurface.
Isometric log ratios (compositional data analysis)	Isometric log ratios allow robust, simultaneous analysis of the ratios of parts and subparts, even where the concentration of subparts are very small <sup>86</sup> . When applied to the reaction pathways for key production and consumption reactions, isometric log ratios satisfy the law of mass balance and allow any rate-limiting effects associated with the availability of reactants to be elucidated.

Table 1. The applications of the combined approaches used to understand the origins and controls on  $\rm CH_4$  in this study.

This paper investigates origins and transport of dissolved  $CH_4$  in a hydrogeological setting where a shallow coal seam gas (CSG) reservoir underlies an important alluvial water resource (Condamine River catchment, Surat and Clarence-Moreton basins, Australia). We use a novel approach, employing a combination of hydrochemical,  $CH_4$  and isotope data with isometric log ratios (reactants and products) and Gibbs free energy calculations from key biological reaction processes to describe the thermodynamic constraints on  $CH_4$  in the alluvium and underlying coal measures. This approach addresses the complexity of  $CH_4$  production and consumption in the subsurface and the range of associated isotopic responses. The approaches/techniques used to address this complexity are outlined in Table 1. Previous interpretations of  $\delta^{13}C$ - $CH_4$  data from a similar area<sup>21</sup> are compared to new data presented here and conclusions regarding  $CH_4$  migration are reviewed. A minimum suite of parameters required to assess  $CH_4$  within and between aquifers is proposed for future monitoring and data collection.

#### Hydrogeological Setting

The Condamine River alluvium (the Condamine alluvium) occurs in the Condamine River catchment, which is a large subcatchment (30,451 km<sup>2</sup>) in the headwaters of the Murray-Darling Basin in southeast Queensland, Australia. This study focusses on the upper, central Condamine alluvium (Fig. 1a). Hydrogeology and hydrochemistry of the alluvium are summarised in previous published work<sup>22–24</sup>. The alluvium overlies the Walloon Coal Measures (the coal measures) and, on the western alluvial flank, parts of the Kumbarilla Beds, which are Jurassic sedimentary features of the Surat and Clarence-Moreton basins (Fig. 1b). At the bedrock-alluvial interface an impervious clay layer (termed: the "transition layer") is proposed to limit interaction with the underlying coal measures; however, the spatial extent of this transition layer is not well known. In some cases, the alluvium has incised the coal measures by up to 130 m within a paleovalley (QWC<sup>25</sup>) (Fig. 1c). Weathered bedrock materials, including coal fragments, occur throughout the alluvium<sup>22</sup>. The alluvium is exploited for water reserves for use in large-scale irrigation, mainly cotton. Higher quality water is generally found in upstream areas of the study area, near Cecil Plains where hydraulic conductivity is higher<sup>22</sup>.

Coal seam gas (CSG) reserves in the underlying coal measures are a significant economic resource and gas production in the study area is focussed on areas in the south west at depths of  $\sim$  300–500 m, (see Fig. 1). Production of coal seam gas requires water to be extracted from the coal seam which has raised concerns about aquifer connectivity. The CH<sub>4</sub> gas in the gas reservoirs is typically biogenic, it tends to be concentrated at geological structures, and coal seams are discontinuous<sup>26-31</sup>. Using  $\delta^{13}$ C-CH<sub>4</sub> of free gas that was collected from degassing alluvial wells during pumping, a recent study concluded that CH4 leakage from the coal measures to the alluvium was occurring in some areas and this was used to infer aquifer connectivity<sup>21</sup>. However, this study did not collect any CH<sub>4</sub> data (free or dissolved) from the underlying coal measures for reference. A previous study that examined  $\delta^7$ Li within and between coal measure and basalt aquifers found very low concentrations of Li in the alluvium when compared to the coal measures<sup>32</sup>. Assuming conservative behaviour of the Li ion<sup>33–37</sup>, these results suggest large-scale solute transport between these aquifers is not occurring. While the gas reservoir that underlies the Condamine catchment is relatively shallow compared to some other areas in the Surat Basin, the commercially viable gas reservoir that directly underlies the Condamine catchment is relatively deep (typically 300-500 m) when compared to the maximum alluvium depth (130 m). In this paper we refer to two areas of the coal measures as follows: (1) the CSG or gas reservoir (200–500 m) of the coal measures where commercial gas reserves are found; and (2) the *shallow coal measures*: shallower zones of the coal measures (<200 m) that are up gradient of the gas reservoir, but which are underlying or adjacent to the alluvium (see Fig. 1b).

### **Results and Discussion**

**Redox and salinity conditions.** The deep gas reservoir is characterised by highly reduced SO<sub>4</sub> (typically less than detection limit (DL) (1 mg/L, or 0.02 meq/L)), and brackish water (Cl = 1000–4500 mg/L or 28–127 meq/L). In the shallower coal measures the SO<sub>4</sub> and Cl concentrations are more variable and show a positive relationship, ranging from <0.1 mg/L to 488 mg/L (10 meq/L) for SO<sub>4</sub> and 82 mg/L (2.3 meq/L) to 4680 mg/L (131.8 meq/L) for Cl. The majority of shallow coal measure samples have SO<sub>4</sub> concentrations below 50 mg/L (1 meq/L). A single coal sample from the shallow coal measures underlying the alluvium at Cecil Plains showed small amounts of pyrite; however, SO<sub>4</sub> concentrations at this site ranged from 8–12 mg/L (0.16–0.25 meq/L),



**Figure 1.** Hydrogeological setting and study area, showing: (a) location of the Condamine River catchment and Surat/Clarence-Moreton basins in eastern Australia; and (b) conceptual cross section of the Condamine alluvium and adjacent sedimentary features. For (a), the basin boundary is defined as the Kumbarilla Ridge<sup>28</sup> and references therein<sup>30</sup>. (a) was prepared using *ArcGIS v 10.1* (www.esri.com) and modified in *Adobe Illustrator CC* 2014. For (b), the land surface and alluvial depth profile is real, as taken from the Condamine Groundwater Visualisation System (GVS)<sup>38</sup>: the outcrops and depth extent of the olivine basalt and sedimentary bedrock features have not been mapped in detail and are represented as conceptualisations based on interpretations of existing literature<sup>22,25,39</sup> by the co-authors.

indicating that SO<sub>4</sub> is not completely reduced at this site. Salinity in the alluvium is also highly variable (Cl ranging from 35–8700 mg/L, or 1–245 meq/L) and also shows a positive relationship with SO<sub>4</sub>, which ranges from <1 mg/L to 988 mg/L (20 meq/L). Peak Cl and SO<sub>4</sub> concentrations are found in shallow (~20 m) wells. This is consistent with the findings of Owen and Cox<sup>23</sup>, which showed higher salinity is related to evapotranspiration processes.

NO<sub>3</sub> is low in all aquifers: typically <0.05 mg/L (0.001 meq/L), with 8 samples having NO<sub>3</sub> below DL (0.01 mg/L or 0.00016 meq/L) for the shallow coal measures, and ranging from 0.02 mg/L (3.23e-04 meq/L) to 2.3 mg/L (3.64e-02 meq/L), with three samples below DL (0.01 mg/L or 1.61e-04 meq/L), for the alluvium. With the exception of one shallow coal measure sample underlying a basalt outcrop (P19) and a shallow (~27 m) alluvial well (ID GM1076), all samples that contained CH<sub>4</sub> had NO<sub>3</sub> concentrations below 0.006 meq/L (0.37 mg/L) which is favourable for methanogenesis<sup>38</sup>. We found no NO<sub>2</sub> above DL (0.01 mg/L) in any aquifer: this shows that significant denitrification is not occurring in these aquifers.

Data on dissolved  $Fe^{2+}/Fe^{3+}$  and Mn species were not available, however, total dissolved concentrations of these ions were low in all aquifers in the study area. In the alluvium, Fe above DL (0.05 mg/L) was found at only 5 sites, (0.11–4.86 mg/L), while Mn concentrations were above DL (0.001 mg/L) at only 12 sites, the majority of which had Mn concentrations <0.01 mg/L. In the shallow coal measures, 8 samples had Fe above DL, with 6 of these being <0.8 mg/L, while only Mn concentrations were <0.09 mg/L at the majority (n = 12) of sites. These low values compare with production water which is highly reduced (SO<sub>4</sub> < 1 mg/L) (see Supplementary Information Table S1).

**Tritium.** Tritium analyses (DL = 0.02 TU) were performed at selected sites: shallow coal measures (n = 5) and alluvium (n = 9). Significant <sup>3</sup>H was observed for only one shallow coal measures well (P12, 0.95 TU): this well occurs on the basalt ranges under a thin basalt outcrop in a recharge area and contained no CH<sub>4</sub>. Only 2 shallow alluvial wells (~27 m (GM1076) and 41 m (GM1338)) were found to have detectable<sup>3</sup> (0.05 TU and 0.22 TU, respectively): these were located ~6 and 16 km from the river, respectively. Only one of these wells contained CH<sub>4</sub>.

(GM1076: 0.05 TU). Alluvial wells with no detectable tritium ranged from 18 m-89 m in depth, including 4 wells <40 m. The absence of tritium in the majority of shallow wells indicates limited to no modern recharge. While river recharge is considered important in this alluvial system<sup>24</sup>, we found no tritium in a 57 m deep alluvial well (GM0057) located approximately 1.4 km from the river.

**Stable isotopes of chlorine** ( $\delta^{37}$ Cl).  $\delta^{37}$ Cl was measured for all samples containing CH<sub>4</sub> within and between aquifers to provide an additional parameter for understanding possible CH<sub>4</sub> migration via a diffusion pathway.  $\delta^{37}$ Cl for these samples ranged from -2.52% to -0.1% in the CSG reservoir; -1.11% to 0.8% in the shallow coal measures; and -0.72% to 0.89% in the alluvium.

**Dissolved organic carbon.** Dissolved organic carbon (DOC) is typically low in coal measures and alluvium aquifers, ranging from 0.3-1.6 mg/L, and 0.1-3.9 mg/L, respectively. The shallowest well (GM1073: 18 m) had the second highest DOC in the alluvial data set 0.4 mg/L, although DOC of all alluvial and coal measure samples could be considered low compared to other studies<sup>19,39,40</sup>. We found no relationship between DOC and CH<sub>4</sub> concentrations in either aquifer.

The two alluvial wells that contained detectable tritium (GM1076 and GM1338) contained 0.2 and 0.4 mg/L of DOC, respectively. Two wells that occur at the alluvial-coal measure interface had DOC concentrations of 0.4 mg/L (GM1193:110 m) and 7 mg/L (IND2:85 m). The shallower sample contained no CH<sub>4</sub> and presented as an anomaly in the dataset.  $\delta^7$ Li for this sample<sup>32</sup> and lithological analysis confirm it is in the upper layer of the coal measures (~7 m below the alluvial basement). The other sample at the alluvial-coal measure transition zone (110 m) contained CH<sub>4</sub> and occurs in an area where the alluvium appears to have incised the coal measures; drill logs show it contains both sand and coal fragments.

**CH<sub>4</sub> within and between aquifers.** Dissolved CH<sub>4</sub> concentrations in the deep CSG reservoir ranged from 2000  $\mu$ g/L-25000  $\mu$ g/L (n = 21). In total, 7 of the 14 shallow coal measure wells contain dissolved CH<sub>4</sub> above DL (10 $\mu$ g/L): concentrations ranged from 95–18000  $\mu$ g/L. Of the 23 wells sampled in the alluvium, only 5 were found to contain CH<sub>4</sub>, with concentrations ranging from 10–535  $\mu$ g/L. All alluvial samples with dissolved CH<sub>4</sub> were found in monitoring wells which were sampled using low flow techniques.

In this study,  $CH_4$  is predominantly the sole hydrocarbon above DL (10µg/L), with only 2 samples in the shallow coal measures (IND1 and IND3) containing small concentrations of ethane and propane above DL (10µg/L).

Figure 2a,b conceptualise the spatial distribution of  $CH_4$  in the shallow coal measures and the alluvium, respectively. The dissolved  $CH_4$  in the alluvium occurred over a large depth range (~20–110 m) and over a large spatial area:  $CH_4$  distribution in the alluvium is relatively sparse, and peak alluvial- $CH_4$  concentrations do not show any spatial relationship with peak  $CH_4$  concentrations in the underlying coal measures.

 $\delta^{13}$ C-CH<sub>4</sub> and  $\delta^{2}$ H-CH<sub>4</sub> within and between aquifers. While thermogenic methane typically has more enriched  $\delta^{13}$ C values, biogenic CH<sub>4</sub> can also have  $\delta^{13}$ C values within what is considered a typical thermogenic range. For example, the dominance of acetoclastic methanogenesis<sup>2,5,13,26,41,42</sup>, shifts in seasonal availability of the substrate<sup>43,44</sup>, enrichment of the CO<sub>2</sub> pool as a result of on-going methanogenesis<sup>3</sup>, and anaerobic (AOM) or aerobic oxidation of CH<sub>4</sub><sup>8,45,46</sup> can all produce CH<sub>4</sub> that is relatively enriched in  $\delta^{13}$ C (see Figure S1 for a summary). Fractionation factors can offer insights into production and consumption pathways:  $\alpha_{\text{DIC-CH4}}$  of ~1.07 and  $\alpha_{\text{H2O-CH4}}$  of ~1.2 are typically indicative of CO<sub>2</sub> reduction pathways, while of  $\alpha_{\text{DIC-CH4}} \sim 1.04$  and  $\alpha_{\text{H2O-CH4}} \sim 1.4$  are typically indicative of acetoclastic methanogenesis or an oxidation pathway<sup>3,5,47</sup>.

In the gas reservoir, the  $\delta^{13}$ C-CH<sub>4</sub> and  $\delta^{2}$ H-CH<sub>4</sub> values ranged from -58% to -49%, and -210% to -198%, respectively, and correlate with positive  $\delta^{13}$ C-DIC values (+9% to +23‰) (Fig. 3a). The  $\alpha_{\text{DIC-CH4}}$  and  $\alpha_{\text{H2O-CH4}}$  of CSG production water are consistently around 1.07 and 1.2, respectively, and there is a positive relationship between the  $\delta^{13}$ C-CH<sub>4</sub> and the  $\delta^{13}$ C-DIC (Fig. 3a,b). This, in combination with no other hydrocarbons above DL, is indicative of a biogenic CO<sub>2</sub>-reduction pathway in a closed system (limited CO<sub>2</sub> pool). This is synonymous with gas trapping on geological structures in closed environments, such as anticlines and synclines. The predominance of biogenic CH<sub>4</sub> in the coal measures in this basin is supported by a number of other studies, with the most recent work suggesting microbial CH<sub>4</sub> in these reservoirs was generated since the late Pleistocene<sup>26-29</sup>.

In the shallow coal measures, the  $\delta^{13}$ C-CH<sub>4</sub> and  $\delta^{2}$ H-CH<sub>4</sub> ranged from 80% to -50%, and -310% to -210%, respectively. In the case of the alluvium, the data showed a similar range of -78% to -49%, and -315% to -186%, for  $\delta^{13}$ C-CH<sub>4</sub> and  $\delta^{2}$ H-CH<sub>4</sub>, respectively. The range of  $\delta^{13}$ C-DIC values was also similar between the shallow coal measures and the alluvium: -15.9% to -3.5%, and -15.3% to -6.6%, respectively. The  $\delta^{13}$ C-CH<sub>4</sub> and  $\delta^{2}$ H-CH<sub>4</sub> and associated fractionation factors indicate CO<sub>2</sub> reduction is the dominant pathway in the shallow coal measures, but variability in this data for the shallow coal measures and alluvium suggest there may be multiple production and/or consumption pathways influencing  $CH_4$  in these aquifers. The enriched  $\delta^{13}$ C-CH<sub>4</sub> (-50‰) and highly depleted  $\delta^{2}$ H-CH<sub>4</sub> (<310‰), and carbon and hydrogen fractionation factors of ~1.4 for a single shallow coal measure (P7) and alluvial sample (IND4) (Fig. 3a,b), are synonymous with acetoclastic methanogenesis<sup>3,48</sup>. These occur in isolation: under basalt sheetwash near Bowenville (shallow coal measures sample), and on the opposite side of the alluvium near Stratheden (alluvial sample) (see Figs 2 and S2). Acetoclastic methanogenesis has not been observed before in the Walloon Coal Measures in the Surat and Clarence-Moreton basins, although evidence of this pathway has been observed at basin margins in other areas<sup>26,27,29,41,49–51</sup>. In both cases these samples are found at relatively higher salinity and  $SO_4$  concentrations: for the shallow coal measures sample, Cl = -1775 mg/L or 50 meq/L, and  $SO_4 = 480 \text{ mg/L}$  or 10 meq/L; for the alluvial sample, Cl = 5990 mg/L or 168 meq/L, and  $SO_4 = 144 \text{ mg/L}$  or 3 meq/L.



**Figure 2.**  $CH_4 \mu g/L$  contours for: (**a**) the coal measures, including the gas reservoir and the shallow coal measures; and (**b**) the alluvium. CSG production sites are marked by arrows in (**a**). Points in (**a**,**b**) represent sample locations. Maps were prepared using *ArcGIS v 10.1* and modified using *Adobe Illustrator CC 2014*. Contours were determined using the spline tool in *ArcGIS v 10.1* (www.esri.com), which interpolates a raster surface from points using a two-dimensional, minimum curvature spline technique and which passes a contour line through each measured point. Contours marked as 0 are based on measured  $CH_4$  below DL ( $<10 \mu g/L$ ) at relevant points. This representation of  $CH_4 \mu g/L$  distribution should be used for conceptual/visualisation purposes for this data set only, as methanogenic and methanotrophic conditions may only be favourable at discrete locations and because contours between points represent a conceptual change in the concentration gradient between measured samples only.

The only coal measure samples that contained hydrocarbons in addition to  $CH_4$  were found at a nested site at Cecil Plains: these samples (IND1 and IND3) contained small concentrations of ethene (70 and  $25 \mu g/L$ ). ethane (30 and  $25 \mu g/L$ ) and propene (24 and  $<10 \mu g/L$ ) which suggests a mixed thermogenic/biogenic gas component.



**Figure 3.** Comparisons between: (a)  $\delta^{13}$ C-CH<sub>4</sub> and  $\delta^{13}$ C-DIC; (b)  $\delta^{2}$ H-H<sub>2</sub>O and  $\delta^{2}$ H-CH<sub>4</sub>; and (c)  $\delta^{18}$ O-H<sub>2</sub>O and  $\delta^{2}$ H-H<sub>2</sub>O for samples that contain CH<sub>4</sub> > 10 µg/L between aquifers in the study area. For (c): GMWL = Global Meteoric Water Line; LMWL = Local Meteoric Water Line at Toowoomba<sup>93</sup>. For (b): long-dashed lines = range of combined hydrogen isotope effects for CO<sub>2</sub>-reduction as reported in Whiticar<sup>3</sup>, being  $\delta^{2}$ H-CH<sub>4</sub> =  $\delta^{2}$ H-H<sub>2</sub>O-165‰ (±15‰); and short-dashed lines = range of combined hydrogen isotope effects for acetoclastic methanogenesis in sulfate-poor systems as reported in Waldron *et al.*<sup>48</sup>, being  $\delta^{2}$ H-CH<sub>4</sub> = 0.675 ×  $\delta^{2}$ H-H<sub>2</sub>O-284‰(±6‰). For (b), arrows represent the range of isotope effects where a combination of methanogenic pathways has potentially influenced isotopes as reported in Whiticar<sup>3</sup>. For (a), Alluvium-Free CH<sub>4</sub><sup>a</sup> = free gas samples taken from the well-head space of irrigation bores after extended pumping (up to 3 months) near Cecil Plains, as reported in Iverach *et al.*<sup>21</sup>.

However, the depleted  $\delta^{13}$ C-CH<sub>4</sub> of these samples (-71‰ and -65‰, respectively) as well as the  $\alpha_{\text{DIC-CH4}}$  and  $\alpha_{\text{H2O-CH4}}$  indicate biogenic CH<sub>4</sub> is dominant at these sites.

 $\delta^{18}$ O and  $\delta^{2}$ H in water. The CSG production water tends to be more isotopically depleted (ranging from – 7.2‰ to –5.2‰, and –44.1‰ to –33.1‰, for  $\delta^{18}$ O and  $\delta^{2}$ H, respectively) than the shallow coal measures (ranging from – 5.5‰ to –4.3‰, and –36.2‰ to –28.2‰, for  $\delta^{18}$ O and  $\delta^{2}$ H, respectively) and the alluvial water (ranging from –5.9‰ to –4.2‰, and –38.2‰ to –26.8‰, for  $\delta^{18}$ O and  $\delta^{2}$ H, respectively). This indicates that these deeper areas of the coal measures were recharged during cooler climates than the shallow coal measures and alluvium (Fig. 3c). These values are within the range previously reported for production water in the Surat Basin, which suggests recharge during the last glacial period in south east Queensland<sup>27,28</sup>. We found no evidence of a spatial relationship between the similarities in the stable isotopes of water from the alluvial sample with depleted stable isotopes of water is found in a shallow well (18 m) located on the north eastern flank of the alluvium and is not related to the gas reservoir. Similarly, the most depleted shallow coal measures sample occurs in the ranges near a basalt outcrop. Some caution needs to be applied to interpretations of the stable isotope of water in CSG production water results because high rates of methanogenesis can influence the  $\delta^{2}$ H-H<sub>2</sub>O in closed systems<sup>41,51,52</sup>.

Assessing potential migration from the CSG reservoir to the shallow coal measures. At the depth interface between the gas reservoir and the shallow coal measures there is a distinct shift in the relatively enriched  $\delta^{13}$ C-CH<sub>4</sub> values of the gas reservoir samples, towards more depleted isotope values for samples from

the shallow coal measures. Diffusion of CH<sub>4</sub> may lead to lighter  $\delta^{13}$ C-CH<sub>4</sub><sup>12</sup> and a depletion of CH<sub>4</sub> along a diffusion pathway<sup>12</sup>. Similarly diffusion of Cl would also lead to a distinct depletion of  $\delta^{13}$ Cl in combination with a decrease in TDS. However, for these data, an upward diffusion scenario from the CSG reservoir to shallower areas is not evidenced from the  $\delta^{13}$ Cl, TDS, CH<sub>4</sub> or  $\delta^{13}$ C-CH<sub>4</sub> data (Fig. 4a–d). This distinct change in the  $\delta^{13}$ C-CH<sub>4</sub> values indicates that there is no evidence of leakage from the deeper gas reservoir to overlying shallow zones in the coal measures, either via diffusion or ebullition/advection. Therefore, the variability of the  $\delta^{13}$ C-CH<sub>4</sub> in the shallow coal measures must be the result of changes in methanogenic pathways and/or consuming processes.

The influence of SO<sub>4</sub> on CH<sub>4</sub> in the shallow coal measures. A decrease in the  $\alpha_{DIC-CH4}$  as the  $\delta^{13}$ C-CH<sub>4</sub> become more depleted in the shallow coal measures suggests influences of different methanogenic pathways (Fig. 5a). This is generally associated with a depletion of  $SO_4$  (Fig. 5b). The presence of  $SO_4$  can limit methanogenic activity, particularly for  $CO_2$  reducers, because  $SO_4$ -reducing organisms are better at accessing both  $H_2$  and acetate<sup>1,4,5,1,4,53,54</sup>. In most cases the  $SO_4$  reducers maintain  $H_2$  levels below a threshold at which  $CO_2$ reducers can compete, resulting in complete inhibition of CO<sub>2</sub> reduction. In contrast, acetoclastic methanogens, despite having a slower growth rate, can compete for acetate with SO<sub>4</sub> reducers to the point where both organisms can co-exist<sup>54</sup>. Therefore, as  $SO_4$  depletes in the coal measures, we can expect changes in the methanogenic community, from one where  $CO_2$  reduction is inhibited by  $SO_4$  reducers and where some acetoclastic methanogenesis can occur, to one where CO<sub>2</sub> reduction dominates over acetogens. This has implications for the  $\delta^{13}$ C-CH<sub>4</sub> and could explain why the  $\alpha_{\text{DIC-CH4}}$  changes as the  $\delta^{13}$ C-CH<sub>4</sub> depletes (Fig. 5a). This hypothesis is supported by a decrease in SO<sub>4</sub> concentrations as the  $CH_4$  increases (Fig. 5e). A positive relationship between  $CH_4$  concentrations and  $\delta^{13}$ C-DIC (Fig. 5c) demonstrates active methanogenesis and, where SO<sub>4</sub> becomes depleted and  $CO_2$  reduction becomes dominant, higher  $CH_4$  concentrations indicate higher rates of methanogenesis via the  $CO_2$ -reduction pathway. Data do not indicate an influence of DO concentrations on  $CH_4$  or associated isotopes (Fig. 5d), although methanogens can tolerate low concentrations of  $DO^{55}$ . Spatially variable  $CH_4$  in the coal measures is supported by other recent studies which suggested variability in recharge as a possible influence<sup>56,57</sup>. In this study, variability of recharge may be contributing  $SO_4$  (either through discharge or pyrite dissolution) and DO, particularly in the shallower zones.

**Thermodynamic controls on CH**<sub>4</sub> in the shallow coal measures. In order to further explore the potential dynamism between CH<sub>4</sub> production pathways, SO<sub>4</sub> reduction, potential anaerobic oxidation of CH<sub>4</sub> (AOM) and their influences on carbon and hydrogen isotopes at these large scales, we use a novel combination of thermodynamic information and changes in the activities of reactive species and isotope data expressed as isometric log ratios. A key aspect of this approach is understanding the behaviour of H<sub>2</sub>, which is a rate-limiting reactant for both CO<sub>2</sub> reduction and SO<sub>4</sub> reduction, while other reactants, such as HCO<sub>3</sub> and SO<sub>4</sub> may also provide favourable/unfavourable conditions for certain microbial pathways in coal seams<sup>4–6,58,59</sup>.

A sequential binary partition is used to calculate each isometric log ratio<sup>60</sup>. The sequential binary partition for each reactant shown in equations (6–8) (CO<sub>2</sub> reduction, SO<sub>4</sub> reduction and AOM, respectively) is shown in Tables 2, 3 and 4, respectively. The activity of H<sub>2</sub>O is ignored in relevant reactions, since it is always ~1. All ilr-coordinates are calculated using equation (1). In all isometric log ratio (ilr) calculations, the first ilr represents the compositional changes in the reaction pathway (products versus reactants). As a result, the first ilr (ilr.1) for each reaction pathway is similar to the reaction quotient (Q) used to calculate the change in Gibbs free energy. Using this approach, the principles of compositional data analysis and the law of mass balance holds, such that changes in the composition of species subsequently change the composition of the reactants and products. Where the reaction pathway is limited by the availability of one or more reactants, the ilr.1 is expected to follow a linear relationship with the changes in Gibbs free energy. The remaining ilr-coordinates partition the reactants into subcompositions, thus describing the availability of reactants for the reaction.

For the CO<sub>2</sub> reduction pathway scenario (Fig. 6a(i–iii)), a decrease of H<sub>2</sub> relative to other reactants (H<sup>+</sup> and HCO<sub>3</sub><sup>-</sup>) (CO<sub>2</sub>-ilr.2) occurs as the reaction pathway proceeds ( $\Delta G/e^-$  become less negative). This can be interpreted as the consumption of H<sub>2</sub> as methanogenesis proceeds and as SO<sub>4</sub> is depleted. The inverse relationship with the CO<sub>2</sub>-ilr.1 (reactants vs products) shows that the availability of H<sub>2</sub> in higher SO<sub>4</sub> environments is limiting CO<sub>2</sub> reduction pathways. A depletion in the relative R- $\delta^{13}$ C-CH<sub>4</sub> and enrichment of R- $\delta^{2}$ H-CH<sub>4</sub> isotope along this pathway support a shift from acetoclastic methanogenesis in higher SO<sub>4</sub> environments where competition from SO<sub>4</sub> reducers is higher to one where CO<sub>2</sub> reduction becomes dominant in lower SO<sub>4</sub> environments. In addition to low SO<sub>4</sub> concentrations, low H<sub>2</sub> and low HCO<sub>3</sub> concentrations can also create more favourable conditions for CO<sub>2</sub> reducers<sup>59</sup>.

For the SO<sub>4</sub> reduction pathway (Fig. 6b(i–iii)), poor relationships between all SO<sub>4</sub>-ilr.2,  $\Delta G/e^-$  and isotopic responses was observed. This indicates different controls on the SO<sub>4</sub> reduction pathway: it does not appear to be limited by the availability of reactants, including H<sub>2</sub> and, with the exception of the obvious acetoclastic sample, does not appear to accompany a distinct carbon or hydrogen isotopic response.

Increases (less negative) in the  $\Delta$ G/e<sup>-</sup> for the AOM pathway are accompanied by an increase in the relative concentration of CH<sub>4</sub> to SO<sub>4</sub> (AOM-ilr.3), which shows that as AOM proceeds, the system moves towards one where the CH<sub>4</sub>/SO<sub>4</sub> ratio increases (Fig. 6c)(i–iii). This indicates that, as the AOM reaction approaches thermodynamic equilibrium, the amount of SO<sub>4</sub> available for the reaction decreases, yet CH<sub>4</sub> must continue to be produced. The availability of SO<sub>4</sub> appears to be a limiting reactant for the AOM pathway. These relationships are accompanied by a relative depletion of R- $\delta^{13}$ C-CH<sub>4</sub> and enrichment of R- $\delta^{2}$ H-CH<sub>4</sub> values as CH<sub>4</sub> concentrations increase. Any CH<sub>4</sub> oxidation in higher SO<sub>4</sub> environments, as well as the slow growth rate of acetoclastic methanogens, is likely to contribute to the relatively lower CH<sub>4</sub> concentrations in higher SO<sub>4</sub> environments. In some cases AOM can occur in tandem with methanogenesis<sup>61</sup>. However, due to generally low S<sub>2</sub><sup>-</sup> and HS<sup>-</sup> being <DL for all



**Figure 4.** CSG groundwater and other groundwater samples that contain  $CH_4 > 10 \mu g/L$ , showing: (a) TDS versus screen depth; (b)  $\delta^{37}$ Cl versus screen depth; (c)  $CH_4$  versus screen depth; and (d)  $\delta^{13}$ C-CH<sub>4</sub> versus screen depth.

samples, we do not expect the influence of AOM to be significant when compared with the influence of SO<sub>4</sub> and shifts from acetoclastic methanogenesis to CO<sub>2</sub>-reduction. At an isolated site underlying a basalt outcrop (P19), NO<sub>3</sub> concentrations were slightly above DL (0.01 mg/L) at 0.02 mg/L (3.23e-04 meq/L), but the highly depleted  $\delta$  <sup>13</sup>C-CH<sub>4</sub> (-80‰) at this site does not suggest oxidation via denitrification is occurring.

**Assessing potential migration of CH**<sub>4</sub> from the shallow coal measures to the alluvium. *Nested sites.* Three (n = 3) nested sites that include wells in the underlying coal measure and overlying alluvium wells were sampled: (1) Cecil Plains; (2) Stratheden; and (3) Dalby (see Figure S2). At all sites CH<sub>4</sub> was observed in the underlying shallow coal measures, or the Kumbarilla Beds, but no CH<sub>4</sub> was found in the alluvial wells.

At the Stratheden nested well site (IND4, IND5, IND6), water levels are similar, indicating there is not a significant pressure gradient to induce groundwater flow, and the absence of  $CH_4$  in the intermediate well does not suggest upward  $CH_4$  at this site. The  $\delta^{13}C$ - $CH_4$  of the alluvial  $CH_4$  at this site is more enriched (-50%) when compared with the deeper Kumbarilla  $CH_4$  sample (-68%): the highly depleted  $\delta^2$ H- $CH_4$  (-315%) of the alluvial sample at this nested site (IND4) indicates acetoclastic methanogenesis<sup>3,48</sup>.

At the Cecil Plains nested site (P20, IND1, IND2, IND3) the sample with peak DOC (7 mg/L) (IND2) occurred in the alluvial-coal measure transition zone (85 m, ~7 m below the alluvial basement), but the DOC of the overlying alluvial sample was significantly lower (0.3 mg/L). On the same note, the CH<sub>4</sub> samples from the two coal measures samples at this site is accompanied by small concentrations of ethene (70 and 25 µg/L). ethane (30 and 25 µg/L) and propene (24 and <10µg/L); yet we found no other hydrocarbons in the alluvial sample (DL for all hydrocarbons = 10µg/L). Similarly, at the Dalby nested site (GM1390, GM1074) no CH<sub>4</sub> was found in the alluvial well (see Figure S2 for details of nested sites). We conclude that no CH<sub>4</sub> migration from the underlying coal measures into the alluvium is occurring at these sites.



**Figure 5.** (a)  $\delta^{13}$ C-CH<sub>4</sub> versus  $\alpha_{\text{DIC-CH4}}$  values for all dissolved and free gas samples (regression line is for shallow coal measures (grey squares) only), as well as dissolved CH<sub>4</sub> samples for shallow coal measures showing: (b) log SO<sub>4</sub> meq/L versus  $\alpha_{\text{DIC-CH4}}$  values; (c)  $\delta^{13}$ C-DIC versus log CH<sub>4</sub> µg/L; (d) DO mg/L versus log CH<sub>4</sub> µg/L; and (e) log SO<sub>4</sub> meq/L versus log CH<sub>4</sub> µg/L. For (a), Alluvium–Free CH<sub>4</sub><sup>a</sup> = free gas samples taken from the well-head space of irrigation bores after extended pumping (up to 3 months) near Cecil Plains, as reported in Iverach *et al.*<sup>21</sup>.

	$[H^+]$	[H <sub>2</sub> ]	[HCO <sub>3</sub> ]	$[CH_4]$
CO2_ilr.1	-1	-1	-1	1
CO <sub>2</sub> _ilr.2	-1	1	-1	
CO <sub>2</sub> _ilr.3	-1		1	

#### Table 2. Sequential binary partition for the CO<sub>2</sub> reduction pathway.

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	[SO4]	[H <sub>2</sub> ]	$[H^+]$	[HS <sup>-</sup> ]
SO <sub>4</sub> _ilr.1	-1	-1	-1	1
SO <sub>4</sub> _ilr.2	-1	1	-1	
SO <sub>4</sub> _ilr.3	1		-1	

#### Table 3. Sequential binary partition for the SO<sub>4</sub> reduction pathway.

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	[HCO <sub>3</sub> ]	$[CH_4]$	[SO <sub>4</sub> ]	[HS <sup>-</sup> ]
AOM_ilr.1	1	-1	-1	1
AOM_ilr.2	1			-1
AOM_ilr.3		1	-1	

#### Table 4. Sequential binary partition for the anaerobic oxidation of CH<sub>4</sub> (AOM) pathway.

Dissolved  $CH_4$  in the alluvium. Results show that the  $CH_4$  in the shallow coal measures that directly underlie the alluvium are depleted in  $\delta^{13}C$ -CH<sub>4</sub> (-80% to -65%), and have  $\delta^{2}H$ -CH<sub>4</sub> between -222% and -209%. Fractionation factors and thermodynamic results indicate CH<sub>4</sub> in the shallow coal measures is generated



**Figure 6.** Microbial reaction pathways in the shallow coal measures (<200 m) for: (**a**) methanogenesis via CO<sub>2</sub> reduction; (**b**) SO<sub>4</sub> reduction; and (**c**) anaerobic oxidation of CH<sub>4</sub> (AOM), showing comparison of isometric log ratios (ilr) derived from the partitioning of reactants and products (ilr.1–solid circles) and the partitioning of reactants (ilr.2–open circles) (see Tables 2–4 for respective SBPs) and: (i) changes in Gibbs free energy standardised to the number of electrons transferred for each reaction (8) ( $\Delta G/e^-$ ); (ii) Rayleigh fractionation of  $\delta^{13}$ C-CH<sub>4</sub>; and (iii) Rayleigh fractionation of  $\delta^{2}$ H-CH<sub>4</sub>. Dashed arrows in (i) represent the direction in which the thermodynamic reaction proceeds (approaches equilibrium). Under acetoclastic methanogenesis the  $\delta^{13}$ C-CH<sub>4</sub> can be relatively enriched, yet should become more depleted as CO<sub>2</sub> reduction proceeds. In the same context, the  $\delta^{2}$ H-CH<sub>4</sub> is highly depleted under acetoclastic methanogenesis, yet more enriched under CO<sub>2</sub> reduction. As a result, a reciprocal response between carbon and hydrogen isotopes is expected as the reaction pathway changes: therefore, the R- $\delta^{13}$ C-CH<sub>4</sub> is defined as R = R<sub>i</sub> f<sup>(1-\alpha)</sup>, while R- $\delta^{2}$ H-CH<sub>4</sub> is defined as R = R<sub>i</sub> f<sup>(\alpha-1)</sup>, where f=0=min CH<sub>4</sub>. The acetoclastic sample is marked with a cross in (i). \*Samples containing ethene (70 and 25 µg/L), ethane (30 and 25 µg/L) and propene (24 and 0µg/L).

predominantly via the CO<sub>2</sub> reduction pathway, with SO<sub>4</sub> concentrations being a major control. Subsequently, the assessment of potential migration of CH<sub>4</sub> from the coal measures to the alluvium must consider this CH<sub>4</sub> of the shallow (underlying) coal measures as the appropriate end member. The lack of evidence of CH<sub>4</sub> leakage from the gas reservoir to the shallow coal measures, and the abrupt shift from enriched  $\delta^{13}$ C-CH<sub>4</sub> (-58‰ to -49‰) of the gas reservoir to the depleted  $\delta^{13}$ C-CH<sub>4</sub> of the shallow coal measures indicate that the migration of CH<sub>4</sub> from the gas reservoir to the alluvium at these sites is not a plausible scenario based on these data. Variability in the TDS and CH<sub>4</sub> concentrations, as well as the  $\delta^{37}$ Cl (Fig. 4), and similar ranges of  $\delta^{13}$ C-CH<sub>4</sub> and  $\delta^{2}$ H-CH<sub>4</sub> between the underlying coal measures and deep alluvial samples (Fig. 7d,e) also do not suggest diffusion of CH<sub>4</sub> from the underlying coal measures to the alluvium.

No relationship between depth and  $CH_4$  concentration in the alluvium was observed at sites sampled in this study, with the highest concentrations occurring at ~60 m (Fig. 7a). Thermodynamic conditions in the alluvium



**Figure 7.** Comparison of various parameters in the alluvial depth profile for sites with dissolved  $CH_4 > DL$  (10µg/L), showing: (a)  $CH_4$  (µg/L),  $SO_4$  meq/L and DO mg/L; (b) Br/Cl ratio,  $NO_3$  mg/L and DOC mg/L; (c)  $\Delta G/e^-$  for  $CO_2$  reduction,  $SO_4$  reduction and AOM pathways; (d) the carbon isotopes of  $CH_4$  and DIC phases and their respective fraction factors; (e) the hydrogen isotopes of  $CH_4$  and  $H_2O$  phases and their respective fraction factors; and (f)  $CO_2$ .ilr-2 (the ilr of  $[H_2]/[reactants of <math>CO_2$  reduction/ $SO_4$  reduction pathways] (see Tables 2 and 3)), and the saturation of indices of kaolinite and gypsum. Circles in (a,d) represent the sample point depth: this corresponds to the sample point for all parameters in all plots. Shaded areas in (d,e) represent the ranges of  $\delta^{13}C-CH_4$ ,  $\delta^2H-CH_4$ , respectively for the coal measures that directly underlie the alluvium.

are suitable for all reaction pathways to occur (Fig. 7c); however, the variability of  $CH_4$  concentration in the alluvium is related to the inverse of  $SO_4$  concentration (Fig. 7a), demonstrating the influence of  $SO_4$  reduction on methanogenic activity. High concentrations of  $SO_4$  accompany high TDS (salinity) and large decreases in the Br/Cl ratio (Fig. 7a,b). This shows that different controls on salinity influence these high  $SO_4$  concentrations. The relatively consistent  $\delta^2$ H-H<sub>2</sub>O at maximum salinity shows that the CH<sub>4</sub> and related hydrochemical conditions

are not necessarily related to different sources of water or simple evaporation processes, and are more likely to reflect the accumulation of salts, including gypsum, or transpiration at a less permeable zone. Interestingly, the high-SO<sub>4</sub> zone coincides with depleted  $\delta^2$ H-CH<sub>4</sub> values (-315‰) that indicate acetoclastic methanogenesis (Fig. 7e). This explains the more enriched  $\delta^{13}$ C-CH<sub>4</sub> for this sample (IND4) and it is a similar scenario to that which we observed in the coal measures.

 $\delta^{13}$ C-CH<sub>4</sub> and  $\delta^{2}$ H-CH<sub>4</sub> values for two deep alluvial samples (GM1193 and GM0057: 110 m and 57 m, respectively) are within a similar range to that of the underlying coal measures (Fig. 7d,e). The deepest site (GM1193) is at/near the alluvial-coal measure transition zone. As stated previously, there are small coal fragments in the sandy alluvial deposits at this site, which could provide a methanogenic substrate. Furthermore, the  $\delta^{18}$ O of this sample is the most enriched of the CH<sub>4</sub> data set and does not suggest a coal measure source (Fig. 3c). Similarly, the  $\delta^{13}$ C-CH<sub>4</sub> and  $\delta^{2}$ H-CH<sub>4</sub>, as well as the  $\alpha_{\text{DIC-CH4}}$  and  $\alpha_{\text{H2O-CH4}}$ , are also consistent with *in situ* CO<sub>2</sub> reduction at the deepest site (Fig. 7d,e), and do not suggest an oxidation pathway or CH<sub>4</sub> sourced from another area/zone.

Where peak CH<sub>4</sub> concentrations occur (~57 m: GM0057), the  $\alpha_{\text{DIC-CH4}}$  values are as low as ~1.04 (Fig. 7d), but the depleted  $\delta^{13}$ C-CH<sub>4</sub> and enriched  $\delta^{2}$ H-CH<sub>4</sub> do not support acetoclastic methanogenesis at this site. While  $\alpha_{\text{DIC-CH4}}$  values of ~1.07 are typical of CO<sub>2</sub> reduction, a fractionation factor of 1.04 is still within the range observed for CO<sub>2</sub> reduction<sup>5,62-64</sup>. These fractionation factors can change between sites and as a function of *in situ* conditions<sup>5,6</sup>. Low  $\Delta$ G/e<sup>-</sup> values at this site may also suggest some AOM has occurred (Fig. 7c). Alternatively the acetate- and H<sub>2</sub>-dependent methanogenesis may also occur concurrently during acetate fermentation in some cases<sup>65,66</sup>. Well GM0057 is located near the river and may also receive some river recharge. This well, and well GM1193, occur in a sandy area of the aquifer where pumping rates and recharge are likely to be relatively higher than areas around Dalby; this could explain slightly higher DO concentrations (Fig. 7a). Methanogenesis may also persist in the presence of low DO concentrations<sup>55</sup>, and mixing of slightly oxygenated water (river recharge) and the dissolution of carbonates may explain the relatively lower  $\alpha_{\text{DIC-CH4}}$  values at GM0057.

In the shallow alluvial zones, fluxes in the type and rate of methanogenesis could be influenced by wetting and drying periods that result in dissolution or precipitation of minerals such as gypsum (Fig. 7f). In addition, clay mineral content has also been shown to influence  $CH_4$  concentrations, with high clay content capable of trapping  $CH_4^{67}$ . Furthermore, some clays (e.g. kaolinite) preserve organic matter better than others<sup>68</sup>. For samples analysed in this study, kaolinite saturation indices are highest in the high-SO<sub>4</sub> zone where acetoclastic methanogenesis dominates (Fig. 7f), which also accompanies a peak in DOC concentrations (Fig. 7b). The presence of kaolinite clay lenses in shallow areas may have a dual effect on methanogenic activity by preventing flushing and promoting salinization that result in higher SO<sub>4</sub>, as well as the preservation of some organic matter that allows fermentation processes to persist. The proportion of [H<sub>2</sub>] to other reactants ([HCO<sub>3</sub>] and [H<sup>+</sup>]) (CO<sub>2</sub>.ilr-2) in the more saline/high SO<sub>4</sub> zone increases (ratio of H<sub>2</sub> to HCO<sub>3</sub> increases), despite  $CH_4$  being low (IND4): this indicates a fermentation process by SO<sub>4</sub> reducers and acetate-dependent methanogens.

While AOM is thermodynamically favourable, slightly more depleted  $\delta^{13}$ C-CH<sub>4</sub> and  $\alpha_{DIC-CH4}$  values ~1.07 in the shallower zones do not suggest significant AOM is occurring (Fig. 7d,e). Mixing processes in the shallow alluvium may create scenarios where water with CH<sub>4</sub> is mixing with water with higher concentrations of redox species. For one sample (GM1076), a small increase in the NO<sub>3</sub> concentration is evident (Fig. 7b) and the oxidation of small amounts of CH<sub>4</sub> via denitrifying bacteria cannot be completely ruled out<sup>69,70</sup>, although we found no NO<sub>2</sub> above DL at any sites, and fractionation factors support a CO<sub>2</sub> reduction pathway. Substrate depletion may also explain enriched  $\delta^{13}$ C-CH<sub>4</sub> in these shallow zones<sup>3</sup>. Some caution should be applied when drawing conclusions for the shallowest alluvial samples (GM1073 and GM1076) because the measured  $\delta^{13}$ C-CH<sub>4</sub> was at or near the limit of quantification (0.8 nanomoles) for the analytical method (this is not the case for  $\delta^2$ H-CH<sub>4</sub>). However, we note that deeper alluvial wells in this area of the alluvium, which is adjacent to the deep gas reservoir, did not contain CH<sub>4</sub> above DL (10µg/L). As a result, CH<sub>4</sub> migration from the underlying coal measures in this area does not seem likely at these sites.

A conceptual model of  $CH_4$  within and between aquifers. A conceptual model (Fig. 8) summarises the major controls on  $CH_4$  within and between aquifers, including:

- (1) Closed system conditions leading to enriched  $\delta^{13}$ C-CH<sub>4</sub> and positive  $\delta^{13}$ C-DIC in the deep gas reservoir (200–500 m); and
- (2) The presence SO<sub>4</sub> concentrations and its influence on methanogenic pathways, including shifts from the acetoclastic pathways in shallow, brackish- and high SO<sub>4</sub>- zones to a dominance of CO<sub>2</sub> reduction in deeper, low SO<sub>4</sub> zones, in both the shallow coal measures and the alluvium.

The inverse relationships between  $CH_4$  and  $SO_4$ , and associated isotopic responses and thermodynamic conditions, in the shallow coal measures and alluvium are consistent with *in situ*  $CH_4$  production in other freshwater and brackish environments<sup>3-6,71</sup>. This, combined with results at nested sites and an absence of  $CH_4 > DL$  (10µg/L) in the alluvium, does not suggest large-scale migration of  $CH_4$  from the underlying coal measures is occurring.

Due to the complexity of methanogenesis and methantrophy in the subsurface, different pathways and sources can result in similar  $\delta^{13}$ C-CH<sub>4</sub> values (see Figure S1), and CH<sub>4</sub> and  $\delta^{13}$ C-CH<sub>4</sub> are not likely to be spatially consistent. Enriched  $\delta^{13}$ C-CH<sub>4</sub> from CSG production water can pertain to gas trapping scenarios at discrete locations<sup>3,26,72</sup> and these values are not necessarily representative of CH<sub>4</sub> in the entire aquifer. For future studies that are concerned with understanding CH<sub>4</sub> behaviour in the subsurface over large areas and/





or associated with CSG, we propose the following parameters as a minimum standard for data collection:  $\delta$  <sup>13</sup>C-CH<sub>4</sub> and  $\delta$ <sup>2</sup>H-CH<sub>4</sub>,  $\delta$ <sup>13</sup>C-DIC, major ions, pH and SO<sub>4</sub> and S<sub>2</sub><sup>-</sup> (other redox species, such as Fe and NO<sub>3</sub>, may also have some value). Researchers are encouraged to prepare comprehensive data sets of a range of parameters that allow conceptual models of the extent, and influences on, CH<sub>4</sub> within and between aquifers to be described and built upon over time. The conceptual model outlined here (Fig. 8) provides a basis for doing this in this catchment. More sampling to identify the presence of methanogenic consortia (culturing studies) within and between aquifers, including the extent of acetoclastic methanogens, would build on the information collected in this study.

**Comparisons with free gas measurements from alluvial wells.** The results presented here are not in agreement with another study in the Cecil Plains area which used  $\delta^{13}$ C-CH<sub>4</sub> of free CH<sub>4</sub> taken from multi-screened irrigation wells during pumping to infer CH<sub>4</sub> leakage from the coal measures to the alluvium at four sites<sup>21</sup>. That study proposed the following be met to infer CH<sub>4</sub> migration from the underlying coal measures:

- (1) DOC > DL, and  ${}^{3}H < QL$  (0.04 TU), where QL is quantification limit (this relationship inferred a potential source of coal measure groundwater/CH<sub>4</sub>); and
- (2) Samples must sit on a mixing line between  $1/CH_4$  and  $\delta^{13}C-CH_4$ , with a *y*-axis intercept with a  $\delta^{13}C-CH_4$  value of -55.9%.

That study assumed that the  $\delta^{13}$ C-CH<sub>4</sub> value of -55.9% used in their mixing line is representative of the CH<sub>4</sub> in entire coal measure aquifer, and that there are only two sources of DOC: river recharge or discharge from the coal measures. A  $\delta^{13}$ C-CH<sub>4</sub> value of -50.8%, based on a single atmospheric measurement downwind of a CSG production water storage pond, was also used as a reference (end-member) value for the coal measures aquifer.

However, that study did not take any samples from the coal measures, either underlying the alluvium or in other areas, for reference.

*Relationships between DOC and CH*<sub>4</sub>. We found DOC in the alluvium (and the coal measures) to be relatively low, yet within a consistent range, regardless of distance from the river or tritium activity. Advanced analytical techniques are required to confidently detect tritium at low TU. We used a highly sensitive tritium analytical technique (DL = 0.02 TU)<sup>73</sup>, yet only found tritium >DL at two alluvial wells (GM1076 and GM1338). Iverach *et al.*<sup>21</sup> suggested that, where tritium was below QL, yet DOC is present, a source of DOC, in addition to river recharge, must be present. These authors proposed that "upwards migration of CH<sub>4</sub> from the coal measures would be the most likely source" of DOC in the alluvium at these sites; however, CH<sub>4</sub> is not part of the DOC pool.

Other studies have indicated that a diffuse recharge component over the alluvium is possible in this catchment<sup>23,74,75</sup>, which may contribute to the DOC pool in the alluvium. In addition, DOC can diffuse through clays and/or be preserved by some clays such as kaolinite, and DOC can also be generated *in situ* in the subsurface from sedimentary sources<sup>39,68,76</sup>. Therefore, DOC and CH<sub>4</sub> are likely to be associated with different sources and transport mechanisms. Furthermore, CO<sub>2</sub> reduction is the dominant methanogenic pathway in the coal measure and alluvial aquifers, and this pathway does not rely on DOC as the energy source, rather it uses  $H_2^2$ (equation 6). We suggest that a more comprehensive research approach is needed to better understand relationships between DOC, age tracers, such as tritium, and methanogenesis within and between aquifers before combinations of these parameters can be used to confidently validate aquifer connectivity, particularly when working at large scales.

Describing the coal measure  $CH_4$  end member. The enriched  $\delta^{13}C$ -CH<sub>4</sub> value (55.9‰) estimated for the regression line used by Iverach *et al.*<sup>21</sup> to infer CH<sub>4</sub> leakage from the coal measures to the alluvium is within the range of the  $\delta^{13}C$ -CH<sub>4</sub> observed for the deeper gas reservoir (>200 m) sampled in our study (-58% to -49%), and other gas reservoirs in the Surat Basin<sup>27,28</sup>. However, it contrasts with the depleted  $\delta^{13}C$ -CH<sub>4</sub> (-80% to -65%) that we observed for the shallow (<200 m) coal measures that underlie the alluvium. In this study area the CSG reservoir occurs in deeper zones (>200 m) of the coal measures where gas trapping occurs on the north-western flank of the alluvium (Figure S2). These conditions produce enriched  $\delta^{13}C$ -CH<sub>4</sub> and high, positive  $\delta^{13}C$ -DIC (Fig. 4a) in the gas reservoir that do not occur in the shallower coal measures directly under the alluvium. High and positive  $\delta^{13}C$ -DIC values were not found in the shallow coal measures or the alluvium, either in this study or by Iverach *et al.*<sup>21</sup>. The only enriched  $\delta^{13}C$ -CH<sub>4</sub> (-50%) we observed for the shallow coal measures was an isolated acetoclastic CH<sub>4</sub> sample (P7) that underlies basalt sheetwash (see Figure S2).

The majority of  $\delta^{13}$ C-CH<sub>4</sub> of free CH<sub>4</sub> measured in Iverach *et al.*<sup>21</sup> are similar to background CH<sub>4</sub> concentrations observed in that study and for ambient air in the southern hemisphere observed in other studies<sup>78-80</sup>. It is possible that most of the CH<sub>4</sub> analysed in Iverach *et al.*<sup>21</sup> were composed of atmospheric CH<sub>4</sub>. Additional sampling (preferably using low flow techniques) to measure the degassing rate/potential from alluvial groundwater would also assist in more accurately describing the proportion of atmospheric versus degassed CH<sub>4</sub> in the well-head spaces measured by Iverach *et al.*<sup>21</sup>. Where mixing with atmospheric and subsurface-derived CH<sub>4</sub> is shown to occur, simple mixing lines may be inadequate to understand mixing of the three theoretical end members that should be considered under these potential inter-aquifer CH<sub>4</sub> migration scenarios: i.e. (1) atmospheric CH<sub>4</sub>; (2) alluvial-derived CH<sub>4</sub>; (3) CH<sub>4</sub> that has migrated from other aquifers.

Hydrogen isotope analyses can reduce uncertainties associated with interpretations that are based solely on  $\delta^{13}$ C-CH<sub>4</sub> values, as presented in Iverach *et al.*<sup>21</sup>. Atmospheric CH<sub>4</sub> tend to be much more enriched in  $\delta^{2}$ H-CH<sub>4</sub> values (-82‰) compared to biogenic CH<sub>4</sub> (-160‰ to >-400‰)<sup>3,48,80</sup>. In addition, CH<sub>4</sub> oxidation could partly explained the enriched  $\delta^{13}$ C-CH<sub>4</sub> values (-47.4‰ to -38.8‰) measured in Iverach *et al.*<sup>21</sup>. Hydrogen isotope data can also provide information about the influence of oxidation, as well as different production pathways, on the isotopic composition of CH<sub>4</sub><sup>3,48</sup>.

#### Conclusions

Using a comprehensive data set (dissolved CH<sub>4</sub>,  $\delta^{13}$ C-CH<sub>4</sub>,  $\delta^{2}$ H-CH<sub>4</sub>,  $\delta^{13}$ C-DIC,  $\delta^{37}$ Cl,  $\delta^{2}$ H-H<sub>2</sub>O,  $\delta^{y}$ O-H<sub>2</sub>O, Na, K, Ca, Mg, HCO<sub>3</sub>, Cl, Br, SO<sub>4</sub>, NO<sub>3</sub> and DO) this study described hydrochemical/thermodynamic controls on CH<sub>4</sub> in a deep coal seam gas (CSG) reservoir (200–500 m), shallower areas of the same coal-bearing formation (the Walloon Coal Measures) (<200 m) and the overlying Condamine River alluvium (Surat/Clarence Moreton basins), eastern Australia. Enriched  $\delta^{13}$ C-CH<sub>4</sub> (–58‰ to –49‰) and positive  $\delta^{13}$ C-DIC (+9‰ to +23‰) in the deep gas reservoir are synonymous with biogenic methanogenesis in closed-system conditions and gas trapping on geological structures. Evidence of leakage from the deep gas reservoir, either via diffusion or ebullution/ advection, was not observed, with  $\delta^{13}$ C-CH<sub>4</sub> of the shallow coal measures underlying the alluvium being depleted (-80‰ to -65‰). Importantly, this study demonstrates that, when evaluating potential CH<sub>4</sub> migration associated with CSG, the enriched  $\delta^{13}$ C-CH<sub>4</sub> of CSG CH<sub>4</sub> is not necessarily the appropriate isotopic end member because the  $\delta^{13}$ C-CH<sub>4</sub> in areas outside of gas reservoirs, yet within the same sedimentary formation, can be distinctly different due to different hydrogeological and microbial conditions. We found the  $\delta^{13}$ C-CH<sub>4</sub> of the alluvium falls within a similar range to that of the shallow coal measures.

Using a novel application of isometric log ratios, this study demonstrated a simple method of providing insight into the microbial controls on  $\delta^{13}$ C-CH<sub>4</sub> and  $\delta^{2}$ H-CH<sub>4</sub> isotopes in the subsurface. The major controls on CH<sub>4</sub> in the shallow coal measures and the alluvium were found to be: (a) the presence of SO<sub>4</sub> and associated

competition between SO<sub>4</sub> reducers and CO<sub>2</sub> reducers; and (b) shifts from acetoclastic methanogenesis in shallow, high-SO<sub>4</sub> zones to the dominance of the CO<sub>2</sub> reduction pathway in low-SO<sub>4</sub> environments. AOM was found to be thermodynamically favourable but there was no evidence of large-scale, significant AOM in the shallow coal measures or the alluvium. Overall, this study did not find conclusive evidence of CH<sub>4</sub> migration to the alluvium from the underlying (shallower <200 m) coal measures, but results do suggest small concentrations of CH<sub>4</sub> are likely to be generated *in situ* in the alluvial aquifer at these sites. This study provides a comprehensive assessment using novel samples. More research and sampling in the area, including culturing studies of methanogenic consortia, will improve our understanding of the nature and extent of CH<sub>4</sub> within and between aquifers.

#### Methods

Sample collection. Samples were collected from 61 wells, including: (a) monitoring wells and stock and domestic wells where there was sufficient space to lower a bladder pump<sup>81</sup>; (b) irrigation/domestic wells that already contained submerged electric pumps; and (c) production water wells. Where a bladder pump could be lowered into a well, the low-flow sampling technique was applied using a flow-through cell<sup>81</sup>. For government monitoring wells, where monitoring wells had multiple screens, the bladder pump was placed at the interval of the lowest screen. Infrastructure at irrigation/domestic wells prevented well dipping: in these cases a conservative water level estimate of ~75% of well depth was applied to consider a purging volume. Sampling coincided with landholders pumping schedules and, as a result, the majority of irrigation/domestic wells had been purged by at least 3 x well volume upon arrival on site. In all cases (low-flow sampling and irrigation/ domestic-well sampling) sampling was only initiated after hydrochemical parameters (pH, temperature, specific conductance and DO) were stabilised<sup>81,82</sup>. Due to limited infrastructure, two operating windmills were sampled in recharge areas on the ranges (P9 and P16): in these cases sampling was conducted after a minimum of 3 days of consistent moderate-strong wind (consistent pumping to purge the well) and after stabilisation of hydrochemical parameters (pH, temperature, specific conductance and DO) were confirmed<sup>83</sup>. Coal seam gas production wells (deep gas reservoir) are constantly pumping and were considered adequately purged upon arrival on-site. Production water from CSG wells was sampled at an outlet of the extraction well prior to the gas-water separator.

Samples were taken from alluvium (n = 23), the Kumbarilla Beds (n = 3), the shallow Walloon Coal Measures (WCM) (<200 m) (n = 14) and from the deeper (200-500 m) gas reservoir in the coal measures (n = 21). Two of the alluvial samples taken were at the alluvial-WCM interface (see section 2). Wells were selected based on drill log information and previous interpretations of hydrogeology in the catchment<sup>23</sup>.

Samples for dissolved CH<sub>4</sub> were collected in glass vials with rubber septums and no headspace (preserved with sulfuric acid). Samples for  $\delta^{13}$ C-CH<sub>4</sub>,  $\delta^{2}$ H-CH<sub>4</sub> and  $\delta^{13}$ C-DIC were filtered through 0.2 micron filters and collected in 12ml entertainer vials with rubber septums and no headspace. Samples for cations, dissolved metals,  $\delta^{37}$ Cl and Br were filtered through high capacity in-line 0.45 µm polyethersulphone filters. Cation and dissolved metal samples were preserved in the field using HNO<sub>3</sub> to pH <2. Samples for <sup>3</sup>H and anions, NO<sub>3</sub>/NO<sub>2</sub>, S<sub>2</sub><sup>-</sup> and unionized HS were collected in unfiltered 1L Nalgene bottles, and HDPE bottles respectively (APHA Table 1060:1). NO<sub>3</sub> and NO<sub>2</sub> samples were preserved in the field using H<sub>2</sub>SO<sub>4</sub> to pH <2. S<sub>2</sub><sup>-</sup> and unionized HS samples were preserved in the field using the field using H<sub>2</sub>SO<sub>4</sub> to pH <2. S<sub>2</sub><sup>-</sup> and unionized HS samples were preserved in the field using the field using H<sub>2</sub>SO<sub>4</sub> to pH <2. S<sub>2</sub><sup>-</sup> and unionized HS samples were preserved in the field using H<sub>2</sub>SO<sub>4</sub> to pH <2. S<sub>2</sub><sup>-</sup> and unionized HS samples were preserved in the field using the field using H<sub>2</sub>SO<sub>4</sub> to pH <2. S<sub>2</sub><sup>-</sup> and unionized HS samples were preserved in the field using the field using the sample water prior to collecting a sample. All samples, with the exception of <sup>3</sup>H, were placed immediately on ice and stored on ice in the field, then in dark cold rooms (<4 °C) until analysis.

**Sample analysis.** Samples were analysed for pH, DO, specific conductivity (conductivity) and temperature using a YSI physico-chemical meter in the field (YSI Professional Plus). Water samples were taken and analysed in the laboratory for major and minor ions (APHA 2320; APHA 3125B) including SO<sub>4</sub> (APHA 4500 SO4-E-laboratory 0.45  $\mu$ m filtered), and Br (APHA 4110 B) as well as unionised HS (APHA 4500-S2-H) and S<sub>2</sub><sup>-</sup> (APHA 4500-S<sup>2</sup>-D), NO<sub>3</sub> and NO<sub>2</sub> (APHA VCl reduction 4500 NO<sub>3</sub><sup>-</sup> + NO<sub>2</sub><sup>-</sup>B), Fe and Mn (APHA 3125B ORP/ICP/MS Octopole Reaction Cell) and dissolved CH<sub>4</sub> concentrations (including C1-C4 gases, DL = 10 $\mu$ g/L: ALS EP033) at the Australian Laboratory Services laboratory, Brisbane, Queensland, and at Queensland Health Scientific and Forensics services laboratory (Br). Bicarbonate values are reported as bicarbonate alkalinity. All major and minor ions, and dissolved C1-C4 hydrocarbons were analysed within ~7 days of collection in the field.

 $\delta^2$ H and  $\delta^{18}$ O were measured using a Los Gatos Research Water Isotope Analyzer (QUT, Institute for Future Environments), with replicate analyses indicating an analytical error of 0.04‰ to 0.45‰, and 0.001‰ to 0.7‰, respectively.

 $\delta^{13}$ C-CH<sub>4</sub> and  $\delta^2$ H-CH<sub>4</sub> were measured using a ThermoScientific PreCon concentration system interfaced to a ThermoScientific Delta V Plus isotope ratio mass spectrometer at the UC Davis Stable Isotope Facility. Standard error of analysed samples was ~0.1‰, for  $\delta^{13}$ C-CH<sub>4</sub> and ranged from 0.9–1.7‰ for  $\delta^2$ H-CH<sub>4</sub>, and limit of quantification = 0.8 and 2 nanomoles respectively.  $\delta^{13}$ C-DIC were also measured at the UC Davis Stable Isotope Facility using a GasBench II system interfaced to a Delta V Plus isotope ratio mass spectrometer. Standard error of  $\delta^{13}$ C-DIC ranged from 0.04–0.09‰: limit of quantification = 150 nanomoles.

 $\delta^{37}$ Cl were measured using a stable isotope ratio mass spectrometer at Isotope Tracer Technologies in Waterloo, Canada. Standard error ranged from 0.03–0.16‰.

The DOC analyses were performed on a Dohrmann DC-190 Total Carbon Analyzer at the Earth and Environmental Sciences at the University of Waterloo, Canada. DOC storage times ( $0.45 \,\mu$ m filtered, dark storage <4 °C) prior to analysis ranged from 260–620 days. Data Tables S1 and S4 report the measured and corrected

	Partition of parts			
Balance	<b>x</b> <sub>1</sub>	x2	<b>X</b> 3	x <sub>4</sub>
z <sub>1</sub>	1	1	-1	-1
z <sub>2</sub>	1	-1		
Z <sub>3</sub>			-1	1

Table 5. Sequential binary partition of a four-part composition  $(x_1, x_2, \dots, x_4)$  deriving three orthonormal coordinates  $(z_1, z_2 \text{ and } z_3)$  for ilr calculation.

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DOC values, as per Peacock *et al.*<sup>84</sup>: these values are broadly similar with modelled loss of DOC being minimal due to low DOC concentrations.

Tritium (<sup>3</sup>H) samples were vacuum distilled and electrolytically enriched prior to liquid scintillation spectrometry analysis by Quantulus ultra-low-level counters at GNS, New Zealand<sup>73</sup>. The sensitivity is now further increased to a lower DL of 0.02 TU (two sigma criterion) via tritium enrichment by a factor of 95, and reproducibility of tritium enrichment of 1% is achieved via deuterium-calibration for every sample. The precision (1 $\sigma$ ) is ~1.8% at 2 TU.

**Data preparation.** All major ion data was above DL (DL) of 1 mg/L, with the exception of SO<sub>4</sub> (n = 24). All S<sub>2</sub><sup>-</sup> measurements with the exception of 1 coal measures sample (well IND3; S<sub>2</sub> = 0.5 mg/L, or ~0.008 meq/L) were below DL (DL = 0.1 mg/L, or 1.56e-03 meq/L). Low SO<sub>4</sub> and S<sub>2</sub><sup>-</sup> concentrations are expected for reduced environments where methanogenesis occurs. Values below DL were imputed using the R package zCompositions via the log ratio Data Augmentation function (lrDA): this function is based on the log-ratio Markov Chain Monte Carlo MC Data Augmentation (DA) algorithm<sup>85</sup>.

**Deriving isometric log ratios.** The isometric log ratio (ilr) uses a sequential binary partition (Table 5) to describe orthonormal bases to which correspond D-1 Cartesian coordinates (ilr-coordinates): these orthonormal coordinates, called balances, are orthogonal<sup>86</sup>. This technique removes potentially spurious correlation caused by scaling and allows the ratios of parts and subparts to be elucidated, even when the concentrations of different parts are relatively small compared to other parts. Here we use ilrs to investigate subcompositional relationships between reactants and products in a number of thermodynamic reaction pathways (CO<sub>2</sub>-reducing methanogenesis, SO<sub>4</sub> reduction and anaerobic oxidation of CH<sub>4</sub> (AOM)). These relationships are compared to isotope fractionation of the  $\delta^{13}$ C-CH<sub>4</sub> and  $\delta^{2}$ H-CH<sub>4</sub> under various conditions. This approach allows subcompositional behaviour of dissolved constituents to be compared to isotopic responses, in order to demonstrate the relationship between methanogenic activity/pathways, thermodynamic conditions and hydrochemistry.

Each partition divides the composition into separate parts ( $x_i$  and  $x_j$ ). For thermodynamic reaction pathways, we use the first partition to separate the activity of the element (represented by [element]) from reactants and products in each reaction, with the following partitions separating the reactants. Once a sequential binary partition is described, the i-th ilr balance is computed as

$$z_{i} = \sqrt{\frac{r_{i}s_{i}}{r_{i} + s_{i}}} \ln \frac{\left(\prod_{i} x_{j}\right)^{\frac{1}{r_{i}}}}{\left(\prod_{i} x_{j}\right)^{\frac{1}{s_{i}}}}$$
(1)

where  $r_i$  and  $s_i$  are the number of parts coded in the sequential binary partition as +1 and -1, respectively<sup>86</sup>. Isometric log ratios were calculated using CodaPak 2.10<sup>87</sup>.

**Describing isotope partitioning (fractionation factors).** The partition of isotopes between phases, e.g. between the dissolved inorganic carbon (DIC) and  $CH_4$  phase, can be described in a number of ways. For simplicity and reproducibility, here we define the isotope partition as the fractionation factor that is simply described as:

$$\alpha = \frac{\delta X + 1000}{\delta CH_4 + 1000} \tag{2}$$

where  $\delta X = \delta^{13}C$ -DIC or  $\delta^{2}H$ -H<sub>2</sub>O and  $\delta CH_{4}$  = the  $\delta^{13}C$ -CH<sub>4</sub> or  $\delta^{2}H$ -CH<sub>4</sub>, respectively: such that  $\alpha_{DIC-CH4}$  = the carbon isotope fractionation factor, and  $\alpha_{H_{2O-CH4}}$  = the hydrogen isotope fractionation factor.

**Rayleigh equations.** Where Rayleigh equations are presented, we use the Rayleigh equation described as:

$$\mathbf{R} = \mathbf{R}_i f^{(\alpha - 1)} \tag{3}$$

where R = change in isotope fractionation relative to the initial value,  $R_i =$  the initial isotope delta values, f = the residual reservoir (e.g. CH<sub>4</sub>). The use of the Rayleigh equation allows simple comparisons of the isotope partitioning as a defined reservoir (e.g. CH<sub>4</sub>) changes. Biogenic methanogenesis tends to operate at thermodynamic equilibrium, rather than being limited by kinetic controls<sup>6</sup> so this approach is considered appropriate here, especially when working at large scales where multiple influences on the carbon and hydrogen isotope may occur. Here we use the Rayleigh equation to simply define the change in carbon and hydrogen isotope partition as CH<sub>4</sub> changes

relative to a CH<sub>4</sub> end member (we do not propose that the  $\delta^{13}$ C-CH<sub>4</sub> or  $\delta^{2}$ H-CH<sub>4</sub> behaviour in the shallow coal measures always follows a simple Rayleigh fractionation process).

**Microbial pathways and Gibbs free energy values.** Gibbs free energy values ( $\Delta G^{\circ}$ ) were calculated for a number of microbial pathways (equations (6–8)) using enthalpy and entropy values for each reaction listed in Stumm and Morgan<sup>88</sup> and corrected for the temperature of each sample (equation 4).

$$\Delta G_{\rm T}^{\rm o} = \Delta H - T \Delta S \tag{4}$$

Where  $\Delta H$  is the change in enthalpy and  $\Delta S$  is the change in entropy for each reaction, and T is the temperature in Kelvin for each sample.

Changes in Gibbs Free Energy  $\Delta G$  were calculated via equation (5).

$$\Delta G = \Delta G_{\rm T}^{\rm o} + RT \ln Q \tag{5}$$

where R is the universal gas constant, T is the temperature in Kelvin and Q is the reaction quotient for each reaction.

For each reaction the activities of reactants and products [activity] were used to calculate Q. The activities of the reactants and products were calculated using *PHREEQC Interactive 3.1.7–9213* using the phreeqc.dat database.

The reaction pathways for  $CO_2$ -reduction (hydrogenotrophic methanogenesis),  $SO_4$  reduction and anaerobic oxidation of  $CH_4$  (AOM) are as follows:

$$4\mathrm{H}_2 + \mathrm{H}^+ + \mathrm{HCO}_3^- \to \mathrm{CH}_4 + 3\mathrm{H}_2\mathrm{O} \tag{6}$$

 $CO_2$  reduction.

 $4H_2 + H^+ + SO_4^{2-} \to HS^- + 4H_2O$ (7)

 $SO_4$  reduction.

$$CH_4 + SO_4^{2-} \rightarrow HS^- + HCO_3^- + H_2O$$
(8)

Anaerobic oxidation of  $CH_4$  (AOM).

Gibbs free energy values ( $\Delta G_T$ ) for the CO<sub>2</sub>-reduction pathways and SO<sub>4</sub> reduction pathways were  $\sim -229 \text{ kJ}$  mol<sup>-1</sup> and -262 kJ mol<sup>-1</sup>, respectively, which is consistent with calculations made for other studies<sup>59,89,90</sup>.

#### References

- 1. Le Mer, J. & Roger, P. Production, oxidation, emission and consumption of methane by soils: A review. *Eur J Soil Biol* **37**, 25–50, doi: http://dx.doi.org/10.1016/S1164-5563(01)01067-6 (2001).
- Kotelnikova, S. Microbial production and oxidation of methane in deep subsurface. *Earth-Sci. Rev.* 58, 367–395, doi: http://dx.doi. org/10.1016/S0012-8252(01)00082-4 (2002).
- Whiticar, M. J. Carbon and hydrogen isotope systematics of bacterial formation and oxidation of methane. *Chem. Geol.* 161, 291–314, doi: http://dx.doi.org/10.1016/S0009-2541(99)00092-3 (1999).
- Cord-Ruwisch, R., Seitz, H.-J. & Conrad, R. The capacity of hydrogenotrophic anaerobic bacteria to compete for traces of hydrogen depends on the redox potential of the terminal electron acceptor. Arch. Microbiol. 149, 350–357, doi: 10.1007/BF00411655 (1988).
- 5. Conrad, R. Quantification of methanogenic pathways using stable carbon isotopic signatures: a review and a proposal. Org. Geochem. 36, 739–752, doi: http://dx.doi.org/10.1016/j.orggeochem.2004.09.006 (2005).
- Conrad, R. Contribution of hydrogen to methane production and control of hydrogen concentrations in methanogenic soils and sediments. FEMS Microbiol. Ecol. 28, 193–202, doi: 10.1111/j.1574-6941.1999.tb00575.x (1999).
- Chanton, J. P., Fields, D. & Hines, M. E. Controls on the hydrogen isotopic composition of biogenic methane from high-latitude terrestrial wetlands. J. Geophys. Res. (G Biogeosci) 111, n/a-n/a, doi: 10.1029/2005JG000134 (2006).
- Kinnaman, F. S., Valentine, D. L. & Tyler, S. C. Carbon and hydrogen isotope fractionation associated with the aerobic microbial oxidation of methane, ethane, propane and butane. *Geochim. Cosmochim. Acta* 71, 271–283, doi: http://dx.doi.org/10.1016/j. gca.2006.09.007 (2007).
- Botz, R., Pokojski, H.-D., Schmitt, M. & Thomm, M. Carbon isotope fractionation during bacterial methanogenesis by CO<sub>2</sub> reduction. Org. Geochem. 25, 255–262, doi: http://dx.doi.org/10.1016/S0146-6380(96)00129-5 (1996).
- Strapoć, D., Schimmelmann, A. & Mastalerz, M. Carbon isotopic fractionation of CH<sub>4</sub> and CO<sub>2</sub> during canister desorption of coal. Org. Geochem. 37, 152–164, doi: http://dx.doi.org/10.1016/j.orggeochem.2005.10.002 (2006).
- 11. Xia, X. & Tang, Y. Isotope fractionation of methane during natural gas flow with coupled diffusion and adsorption/desorption. *Geochim. Cosmochim. Acta* 77, 489–503, doi: http://dx.doi.org/10.1016/j.gca.2011.10.014 (2012).
- Prinzhofer, A. & Pernaton, É. Isotopically light methane in natural gas: bacterial imprint or diffusive fractionation? *Chem. Geol.* 142, 193–200, doi: http://dx.doi.org/10.1016/S0009-2541(97)00082-X (1997).
- Whiticar, M. J., Faber, E. & Schoell, M. Biogenic methane formation in marine and freshwater environments: CO<sub>2</sub> reduction vs. acetate fermentation–Isotope evidence. *Geochim. Cosmochim. Acta* 50, 693–709, doi: http://dx.doi.org/10.1016/0016-7037(86)90346-7 (1986).
- Heimann, A., Jakobsen, R. & Blodau, C. Energetic Constraints on H2-Dependent Terminal Electron Accepting Processes in Anoxic Environments: A Review of Observations and Model Approaches. *Environ. Sci. Technol.* 44, 24–33, doi: 10.1021/es9018207 (2010).
- Penger, J., Conrad, R. & Blaser, M. Stable carbon isotope fractination by methyltrophic methanogenic archea. Applied Environmental Microbiology 78, 7596–7602, doi: 10.1128/AEM.01773-12 (2012).
- Roy, R., Klüber, H. D. & Conrad, R. Early initiation of methane production in anoxic rice soil despite the presence of oxidants. FEMS Microbiol. Ecol. 24, 311–320, doi: http://dx.doi.org/10.1016/S0168-6496(97)00072-X (1997).
- 17. Etiope, G. Natural Gas Seepage. Vol. 1, Ch. 3, 50-52 (Springer International Publishing 2015).

- 18. McIntosh, J. S., M. & Bates, B. In *Technical Workshops for the hydraulic fracturing study: US EPA*, Feb 24–25, 2011. (United States Environmental Protection Agency).
- Aravena, R., Harrison, S. M., Barker, J. F., Abercrombie, H. & Rudolph, D. Origin of methane in the Elk Valley coalfield, southeastern British Columbia, Canada. Chem. Geol. 195, 219–227, doi: 10.1016/s0009-2541(02)00396-0 (2003).
- Hansen, L. K., Jakobsen, R. & Postma, D. Methanogenesis in a shallow sandy aquifer, Rømø, Denmark. Geochim. Cosmochim. Acta 65, 2925–2935, doi: http://dx.doi.org/10.1016/S0016-7037(01)00653-6 (2001).
- Iverach, C. P. et al. Assessing Connectivity Between an Overlying Aquifer and a Coal Seam Gas Resource Using Methane Isotopes, Dissolved Organic Carbon and Tritium. Sci Rep. 5, 15996, doi: 10.1038/srep15996, http://www.nature.com/articles/ srep15996#supplementary-information (2015).
- 22. Huxley, W. J. The hydrogeology, hydrology and hydrochemistry of the Condamine River Valley Alluvium Masters thesis, Queensland Institute of Technology (1982).
- Owen, D. D. R. & Cox, M. E. Hydrochemical evolution within a large alluvial groundwater resource overlying a shallow coal seam gas reservoir. *Sci. Total Environ.* 523, 233–252, doi: http://dx.doi.org/10.1016/j.scitotenv.2015.03.115 (2015).
- Dafny, E. & Silburn, D. M. The hydrogeology of the Condamine River Alluvial Aquifer, Australia: a critical assessment. *Hydrogeol. J.*, 1–23, doi: 10.1007/s10040-013-1075-z (2013).
- 25. QWC. (ed Queensland Water Commission) (Brisbane, 2012).
- Golding, S. D., Boreham, C. J. & Esterle, J. S. Stable isotope geochemistry of coal bed and shale gas and related production waters: A review. Int. J. Coal Geol. 120, 24–40, doi: http://dx.doi.org/10.1016/j.coal.2013.09.001 (2013).
- Baublys, K. A., Hamilton, S. K., Golding, S. D., Vink, S. & Esterle, J. Microbial controls on the origin and evolution of coal seam gases and production waters of the Walloon Subgroup; Surat Basin, Australia. *Int. J. Coal Geol.* 147–148, 85–104, doi: http://dx.doi. org/10.1016/j.coal.2015.06.007 (2015).
- Hamilton, S. K., Golding, S. D., Baublys, K. A. & Esterle, J. S. Stable isotopic and molecular composition of desorbed coal seam gases from the Walloon Subgroup, eastern Surat Basin, Australia. *Int. J. Coal Geol.* 122, 21–36, doi: http://dx.doi.org/10.1016/j. coal.2013.12.003 (2014).
- 29. Draper, J. J. & Boreham, C. J. Geological controls on exploitable coal seam gas distribution in Queensland. *APPEA Journal* 46, 343–366. (2006).
- Cook, A. G. & Draper, J. J. In *Geology of Queensland* (ed. P. A. Jell) Ch. 7, 533–539 (Geological Survey of Queensland, Brisbane, QLD, 2013).
- 31. Jell, P. A., McKellar, J. L. & Draper, J. J. Geology of Queensland: 7.10 Clarence-Moreton Basin. 5 (2013).
- 32. Owen, D. D. R., Millot, R., Négrel, P., Meredith, K. & Cox, M. E. Stable Isotopes of Lithium as Indicators of Coal Seam Gas-bearing Aquifers. *Procedia Earth Planet. Sci.* 13, 278–281, doi: http://dx.doi.org/10.1016/j.proeps.2015.07.065 (2015).
- Richards, L. A., Magnone, D., van Dongen, B. E., Ballentine, C. J. & Polya, D. A. Use of lithium tracers to quantify drilling fluid contamination for groundwater monitoring in Southeast Asia. *Appl. Geochem.* 63, 190–202, doi: http://dx.doi.org/10.1016/j. apgeochem.2015.08.013 (2015).
- 34. Murray, J. P., Rouse, J. V. & Carpenter, A. B. Groundwater contamination by sanitary landfill leachate and domestic wastewater in carbonate terrain: Principal source diagnosis, chemical transport characteristics and design implications. Water Res. 15, 745–757, doi: http://dx.doi.org/10.1016/0043-1354(81)90168-8 (1981).
- Carrillo-Rivera, J. J., Cardona, A. & Edmunds, W. M. Use of abstraction regime and knowledge of hydrogeological conditions to control high-fluoride concentration in abstracted groundwater: San Luis Potosi basin, Mexico. J Hydrol 261, 24–47, doi: 10.1016/ S0022-1694(01)00566-2 (2002).
- 36. Hem, J. D. Study and interpretation of the chemical characteristics of natural water: USGS Water-Supply Paper 2254. Third edn (United States Geological Survey, 1985).
- Wrenn, B. A. et al. Nutrient transport during bioremediation of contaminated beaches: Evaluation with lithium as a conservative tracer. Water Res. 31, 515–524, doi: http://dx.doi.org/10.1016/S0043-1354(96)00304-1 (1997).
- Humez, P., Mayer, B., Nightingale, M., Becker, V., Kingston, A., Taylor, S., Bayegnak, G., Millot, R. & Kloppmann, W. Redox controls on methane formation, migration and fate in shallow aquifers. *Hydrol. Earth Syst. Sci. Discuss.*, doi: 10.5194/hess-2016-85, in review, 2016 (2016).
- Aravena, R. & Wassenaar, L. I. Dissolved organic carbon and methane in a regional confined aquifer, southern Ontario, Canada: Carbon isotope evidence for associated subsurface sources. *Appl. Geochem.* 8, 483–493, doi: http://dx.doi.org/10.1016/0883-2927(93)90077-T (1993).
- Wassenaar, L., Aravena, R., Hendry, J. & Fritz, P. Radiocarbon in Dissolved Organic Carbon, A Possible Groundwater Dating Method: Case Studies From Western Canada. Water Resour. Res. 27, 1975–1986, doi: 10.1029/91WR00504 (1991).
- Flores, R. M., Rice, C. A., Stricker, G. D., Warden, A. & Ellis, M. S. Methanogenic pathways of coal-bed gas in the Powder River Basin, United States: The geologic factor. *Int. J. Coal Geol.* 76, 52–75, doi: http://dx.doi.org/10.1016/j.coal.2008.02.005 (2008).
- Blair, N. E. & Carter, W. D. Jr. The carbon isotope biogeochemistry of acetate from a methanogenic marine sediment. Geochim. Cosmochim. Acta 56, 1247–1258, doi: http://dx.doi.org/10.1016/0016-7037(92)90060-V (1992).
- Riveros-Iregui, D. A. & King, J. Y. Isotopic evidence of methane oxidation across the surface water-ground water interface. Wetlands 28, 928–937, doi: 10.1672/07-191.1 (2008).
- 44. Moura, J. M. S. *et al.* Spatial and seasonal variations in the stable carbon isotopic composition of methane in stream sediments of eastern Amazonia. *Tellus B* **60**, 21–31, doi: 10.1111/j.1600-0889.2007.00322.x (2008).
- Tsunogai, U., Yoshida, N. & Gamo, T. Carbon isotopic evidence of methane oxidation through sulfate reduction in sediment beneath cold seep vents on the seafloor at Nankai Trough. *Mar. Geol.* 187, 145–160, doi: http://dx.doi.org/10.1016/S0025-3227(02)00263-3 (2002).
- Ahmed, M. & Smith, J. W. Biogenic methane generation in the degradation of eastern Australian Permian coals. Org. Geochem. 32, 809–816, doi: http://dx.doi.org/10.1016/S0146-6380(01)00033-X (2001).
- Chanton, J. P., Chasar, L. C., Glaser, P. & Siegel, D. In Stable Isotopes and Biosphere-Atmosphere Interactions, Physiol. Ecol. Ser. (eds L. B. Flanagan, J. R. Ehleringer & D. E. Pataki) Ch. 6, 85–105 (Elsevier, 2005).
- Waldron, S., Lansdown, J. M., Scott, E. M., Fallick, A. E. & Hall, A. J. The global influence of the hydrogen iostope composition of water on that of bacteriogenic methane from shallow freshwater environments. *Geochim. Cosmochim. Acta* 63, 2237–2245, doi: http://dx.doi.org/10.1016/S0016-7037(99)00192-1 (1999).
- Papendick, S. L. et al. Biogenic methane potential for Surat Basin, Queensland coal seams. Int. J. Coal Geol. 88, 123–134, doi: http:// dx.doi.org/10.1016/j.coal.2011.09.005 (2011).
- McIntosh, J. C., Grasby, S. E., Hamilton, S. M. & Osborn, S. G. Origin, distribution and hydrogeochemical controls on methane occurrences in shallow aquifers, southwestern Ontario, Canada. *Appl. Geochem.* 50, 37–52, doi: http://dx.doi.org/10.1016/j. apgeochem.2014.08.001 (2014).
- Quillinan, S. A. & Frost, C. D. Carbon isotope characterization of powder river basin coal bed waters: Key to minimizing unnecessary water production and implications for exploration and production of biogenic gas. *Int. J. Coal Geol.* 126, 106–119, doi: http://dx.doi.org/10.1016/j.coal.2013.10.006 (2014).
- 52. Clark, I. & Fritz, P. Environmental Isotopes in Hydrogeology. (CRC Press LLC, 1997).
- Lovley, D. R. & Klug, M. J. Sulfate Reducers Can Outcompete Methanogens at Freshwater Sulfate Concentrations. Appl. Environ. Microbiol. 45, 187–192 (1983).

- Stams, A. J. M. *et al.* Metabolic interactions in methanogenic and sulfate-reducing bioreactors. *Water Sci. Technol.* 52, 13–20 (2005).
   Kato, M. T., Field, J. A. & Lettinga, G. The anaerobic treatment of low strength wastewaters in UASB and EGSB reactors. *Water Sci.*
- Technol. **36**, 375–382, doi: http://dx.doi.org/10.1016/S0273-1223(97)00545-3 (1997). 56. Walker, G. R. & Mallants, D. Methodologies for Investigating Gas in Water Bores and Links to Coal Seam Gas Development.
- (Australia, 2014). 57. Feitz, A. J. *et al.* Geoscience Australia and Geological Survey of Queensland Surat and Bowen Basins Groundwater Surveys
- Hydrochemistry Dataset (2009–2011). (Canberra Australia, 2014).
- Lovley, D. R. & Goodwin, S. Hydrogen concentrations as an indicator of the predominant terminal electron-accepting reactions in aquatic sediments. *Geochim. Cosmochim. Acta* 52, 2993–3003, doi: http://dx.doi.org/10.1016/0016-7037(88)90163-9 (1988).
- Strapoć, D. et al. Methane-producing microbial community in a coal bed of the Illinois basin. Appl. Environ. Microbiol. 74, 2424–2432, doi: 10.1128/AEM.02341-07 (2008).
- Egozcue, J. J. & Pawlowsky-Glahn, V. Groups of Parts and Their Balances in Compositional Data Analysis. Math. Geol. 37, 795–828, doi: 10.1007/s11004-005-7381-9 (2005).
- Smemo, K. A. & Yavitt, J. B. Evidence for Anaerobic CH<sub>4</sub> Oxidation in Freshwater Peatlands. *Geomicrobiol. J.* 24, 583–597, doi: 10.1080/01490450701672083 (2007).
- Games, L. M., HayesRobert, J. M. & Gunsalus, P. Methane-producing bacteria: natural fractionations of the stable carbon isotopes. Geochim. Cosmochim. Acta 42, 1295–1297, doi: http://dx.doi.org/10.1016/0016-7037(78)90123-0 (1978).
- Krzycki, J. A., Kenealy, W. R., DeNiro, M. J. & Zeikus, J. G. Stable Carbon Isotope Fractionation by Methanosarcina barkeri during Methanogenesis from Acetate, Methanol, or Carbon Dioxide-Hydrogen. *Appl. Environ. Microbiol.* 53, 2597–2599 (1987).
- Balabane, M., Galimov, E., Hermann, M. & Létolle, R. Hydrogen and carbon isotope fractionation during experimental production of bacterial methane. Org. Geochem. 11, 115–119, doi: http://dx.doi.org/10.1016/0146-6380(87)90033-7 (1987).
- Alperin, M. J., Blair, N. E., Albert, D. B., Hoehler, T. M. & Martens, C. S. Factors that control the stable carbon isotopic composition of methane produced in an anoxic marine sediment. *Global Biogeochem. Cycles* 6, 271–291, doi: 10.1029/92GB01650 (1992).
- 66. Bilek, R. S., Tyler, S. C., Sass, R. L. & Fisher, F. M. Differences in CH<sub>4</sub> oxidation and pathways of production between rice cultivars deduced from measurements of CH<sub>4</sub> flux and δ13C of CH<sub>4</sub> and CO<sub>2</sub>. *Global Biogeochem. Cycles* 13, 1029–1044, doi: 10.1029/1999GB900040 (1999).
- Sass, R. L., Fisher, F. M., Lewis, S. T., Jund, M. F. & Turner, F. T. Methane emissions from rice fields: Effect of soil properties. *Global Biogeochem. Cycles* 8, 135–140, doi: 10.1029/94GB00588 (1994).
- 68. Oades, J. M. The Retention of Organic Matter in Soils. Biogeochemistry 5, 35-70 (1988).
- Kits, K. D., Klotz, M. G. & Stein, L. Y. Methane oxidation coupled to nitrate reduction under hypoxia by the Gammaproteobacterium Methylomonas denitrificans, sp. nov. type strain FJG1. *Environ. Microbiol.* 17, 3219–3232, doi: 10.1111/1462-2920.12772 (2015).
- 70. Ettwig, K. F. *et al.* Denitrifying bacteria anaerobically oxidize methane in the absence of Archaea. *Environ. Microbiol.* **10**, 3164–3173, doi: 10.1111/j.1462-2920.2008.01724.x (2008).
- Conrad, R., Klose, M., Claus, P. & Enrich-Prast, A. Methanogenic pathway, 13C isotope fractionation, and archaeal community composition in the sediment of two clear-water lakes of Amazonia. *Limnol. Oceanogr.* 55, 689–702, doi: 10.4319/lo.2010.55.2.0689 (2010).
- Schlegel, M. E., McIntosh, J. C., Bates, B. L., Kirk, M. F. & Martini, A. M. Comparison of fluid geochemistry and microbiology of multiple organic-rich reservoirs in the Illinois Basin, USA: Evidence for controls on methanogenesis and microbial transport. *Geochim. Cosmochim. Acta* 75, 1903–1919, doi: http://dx.doi.org/10.1016/j.gca.2011.01.016 (2011).
- Morgenstern, U. & Taylor, C. B. Ultra low-level tritium measurement using electrolytic enrichment and LSC. Isotopes Environ. Health Stud. 45, 96–117, doi: 10.1080/10256010902931194 (2009).
- Hocking, M. & Kelly, B. F. J. Groundwater recharge and time lag measurement through Vertosols using impulse response functions. J Hydrol 535, 22–35, doi: http://dx.doi.org/10.1016/j.jhydrol.2016.01.042 (2016).
- 75. KCB. Central Condamine alluvium, stage III: detailed water balance: Final Report. (Toowoomba, Queensland, 2011).
- Hendry, M. J., Ranville, J. R., Boldt-Leppin, B. E. J. & Wassenaar, L. I. Geochemical and transport properties of dissolved organic carbon in a clay-rich aquitard. *Water Resour. Res.* 39, n/a-n/a, doi: 10.1029/2002WR001943 (2003).
- Sharma, M. L. & Hughes, M. W. Groundwater recharge estimation using chloride, deuterium and oxygen-18 profiles in the deep coastal sands of Western Australia. J Hydrol 81, 93–109, doi: http://dx.doi.org/10.1016/0022-1694(85)90169-6 (1985).
- Dlugokencky, E. J., Nisbet, E. G., Fisher, R. & Lowry, D. Global atmospheric methane: budget, changes and dangers. *Philos. Trans. Roy. Soc. London Ser. A* 369, 2058–2072, doi: 10.1098/rsta.2010.0341 (2011).
- 79. Stalker, L. Methane origins and behaviour (Commonwealth Scientific and Industrial Research Organisation, Australia, 2013).
- Khalil, M. A. K. Atmospheric Methane: Sources, Sinks, and Role in Global Change. 199–229 (Springer-Verlag, 1991).
   Puls, R. W. & Barcelona, M. J. LOW-FLOW (MINIMAL DRAWDOWN) GROUND-WATER SAMPLING PROCEDURES (United
- St. Puis, R. W. & Barceiona, M. J. LOW-FLOW (MINIMAL DRAWDOWN) GROUND-WATER SAMPLING PROCEDORES (United States Environmental Protection Agency, 1996).
- Barcelona, M. J., Varljen, M. D., Puls, R. W. & Kaminski, D. Ground water purging and sampling methods: History vs. hysteria. Ground Water Monitoring & Remediation 25, 52–62, doi: 10.1111/j.1745-6592.2005.0001.x (2005).
- 83. Noble, R. R. P., Gray, D. J. & Gill, A. J. Field guide for mineral exploration using hydrogeochemical analysis (Bentley, Western Australia, 2011).
- Peacock, M., Freeman, C., Gauci, V., Lebron, I. & Evans, C. D. Investigations of freezing and cold storage for the analysis of peatland dissolved organic carbon (DOC) and absorbance properties. *Environmental Science: Processes & Impacts* 17, 1290–1301, doi: 10.1039/C5EM00126A (2015).
- Palarea-Albaladejo, J. & Martín-Fernández, J. A. zCompositions—R package for multivariate imputation of left-censored data under a compositional approach. *Chemometrics Intellig. Lab. Syst.* 143, 85–96, doi: http://dx.doi.org/10.1016/j.chemolab.2015.02.019 (2015).
- Egozcue, J. J., Pawlowsky-Glahn, V., Mateu-Figueras, G. & Barceló-Vidal, C. Isometric Logratio Transformations for Compositional Data Analysis. *Math. Geol.* 35, 279–300, doi:10.1023/A:1023818214614 (2003).
- Comas-Cufí, M. & Thió-Henestrosa, S. In CoDaWork'11: 4th International Workshop on Compositional Data (eds Egozcue J. J., Tolosana-Delgado R. & Ortego M. I.) (2011).
- Stumm, W. & Morgan, J. J. Aquatic chemistry: chemical equilibria and rates in natural waters. Third Edition edn, 1022 (John wiley and Sons, Inc, 1996).
- Lin, H.-T. et al. Dissolved hydrogen and methane in the oceanic basaltic biosphere. Earth. Planet. Sci. Lett. 405, 62–73, doi: http:// dx.doi.org/10.1016/j.epsl.2014.07.037 (2014).
- Ozuolmez, D. et al. Methanogenic archaea and sulfate reducing bacteria co-cultured on acetate: teamwork or coexistence? Frontiers in Microbiology 6, 492, doi: 10.3389/fmicb.2015.00492 (2015).
- Cox, M. E., James, A., Hawke, A. & Raiber, M. Groundwater Visualisation System (GVS): A software framework for integrated display and interrogation of conceptual hydrogeological models, data and time-series animation. J Hydrol 491, 56–72, doi: http:// dx.doi.org/10.1016/j.jhydrol.2013.03.023 (2013).
- 92. Lane, W. B. Progress Report on Condamine Underground Investigation to December 1978 (Brisbane, Queensland, 1979).
- 93. Crosbie, R. S. et al. (ed. CSIRO Water for a Healthy Country Flagship) (Australia 2012).

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#### **Author Contributions**

Experimental design regarding CH<sub>4</sub>, chlorine isotopes and DOC was carried out by D.D.R.O., O.S. and R.A. Data interpretation and manuscript preparation were performed via contributions from all co-authors.

#### Additional Information

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