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Health risk assessment for heavy metal accumulation in leafy vegetables grown on tannery effluent contaminated soil

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ABSTRACT

Accumulation of metals (Cr, Zn, Ni, Cd, and Cu) in leafy vegetables cultivated on tannery effluent contaminated soil and agricultural land soil were determined with an Atomic Absorption Spectrophotometer (AAS). The values of risk factors for the human population were studied, where metals were transferred from tannery effluent to plants via effluent contaminated soil and finally, transmitted to human body through the consumption of these metal accumulated leafy vegetables. Leafy vegetables, namely Stem amaranths (Amaranthus lividus), Spinach (Spinacia oleracea), Red amaranths (Amaranthus gangeticus), Jute mallows (Corchorus capsularis), Water spinach (Ipomoea aquatica), and Malabar spinach (Basella alba) were cultivated on the soils collected from downstream of Hazaribagh tannery area and Keraniganj agricultural land. The study revealed that the metal contents in contaminated soil exceeded the permissible limits recommended by WHO/DoE. Tannery effluent contaminated soil was found more polluted than the agricultural land soil. Metal contents in leafy vegetables cultivated on contaminated soil were higher than that of agricultural soil and exceeded the permissible limit, particularly in the case of Cr (125.50-168.99 mg/kg Dw) and Cd (0.19-0.83 mg/kg Dw). Metal content order was found as Cr>Zn>Ni>Cu>Cd for contaminated soil and Zn>Cr>Cu>Ni>Cd for agricultural land soil. The metal accumulation and translocation were found in vegetables in the order of Spinach>Water spinach>Malabar spinach>Jute mallows>Red amaranths>Stem amaranths. The analyses also revealed that the metal translocation rate in the plants of contaminated soil was higher than that of non-contaminated agricultural soil. The values of each risk index exceeded 1 in case of vegetables cultivated in contaminated soil. Therefore, the possible threat of chronic and carcinogenic diseases emerged if those polluted vegetables would be consuming as daily diet.

1. Introduction

The desire for a luxurious life with the emerging benefits from industrialization might bring comfort in our daily lifestyle, however it has led the environment towards successive deterioration. Heavy metals discharged from various industries play a critical role in this context and bring about a serious threat to public health [1,2]. Tanneries have been recognized as a possible source of producing heavy metal load to the environment even though having a significant contribution to the economic development and employment in many countries [3,4]. Leather processing from animal hides/skins includes a series of mechanical and chemical operations such as soaking, liming/unhairing, de-liming, bating, pickling, skin degreasing, tanning, post-tanning, and dyeing of

raw hides/skins using different types of chemicals e.g., sodium hydroxide, sodium hypochlorite, enzymes, lime, chlorides, sulfuric acid, formic acid, ammonium salts, different metallic salts, and organic chemicals, etc. [5]. A large amount of chromium (Cr) salt named basic chromium sulfate (BCS) is used in the tanning process to provide better hydrothermal stability and some unique properties to the tanned leathers [6]. Heterocomplex iron (Fe) salts are used instead of basic chromium sulfate (BCS) to achieve sufficient waterproofing effect and strengthen the loose fibers through higher exhaustion of tannage within a short tanning span [7]. Furthermore, organometallic salts of different heavy metals such as chromium (Cr) cadmium (Cd), copper (Cu), lead (Pb), etc. are extensively used as coloring agents and mordant during leather dyeing and in other post-tanning and finishing operations to

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Fig. 1. Locations of the (a) Tannery effluent contaminated and (b) Non-contaminated agricultural reference soil sites. Adapted upon modification from Ali et al. [43].

make the leather suitable for making various leather products such as bags, shoes, wallets, upholstery furniture, etc. [8,9]. As a result, a substantial amount of different heavy metals like Cr, Cd, Cu, Ni, Pb, Ba, As, Fe, etc. are discharged with tannery wastewaters [10]. Besides, the dumping of tannery solid wastes in land filling, use of treated tannery effluents as irrigation water, and burning of leather wastes are some common pathways of heavy metal transmission into natural resources [11–13].

The ultimate fate of these metals is to deposit into the soil and water, might get transferred to different plant tissues (stem, root, leaf, etc.) as some of them act as essential micro-nutrients for plant's growth, and finally introduced to human food chain if those plants are consumed [14, 15]. Previous studies have stated that heavy metal-containing industrial wastewaters are responsible for bringing about the toxicological effect on natural elements (soil, water) and other living species [16-18]. Singh and Kalamdhad [19] showed that heavy metals (e.g., Cu, Ni, Cd, Zn, Cr, and Pb) reduce the physicochemical parameters (e.g., organic matter, clay contents, and pH) of soil and show toxic symptoms towards soil living biota by decreasing soil microorganism (leguminous bacteria) number that hamper their enzymatic and metabolic activities, and stimulate the formation of reactive oxygen species (ROS), which causes lethal consequences on other non-vertebrates (earthworms) and reptiles (snakes, lizards, rats) living in soil, and in fishes and other aquatic species when they mix up with surface water through agricultural runoff. Moreover, the bioavailability of some heavy metals (Cu, Zn) causes a degradation in the aquaculture population through physical deformation in water species and cellular damages in both aquatic flora

and fauna [20]. Apart from environmental pollution, some of the heavy metals are also reported as the topmost hazardous elements that result in deleterious influences on plant materials and human health by hampering normal cellular activities. Scholars have observed that the plants grown around industrial zones or waste dumping yards uptake more heavy metals or trace elements as compared to those plants grown around nonindustrial areas; store them in their vegetative and reproductive parts; and exhibit toxicity symptoms e.g., necrotic lesions, and chlorosis as a result of prolonged exposure [21,22]. These heavy metals can easily accumulate to different plant organs (e.g., roots, leaves, seeds) and show toxic effects in long-term exposure. For instance, the toxicity of chromium and cadmium to plant brings about less seed germination, alteration in different organs (root, stem, leaves), shackles in development and metabolic performance, inhibit photosynthesis ratio through chlorosis, hinder the transportation of nutrients and water, and advances cellular decay [23,24]. Although copper, zinc, lead is documented as essential micronutrients for plant's development, the excess deposition of these metals causes extreme phytotoxic effects like stunting, leaf discoloration, necrosis, curling of young leaves, death of leaf tips, lipid peroxidation, inhibition in ATP production during respiration, DNA damage due to excessive ROS generation and many more [25-27]. Consequently, consumption of metal contaminated vegetables can cause a negative impact in the food chain and may extend a negative influence on human health which can be an obstacle to ensuring safer foods and also achieving the UN's Sustainable Development Goal-2 (Zero hunger), Goal-3 (Good health and well-being) and Goal-12 (Responsible consumption and production) for many countries [28].

Many toxicological reports have proclaimed that the daily intake of metal contaminated vegetables at an excess rate can cause carcinogenic, mutagenic, neurotoxic, and teratogenic effects in the human body that result in severe health problems; like anemia, hypertension, malnutrition, cancer, organ disability, infertility, cardiovascular diseases, metabolic and psychological disorder which may lead to death at prolonged exposure [29-31]. The risk factors related to heavy metals' contamination in food mainly arise from agro-based products like rice, crops, and vegetables and have affected many regions around the globe, exclusively in developing countries like Bangladesh [32,33]. Although vegetables are recognized as one of the most popular daily diets due to their enrichment with nutritious elements (e.g., fiber, minerals, and vitamins), the metal uptake rate of leafy vegetables was found to be the highest among all crops and act as a bio-indicator of soil pollution [34, 35]. Therefore, particular studies on metal contamination in leafy vegetables must be prioritized to draw attention to such health risks and raise public awareness for its remedy since only a few publications are available on this issue in the context of Bangladesh [29,36–38].

Concerning this susceptible situation, an attempt had been taken to determine the amount of selected heavy metals in common leafy vegetables cultivated on tannery effluents contaminated soil and agricultural land soil to investigate the gateway of heavy metal accumulation in the human body. Specifically, this study was envisioned at

- a) Assessment of metal content in both tannery effluent affected soil and agricultural land soil.
- b) Measurement of the metal translocation rate from soil to plant.
- c) Evaluation of the potential health risk factors of metal uptake through vegetable consumption with respect to risk indices (e.g., daily metal intake, health risk index, non-carcinogenic risk).

2. Experimental

2.1. Study area

The study was conducted with soils collected from canal side at the downstream of Hazaribagh tannery area under Dhaka South City Corporation (DSCC) and Keraniganj agricultural area, Bangladesh. The areas are located at a latitude in between 2343' and 2344' N and longitude in between 9021' and 9022.5' E [39]. Approximately 185 of the total 220 tanneries had been situated around 4 km² areas, with an average daily production of 220 metric tons of finished leathers after their foundation (1965) to continuous development, before the tanneries were completely relocated to Hemayetpur, Savar by the fiscal year 2017–18 [40]. During production days, the tanneries daily produced around 20000 m³ of tannery effluents from various chemical processes, and they were directly discharged to the Buriganga River that contributed hazardous impact on river ecosystem and water quality for several decades [41,42].

The sampling points for soil collection are shown in Fig. 1. The main discharge point of tannery effluents located near the Beribadh (a) was selected for collecting contaminated soil samples. It was under the direct influence of tannery effluents for several decades and directly linked with the Buriganga River by a narrow canal. In the same way, reference soil samples were collected from an agricultural land located at Keraniganj (b), outskirts of the city which was comparatively contamination

| Table 1 |
|--|
| Instrumental operating condition for metal analysis. |

free than the site (a).

2.2. Preparation of soil bed and cultivation of vegetables

Ten soil samples (n = 10) were collected from each of the tannery effluent contaminated area and non-contaminated agricultural area. About 10 kg soil was collected from each point up to 15 cm depth (0-15 cm) with a stainless-steel auger. A minimum distance of 5 m was maintained between two sample points. All the soil samples from each category i.e., tannery effluent contaminated area and non-contaminated agricultural area were dried, grinded, mixed, and homogenized separately before preparing the seed beds. Thirty parallel seed beds were prepared with those tannery effluent contaminated soil and noncontaminated agricultural soils in labeled plastic pots prior to cultivating the leafy vegetables (number of set, n = 5/vegetable). Seeds of different vegetable plants were collected from Bangladesh Agricultural Development Corporation (BADC) in labeled plastic bottles. The collection bottles were dipped overnight into diluted nitric acid solution, washed with running water for several times to reduce surplus acid, and dried properly to disinfect them before collecting the seeds.

Six leafy vegetables viz., Stem amaranths (*Amaranthus lividus*), Spinach (*Spinacia oleracea*), Red amaranths (*Amaranthus gangeticus*), Jute mallows (*Corchorus capsularis*), Water spinach (*Ipomoea aquatica*), and Malabar spinach (*Basella alba*) were cultivated in the collected soils in labeled pots following pot farming method, suggested by Dobson [44]. No chemical fertilizer or plant protection product was used in cultivation of vegetables to avoid external heavy metal contamination. In addition, only harvested rainwater was used during watering plants. All plants were cultivated in the same manner under natural environment. A proper care was taken during cultivation of the vegetables and preparing edible parts of vegetable for analysis.

2.3. Sample preparation

Non-soil particles like stones, wooden pieces, debris, etc. were removed from each soil sample manually. About 100 g of soil sample was dried at 95 °C in an air woven for 48 h and ground into powders. Soil powders were sieved with a nylon mesh (2 mm) to remove non-granular matters. The samples were then homogenized and stored in a labeled polythene sampling bags for wet digestion at -15 °C and to prevent any chemical change.

300 g edible parts of cultivated vegetable species were collected randomly and packed in polyethylene bags. The collected edible top parts and leaves were then washed with running water and subsequently dipped into diluted HCl (0.01 N) for few minutes. Then the samples were thoroughly washed with deionized water for removing airborne pollutants and cut into small pieces with a sharp stainless-steel knife. The cut pieces were completely oven dried at 70–80 °C, crushed, sieved, and stored in airtight labeled plastic containers until experiment.

2.4. Sample digestion

The digestion of each vegetable and soil sample was carried out following the standard method. A known amount (0.5 g) of each vegetable and soil samples was digested on a hot plate (Jeo Tech TM-14SB) at 80 °C with 14 ML concentrated acid mixture (HNO₃: H₂SO₄: HClO₄ = 5:

| Metals | Wavelength (nm) Slit width (nm) | | Wavelength (nm) Slit width (nm) Lamp type Lamp current (nm) | | Detection Level (mg/L) | Range | |
|--------|---------------------------------|-----|---|---|------------------------|-------|--|
| Cr | 357.87 | 0.7 | HCL | 7 | 0.31 | 5.0 | |
| Zn | 213.86 | 0.7 | EDL/HCL | 5 | 0.084 | 1.0 | |
| Ni | 232.00 | 0.2 | Iron containing ML | 4 | 1.7 | 2.0 | |
| Cd | 228.80 | 0.7 | EDL/HCL | 4 | 0.11 | 2.0 | |
| Cu | 324.75 | 0.7 | Nickel/Iron containing ML | 4 | 0.45 | 5.0 | |

Table 2

Analytical results obtained on certified reference material IAEA 433.

| Metals | Certificate Value (mg/kg) | Observed Value (mg/kg) | Precision CV (%) | Recovery (%) |
|--------|------------------------------|---------------------------|---------------------|-----------------|
| Cr | 136 | 133 | 2.205882 | 97.79412 |
| Zn | 101 | 93 | 7.920792 | 92.07921 |
| Cu | 30.8 | 29 | 5.844156 | 94.15584 |
| Ni | 39.4 | 36.01 | 8.604061 | 91.39594 |
| Cd | 0.153 | 0.145 | 5.228758 | 94.77124 |

1: 1) to achieve a transparent solution. The solutions were filtered with Whattman No. 41 filter paper and the further diluted to 25 ML with DI water and stored at ambient temperature.

2.5. Metal analysis

2.5.1. Instrument and operating condition

An Atomic Absorption Spectrophotometer (Model: Varian AA240FS) was subjected to measure the concentration of heavy metals in soil and vegetable samples. The functional parameters viz. alignment of lamp and burner, adjustment of slit width, and wavelength were optimized by following the instrument instruction to obtain results at the highest signal intensity. The flame condition was kept favorable by using acetylene and controlling the airflow rate. The operating conditions for the quantification of Cr, Zn, Ni, Cd, and Cu in AAS were recorded as showed in Table 1.

2.5.2. Machine calibration

The metal assessment in examined samples was conducted by using the calibrated curves for each heavy metal. Standard solution of each metal was procured from Merck, India. Standard solutions were diluted with deionized water to make appropriate working standard solution for metal detection. Samples were aspirated through a nebulizer and the absorbance was measured with a blank reference. The calibration curve was obtained using standard samples (containing 0.5, 1.0, 1.5, 2.0, and 2.5 mg/L standard solutions for each metal). The correlation coefficient had determined, and the samples had to be diluted many folds to keep the results in the analytical range.

2.5.3. Quality assurance and quality control

Assessment of contamination and reliability of data was done as a part of quality control measure. Blank, quality control standard samples were examined after every ten samples for ensuring that obtained results are within the range. The Certified Reference Materials (CRM) for every metal obtained from the International Atomic Energy Agency (IAEA) were used to check for accuracy and precision of analysis of each metal in soil matrix and for vegetables samples spike samples are used to check recovery. A good agreement was observed between analytical results and certified values of the CRM, with the percentage recoveries ranging from 91% to 97%. The precision of the analytical procedures expressed as the coefficient of variation (CV) was within 10% for all the metals analyzed by AAS (Table 2). All analyses were carried out in triplicate, and the results were expressed by the mean value.

2.6. Data analysis

2.6.1. Pollution load index

The pollution load index (PLI) is used to determine the soil contamination factor by each heavy metal [45]. The following equation was used to calculate PLI value of the contaminated site [46]:

$$PLI = \frac{C_c}{C_r}$$
(1)

Here, C_c = heavy metal's concentration in contaminated soil and C_r = heavy metal's concentration in reference soil. The categorization of

PLI values is given below [46]:

PLI 0–1 = no pollution; PLI 1–2 = uncontaminated to moderately contaminated; PLI 2–3 = moderately contaminated; PLI 3-4 = moderately contaminated to highly contaminated; PLI 4-5 = highly contaminated; and PLI > 5 = very highly contaminated.

2.6.2. Metal transfer factor

The metal transfer factor (MTF) represents the possible bioavailability of a particular metal at a specific plant part [47]. It is the ratio of metal concentration in plant tissue (mg/kg Dw) e.g., stem, leaf, root, etc. to metal concentration in soil (mg/kg Dw) [48]. Metal transfer factor of Cr, Zn, Ni, Cd, and Cu was calculated by using the following equation [49]:

$$MTF = \frac{C_{Plant}}{C_{Soil}}$$
(2)

Here, C_{plant} = metal concentration in plant tissue (mg/kg Dw), C_{soil} = metal concentration in soil (mg/kg Dw). The translocation of each metal from soil to plant was measured for all vegetable samples that were grown in both the contaminated and non-contaminated agricultural soil.

2.6.3. Metal pollution index

The metal pollution index (MPI) is the assessment of the overall metal load in each vegetable growing in at a particular site. It is also known as the geometric mean of the total concentration of all metals in edible parts of the vegetables. To calculate MPI of the vegetable samples (grown in both the contaminated and non-contaminated agricultural soil), the following equation was used [13]:

$$MPI = (C_1 \times C_2 \times . \times C_n)^{\frac{1}{n}}$$
(3)

Here, $C_n =$ concentration of 'n' number of metals in the samples (Here, n = 5).

2.6.4. Health risk assessment

2.6.4.1. Daily Intake of Metals. Daily intake of metals (DIM) represents the relative phyto-availability of metals [50]. The following equation was used to determine the daily intake of metals through vegetable consumption [51]:

$$DIM = \frac{\left(C_{metal} \times C_{factor} \times C_{int \ ake}\right)}{B_{weight}} \tag{4}$$

Where, C_{metal} = the concentration of heavy metals in vegetables (mg/kg), C_{factor} = conversion factor to convert the fresh weight of vegetables to dry weight, C_{intake} = ingestion rate.

In this study, the ingestion rate for leafy vegetables was considered 0.0385 kg/person/day [52], conversion factor (0.085) was used for the selected vegetable samples [53], and B_{weight} = the average body weight, considered 70 kg (adults) [54].

2.6.4.2. *Health Risk Index*. Health risk index (HRI) represents the susceptible health risk through continuous exposure to heavy metal that further leads to a chronic hazard [55]. The index is a numerical value where higher numbers represent higher risk. The following equation was used to calculate HRI for each leafy vegetable [56]:

$$HRI = \frac{DIM}{R_f D}$$
(5)

The R_fD values for Cr, Zn, Ni, Cd and Cu were 1.5, 0.3, 0.02, 0.001, 0.04 mg/kg/day respectively [57]. The HRI value less than 1 indicates population safety whereas it may turn to a concerning matter if HRI value is in between 1 and 5 [58].

Health risk index may be carcinogenic and non-carcinogenic risk.

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Table 3

Concentration of heavy metals (Mean \pm SD) in soil samples (mg/kg Dw).

| Metals | Sites | | Permissible levels in soil* | Permissible levels in soil** | t-test*** | PLI |
|--------|-------------------------------------|------------------------------------|-----------------------------|------------------------------|-----------|-------|
| | Effluent contaminated soil | Non-contaminated agricultural soil | | | | |
| Cr | 738.62 ± 0.06 | $\textbf{48.96} \pm \textbf{0.11}$ | 100 | 100 | 11.4 | 15.09 |
| Zn | 99.13 ± 0.003 | 64.21 ± 0.09 | 300 | 200 | 3.29 | 1.543 |
| Ni | 17.62 ± 0.40 | 10.35 ± 0.5 | 50 | 50 | 1.88 | 1.702 |
| Cd | $\textbf{7.21} \pm \textbf{0.10}$ | 3.86 ± 0.02 | 3 | 1.5 | 2.04 | 1.87 |
| Cu | $\textbf{22.28} \pm \textbf{0.005}$ | 17.20 ± 1.08 | 140 | 60 | 1.33 | 1.29 |

* WHO (2001) [63],

** DOE (2015) [64],

Non-carcinogenic risk is represented as Hazard quotient (HQ) [59]. It is the ratio of the potential exposure to a substance and the level at which no adverse effects are expected. The HQ is calculated according to the following equation [60]:

$$HQ = \frac{EDI}{R_f D} \tag{6}$$

$$EDI = \frac{C \times IR \times EF \times ED}{BW \times AT}$$
(7)

Here, C= metal concentration in vegetable (mg/kg Dw), IR= Ingestion rate of leafy vegetables (0.0385 kg/person/day), EF=Exposure frequency (365 days/year), ED= Exposure duration= 72 years (adults) [61]. R_fD= oral reference dose, BW= average body weight (70 kg), AT= average exposure time for non-carcinogenic effects= (365 days/year × number of exposure years (72)) = 26280 days.

The following equation was used to estimate hazard index (HI) [60]:

$$HI = \sum_{k=1}^{n} HQ_k \tag{8}$$

The HI value <1 indicates less health hazard, while the value of 1–5 indicates a concerned health hazard level and HI>10 indicates chronic health hazard [62].

2.7. Statistical analysis

The recorded data were statistically analyzed by SPSS version 12. Descriptive analysis was represented to narrate the scenario of soil pollution and metal translocation in plants as compared to standards. Paired t-tests were carried out to determine the significant differences between heavy metals concentration in soil and vegetable samples of both sites. A one-way analysis of variance (ANOVA-1) test was conducted to estimate the homogeneity of the recorded data and validation of results.

3. Results and discussion

3.1. Heavy metals in soil

Heavy metal concentrations in soil collected from Keranigonj agricultural land and downstream of Hazaribagh tannery area (Veribadh canal side) were presented in Table 3. The metal content in the soil of Hazaribagh canal side was higher than that of Keranigonj agricultural land site and the concentration of Cr (738.62 \pm 0.06 mg/kg Dw) and Cd (7.21 \pm 0.10 mg/kg Dw) was beyond the permissible limit. However, the metal content in agricultural soil was found less than the threshold limit except Cd which was slightly higher [63,64]. The descending order of metal contents was Cr>Zn>Ni>Cu>Cd for contaminated soil and Zn>Cr>Cu>Ni>Cd for agricultural soil. The PLI pointed out a significant variation in the accumulation of chromium in contaminated soil (Table 3) and exceeded the tolerable limit in following order Cr>Cd>Ni>Zn>Cu. The quantity of chromium and cadmium, including

Table 4

| Heavy metal values | (Mean \pm SD) in | vegetables | (mg/kg Dw). |
|--------------------|--------------------|------------|-------------|
|--------------------|--------------------|------------|-------------|

| Vegetables | Cr | Zn | Ni | Cd | Cu | | |
|-------------------------------------|----------------|------------|------------|------------|------------|--|--|
| Tannery effluents contaminated soil | | | | | | | |
| Stem | 125.20 | 29.73 | 4.28 | 0.19 | 2.95 | | |
| Amaranths | ± 0.76 | ± 0.39 | ± 0.56 | ± 0.69 | \pm 0.77 | | |
| Spinach | 168.99 | 36.24 | 5.49 | 0.83 | 4.03 | | |
| | ± 0.26 | ± 0.33 | ± 0.82 | ± 0.57 | ± 0.37 | | |
| Red | 132.96 | 30.93 | 4.34 | 0.26 | 3.39 | | |
| Amaranths | ± 0.33 | ± 0.45 | ± 0.95 | ± 0.22 | ± 0.92 | | |
| Jute Mallows | 139.65 | 33.36 | 4.79 | 0.46 | 3.67 | | |
| | ± 0.70 | ± 0.48 | \pm 0.78 | ± 0.28 | ± 0.66 | | |
| Water Spinach | 154.96 | 34.48 | 5.14 | 0.42 | 3.54 | | |
| | \pm 0.88 | \pm 0.27 | ± 0.29 | ± 0.15 | ± 0.84 | | |
| Malabar | 159.32 | 34.64 | 5.27 | 0.32 | 3.83 | | |
| Spinach | ± 0.93 | ± 0.64 | ± 0.38 | ± 0.71 | ± 0.76 | | |
| Non-contaminated | agricultural s | oil | | | | | |
| Stem | 1.26 | 13.12 | 2.12 | 0.0051 | 2.83 | | |
| Amaranths | ± 0.39 | ± 0.68 | ± 0.60 | ± 0.88 | ± 0.66 | | |
| Spinach | 2.03 | 14.51 | 3.13 | 0.013 | 4.12 | | |
| | ± 0.20 | ± 0.50 | ± 0.31 | ± 0.26 | ± 0.23 | | |
| Red | 1.41 | 13.31 | 2.45 | 0.005 | 3.13 | | |
| Amaranths | ± 0.65 | ± 0.79 | ± 0.62 | ± 0.11 | ± 0.59 | | |
| Jute Mallows | 1.64 | 15.72 | 2.57 | 0.007 | 3.49 | | |
| | ± 0.89 | ± 0.62 | \pm 0.77 | ± 0.95 | ± 0.76 | | |
| Water Spinach | 1.91 | 14.81 | 3.04 | 0.008 | 4.03 | | |
| | $\pm \ 0.72$ | ± 0.65 | ± 0.56 | \pm 0.77 | ± 0.85 | | |
| Malabar | 1.75 | 15.38 | 2.76 | 0.007 | 3.97 | | |
| Spinach | $\pm \ 0.61$ | \pm 0.28 | ± 0.37 | ± 0.84 | ± 0.30 | | |
| Threshold | 2.30 | 50 | 10 | 0.02 | 40 | | |
| limits[63] | | | | | | | |
| F value* | 9.91 | 6.01 | 4.99 | 1.667 | 2.96 | | |

 * p < 0.01, that depicted that the variables show no significant difference.

their PLI in the soil of Hazaribagh area was higher than any other metal which was similar to the findings of previous research works [65–67]. This is because of the deposition of abundant chromium from the discharged tanning effluents, as 90% tanning industries use basic chromium sulfate for tanning process [6,68]. The extensive application of cadmium containing dyes and pigments in leather dyeing and finishing is the main basis of cadmium deposition in the soil of Hazaribagh, since these coloring components impart a vibrant impression on dyed substrates [69]. It can be understood from t-test values, that the metal content values obtained from both sides are not significantly different.

3.2. Heavy metals in vegetables

Heavy metal contents in vegetables cultivated on the soil collected from the contaminated and non-contaminated agricultural sites were presented in Table 4. Analysis of heavy metal contents showed a noticeable variation in the vegetable samples. Metal accumulation in vegetables grown in contaminated soil was found higher than that of the reference agricultural soil. Total heavy metals contents in vegetables cultivated in both soil was found in descending order as follows: spinach>water spinach>Malabar spinach>jute mallows>red amaranths>stem amaranths.

^{***} p < 0.001.



Fig. 2. Metal Transfer Factor of (a) Cr, (b) Zn, (c) Ni, (d) Cd and (e) Cu in vegetables grown in both the tannery effluent contaminated soil and non-contaminated agricultural soil.

Chromium and Cadmium contents of the vegetables grown in contaminated soil were higher than the threshold limit (2.30 mg/kg Dw for chromium and 0.02 mg/kg Dw for cadmium) value recommended by World Health Organization (WHO) [63]. Among these vegetables the highest chromium content was found in spinach (168.99 \pm 0.26 mg/kg Dw) followed by water spinach>Malabar spinach>jute mallows>red amaranths>stem amaranths. The cadmium content was found highest in jute mallows (0.46 \pm 0.28 mg/kg Dw) followed by water spinach>Malabar spinach>stem amaranths. The overall metal content in vegetables cultivated in contaminated source was in the order of Cr>Zn>Ni>Cd>Cu.

On the other hand, the accumulation of Zn in the vegetables cultivated in the reference soil was found higher compared to other metals. The highest Zn content was found in jute Mallows ($15.72 \pm 0.62 \text{ mg/kg}$ Dw) followed by Malabar spinach>water spinach>spinach>red amaranths>stem amaranths. The accumulation of heavy metals in vegetables

grown on reference agricultural soil were in the order of Zn>Cr>Cu>Ni>Cd and were within the permissible limit.

3.3. Metal translocation in plants from soil

Metal transfer from soil to plants is a principal factor of heavy metal exposure in human food chain and describes the possible availability of any specific trace element in any plant part. The Metal Transfer Factor (MTF) of heavy metals (Cr, Zn, Ni, Cd and Cu) of cultivated vegetables in both the contaminated and agricultural soil was represented in Fig. 2.

The metal translocation was observed higher in vegetables grown in contaminated soil than those of the referenced soil from the vivid analysis. Among these vegetables, the MTF was higher in spinach and Malabar spinach. The MTF in spinach was found in the order of Cr (0.41)>Zn (0.37)>Cd (0.32)>Ni (0.31)>Cu (0.18) and in Malabar spinach the MTF was in the order Cr (0.40)>Zn (0.35)>Cd (0.35)>Ni



Fig. 3. Metal Pollution Index of vegetables grown in both the tannery effluent contaminated and non-contaminated soil.

(0.29)>Cu (0.17). However, MTF in other vegetables was in the order: water spinach>jute mallows>red amaranths>stem amaranths. The translocation of metals in vegetables was observed in the following order: Cr>Zn>Cd>Ni>Cu.

The MTF in vegetables grown in reference soil was found highest in spinach followed by water spinach>Malabar spinach>jute mallows>red amaranths>stem amaranths in the following metal accumulation order: Zn>Cr>Cu>Ni>Cd.

3.4. Metal pollution index

Metal Pollution Index (MPI) is a precise and reliable method that is used for the regular monitoring of heavy metal pollution by quantification of metal contents in crops grown in contaminated sites [38]. MPI value of each vegetable was measured that showed difference in plant species along with sampling sites.

The vegetables grown in contaminated soil had the higher metal accumulation and uptake capacity than those grown in noncontaminated soil. Hence, the MPI values will be higher in the vegetables cultivated on tannery effluent affected soil (Fig. 3). Among the vegetables grown in contaminated soil, the highest MPI was found in spinach (10.24), followed by Malabar spinach (8.36), water spinach (8.13) the least was found in stem amaranths (6.16). Since each vegetable possesses higher MPI values (>5), they are remarked as highly contaminated foodstuffs which may be responsible for sever, acute and chronic diseases upon long-term consumption [70,71].

On the other hands, the MPI found among the vegetables grown in non-contaminated soil was in the order of spinach (1.37)>water spinach (1.25)>Malabar spinach (1.17)>jute mallows (1.09)>red amaranths (0.94)>stem amaranths (0.87). The observed MPI values showed consistency in variance among vegetables grown in both soil samples.

This study was conducted in the same location and no pesticides were applied. Therefore, it was found a very good agreement between the higher MPI value and excess heavy metal load in the vegetables grown in contaminated soil.

3.5. Health risk assessment

3.5.1. Daily intake of metal

Heavy metals contamination in vegetable is an outcome of soil contamination and an obstacle of having quality foods for life. The risk of metal exposure in human body depends on the metal concentration in the respective food. Daily intake of metal (DIM) is an index to determine the exposure of heavy metal in body by feeding [72]. DIM values were

Table 5

Daily intake of metals (mg/kg/Dw-day) from vegetables grown in both tannery effluent contaminated soil and non-contaminated agricultural soil.

| Vegetables | Cr | Zn | Ni | Cd | Cu |
|-----------------------------|-------------------|--------------------|------------------|--------------------|-----------------|
| Tannery effluents contamina | ited soil | | | | |
| Stem Amaranths | 585E- | 139E- | 20E-05 | 8.9E-05 | 14E- |
| | 05 | 05 | | | 05 |
| Spinach | 790E- | 169E- | 26E-05 | 3.9E-05 | 19E- |
| | 05 | 05 | | | 05 |
| Red Amaranths | 622E- | 145E- | 20E-05 | 1.2E-05 | 16E- |
| | 05 | 05 | | | 05 |
| Jute mallows | 653E- | 156E- | 22E-05 | 2.2E-05 | 17E- |
| | 05 | 05 | | | 05 |
| Water Spinach | 724E- | 161E- | 24E-05 | 2.0E-05 | 17E05 |
| | 05 | 05 | | | |
| Malabar spinach | 745E- | 162E- | 25E-05 | 1.5E-05 | 18E- |
| | 05 | 05 | | | 05 |
| Non-contaminated agricultu | ral soil | | | | |
| Stem Amaranths | 5.9E-05 | 61E-05 | 9.9E- | 240E- | 13E- |
| | | | 05 | 05 | 05 |
| Spinach | 9.5E-05 | 68E-05 | 15E-05 | 610E- | 19E- |
| | | | | 05 | 05 |
| Red Amaranths | 6.6E-05 | 62E-05 | 11E-05 | 250E- | 15E- |
| | | | | 05 | 05 |
| Jute mallows | 7.7E-05 | 73E-05 | 12E-05 | 300E- | 16E- |
| | | | | 05 | 05 |
| Water Spinach | 8.9E-05 | 69E-05 | 14E-05 | 410E- | 19E- |
| | | | | 05 | 05 |
| Malabar spinach | 8.2E-05 | 72E-05 | 13E-05 | 340E- | 19E- |
| | | | | 05 | 05 |
| Maximum tolerable limit | 0.02 ^a | 11(8) ^b | 0.5 ^b | 0.068 ^c | 10 ^c |
| | | | | | |

^a [75]

^b [73];

^c [74];

Table 6

HRI of HM for adults via vegetables consumption grown in both the tannery effluents contaminated soil and non-contaminated agricultural soil.

| Vegetables | Cr | Zn | Ni | Cd | Cu | | |
|-------------------------------------|-------------------|---------|---------|---------|---------|--|--|
| Tannery effluents contaminated soil | | | | | | | |
| Stem | 19500E- | 46.33E- | 100E-04 | 80E-04 | 37E-04 | | |
| Amaranths | 04 | 04 | | | | | |
| Spinach | 26300E- | 56.47E- | 128E-04 | 390E-04 | 51E-04 | | |
| | 04 | 04 | | | | | |
| Red Amaranths | 20700E- | 48.20E- | 102E-04 | 120E-04 | 43E-04 | | |
| | 04 | 04 | | | | | |
| Jute mallows | 21800E- | 51.99E- | 112E-04 | 220E-04 | 46E-04 | | |
| | 04 | 04 | | | | | |
| Water Spinach | 24200E- | 53.73E- | 120E-04 | 200E-04 | 45E-04 | | |
| | 04 | 04 | | | | | |
| Malabar | 24800E- | 53.98E- | 123E-04 | 150E-04 | 48E-04 | | |
| spinach | 04 | 04 | | | | | |
| Non-contaminated | l agricultural so | oil | | | | | |
| Stem | 39.3E-04 | 20.45E- | 49.56E- | 2.40E- | 35.76E- | | |
| Amaranths | | 04 | 04 | 04 | 04 | | |
| Spinach | 63.3E-04 | 22.61E- | 73.16E- | 6.10E- | 52.06E- | | |
| | | 04 | 04 | 04 | 04 | | |
| Red Amaranths | 44.0E-04 | 20.74E- | 57.27E- | 2.50E- | 39.55E- | | |
| | | 04 | 04 | 04 | 04 | | |
| Jute mallows | 51.1E-04 | 24.50E- | 60.07E- | 3.00E- | 44.10E- | | |
| | | 04 | 04 | 04 | 04 | | |
| Water Spinach | 59.6E-04 | 23.08E- | 71.06E- | 4.10E- | 50.92E- | | |
| | | 04 | 04 | 04 | 04 | | |
| Malabar | 54.6E-04 | 23.97E- | 64.52E- | 3.40E- | 50.16E- | | |
| spinach | | 04 | 04 | 04 | 04 | | |

observed higher in vegetables grown in tannery effluent contaminated soil than those of non-contaminated agricultural soil (Table 5). Among vegetables cultivated in contaminated soil, the DIM rate for metals was observed as Cr>Zn>Ni>Cd>Cu. DIM was found highest for Cr (7.90E-03) in spinach followed by Malabar spinach (7.45E-03). In spinach the DIM values for metals were found to be in the order of Cr (7.90E-3)>Zn (1.69E-3)>Ni (0.26E-3)>Cu (0.19E-3)>Cd (3.9E-05) and

Table 7

HQ and HI of heavy metals for adults via vegetables consumption grown in both the tannery effluent contaminated soil and non-contaminated agricultural soil.

| Vegetables | Hazard quotient (HQ) | | | | | |
|----------------------------|----------------------|---------|---------|--------|----------|--------|
| | Cr | Zn | Ni | Cd | Cu | |
| Tannery effluents contamir | nated soil | | | | | |
| Stem Amaranths | 22.95 | 0.05451 | 0.1177 | 0.182 | 0.040563 | 23.35 |
| Spinach | 30.98 | 0.06644 | 0.15098 | 0.152 | 0.055413 | 31.41 |
| Red Amaranths | 24.38 | 0.05671 | 0.11935 | 0.048 | 0.046613 | 24.65 |
| Jute mallows | 25.60 | 0.06116 | 0.13173 | 0.084 | 0.050463 | 25.93 |
| Water Spinach | 28.41 | 0.06321 | 0.14135 | 0.077 | 0.048675 | 28.74 |
| Malabar spinach | 29.21 | 0.06351 | 0.14493 | 0.059 | 0.052663 | 29.53 |
| Non-contaminated agricult | ural soil | | | | | |
| Stem Amaranths | 0.00046 | 0.024 | 0.058 | 0.0009 | 0.039 | 0.1227 |
| Spinach | 0.00074 | 0.027 | 0.086 | 0.0024 | 0.057 | 0.1725 |
| Red Amaranths | 0.00051 | 0.025 | 0.068 | 0.001 | 0.043 | 0.1363 |
| Jute mallows | 0.00060 | 0.029 | 0.071 | 0.0012 | 0.048 | 0.1493 |
| Water Spinach | 0.00070 | 0.027 | 0.084 | 0.0016 | 0.055 | 0.1685 |
| Malabar spinach | 0.00064 | 0.028 | 0.076 | 0.0013 | 0.054 | 0.1607 |

comparatively lower DIM values were found in stem amaranths in the order of Cr (5.85E-03)>Zn (1.39E-03)>Ni (0.20 E-03)>Cu (0.14E-03)>Cd (4.6E-05).

Besides, the DIM values for spinach grown in referenced soil was observed in the order Zn (68E-54)>Cr (24E-5)>Cu (19E-5)>Ni (15E-5)>Cd (6.1E-6). However, the DIM was found in such order Malabar spinach>water spinach>jute mallows>red amaranths>stem amaranths. The order of metal ingestion rate from vegetables grown in non-contaminated agricultural soil was found Zn>Cr>Cu>Ni>Cd. The variance of TM concentration in vegetable samples is the main cause for the difference of DIM values that is influenced by factors like metal accumulation, HM concentration in soil, plant speciation, etc.

3.5.2. Health risk index

Health risk index evaluates the possibility of any health hazard due to consumption of contaminated food. HRI values of heavy metals for adults via vegetables consumption which was grown in the contaminated soil were higher than those of grown in non-contaminated agricultural soil (Table 6). Among these vegetables, the contribution to highest HRI value was found for chromium in spinach (2.63). However, the HRI values of chromium for each vegetable grown in contaminated soil exceeded 1 since excess accumulation of chromium from tannery effluents to vegetable plants. Therefore, consumption of these vegetables could pose threat to human health.

HRI values for all studied metals in the vegetables grown noncontaminated agricultural soil did not exceed1, which ensure no further harm in consumption of these vegetables as daily diet.

HQ and HI values were measured to determine the possibility of noncarcinogenic health risk of the dwellers of the study sites upon vegetable consumption and the values were shown in Table 7. Based on the analyses, it was found that the HQ value of chromium crossed 10 for the vegetables grown in polluted soil, which revealed the possibility of chronic hazard symptoms in human body upon vegetable consumption. On the other hand, the HQ values of each heavy metal in the vegetables that were cultivated in non-contaminated agricultural soil were found less than 1. Therefore, there might be no harm to consume those vegetables.

Carcinogenicity depends on the changes in cell's DNA due to genetic motive and environmental influences, e.g., lifestyle factors (nutrition, tobacco and alcohol use, physical inactivity, etc.), naturally occurring exposures (ultraviolet light, radon gas, infectious agents, etc.), medical treatments (radiation and medicines including chemotherapy, hormone drugs, drugs that suppress the immune system, etc.), workplace exposures, household exposures, pollution, etc. In this research carcinogenic risk was not evaluated due to lack of time for further extensive research. Authors would like to suggest for further investigations to evaluate the carcinogenic risks for the consumption of heavy metals contaminated vegetables.

4. Conclusion

Metal contamination in vegetable has become a matter of great concern in many countries as the assurance to have good quality and safer foods are getting interest among people. In this study, the possible scenario of metal contamination by tannery effluents and their adverse effect on human health was illustrated. The article was particularly focused on non-carcinogenic health risk assessment. It was found that among the metal contents, chromium content is the highest in the soil around the industrial zone. Cultivation of leafy vegetables in tannery effluent contaminated soil results in crucial contamination with heavy metals. It was also observed that the leafy vegetables have the high tendency to accumulate heavy metals that is quite alarming for human consumption. Therefore, these vegetables are not suitable for cultivating in tannery effluent contaminated soil. The research inspires and enables the adoption of suitable strategies to manage and mitigate heavy metals from tannery effluents to the safer level before discharge to the environment for the safety of human health.

CRediT authorship contribution statement

Sobur Ahmed: Conceptualization, Methodology, Investigation, Editing, Writing – original draft. **Fatema-Tuj-Zohra:** Writing- Introduction, Literature review and Conclusions. **Meem Muhtasim Mahdi:** Conceptualization, Methodology, Formal analysis, Data collection, Writing – original draft. **Dr. Md. Nurnabi:** Resources, Supervision, Writing – review & editing. **Dr. Md. Zahangir Alam:** Supervision, Resources, Methodology, Writing – review & editing. **Dr. Tasrina Rabia Choudhury:** Visualization, Data analysis, Writing – review & editing. All authors have read and agreed to the published version of the manuscript.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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References

- A. Mudhoo, et al., An analysis of the versatility and effectiveness of composts for sequestering heavy metal ions, dyes and xenobiotics from soils and aqueous milieus, Ecotoxicol. Environ. Saf. 197 (2020), 110587.
- [2] J. Pandey, et al., Palmarosa [*Cymbopogon martinii* (Roxb.) Wats.] as a putative crop for phytoremediation, in tannery sludge polluted soil, Ecotoxicol. Environ. Saf. 122 (2015) 296–302.
- [3] N.K. Mondal, S. Chakraborty, Adsorption of Cr (VI) from aqueous solution on graphene oxide (GO) prepared from graphite: equilibrium, kinetic and thermodynamic studies, App. Water Sci. 10 (2) (2020) 1–10.
- [4] A. Shahriar, A.W. Murad, S. Ahmed, Enhancement of waterproofing properties of finished upper leather produced from bangladeshi cow hides, Euro. J. Eng. Technol. Res. 4 (7) (2019) 63–71.
- [5] Z. Ali, et al., Enrichment, risk assessment, and statistical apportionment of heavy metals in tannery-affected areas, Int. J. Environ. Sci. Technol. 12 (2) (2015) 537–550.
- [6] C.K. Ozkan, H. Ozgunay, H. Akat, Possible use of corn starch as tanning agent in leather industry: Controlled (gradual) degradation by H₂O₂, Int. J. Biol. Macromol. 122 (2019) 610–618.
- [7] C. Gaidau, F. Platon, N. Badea, Investigation into iron tannage, J. Soc. Leather Technol. Chem. 82 (4) (1998) 143–146.
- [8] M.M. Mahdi, F.T. Zohra, S. Ahmed, Dyeing of shoe upper leather with extracted dye from *Acacia nilotica* plant bark-An eco-friendly initiative, Prog. Color, Color. Coat. 14 (4) (2021) 241–258.
- [9] N. Sivaram, D. Barik, Toxic waste from leather industries, in: D. Barik (Ed.), Energy from Toxic Organic Waste for Heat and Power Generation, Elsevier, Amsterdam, the Netherlands, 2019, pp. 55–67.
- [10] H.I. Abdel-Shafy, W. Hegemann, E. Genschow, Fate of heavy metals in the leather tanning industrial wastewater using an anaerobic process, Environ. Manage. Health 6 (2) (1995) 28–33.
- [11] J. Manzoor, M. Sharma, K.A. Wani, Heavy metals in vegetables and their impact on the nutrient quality of vegetables: a review, J. Plant Nutr. 41 (13) (2018) 1744–1763.
- [12] R.O. Oruko, et al., Investigating the chromium status, heavy metal contamination, and ecological risk assessment via tannery waste disposal in sub-Saharan Africa (Kenya and South Africa), Environ. Sci. Pollution Res. 28 (2021) 42135–42149.
- [13] A. Sharma, J.K. Katnoria, A.K. Nagpal, Heavy metals in vegetables: screening health risks involved in cultivation along wastewater drain and irrigating with wastewater, SpringerPlus 5 (1) (2016) 1–16.
- [14] S. Ahmed, et al., Chromium from tannery waste in poultry feed: A potential cradle to transport human food chain, Cogent Environ. Sci. 3 (1) (2017), 1312767.
- [15] H. Chen, et al., Assessments of chromium (and other metals) in vegetables and potential bio-accumulations in humans living in areas affected by tannery wastes, Chemosphere 112 (2014) 412–419.
- [16] G.A. Engwa, et al., Mechanism and health effects of heavy metal toxicity in humans, in: O. Karcioglu, B. Arslan (Eds.), Poisoning in the Modern World-new Tricks for an Old Dog??, 5, IntechOpen, London, United Kingdom, 2019, pp. 1–23.
- [17] G.K. Kinuthia, et al., Levels of heavy metals in wastewater and soil samples from open drainage channels in Nairobi, Kenya: community health implication, Sci. Rep. 10 (1) (2020) 1–13.
- [18] P.B. Tchounwou, et al., Heavy metal toxicity and the environment. in Molecular, clinical and environmental toxicology, in: A. Luch (Ed.), Experientia Supplementum, Springer Nature, Basel, Switzerland, 2012, pp. 133–164.
- [19] J. Singh, A.S. Kalamdhad, Effects of heavy metals on soil, plants, human health and aquatic life, Int. J. Res. Chem. Environ. 1 (2) (2011) 15–21.
- [20] S.S. Sonone, et al., Water contamination by heavy metals and their toxic effect on aquaculture and human health through food Chain, Lett. Appl. NanoBioSci 10 (2) (2020) 2148–2166.
- [21] S. Sinha, A.K. Gupta, K. Bhatt, Uptake and translocation of metals in fenugreek grown on soil amended with tannery sludge: involvement of antioxidants, Ecotoxicol. Environ. Saf. 67 (2) (2007) 267–277.
- [22] M. Xaba, J. Olowoyo, G. Scott, Trace metal deposition on soil and accumulation in plants around a coal power station in Pretoria, South Africa, J. Environ. Sci. Manage. 21 (2) (2018) 23–29.
- [23] S. Qadir, et al., Modulation of plant growth and metabolism in cadmium-enriched environments, Rev. Environ. Contam. Toxicol (2014) 51–88.
- [24] H.P. Singh, et al., Chromium toxicity and tolerance in plants, Environ. Chem. Lett. 11 (3) (2013) 229–254.
- [25] V. Kumar, et al., Copper bioavailability, uptake, toxicity and tolerance in plants: a comprehensive review, Chemosphere 262 (2021), 127810.
- [26] B. Pourrut, et al., Lead uptake, toxicity, and detoxification in plants, Rev. Environ. Contam. Toxicol 213 (2011) 113–136.
- [27] G.R. Rout, P. Das, Effect of metal toxicity on plant growth and metabolism: I. Zinc, in: E. Lichtfouse, M. Navarrete, P. Debaeke, S. Véronique, C. Alberola (Eds.), Sustainable Agriculture, Springer Science + Business Media, Dordrecht, the Netherlands, 2009, pp. 873–884.
- [28] Resolution A/RES/70/1, Transforming our world: the 2030 agenda for sustainable development, in: Seventieth United Nations General Assembly, New York, United Nations, New York, 25 2015. (http://www.un.org/ga/search/view_doc.asp? symbol=A/RES/70/1&Lang=E).
- [29] M.S. Islam, et al., The concentration, source and potential human health risk of heavy metals in the commonly consumed foods in Bangladesh, Ecotoxicol. Environ. Saf. 122 (2015) 462–469.
- [30] Q. Li, et al., Toxic effects of heavy metals and their accumulation in vegetables grown in a saline soil, Ecotoxicol. Environ. Saf. 73 (1) (2010) 84–88.

- [31] R.K. Sharma, M. Agrawal, F. Marshall, Heavy metal contamination of soil and vegetables in suburban areas of Varanasi, India, Ecotoxicol. Environ. Saf. 66 (2) (2007) 258–266.
- [32] A.H. Baghaie, M. Fereydoni, The potential risk of heavy metals on human health due to the daily consumption of vegetables, Environ. Health Eng. Manage. 6 (1) (2019) 11–16.
- [33] R.A. Wuana, F.E. Okieimen, Heavy metals in contaminated soils: a review of sources, chemistry, risks and best available strategies for remediation, Int. Sch. Res. Notices 2011 (2011), 402647.
- [34] F.A. Jan, et al., A comparative study of human health risks via consumption of food crops grown on wastewater irrigated soil (Peshawar) and relatively clean water irrigated soil (lower Dir), J. Hazard. Mater. 179 (1–3) (2010) 612–621.
- [35] A. Singh, S.M. Prasad, Effect of agro-industrial waste amendment on Cd uptake in *Amaranthus caudatus* grown under contaminated soil: an oxidative biomarker response, Ecotoxicol. Environ. Saf. 100 (2014) 105–113.
- [36] M. Alam, E. Snow, A. Tanaka, Arsenic and heavy metal contamination of vegetables grown in Samta village, Bangladesh, Sci. Total Environ. 308 (1–3) (2003) 83–96.
- [37] M.S. Islam, et al., Assessment of trace metals in foodstuffs grown around the vicinity of industries in Bangladesh, J. Food Composit. Anal. 42 (2015) 8–15.
- [38] M.S. Sultana, et al., Transfer of heavy metals and radionuclides from soil to vegetables and plants in Bangladesh, in: K.R. Hakeem, M. Sabir, M. Öztürk, A. R. Mermut (Eds.), Soil Remediation and Plants: Prospects and Challenges, Academic Press, Cambridge, Massachuesetts, USA, 2015, pp. 331–366.
- [39] A. Khan, et al., Processes controlling the extent of groundwater pollution with chromium from tanneries in the Hazaribagh area, Dhaka, Bangladesh, Sci. Total Environ. 710 (2020), 136213.
- [40] S.C. Hong, Developing the Leather Industry in Bangladesh (2018).
- [41] H. Paul, et al., Bangladeshi leather industry: an overview of recent sustainable developments, J. Soc. Leather Technol. Chem. 97 (1) (2013) 25–32.
- [42] P. Whitehead, et al., Modelling heavy metals in the Buriganga River System, Dhaka, Bangladesh: impacts of tannery pollution control, Sci. Total Environ. 697 (2019), 134090.
- [43] M. Ali, et al., Investigation on physicochemical parameters of tannery effluent, Univers. J. Environ. Res. Technol. 5 (3) (2015) 122–130.
- [44] L.R. Dobson, Home Gardening as Part of the Food Security Safety Net in Rural Appalachia. ASA Annual Conference, Marshall University, Huntington, West Virginia, USA, 2017, p. 106.
- [45] A.M. Rabee, Y.F. Al-Fatlawy, M. Nameer, Using pollution load index (PLI) and geoaccumulation index (I-Geo) for the assessment of heavy metals pollution in Tigris river sediment in Baghdad Region, Al-Nahrain J. Sci 14 (4) (2011) 108–114.
- [46] H.S. Shehata, T.M. Galal, Trace metal concentration in planted cucumber (*Cucumis sativus L.*) from contaminated soils and its associated health risks, J. Consum. Protect. Food Saf 15 (2020) 205–217.
- [47] A. Gebrekidan, et al., Toxicological assessment of heavy metals accumulated in vegetables and fruits grown in Ginfel river near Sheba Tannery, Tigray, Northern Ethiopia, Ecotoxicol. Environ. Saf. 95 (2013) 171–178.
- [48] Z. Leblebici, M. Kar, Heavy metals accumulation in vegetables irrigated with different water sources and their human daily intake in Nevsehir, J. Agric. Sci. Technol. 20 (2) (2018) 401–415.
- [49] Z.J. Shen, Y.S. Chen, Z. Zhang, Heavy metals translocation and accumulation from the rhizosphere soils to the edible parts of the medicinal plant Fengdan (*Paeonia ostii*) grown on a metal mining area, China, Ecotoxicol. Environ. Saf. 143 (2017) 19–27.
- [50] N. Gupta, et al., Trace elements in soil-vegetables interface: translocation, bioaccumulation, toxicity and amelioration-a review, Sci. Total Environ. 651 (2019) 2927–2942.
- [51] N.S. Chary, C. Kamala, D.S.S. Raj, Assessing risk of heavy metals from consuming food grown on sewage irrigated soils and food chain transfer, Ecotoxicol. Environ. Saf. 69 (3) (2008) 513–524.
- [52] Consumption of food: Final Report on Household Income and Expenditure Survey 2016, Bangladesh Bureau of Statistics, Statistics and Informatics Division, Ministry of Planning, Bangladesh, (Report Published on June, 2019), p. 45.
- [53] K. ur Rehman, et al., Ecological risk assessment of heavy metals in vegetables irrigated with groundwater and wastewater: the particular case of Sahiwal district in Pakistan, Agric. Water Manag. 226 (2019), 105816.
- [54] T. Sarkar, et al., Assessment of heavy metals contamination and human health risk in shrimp collected from different farms and rivers at Khulna-Satkhira region, Bangladesh, Toxicol. Rep. 3 (2016) 346–350.
- [55] A. Singh, et al., Health risk assessment of heavy metals via dietary intake of foodstuffs from the wastewater irrigated site of a dry tropical area of India, Food Chem. Toxicol. 48 (2) (2010) 611–619.
- [56] J.M. Antoine, L.A.H. Fung, C.N. Grant, Assessment of the potential health risks associated with the aluminium, arsenic, cadmium and lead content in selected fruits and vegetables grown in Jamaica, Toxicol. Rep. 4 (2017) 181–187.
- [57] US EPA (United States Environmental Protection Agency, Human Health Risk Assessment, Regional Screening Level (RSL) - Summary Table, Available from: https://www.epa.gov/risk/regional-screening-levels-rsls-generic-tables,2020, Accessed on: August 25, 2021.
- [58] M.S. Sultana, et al., Health risk assessment for carcinogenic and non-carcinogenic heavy metal exposures from vegetables and fruits of Bangladesh, Cogent Environ. Sci. 3 (1) (2017), 1291107.
- [59] N. Abdu, J.O. Agbenin, A. Buerkert, Phytoavailability, human risk assessment and transfer characteristics of cadmium and zinc contamination from urban gardens in Kano, Nigeria, J. Sci. Food Agric. 91 (15) (2011) 2722–2730.

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- [60] M. Goumenou, A. Tsatsakis, Proposing new approaches for the risk characterisation of single chemicals and chemical mixtures: The source related Hazard Quotient (HQS) and Hazard Index (HIS) and the adversity specific hazard index (HIA), Toxicol. Rep. 6 (2019) 632–636.
- [61] Life expectancy at birth by sex, Monitoring the Situation of Vital Statistics of Bangladesh (MSVSB) 2017, Bangladesh Statistics, Bangladesh Bureau of Statistics, Statistics and Informatics Division, Ministry of Planning, Bangladesh, 2018, p. 11.
- [62] X. Li, et al., Health risks of heavy metal exposure through vegetable consumption near a large-scale Pb/Zn smelter in central China, Ecotoxicol. Environ. Saf. 161 (2018) 99–110.
- [63] Z. Kabir, I. Khan, Environmental impact assessment of waste to energy projects in developing countries: general guidelines in the context of Bangladesh, Sustain. Energy Technol. Assess. 37 (2020), 100619.
- [64] Trace Elements in Human Nutrition and Health, World Health Organization, Geneva, Switzerland, 1996, pp. 72–210.
- [65] Bangladesh standards and guidelines for sludge management, Department of Environment, Ministry of Environment and Forest, Government of the Peoples' Republic of Bangladesh, Dhaka, Bangladesh, 2015, p. 30.
- [66] M.M. Islam, et al., Heavy metal and metalloid pollution of soil, water and foods in bangladesh: a critical review, Int. J. Environ. Res. Public Health 15 (12) (2018) 2825.
- [67] M.S. Islam, et al., Human and ecological risks of metals in soils under different land-use types in an urban environment of Bangladesh, Pedosphere 30 (2) (2020) 201–213.

- [68] M.A. Hashem, et al., Water hyacinth biochar for trivalent chromium adsorption from tannery wastewater, Environ. Sustain. Indic 5 (2020), 100022.
- [69] P.K. Bayannavar, et al., Synthesis of metal free organic dyes: Experimental and theoretical approach to sensitize one-dimensional cadmium sulphide nanowires for solar cell application, J. Mol. Liq. 336 (2021), 116862.
- [70] D. Kumar, et al., Metal pollution index and daily dietary intake of metals through consumption of vegetables, Int. J. Environ. Sci. Technol. 17 (6) (2020) 3271–3278.
- [71] L. Zote, et al., Macro-, micro-, and trace element distributions in areca nut, husk, and soil of northeast India, Environ. Monit. Assess. 193 (2) (2021) 1–12.
 [72] L. Daniele, et al., Chemical composition of Chilean bottled waters: Anomalous
- values and possible effects on human health, Sci. Total Environ. 689 (2019) 526–533.
- [73] M. Harmanescu, et al., Heavy metals health risk assessment for population via consumption of vegetables grown in old mining area; a case study: Banat County, Romania. Chem. Cent. J. 5 (1) (2011) 1–10.
- [74] L. García-Rico, J. Leyva-Perez, M.E. Jara-Marini, Content and daily intake of copper, zinc, lead, cadmium, and mercury from dietary supplements in Mexico, Food Chem. Toxicol. 45 (9) (2007) 1599–1605.
- [75] P. Trumbo, et al., Dietary reference intakes: vitamin A, vitamin K, arsenic, boron, chromium, copper, iodine, iron, manganese, molybdenum, nickel, silicon, vanadium, and zinc, J. Am. Diet. Assoc. 101 (3) (2001) 294–301.