

RESEARCH

Open Access



Home ranges and movements of an arboreal folivore after wildfire: comparing rehabilitated and non-rehabilitated animals in burnt and unburnt woodlands

Murraya R. Lane^{1*}, Kara N. Youngentob², Robert G. Clark³, James D. Skewes⁴ and Karen J. Marsh¹

Abstract

Background Wildfires can have complex effects on wildlife populations. Understanding how post-fire conditions affect the movement ecology of threatened species can assist in better conservation and management, including informing the release of rescued and rehabilitated animals. The 2019–2020 megafires in Australia resulted in thousands of animals coming into care due to injury or concerns over habitat degradation. This included hundreds of koalas (*Phascolarctos cinereus*), for which relatively little was known about how fire affected habitat suitability, or when rehabilitated animals could be returned to burnt areas.

Methods We compared the movements of koalas across three experimental groups—non-rehabilitated koalas in burnt habitat, non-rehabilitated koalas in nearby unburnt habitat, and rehabilitated koalas returned to their rescue location in burnt habitat in New South Wales, Australia. We GPS-tracked 32 koalas for up to nine months and compared, across treatment groups, home ranges, mean nightly distance moved, the farthest distance moved from their release site and total displacement distance.

Results We found no differences in koala movements and home range size between non-rehabilitated koalas in burnt and unburnt habitat. However, rehabilitated koalas moved farther from their release site, had larger displacement distances, and larger home ranges than non-rehabilitated individuals. Regardless of their experimental group, we also found that males moved further than females each night. Additionally, our resource selection analysis showed that, koalas preferred low and moderately burnt habitats over all other fire severity classes.

Conclusions Experimental frameworks that incorporate “treatment” and “control” groups can help isolate disturbance effects on animal movements. Encouragingly, despite catastrophic wildfires, burnt woodlands provided adequate resources for koalas to persist and recover. Furthermore, rehabilitated koalas re-integrated into the burnt landscape despite moving farther from their release sites than non-rehabilitated individuals. Studies like this improve our understanding of the ecological impacts of fire on species and their habitats, and will be instrumental in informing wildlife management and conservation efforts as wildfires increase in frequency and severity worldwide in response to climate change.

Keywords Wildlife rehabilitation, Habitat, Koala, Landscape disturbance, Endangered species, Conservation, Radio-tracking

*Correspondence:

Murraya R. Lane

murraya.lane@anu.edu.au

Full list of author information is available at the end of the article



© The Author(s) 2024. **Open Access** This article is licensed under a Creative Commons Attribution-NonCommercial-NoDerivatives 4.0 International License, which permits any non-commercial use, sharing, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons licence, and indicate if you modified the licensed material. You do not have permission under this licence to share adapted material derived from this article or parts of it. The images or other third party material in this article are included in the article's Creative Commons licence, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons licence and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this licence, visit <http://creativecommons.org/licenses/by-nc-nd/4.0/>.

Introduction

Animals move through their environment to find and defend resources, disperse, breed, avoid predation, interact with conspecifics, and seek shelter [35, 109]. These essential movements normally occur within a defined area, which is often described as a home range [17, 23, 33]. Several intrinsic and extrinsic factors can influence home range size, including body size [76, 90], age [8, 121], sex [21, 44], food availability and quality [51, 89, 121, 125], climate [94, 117], and landscape fragmentation [44]. Understanding the factors that affect home range size can provide important information about the space and resources needed by species and populations, explain the drivers of movement patterns [130], and assist with conservation and management planning.

Landscape disturbance from natural and anthropogenic sources (e.g., fire, floods, land clearing, logging, etc.) can influence the structure of faunal and floral communities and can fragment landscapes by creating physical barriers to movement [37, 129, 155]. These can directly and indirectly affect how animals utilise and move within their environment [12, 100, 129, 146]. Fire is a common stochastic disturbance process in many terrestrial environments and can result in short-term effects on animal movement patterns due to the immediate emigration of individuals out of a landscape that are trying to escape fire fronts, and cause longer-term shifts in movement patterns within burnt areas due to changes in habitat suitability and altered competition and predation dynamics [42, 100, 146]. Changes in habitat suitability after fire can include the destruction or modified pattern of food and shelter availability. This can place substantial physiological stress on animals, which can negatively impact reproductive success, thermoregulation, and susceptibility to other threats such as predation and disease [26, 29, 31, 34, 57, 91, 97]. The ability to move and find resources in a burnt landscape can therefore affect the survival and recovery of individuals [59, 99, 100, 103, 144].

Past research into the effects of fire on home ranges shows considerable variation between mammal species. Some species show no change in home range size following wildfire [22, 82, 134, 138], while others expand their home ranges following fire [50, 54] or reduce them [105]. Increases in home range sizes have been attributed to easier movement through landscapes to find resources in the absence of understorey vegetation [54] or the need to move farther to find food if its availability is reduced [82, 133]. Where home range sizes were similar after fire despite a measurable change in resource availability, this has been attributed to individuals accessing one resource in burnt areas and another in adjacent unburnt areas [82]. The different responses reinforce the complex effects of fire on animal movements and home ranges

and highlight the need to improve our understanding of the mechanisms that underly the varied patterns in wildlife population responses to fire.

The temperate forests of the south-east region of Australia are dominated by tree species of the *Eucalyptus* genus and are incredibly fire-prone [7, 14, 79, 119]. This region is also subjected to lower rainfall and hot dry summers, which are significant contributors to major wildfires [7]. From August 2019 to March 2020, extensive bushfires burned over 12 million hectares of eastern Australia [149]. These fires were unprecedented in both scale and severity [32, 73, 149]. Throughout this time, there were extensive wildlife rescue and rehabilitation efforts [38, 104]. One species that received particular attention was the koala. During the 2019–2020 fires, an estimated 60,000 koalas were killed or injured [136], and over 900 taken into wildlife care triage units [25, 80, 104].

Fire affects the distribution and abundance of koalas [26, 91], and could also influence short-term behavioural shifts in landscape use due to changes in population density, food quality and availability. Koalas, like many other animal species, form distinct home ranges, with males and females tending to overlap [28, 39, 56, 148]. Koala home range sizes can be influenced by a number of factors, such as the sex and age of the animal, population density, the availability and quality of feed trees, and by disturbance like fire [71, 78, 85, 113, 114]. Very few studies have investigated the effects of fire and rehabilitation on movements and home range sizes of the koala (but for exceptions see [81, 88]). Evaluating home range size after fire can provide insight into disturbance related changes in habitat quality, population health, and resilience, which can help to improve landscape and wildlife management [140, 141].

Understanding how rehabilitation in captivity may affect the longer-term recovery of individuals after fire is also important for decision-making around this type of intervention. Rehabilitation success is often determined by recovery and survival of animals in care to the point of release [81]; however, a better metric is survival post-release, which is not always monitored or recorded [53, 98]. Some studies have found that, after release, rehabilitated animals move farther than non-rehabilitated individuals, which may be due to several factors, including being displaced or even losing the skills to compete and survive following time in care [45, 47, 135]. Increased movements and dispersal can be energetically costly [13, 118], lead to increased exposure to predators [154] and mortality [47, 77, 118]. Post-release effects such as these could be exacerbated when the release environment has been affected by disturbance, such as fire [25]. For example, there could be greater conflict with non-rehabilitated individuals due to increased competition for limited food

resources. Since fire severity and frequency in many parts of the world are predicted to increase with climate change [79], we need to improve our understanding of these post-release issues to effectively manage the reintegration of rehabilitated individuals into surviving populations of threatened species.

This study addresses identified knowledge gaps in managing wildlife living in landscapes impacted by fire, including the success of post-rehabilitation release [25]. The specific aim was to determine whether fire and/or rehabilitation affect the home range sizes and movements of koalas. We addressed this aim within an experimental framework, which is commonly lacking in these types of studies due to the stochastic nature of fire and wildlife rehabilitation [25]. We monitored koalas fitted with GPS collars across three treatment groups: 1) non-rehabilitated koalas in burnt habitat, 2) non-rehabilitated koalas in unburnt habitat, and 3) rehabilitated koalas released into their original burnt habitat. Each koala was monitored for up to nine months. Since severe fire can influence the availability of foliage biomass [58, 127], we expected that food resources would be diminished for koalas in the burnt landscape, resulting in them moving farther each day to obtain adequate nutrition and have larger home ranges compared to individuals in unburnt habitat. Because rehabilitated koalas had to reintegrate into the population in burnt habitat, where competition for resources may be increased, we also expected that they would have larger home ranges and movements compared to non-rehabilitated koalas.

Materials and methods

Study sites

The study was located in the New South Wales (NSW) Snowy Monaro Shire, a local government area in Australia spanning just over 15,000 km² and located approximately 130 km south of the Australian Capital Territory [49]. Much of the koala tracking work was located on the Macanally-Numeralla ranges. The area is on steep terrain, dominated by north-to-south facing ridges [102]. The elevation ranges from 800 to 1233 m above sea level [72] and the underlying geology is comprised of Palaeozoic Era sediments, with fluvial sandstones and lacustrine sediments amongst granite plutons and alkali basalts [86]. Overstorey vegetation is dominated by red stringybark

(*Eucalyptus macrorhyncha*) and scribbly gum (*Eucalyptus rossii*), with ribbon gum (*Eucalyptus viminalis*) and candlebark (*Eucalyptus rubida*) more prominent in creek lines, and snow gums (*Eucalyptus pauciflora*) at higher elevations [102]. Brittle gum (*Eucalyptus mannifera*), broad-leaved peppermint (*Eucalyptus dives*), and apple box (*Eucalyptus bridgesiana*) were also common in some areas [1].

The Monaro Tableland, which encompasses the study area, is within a rain shadow zone [62, 102]. Prior to the fires in 2019, annual rainfall in the nearest city, Cooma, was 318.6 mm, which is well below the annual average of 528.1 mm [15]. Mean daily minimum and maximum temperatures in the Snowy Monaro Shire at these sites average 9 °C to 27 °C in summer, and −3 °C to 13 °C in winter [86].

The “burnt” site was burnt by wildfire in January 2020, with large areas of moderate to complete canopy loss (Fig. 3). The “unburnt” site was located approximately 15 km to the south and was not directly impacted by fire. Both sites were connected by continuous forest cover, and had similar tree species composition.

Koala groups

We used three treatment groups of koalas: 1) non-rehabilitated koalas in burnt habitat, 2) non-rehabilitated koalas in unburnt habitat, and 3) rehabilitated koalas released into the burnt habitat where they were originally rescued. The number of koalas in each group was a combined function of the available numbers of individuals and logistical considerations regarding tracking larger numbers of animals in multiple areas. The rehabilitated koalas comprised ten individuals that had been rescued from burnt habitat in the study area between early January 2020 and early March 2020 due to poor body condition and bushfire-related injuries. Rehabilitated koalas were released between June 2020 and November 2020. All rehabilitated koalas apart from a young, orphaned male were released at their capture location. The young male was housed and released with an adult female and her joey ~4.6 km from his capture location. Release timing depended on the overall health of each koala (subject to a health check—see “[Health checks, collar fitting and release](#)”) and evidence that there was food available in

(See figure on next page.)

Fig. 1 Initial capture locations of all GPS-tracked rehabilitated koalas (green), and non-rehabilitated koalas in burnt habitat (blue) and unburnt habitat (red) (map data ©Google). Fire extent and severity mapping (FESM) [131] is also shown, with 2 representing the lowest severity (burnt understorey with unburnt canopy), 3 representing moderate (partial canopy scorch), 4 representing high (completed canopy scorch, ±partial canopy consumption) and 5 representing extreme (complete canopy consumption). The inset shows the location of the study area within south-eastern Australia

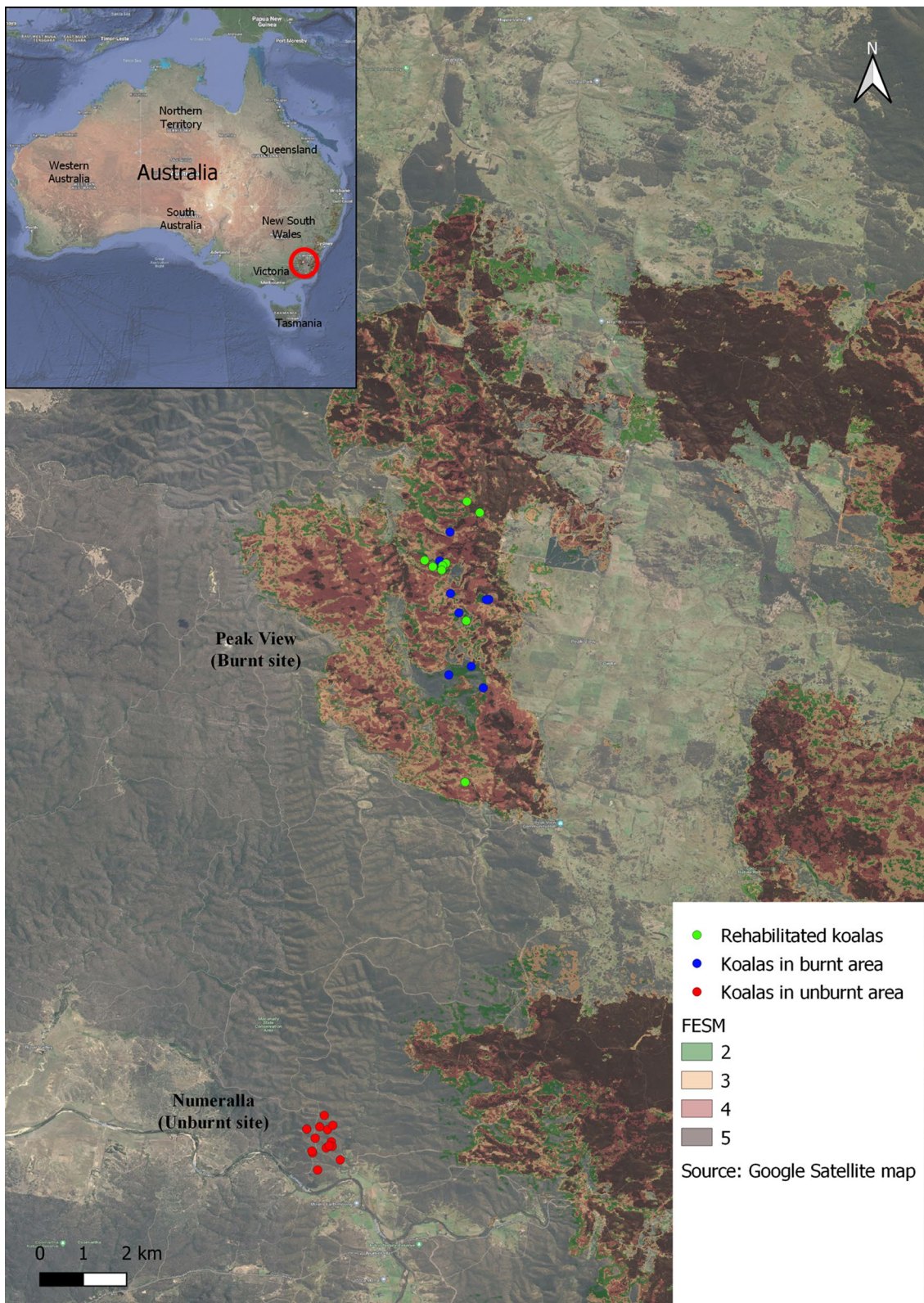


Fig. 1 (See legend on previous page.)

the area from which they were rescued (trees with intact canopy or producing new growth).

Non-rehabilitated koalas in burnt ($n=9$) and unburnt landscapes ($n=13$) were located between May 2020 and October 2020 using a combination of daytime ground-based searches and night-time thermal drone searches [150]. These koalas were captured using the noose and flag method [83], placed into canvas bags and transported by car (usually less than 1 km) for a health check by a wildlife veterinarian.

Health checks, collar fitting and release

Koalas in all three groups were given a health assessment under sedation by an experienced koala veterinarian immediately prior to being fitted with a GPS tracking collar and released [64]. The koalas ranged in age from <2 to >10 years old (based on tooth wear; [84]) and included 15 males and 17 females (Additional file 1).

While sedated, we fitted koalas with a Lotek LiteTrack 60 collar (Lotek Havelock North, New Zealand) with VHF and GPS capabilities. These collars weighed approximately 80 g each and had a “weak link” made of elastic that was designed to break under strain. Koalas were placed in a pet carrier covered with a towel in a quiet room to recover from sedation. All of the non-rehabilitated koalas from the burnt and unburnt landscapes were released on the same day of their capture, at their capture location.

Koala tracking and monitoring

Given that koalas are nocturnal, collars collected five GPS points per 24 h period at three-hourly intervals between 1900 and 0700 h for the duration of the study. This data was stored on the collars and could be downloaded remotely using a paired PinPoint Commander (Lotek, Havelock North, New Zealand). If there was no movement detected for 24 h, a mortality signal would activate.

We visually checked koalas daily for the first two weeks following the collar fitting, twice per week for the following two weeks, and then weekly when possible for the remainder of the study. Koalas were located using their unique VHF signal. Once a collared koala was located, we observed them with binoculars to check for any signs of injury, illness or decline of health. Occasionally, there were times when a koala could not be visually accessed at a scheduled point in time due to, for example, inaccessible roads or the koala had moved to a property that we did not have permission to access. When this occurred, we remotely downloaded their GPS data to check that their daily movements were normal (i.e., that they were moving around at night).

Collar removal

We removed collars from koalas after approximately nine months, unless the collar had already fallen off due to breakage of the weak link or slipping over the koala's head. Koalas were captured using the same noose and flag method, and received another veterinary health check under sedation [64]. Once recovered, koalas were released back into the same tree from which they were captured.

Home range calculations

In the context of this study, home ranges were defined as the area used by koalas (50% and 95% kernel utilisation distributions) across the entire period over which they were tracked before calculating home ranges, we cleaned the GPS data to remove points with poor spatial accuracy. We removed any points where the recorded altitude fell outside the range available at the sites because preliminary inspection revealed that these locations were invariably inaccurate. We then calculated the straight-line distance and difference in altitude between each GPS fix. If either difference was unlikely given the site and terrain, we removed the point. Finally, if the distance moved between consecutive points was greater than 200 m, we visually checked the location on a map relative to the previous and next point. In this case, we removed any points where a koala appeared to have moved to another location and then returned to the first at the following fix, because this scenario was unlikely. After this process, we further thinned the data to two GPS points per koala per night (22:00 and 07:00), to get the best spread and timing where koalas were most likely to have moved.

We calculated 50% and 95% kernel utilisation distributions (KUDs) for each koala using a bivariate normal kernel method in the *adehabitatHR* package (v0.4.19) [18] in R Statistical Software (v3.6.3) [112]. The h value (smoothing parameter), which controls the “width” or size of the kernel function around each point was set to the “reference” bandwidth defined as “ $href$ ” [20]. We generated asymptotes from the *hrBootstrap* function in the *adehabitatHR* package to determine which koalas had a home range and which did not. Those koalas that did not have a home range were still moving significant distances, which suggest they had not ‘settled’ or were not in a defined ‘range’. These koalas were not used in statistical comparisons of home ranges. Using the output from the analysis, we generated maps of home ranges in QGIS v.3.22.8 [110].

Calculation of movements

We calculated the nightly movement distance for each koala by adding the straight-line distances between each

sequential GPS point between 1900 and 0700 h the following morning. We used these values to calculate the average nightly distance moved by koalas throughout the study, as well as the average nightly distance moved during different seasons. We calculated the displacement distance for each koala by measuring the straight-line distance between the point at which koalas were released after being fitted with a tracking collar and the location at which they either lost their collar or were recaptured. We calculated the greatest distance moved by measuring the farthest distance a koala was recorded from its initial capture location.

Resource use analysis

We used resource selection analysis to investigate whether koalas use habitat in different fire severity classes in proportion to its availability within their home ranges.

Using the *raster* package [55] in R and the NSW fire extent and severity mapping spatial dataset [131] we calculated the proportional availability of each fire severity class (as described by the State Government of NSW and Department of Planning and Environment 2020) within the home range of each koala that spent time in burnt habitat ($n=18$). We also extracted the fire severity class at every GPS point used to generate home ranges, and calculated the proportion that fell within each of the five fire severity classes for each koala. We used the *compana* function in the *adehabitatHS* package [19] to perform a compositional analysis of habitat use [2], comparing proportional fire severity class use by each koala relative to the proportional availability of each fire severity class within their home range.

Statistical analyses

All analyses were performed using R Statistical Software (v3.6.3) [112]. We investigated whether home range size and movement data (displacement distance, greatest distance moved and mean nightly distance) differed between koala groups (rehabilitated, or non-rehabilitated in burnt or unburnt habitat), sexes or their interaction using ANOVAs. We also used an ANOVA to investigate whether mean nightly distance differed seasonally between male and female koalas. For greatest distance moved and displacement distance, all koalas tracked less than 50 days were excluded from analyses to avoid potential correlations between distance moved and the length of time tracked. We also excluded two female koalas from the non-rehabilitated unburnt group ('Rosalie' and 'Brandy') from all analyses because they moved from the unburnt habitat into burnt habitat during the study. Thus, their movements may have been influenced by characteristics of both unburnt and burnt habitat, and it was not appropriate to consider them within the unburnt group

for comparative purposes. Despite this, their data are still interesting and could be informative, so we have included them descriptively in the results. Displacement distance, greatest distance moved, and 50% and 95% KUDs were all logged prior to analysis to achieve normal distribution. If results of the ANOVA were statistically significant ($p < 0.05$), we used the *emmeans* package (v.1.6.3) [75] to conduct pairwise t-tests to determine which treatments differed from the others ($p < 0.05$).

Results

Thirty-two koalas across the three groups were followed for between 9 and 256 days each (Additional file 1). Both rehabilitated and non-rehabilitated koalas from the burnt landscape used burnt habitat extensively (Fig. 2). The average nightly distance moved by each koala ranged from 74 to 448 m (Additional file 1), although 14 koalas moved more than 1 km in a night at least once during the study. There was no difference in the mean nightly distances moved by koalas between the three groups ($F_{2, 26} = 1.35$, $p = 0.28$; Fig. 3a). However, male koalas moved farther each night than females ($F_{1, 26} = 24.35$, $p < 0.001$), especially during spring ($F_{1, 24} = 33.03$, $p < 0.001$) and summer ($F_{1, 17} = 34.21$, $p < 0.001$; Fig. 4).

The maximum distance from the release location varied between koalas (405 m to 8,362 m for koalas included in statistical analyses; Additional file 1). Similarly, the displacement distance between the first and last locations recorded for each included koala varied, ranging from 32 to 6,567 m. For koalas tracked more than 50 days, both displacement distance, and greatest distance moved differed significantly between koala groups ($F_{2, 19} = 4.93$, $p = 0.019$; $F_{2, 19} = 5.94$, $p = 0.010$, respectively; Fig. 3b and c). On average, rehabilitated koalas had a displacement distance over three times that of non-rehabilitated koalas in both the burnt ($t(19) = 2.345$, $p = 0.03$) and unburnt landscapes ($t(19) = 2.615$, $p = 0.017$; Fig. 3b). We found a similar pattern for the greatest distance moved. The greatest distance moved was 2.5 times higher in rehabilitated koalas compared to non-rehabilitated in the burnt area ($t(19) = 2.240$, $p = 0.037$) and was three times higher compared to non-rehabilitated in the unburnt ($t(19) = 3.032$, $p = 0.007$; Fig. 3c). Displacement distance and greatest distance moved did not significantly differ between males and females ($F_{1, 19} = 0.295$, $p = 0.594$; $F_{2, 19} = 0.392$, $p = 0.539$, respectively).

We determined home-range sizes for 26 koalas (8 rehabilitated koalas, 8 non-rehabilitated koalas in the burnt landscape, 8 non-rehabilitated koalas in the unburnt landscape and for the two koalas that moved from unburnt area into burnt (Additional file 1). For the six koalas that did not reach an asymptote, five were tracked for less than 50 days (one rehabilitated, one

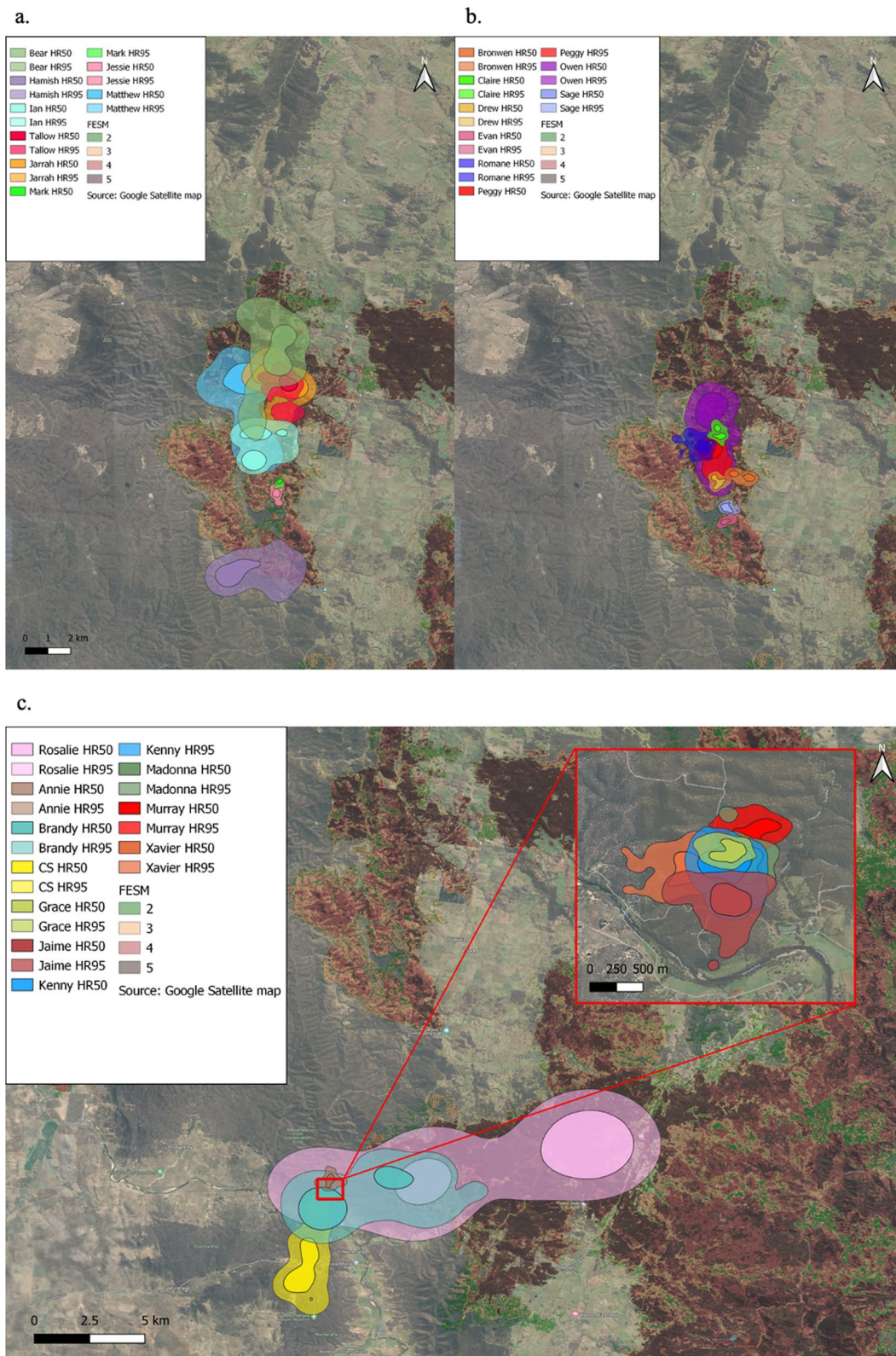


Fig. 2 Home ranges at KUD 50% (darker colours) and 95% (lighter colours) for rehabilitated koalas **(a)**, non-rehabilitated koalas in burnt habitat **(b)** and non-rehabilitated koalas in unburnt habitat **(c)** (map data ©Google). Part c also shows two koalas that moved from unburnt into burnt habitat (Brandy and Rosalie). The inset on Fig. 2c shows the home ranges of six koalas in the unburnt area with comparatively smaller home ranges. Note that the scale on Fig. 2c differs from that of 2a and b

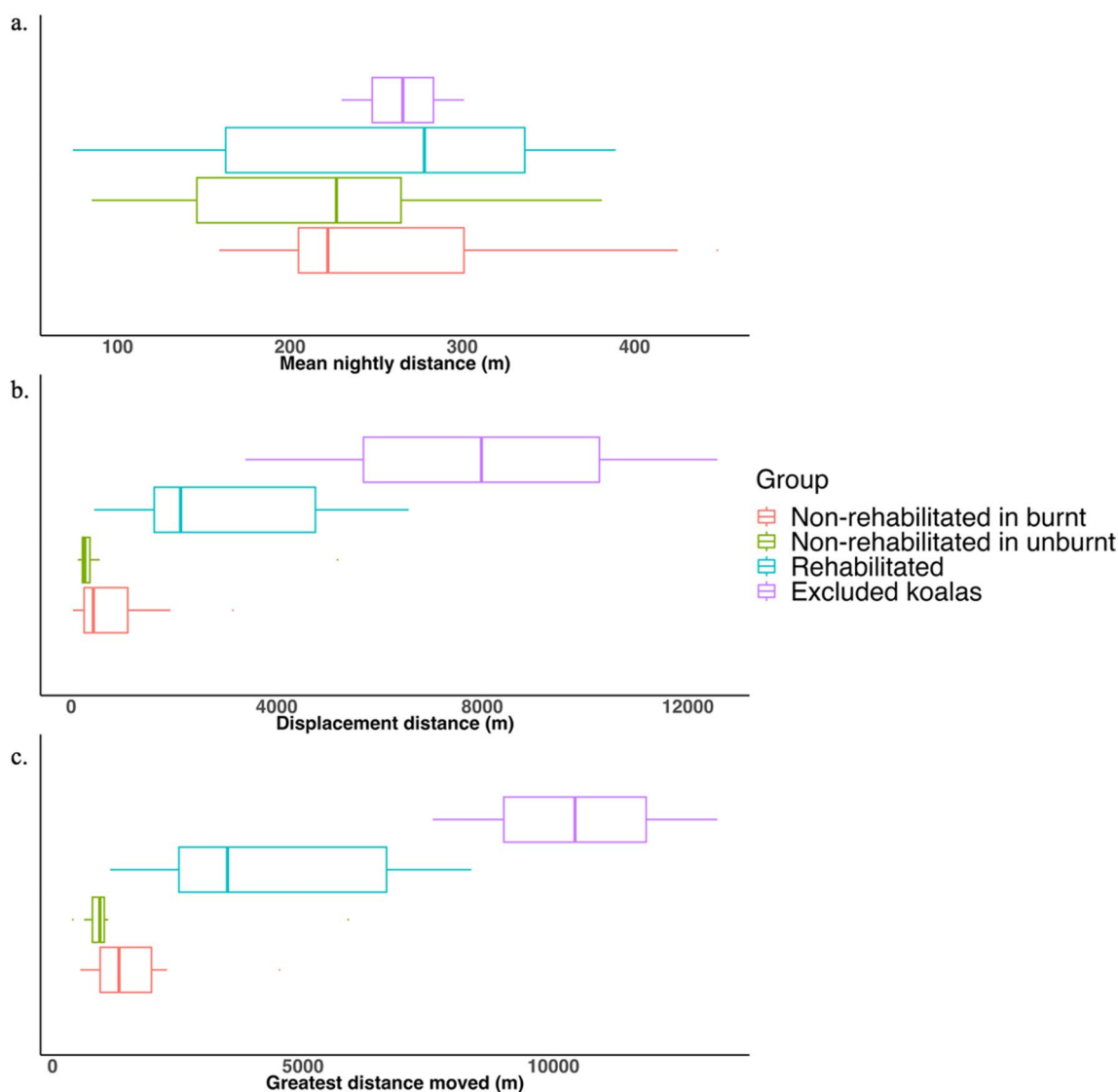


Fig. 3 Mean nightly distance moved by all koalas ($n=30$) (a), displacement distance for koalas tracked > 50 days ($n=23$) (b), and greatest distance moved from release point for koalas tracked > 50 days (c) in each study group. Excluded non-rehabilitated koalas ($n=2$) moved from unburnt habitat into burnt habitat, and therefore did not fit the criteria for any group. Note that the lines within each bar are the median value

non-rehabilitated in the burnt, and three non-rehabilitated in the unburnt). The sixth koala was tracked for 120 days (a rehabilitated female), but she moved consistently in one direction. Home range size for the other 24 koalas in the treatment groups (11 females and 13 males) varied from 4.8 to 279.8 hectares at 50% KUD and 18.5 to 1,543.1 hectares at 95% KUD (Fig. 2; Additional file 1). At 95% KUD, mean home range was significantly different between koala groups ($F_{2, 20}=3.639, p=0.045$), with post-hoc analyses showing larger home range sizes for rehabilitated koalas compared to non-rehabilitated individuals in the unburnt landscape ($t(20)=2.291, p=0.033$; Table 1). 50% KUD, koalas in the rehabilitated group

tended to have larger home ranges than koalas in the non-rehabilitated groups ($F_{2, 20}=3.041, p=0.070$; Table 1). On average, home range sizes for males and females were not significantly different at both 50% KUD ($F_{1, 20}=0.121, p=0.732$) and 95% KUD ($F_{1, 20}=0.114, p=0.739$), with large variation between individuals of both sexes. There was no group by sex interaction observed at either 50% or 95% KUD.

There was substantial overlap in the home ranges of tracked koalas at the 95% KUD, particularly in the unburnt area, but less so at the 50% KUD. For koalas living in the burnt landscape ($n=16$), the core home range (50% KUD) of seven koalas did not overlap with any

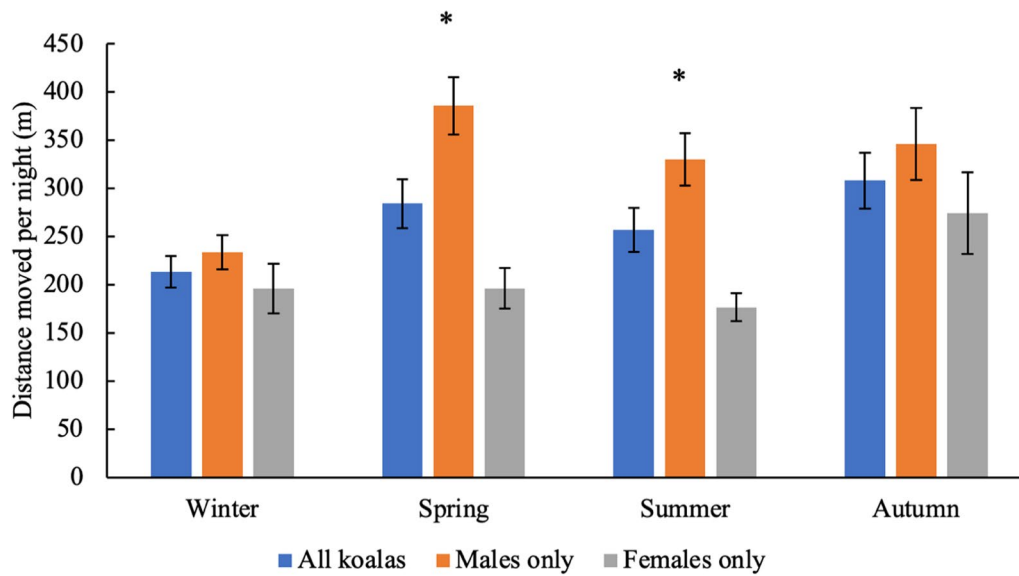


Fig. 4 Mean ± SE nightly distances moved by koalas during different seasons. Seasons with an asterisk above them indicate that the distances moved are significantly different ($p < 0.05$) between males and females

Table 1 Mean ± SE home range size at 50% and 95% kernel utilisation distribution (KUD) for koala groups and excluded non-rehabilitated koalas that used both habitat types

Group	50% KUD (ha)	95% KUD (ha)
Rehabilitated koalas (n=8)	151.5 ± 39.8	773.8 ± 206.5
Non-rehabilitated in burnt habitat (n=8)	58.0 ± 28.0	250.8 ± 123.3
Non-rehabilitated in unburnt habitat (n=8)	45.1 ± 33.6	194.7 ± 142.7
Excluded koalas (n=2)	1178.1 ± 598.8	5880.4 ± 2741.7

other tracked individuals. However, for the remaining nine koalas, overlap occurred with at least one individual and up to a maximum of three individuals. For koalas in the unburnt landscape, only one individual had a core home range that did not overlap with any other koalas.

There was more extensive overlap for the remaining nine koalas in the unburnt landscape, with overlap observed with at least one other individual, and maximum overlap occurring with eight other tracked individuals.

Relative to their availability, koalas showed a preference for low and moderately burnt habitat over all other severity classes, including unburnt ($\lambda = 0.228, p = 0.002$) (Table 2). They also used habitat in the high fire severity class more often relative to its availability than the extreme or unburnt classes (Table 2). Fire severity classes were ranked in the following order of preference: low severity (class 2), moderate severity (3), high severity (4), extreme severity (5) and unburnt (0).

Discussion

Understanding how environmental disturbances such as wildfire affect animal movements and home ranges is important for assessing and guiding the conservation

Table 2 Relative preference for fire severity categories by koalas considering availability. A description of the categories is provided as per [131]. Symbols indicate whether koalas demonstrated a slight (+) or strong preference (+++) for, or avoidance (- or --) of different fire severities

Severity class	Description	0	2	3	4	5
0	Unburnt	0	---	---	---	-
2	Low		0	+	+++	+++
3	Moderate			0	+++	+++
4	High				0	+++
5	Extreme					0

and management of endangered species. In our study, the use of “control” groups within an experimental framework was particularly valuable because it allowed us to compare movement distances and home range sizes of rehabilitated and non-rehabilitated koalas in burnt and unburnt eucalypt woodland to understand how both bushfire and rehabilitation influenced movement behaviour. Our key findings were that: 1) contrary to expectations, the movements and home ranges of non-rehabilitated koalas were similar in burnt and unburnt habitat, and 2) as anticipated, displacement distances (distance between release and recapture points), the farthest distance moved from the release point, and home ranges were largest on average in rehabilitated koalas compared to their non-rehabilitated counterparts. Interestingly, our compositional analysis of habitat use showed that koalas preferred low to moderate burn severity, relative to all other burn severity categories. These findings have important implications for our current understanding of the value of burnt habitat for koalas and other marsupial folivores, and for refining intervention and release guidelines for koalas that are rescued from fire-affected areas.

Effects of fire on home ranges and movement

Our predictions of larger home ranges and longer movements in burnt habitat were based on the idea that fire reduces habitat quality (e.g. less food available; [87]) and competitive pressure from conspecifics due to fire-related mortality [27], forcing individuals to travel greater distances to find adequate food and/or allowing them to expand into unoccupied habitat immediately after a fire event. Unquestionably, leaf biomass, and therefore food availability, is substantially reduced in eucalypt forests after high severity fire [111]. Likewise, moderate to severe wildfire can substantially reduce koala abundance in burnt compared to unburnt areas [26, 71, 107]. Despite these changes, movement distances and home range sizes of non-rehabilitated koalas in burnt habitat were similar to those in unburnt habitat within our study timeframe, which spanned five to sixteen months post fire.

By the time our study commenced, food availability may not have been limiting for the remaining animals. Eucalypts are well-adapted to fire, and many species produce new growth from epicormic buds within days or weeks [16, 106]. In addition, a reduction in the availability of canopy foliage may be offset by an increase in the nutritional quality of epicormic leaves in a relatively short time after fire. This is further supported by our compositional analysis of habitat use, which demonstrated that koalas showed a preference for low to high burn severity over unburnt habitat, which could be linked to the epicormic foliage in these areas being more nutritious.

Recent studies have shown that epicormic leaves are higher in nutritional quality in some eucalypt species than the adult-phase leaves that are present in unburnt areas [65]. This may have contributed to our findings, since many trees in the burnt landscape had at least some epicormic growth at the time we initiated our study.

It was interesting to note that two koalas in our study who were initially captured as part of the group in unburnt habitat moved into burnt areas. Although they were excluded from statistical analyses, this behaviour is worth discussing. At a minimum, the movement from unburnt to burnt habitat reinforces the idea that burnt landscapes can provide habitat for koalas. Additionally, the compositional analysis showed that koalas preferred low, moderate and highly burnt areas over extremely burnt areas and even unburnt areas, demonstrating that burnt areas can support koalas, and may even be favoured if the canopy has not been completely consumed. Other studies have observed similar results, with some grazing and browsing mammals attracted to burnt areas for the new growth of plants recovering from fire [4, 5, 63, 92]. The movement of individuals from unburnt to burnt areas should also be considered in the context of post-fire recovery, whereby populations in unburnt areas can drive the recolonisation of burnt habitat [128, 143]. It would be valuable to design future studies to investigate the emigration of koalas from unburnt to burnt habitat to better understand how this may contribute to population recovery after wildfire.

Effects of rehabilitation on home ranges and movement

Rescuing and rehabilitating wildlife can be stressful for animals [139], and the risks to animal welfare need to be weighed against the potential benefits for both the individual and the population [24]. Some species have a low survival rate after release following rehabilitation [11], while other species reintegrate back into the wild with minimal expense to fitness [95]. The effects of rehabilitation and release can be complex, with various factors contributing to outcomes [74, 93]. We found that rehabilitated koalas had larger home ranges and larger displacement distances from their release site than non-rehabilitated koalas in both the burnt and unburnt landscape. While several studies have looked at koala movements following release from rehabilitation (e.g., [40, 48, 74, 88]), only Matthews et al. [88] also compared their findings to non-rehabilitated koalas in the same area. In contrast to us, they found that rehabilitated koalas in Port Stephens had similar home range sizes to non-rehabilitated koalas. Future research is needed to better understand which factors drive these differences and could provide insights into when koalas will reintegrate most easily into their habitat after rehabilitation.

The larger movements of rehabilitated koalas are unlikely to be due to their release into burnt habitat per se, because their movements differed from non-rehabilitated koalas in the same area. Previous studies have found that rehabilitated koalas that are released away from their capture location often disperse to other areas [9, 74]. However, this does not explain our findings either because koalas in all groups were released at the location from which they were captured. Instead, there are other possible explanations, such as larger movements to refamiliarise themselves with the landscape, or competitive exclusion by established, non-rehabilitated koalas. For example, rehabilitated red foxes (*Vulpes vulpes*) had larger home ranges and travelled farther from their release site than wild individuals, which was attributed to captivity causing territorial displacement [135]. To address this, future research could look at interactions among rehabilitated and non-rehabilitated koalas in the same landscape to better understand their social dynamics, particularly when male–female and male–male interactions are likely to be high, such as during the breeding season [142]. This could guide more effective management of rescued koalas and aid rehabilitation and release practices.

Despite the greater distance moved and larger area used by rehabilitated individuals, measured health parameters were similar between koalas in all groups throughout the study [64]. Furthermore, there was no change in the chlamydia status of individuals (a disease which can be triggered by stress; [145]), and koalas in all groups improved in body condition over the study time-frame and had high rates of survival [64]. This suggests that any costs associated with moving away from release locations did not have substantial longer-term (i.e. at nine months post-release) effects on the health of individuals. In combination with our findings, there is strong evidence from multiple studies that rehabilitated koalas are capable of integrating back into the landscape, including when it is burnt, and they can successfully form home ranges, although they may initially need to move farther from their release site to re-establish their territory [48, 74, 88].

Based on our findings, koalas in fire-affected landscapes that are injured or in poor body condition should be taken into care, but otherwise, uninjured koalas should be allowed to remain in burnt landscapes, as long as some browse is available. For animals that are taken into care, our study demonstrates that rehabilitated individuals can be safely released at their rescue location rather than moving them to areas that are less disturbed, a practice that is sometimes undertaken [74]. Our findings should give wildlife carers confidence that rehabilitated koalas can be successfully released into burnt

habitat, and to incorporate this practice into release policy. More research like this is needed across the range of the koala, since tree species composition can have substantial effects on food quality for koalas before and after fire [65]. In addition, climatic conditions can vary considerably across the koala's large distribution and reduced canopy cover after fire may impact thermoregulation more substantially in some regions than others. The ability to develop region specific guidelines would assist policy-makers, landscape managers and the wildlife care sector to more effectively conserve koalas and their habitat in the face of increasingly severe and frequent fires from anthropogenic climate change [116].

Challenges of interpreting animal movements

Studying animal movements and interpreting the outcomes of home range analyses can be challenging, as many animal species do not use their home range, or move, in a uniform way [52]. It is often assumed that animal movement data captured at a particular time adequately reflects normal behaviour, however it fails to recognise dispersal and shifts in habitat use due to changes in the landscape and environmental conditions [108]. Core home range measurements (i.e., 50% KUD in our study) focus on where animals spend the majority of their time, and they remove longer and more directional movements that are considered for the 95% KUD. We observed that many of the rehabilitated koalas gradually moved away from their release sites until they remained in one area more consistently. These gradual directional movements were captured in the 95% KUD, whereas the 50% KUD was more concentrated around the area where koalas completed the study (see Fig. 2).

Similar to some of the koalas in our study, Ream et al. [115] observed a lone female wolf dispersing a large distance and then settling into a well-defined home range, and intensively using two core areas. Spencer [130] suggested that the reason for this type of behaviour may simply be exploration that resulted in finding a good place to stay and settle down. While our KUD analyses represent the area used by koalas over the study period, the movements of some koalas (e.g. moving from unburnt to burnt habitat or dispersing after release from rehabilitation) may not be typical of home range use and extent at other times. Researchers should be mindful of these issues when interpreting the output from home range analyses [43, 67, 151]. In addition, end users, such as policy makers and conservation practitioners, need to use caution when translating this type of output into management recommendations. For example, there may be limited value in using “home range” size from rehabilitated koalas to understand the typical amount of area needed to support koalas in the Monaro region.

Our study, like many tracking studies, was restricted in scope by the logistical limitations of tracking many different individual animals across rugged terrain [123]. However, a larger number in each group would have provided more statistical power, particularly given the large variation between koalas in core home range sizes in our study area (5 ha to 280 ha). The “trend” observed in 50% KUD between rehabilitated and non-rehabilitated koalas ($p=0.07$) may have been significant with additional data. Notably, the variation that we observed between individual koala home ranges is not unusual and similar variability has been reported in other koala tracking studies [6, 39, 41, 69, 85, 88]. This demonstrates that, in any study, there is a need to consider how large variations in movement behaviour between animals might influence the capacity to address project aims when selecting the number of individuals to monitor.

Other findings

Other studies have found that movement parameters differ between males and females [30, 88, 120]. We found that male koalas moved farther each night on average than females, particularly during spring and summer. This is likely because spring and summer are the breeding season for koalas [70]. Male koalas also often have larger home ranges than females [41, 66, 88, 148] and tend to disperse farther [28]. In our study, home range sizes were similar between the sexes, and the longest movements or dispersals away from the release site were undertaken by females. This demonstrates that variability in movements is not explained by sex alone, but also by differences between individuals.

The home ranges observed in this population were generally larger than other studied koala populations. Other studies in NSW have reported home range sizes ranging from 4.9 to 58.9 ha using the fixed kernel method (FK) at 95% [48, 61, 66, 88], which is considerably smaller than the 95% KUD reported for the non-rehabilitated koalas in our study (250.8 hectares in the burnt and 194.7 hectares in the unburnt). Some studies in Queensland, however, have reported larger home ranges, including up to 135 ha in Blair Athol [39]. It is important to acknowledge that home range size estimates will vary depending on the method or technique used. Kernel methods are used to calculate utilisation distributions (i.e. distribution of an animal's position) by smoothing locational data and creating a density estimate [137, 152], and are thought to be more accurate measures compared to minimum convex polygons (MCPs) and the harmonic mean [36, 151, 153]. MCPs are thought to either under- or over-estimate home range size, depending on sample size [36] and over-estimates have also been documented for the harmonic mean [96]. For example, Goldingay and

Dobner [48] found larger home ranges when using MCPs (e.g., average of 37.4 ha compared to the 8 ha at FK 95%). Despite this, it is likely that the larger home ranges and movement distances in our study are not just an artefact of using different home range methodologies, but demonstrate that some koalas in the NSW Monaro use larger areas than koalas that have been studied elsewhere.

There are a number of reasons why home range sizes differ between populations in other mammal species. These include differences in the availability of resources [126], disturbance [101], environmental conditions (such as rainfall; [28]) and population densities [46, 124], with populations at lower density having larger home ranges than areas with higher densities [10]. Several of these factors likely contribute to the larger home ranges of koalas recorded in our study. For example, the Snowy Monaro region has a relatively low density koala population compared to some other areas, with Cristescu et al. [26] estimating 0.032 koalas per hectare in burnt habitat at Peak View, NSW and 0.041 koalas per hectare in unburnt habitat in Numeralla, NSW. Higher density populations in other parts of NSW are typically around 0.3 koalas per hectare [68], and can be much higher in other states (e.g., 10.1–18.4 koalas per hectare recorded in Cape Otway, Victoria between 2011 and 2013; [147]). The quality of food resources may be lower or more dispersed in our study area compared to other areas [132]. However, this region is one of the few areas in Australia where koala populations appear to be stable rather than declining [3]. Our study provides the first information about the movements and home ranges of this understudied population of koalas, which will assist with the development of local management plans.

Conclusions

Studies such as this one that document the movement ecology of species after fire play an important role in understanding the ecological implications of wildfire, and can help guide conservation efforts by evaluating how burnt landscapes are used. Incorporating an experimental design with “treatment” and “control” groups in the same study can help identify disturbance effects on animal movements and landscape use with more certainty than comparing data collected from different areas, with different populations and/or at different times. Our findings fill important knowledge gaps about post-rehabilitation release success and the value of burnt habitat to rehabilitated and non-rehabilitated animals, and can assist with decisions about the rescue and release of wildlife in burnt landscapes.

Encouragingly, despite the catastrophic wildfires of 2019–2020, our study found that burnt woodlands in the Monaro region of the NSW Southern Tablelands

provided adequate resources for koalas to persist and recover, which is further supported by the high body condition of all koalas at the end of the study [64]. In addition, apart from habitat that burnt in the extreme category, koalas spent proportionally more time in burnt than unburnt areas within their home ranges. The relatively mild conditions and higher than average rainfall from multiple years of La Niña after the fires also may have contributed positively to landscape and wildlife recovery. With wildfires predicted to increase in frequency and severity worldwide in response to climate change [60], the success of wildlife management and conservation efforts will likely depend on improving our understanding of the ecological impacts of fire on species and their habitats [122].

Supplementary Information

The online version contains supplementary material available at <https://doi.org/10.1186/s40462-024-00519-0>.

Additional file1 (DOCX 24 KB)

Acknowledgements

We thank the landholders around Peak View and Numeralla, NSW who allowed us to access their properties to track koalas and provided additional help and support, particularly James Fitzgerald and Matthew and Madeline Boucher. Rod Pietsch was instrumental in enabling this project to take place. We would also like to thank Georgina Andersen and Lachlan Wilmott for their guidance with home range analyses. Many other people, including Kimberley Schiphof, also assisted with various aspects of this research.

Author contributions

Conceptualisation, K.J.M., M.R.L. and K.N.Y.; methodology, K.J.M., M.R.L., K.N.Y. and R.G.C.; fieldwork and data collection, M.R.L., K.J.M. and J.S.; formal analysis, M.R.L., R.G.C. and K.J.M.; investigation, K.J.M., M.R.L., K.N.Y. and J.S.; data curation, M.R.L., K.J.M. and J.S.; writing—original draft preparation, M.R.L.; writing—review and editing, all authors; supervision, K.J.M., R.G.C. and K.N.Y.; project administration, K.J.M. and K.N.Y.; funding acquisition, K.J.M. and K.N.Y. All authors have read and agreed to the published version of the manuscript.

Funding

The project began with a \$100,000 donation from Two Thumbs Wildlife Trust in memory of Captain Ian McBeth, First Officer Paul Hudson and Flight Engineer Rick DeMorgan Jr, who lost their lives during firebombing operations on Thursday 23rd January 2020 when their C-130 Large Air Tanker crashed after dropping fire retardant on Two Thumbs Wildlife Trust's koala sanctuary at Peak View, NSW, Australia. Additional funding was provided by the Minderoo Foundation (TI_ANU001) and the NSW Government under the NSW Koala Strategy and the Australian Government Wildlife and Habitat Bushfire Recovery Program (DOC20/500625).

Availability of data and materials

All data available at: <https://datacommons.anu.edu.au/DataCommons/>

Declarations

Ethics approval

This research was approved by the NSW Government Department of Planning, Industry and Environment Animal Ethics Committee (permit number 200421/04) and was conducted in accordance with the Australian Code for the care and use of animals for scientific purposes.

Competing interests

The authors declare no competing interests.

Author details

¹Research School of Biology, The Australian National University, Canberra, ACT 2601, Australia. ²Fenner School of Environment and Society, The Australian National University, Canberra, ACT 2601, Australia. ³Research School of Finance, Actuarial Studies and Statistics, The Australian National University, Canberra, ACT 2601, Australia. ⁴Earthy Inc., Brisbane, QLD, Australia.

Received: 11 July 2024 Accepted: 22 November 2024

Published online: 03 December 2024

References

- Adams, R (2023) Investigating the effect of sampling method on assessments of koala (*Phascolarctos cinereus*) habitat. In: Honours Thesis, Australian national university.
- Aebischer NJ, Robertson PA, Kenward RE. Compositional analysis of habitat use from animal radio-tracking data. *Ecology*. 1993;74(5):1313–25.
- Allen, C (2010) estimating koala numbers & assessing population trends in south eastern NSW—information prepared for the threatened species scientific committee to assist its assessment on the listing of the koala as a threatened species under the EPBC Act. Report.
- Allred BW, Fuhlendorf SD, Engle DM, Elmore RD. Ungulate preference for burned patches reveals strength of fire–grazing interaction. *Ecol Evol*. 2011;1(2):132–44.
- Archibald S, Bond W, Stock W, Fairbanks D. Shaping the landscape: fire–grazer interactions in an African savanna. *Ecol Appl*. 2005;15(1):96–109.
- Ashman KR, Page NR, Whisson DA. Ranging behavior of an arboreal marsupial in a plantation landscape. *J Wildl Manag*. 2020;84(6):1091–9.
- Attwill PM, Adams MA. Mega-fires, inquiries and politics in the eucalypt forests of Victoria, south-eastern Australia. *For Ecol Manage*. 2013;294:45–53.
- Badyaev AV, Etges WJ, Martin TE. Ecological and behavioral correlates of variation in seasonal home ranges of wild turkeys. *J Wildl Manag*. 1996;60(1):154–64.
- Beamant JE, Mulligan C, Moore C, Mitchell D, Narayan E, Burke da Silva K. Resident wild koalas show resilience to large-scale translocation of bushfire-rescued koalas. *Conser Physiol*. 2023;11(1):1–12.
- Bengsen A, Algar D, Ballard G, Buckmaster T, Comer S, Fleming PJ, Friend J, Johnston M, McGregor H, Moseby K. Feral cat home-range size varies predictably with landscape productivity and population density. *J Zool*. 2016;298(2):112–20.
- Beringer J, Mabry P, Meyer T, Wallendorf M, Eddleman WR. Post-release survival of rehabilitated white-tailed deer fawns in Missouri. *Wildl Soc Bull*. 2004;32(3):732–8.
- Berry L, Lindenmayer D, Dennis T, Driscoll D, Banks S. Fire severity alters spatio-temporal movements and habitat utilisation by an arboreal marsupial, the mountain brushtail possum (*Trichosurus cunninghami*). *Int J Wildland Fire*. 2016;25(12):1291–302.
- Bonte D, Van Dyck H, Bullock JM, Coulon A, Delgado M, Gibbs M, Lehoucq V, Matthysen E, Mustin K, Saastamoinen M. Costs of dispersal. *Biol Rev*. 2012;87(2):290–312.
- Bradstock RA, Williams RJ, Gill AM. Flammable Australia: fire regimes, biodiversity and ecosystems in a changing world. Melbourne: CSIRO publishing; 2012.
- Bureau of Meteorology (2020) Annual climate summary for new south wales—new south wales in 2019: record warm and record dry. Available at <http://www.bom.gov.au/climate/current/annual/nsw/archive/2019.summary.shtml> [Verified June 8]
- Burrows G. Epicormic strand structure in *Angophora*, *Eucalyptus* and *Lophostemon* (*Myrtaceae*): implications for fire resistance and recovery. *New Phytol*. 2002;153(1):111–31.
- Burt WH. Territoriality and home range concepts as applied to mammals. *J Mammal*. 1943;24(3):346–52.

18. Calenge C. The package “adehabitat” for the R software: a tool for the analysis of space and habitat use by animals. *Ecol Model.* 2006;197(3–4):516–9.
19. Calenge C, and Basille M (2024) adehabitatHS: analysis of habitat selection by animals (version 0.3.18). Available at <https://cran.r-project.org/web/packages/adehabitatHS/index.html>
20. Calenge C, and Fortmann-Roe S (2011) Home range estimation in R: the adehabitatHR package. Available at <https://CRAN.R-project.org/package=adehabitatHR>
21. Cederlund G, Sand H. Home-range size in relation to age and sex in moose. *J Mammal.* 1994;75(4):1005–12.
22. Cook, A (2010) Habitat use and home-range of the northern quoll, *Dasyurus hallucatus*: effects of fire. In: Master of science in natural resource management thesis, University of western Australia.
23. Cooper WE Jr. Home range criteria based on temporal stability of areal occupation. *J Theor Biol.* 1978;73(4):687–95.
24. Cope HR, McArthur C, Dickman CR, Newsome TM, Gray R, Herbert CA. A systematic review of factors affecting wildlife survival during rehabilitation and release. *PLoS ONE.* 2022;17(3): e0265514.
25. Cope HR, McArthur C, Gray R, Newsome TM, Dickman CR, Sriram A, Haering R, Herbert CA. Trends in rescue and rehabilitation of marsupials surviving the Australian 2019–2020 bushfires. *Animals.* 2024;14(7):1019.
26. Cristescu RH, Gardiner R, Terrabea J, McDonald K, Powell D, Levensgood AL, Frère CH. Difficulties of assessing the impacts of the 2019–2020 bushfires on koalas. *Austral Ecol.* 2023;48(1):12–8.
27. Crouner AW, Barrett GW. Effects of fire on the small mammal component of an experimental grassland community. *J Mammal.* 1979;60(4):803–13.
28. Davies N, Gramotnev G, Seabrook L, Bradley A, Baxter G, Rhodes J, Lunney D, McAlpine C. Movement patterns of an arboreal marsupial at the edge of its range: a case study of the koala. *Mov Ecol.* 2013;1(1):1–15.
29. Davies NA, Gramotnev G, McAlpine C, Seabrook L, Baxter G, Lunney D, Rhodes JR, Bradley A. Physiological stress in koala populations near the arid edge of their distribution. *PLoS ONE.* 2013;8(11): e79136.
30. De Oliveira S, Murray P, De Villiers D, Baxter G. Ecology and movement of urban koalas adjacent to linear infrastructure in coastal south-east Queensland. *Aust Mammal.* 2013;36(1):45–54.
31. Degabriele R. A relative shortage of nitrogenous food in the ecology of the koala (*Phascolarctos cinereus*). *Aust J Ecol.* 1981;6(2):139–41.
32. Dickman CR, Hutchings P, Law B, Lunney D. Raking over the ashes: assessing the impact of fire on native fauna in the aftermath of Australia’s 2019–2020 fires. *Aust Zool.* 2022;42(2):643–53.
33. Dingle H. Migration: the biology of life on the move. 2nd ed. Oxford: Oxford University Press; 2014.
34. Doherty TS, Davis RA, van Etten EJ, Collier N, Virens E. Response of a shrubland mammal and reptile community to a history of landscape-scale wildfire. *Int J Wildland Fire.* 2015;24(4):534–43.
35. Doherty TS, Fist CN, Driscoll DA. Animal movement varies with resource availability, landscape configuration and body size: a conceptual model and empirical example. *Landsc Ecol.* 2019;34:603–14.
36. Downs JA, Horner MW. Effects of point pattern shape on home-range estimates. *J Wildl Manag.* 2008;72(8):1813–8.
37. Driscoll DA, Armenteras D, Bennett AF, Brotons L, Clarke MF, Doherty TS, Haslem A, Kelly LT, Sato CF, Sitters H. How fire interacts with habitat loss and fragmentation. *Biol Rev.* 2021;96(3):976–98.
38. Dunstan E, Funnell O, McLelland J, Stoeckeler F, Nishimoto E, Mitchell D, Mitchell S, McLelland DJ, Kalvas J, Johnson L. An analysis of demographic and triage assessment findings in bushfire-affected koalas (*Phascolarctos cinereus*) on Kangaroo Island, South Australia, 2019–2020. *Animals.* 2021;11(11):3237.
39. Ellis W, Melzer A, Carrick F, Hasegawa M. Tree use, diet and home range of the koala (*Phascolarctos cinereus*) at Blair Athol, central Queensland. *Wildl Res.* 2002;29(3):303–11.
40. Ellis W, White N, Kunst N, Carrick F. Response of koalas (*Phascolarctos-Cinereus*) to reintroduction to the wild after rehabilitation. *Wildl Res.* 1990;17(4):421–6.
41. Ellis WA, Melzer A, Bercovitch FB. Spatiotemporal dynamics of habitat use by koalas: the checkerboard model. *Behav Ecol Sociobiol.* 2009;63:1181–8.
42. Engstrom RT. First-order fire effects on animals: review and recommendations. *Fire Ecology.* 2010;6:115–30.
43. Fieberg J, Börger L. Could you please phrase “home range” as a question? *J Mammal.* 2012;93(4):890–902.
44. Gardiner R, Proft K, Comte S, Jones M, Johnson CN. Home range size scales to habitat amount and increasing fragmentation in a mobile woodland specialist. *Ecol Evol.* 2019;9(24):14005–14.
45. Gaydos JK, Ignacio Vilchis L, Lance MM, Jeffries SJ, Thomas A, Greenwood V, Harner P, Ziccardi MH. Postrelease movement of rehabilitated harbor seal (*Phoca vitulina richardii*) pups compared with cohort-matched wild seal pups. *Mar Mamm Sci.* 2013;29(3):E282–94.
46. Glessner KD, Britt A. Population density and home range size of *Indri indri* in a protected low altitude rain forest. *Int J Primatol.* 2005;26:855–72.
47. Goldenberg SZ, Parker JM, Chege SM, Greggor AL, Hunt M, Lamberski N, Leigh KA, Nollens HH, Ruppert KA, Thouless C. Revisiting the 4 R’s: Improving post-release outcomes for rescued mammalian wildlife by fostering behavioral competence during rehabilitation. *Front Conserv Sci.* 2022;3: 910358.
48. Goldingay RL, Dobner B. Home range areas of koalas in an urban area of north-east New South Wales. *Aust Mammal.* 2013;36(1):74–80.
49. Gould L, Mackenzie J, and Lynch J (2019) Snowy Monaro biodiversity study. In: Molonglo conservation consulting, report.
50. Green K. Winter home range and foraging of common wombats (*Vombatus ursinus*) in patchily burnt subalpine areas of the Snowy Mountains. *Aust Wildl Res.* 2005;32(6):525–9.
51. Hanya G, Kiyono M, Yamada A, Suzuki K, Furukawa M, Yoshida Y, Chijiwa A. Not only annual food abundance but also fallback food quality determines the Japanese macaque density: evidence from seasonal variations in home range size. *Primates.* 2006;47:275–8.
52. Harris S, Cresswell W, Forde P, Trehwella W, Woollard T, Wray S. Home-range analysis using radio-tracking data—a review of problems and techniques particularly as applied to the study of mammals. *Mammal Rev.* 1990;20(2–3):97–123.
53. Hernandez SM. Postrehabilitation release monitoring of wildlife. In: Hernandez SM, Barron HW, Miller EA, Aguilar RF, Yabsley MJ, editors. Medical management of wildlife species: a guide for practitioners. Hoboken: John Wiley & Sons; 2019. p. 123–7.
54. Herzog NM, Parker CH, Keefe ER, Coxworth J, Barrett A, Hawkes K. Fire and home range expansion: a behavioral response to burning among savanna dwelling vervet monkeys (*Chlorocebus aethiops*). *Am J Phys Anthropol.* 2014;154(4):554–60.
55. Hijmans RJ (2024) raster: geographic data analysis and modeling, version 3.6–30. Available at <https://cran.r-project.org/web/packages/raster/index.html>
56. Hindell MA, Lee AK. Tree use by individual koalas in a natural forest. *Wildl Res.* 1988;15(1):1–7.
57. Hradsky BA. Conserving Australia’s threatened native mammals in predator-invaded, fire-prone landscapes. *Wildl Res.* 2020;47(1):1–15.
58. Jhariya MK, Raj A. Effects of wildfires on flora, fauna and physico-chemical properties of soil—An overview. *J Appl Nat Sci.* 2014;6(2):887–97.
59. Joly K, Duffy PA, Rupp TS. Simulating the effects of climate change on fire regimes in Arctic biomes: implications for caribou and moose habitat. *Ecosphere.* 2012;3(5):1–18.
60. Jones MW, Smith A, Betts R, Canadell JG, Prentice IC, Le Quére C. Climate change increases the risk of wildfires. *ScienceBrief Rev.* 2020;116:117.
61. Kavanagh RP, Stanton MA, Brassil TE. Koalas continue to occupy their previous home-ranges after selective logging in *Callitris-Eucalyptus* forest. *Wildl Res.* 2007;34(2):94–107.
62. Keith DA, Bedward M. Native vegetation of the South East Forests region. Eden New South Wales Cunninghamia. 1999;6(1):1–218.
63. Klop E, van Goethem J, de longh HH. Resource selection by grazing herbivores on post-fire regrowth in a West African woodland savanna. *Wildl Res.* 2007;34(2):77–83.
64. Lane MR, Lowe A, Vukcevic J, Clark RG, Madani G, Higgins DP, Silver L, Belov K, Hogg CJ, Marsh KJ. Health assessments of koalas after wildfire: a temporal comparison of rehabilitated and non-rescued resident individuals. *Animals.* 2023;13(18):2863.
65. Lane MR, Youngentob KN, Clark RG, Marsh KJ. The nutritional quality of post-fire eucalypt regrowth and its consumption by koalas in the

- new south wales southern Tablelands. *Aust J Zool.* 2024. <https://doi.org/10.1071/ZO23024>.
66. Lassau SA, Ryan B, Close R, Moon C, Geraghty P, Coyle A, Pile J. Home ranges and mortality of a roadside Koala *Phascolarctos cinereus* population at Bonville, New South Wales. *Aust Zool.* 2008;34:127–36.
 67. Laver PN, Kelly MJ. A critical review of home range studies. *J Wildl Manag.* 2008;72(1):290–8.
 68. Law B, Gonsalves L, Burgar J, Brassil T, Kerr I, Wilmott L, Madden K, Smith M, Mella V, Crowther M. Estimating and validating koala *Phascolarctos cinereus* density estimates from acoustic arrays using spatial count modelling. *Wildl Res.* 2021;49(5):438–48.
 69. Law B, Gonsalves L, Slade C, Brassil T, Flanagan C. GPS tracking reveals koalas *Phascolarctos cinereus* use mosaics of different forest ages after environmentally regulated timber harvesting. *Austral Ecol.* 2024;49(4): e13518.
 70. Law BS, Brassil T, Gonsalves L, Roe P, Truskinger A, McConville A. Passive acoustics and sound recognition provide new insights on status and resilience of an iconic endangered marsupial (koala *Phascolarctos cinereus*) to timber harvesting. *PLoS ONE.* 2018;13(10): e0205075.
 71. Law BS, Gonsalves L, Burgar J, Brassil T, Kerr I, O'Loughlin C. Fire severity and its local extent are key to assessing impacts of Australian megafires on koala (*Phascolarctos cinereus*) density. *Glob Ecol Biogeogr.* 2022;31(4):714–26.
 72. Leavesley, A, and Rhind S (2008) Vertebrate fauna of monaro plains conservation reserves. In: Report, department of environment and climate change, queanbeyan, New South Wales, Australia.
 73. Legge S, Rumpff L, Woinarski JC, Whiterod NS, Ward M, Southwell DG, Scheele BC, Nimmo DG, Lintermans M, Geyle HM. The conservation impacts of ecological disturbance: time-bound estimates of population loss and recovery for fauna affected by the 2019–2020 Australian megafires. *Glob Ecol Biogeogr.* 2022;31(10):2085–104.
 74. Leigh KA, Hofweber LN, Sloggett BK, Inman VL, Pettit LJ, Sriram A, Haering R. Outcomes for an arboreal folivore after rehabilitation and implications for management. *Sci Rep.* 2023;13(1):6542.
 75. Lenth, R (2021) 'emmeans: estimated marginal means, aka Least-Square Means. R package 687 version 1.6.3.' Available at <https://CRAN.R-project.org/package=emmeans>
 76. Lindstedt SL, Miller BJ, Buskirk SW. Home range, time, and body size in mammals. *Ecology.* 1986;67(2):413–8.
 77. Linnell JD, Aanes R, Swenson JE, Odden J, Smith ME. Translocation of carnivores as a method for managing problem animals: a review. *Biodivers Conserv.* 1997;6:1245–57.
 78. Logan M, Sanson GD. The effects of tooth wear on the activity patterns of free-ranging koalas (*Phascolarctos cinereus* Goldfuss). *Aust J Zool.* 2002;50(3):281–92.
 79. Lucas, C, Hennessy K, Mills G, and Bathols J (2007) Bushfire weather in southeast australia: recent trends and projected climate change impacts. In: Report, bushfire cooperative research centre, Melbourne, Victoria, Australia.
 80. Lunney D, Cope H, Sonawane I, Haering R. A state-wide picture of koala rescue and rehabilitation in New South Wales during the 2019–2020 bushfires. *Aust Zool.* 2022;42(2):243–55.
 81. Lunney D, Gresser SM, Mahon PS, Matthews A. Post-fire survival and reproduction of rehabilitated and unburnt koalas. *Biol Cons.* 2004;120(4):567–75.
 82. MacGregor CI, Wood JT, Dexter N, Lindenmayer DB. Home range size and use by the long-nosed bandicoot (*Perameles nasuta*) following fire. *Aust Mammal.* 2013;35(2):206–16.
 83. Madani G, Ashman K, Mella V, Whisson D. A review of the 'noose and flag' method to capture free-ranging koalas. *Aust Mammal.* 2020;42(3):341–8.
 84. Martin R. Age-specific fertility in three populations of the koala, *Phascolarctos cinereus* Goldfuss. *Vic Wildl Res.* 1981;8(2):275–83.
 85. Martin RH, Ann K. The koala: natural history, conservation and management. Kensington: UNSW press; 1999.
 86. Martin S, Youngentob KN, Clark RG, Foley WJ, Marsh KJ. The distribution and abundance of an unusual resource for koalas (*Phascolarctos cinereus*) in a sodium-poor environment. *PLoS ONE.* 2020;15(6): e0234515.
 87. Masters P. The effects of fire-driven succession and rainfall on small mammals in spinifex grassland at Uluru national park. *North Territ Wildl Res.* 1993;20(6):803–13.
 88. Matthews A, Lunney D, Gresser S, Maitz W. Movement patterns of koalas in remnant forest after fire. *Aust Mammal.* 2016;38(1):91–104.
 89. McGregor D, Nordberg E, Yoon H-J, Youngentob K, Schwarzkopf L, Krockenberger A. Comparison of home range size, habitat use and the influence of resource variations between two species of greater gliders (*Petauroides minor* and *Petauroides volans*). *PLoS ONE.* 2023;18(10): e0286813.
 90. McNab BK. Bioenergetics and the determination of home range size. *Am Nat.* 1963;97(894):133–40.
 91. Melzer A, Carrick F, Menkhorst P, Lunney D, John BS. Overview, critical assessment, and conservation implications of koala distribution and abundance. *Conserv Biol.* 2000;14(3):619–28.
 92. Moe SR, Wegge P, Kapela EB. The influence of man-made fires on large wild herbivores in Lake Burungi area in northern Tanzania. *Afr J Ecol.* 1990;28(1):35–43.
 93. Molony S, Baker P, Garland L, Cuthill I, Harris S. Factors that can be used to predict release rates for wildlife casualties. *Anim Welf.* 2007;16(3):361–7.
 94. Morellet N, Bonenfant C, Börger L, Ossi F, Cagnacci F, Heurich M, Kjellander P, Linnell JD, Nicoloso S, Sustr P. Seasonality, weather and climate affect home range size in roe deer across a wide latitudinal gradient within Europe. *J Anim Ecol.* 2013;82(6):1326–39.
 95. Myers PJ, Young JK. Post-release activity and habitat selection of rehabilitated black bears. *Human Wildl Interact.* 2018;12(3):322–37.
 96. Naef-Daenzer B. A new transmitter for small animals and enhanced methods of home-range analysis. *J Wildl Manag.* 1993;57(4):680–9.
 97. Narayan E. Physiological stress levels in wild koala sub-populations facing anthropogenic induced environmental trauma and disease. *Sci Rep.* 2019;9(1):1–9.
 98. Nichols JD, Armstrong DP. Monitoring for reintroductions. In: Ewen JG, Armstrong DP, Parker KA, Seddon PJ, editors. *Reintroduction biology: integrating science and management.* Oxford: Blackwell Publishing Ltd; 2012. p. 223–55.
 99. Nimmo D, Kelly L, Farnsworth L, Watson S, Bennett A. Why do some species have geographically varying responses to fire history? *Ecography.* 2014;37(8):805–13.
 100. Nimmo DG, Avitabile S, Banks SC, Bliege Bird R, Callister K, Clarke MF, Dickman CR, Doherty TS, Driscoll DA, Greenville AC. Animal movements in fire-prone landscapes. *Biol Rev.* 2019;94(3):981–98.
 101. O'Donnell K, delBarco-Trillo J. Changes in the home range sizes of terrestrial vertebrates in response to urban disturbance: a meta-analysis. *J Urban Ecol.* 2020. <https://doi.org/10.1093/jue/juaa014>.
 102. Office of Environment and Heritage NSW National Parks and Wildlife Services (2012) Plan of management: northern Monaro reserves. In: Report, office of environment and heritage NSW.
 103. Palm EC, Sutor MJ, Joly K, Herriges JD, Kelly AP, Hervieux D, Russell KL, Bentzen TW, Larter NC, Hebblewhite M. Increasing fire frequency and severity will increase habitat loss for a boreal forest indicator species. *Ecol Appl.* 2022;32(3): e2549.
 104. Parrott ML, Wicker LV, Lamont A, Banks C, Lang M, Lynch M, McMeekin B, Miller KA, Ryan F, Selwood KE. Emergency response to Australia's black summer 2019–2020: the role of a zoo-based conservation organisation in wildlife triage, rescue, and resilience for the future. *Animals.* 2021;11(6):1515.
 105. Pasch B, Koprowski JL. Impacts of fire suppression on space use by Mexican fox squirrels. *J Mammal.* 2011;92(1):227–34.
 106. Pausas JG, Keeley JE. Epicormic resprouting in fire-prone ecosystems. *Trends Plant Sci.* 2017;22(12):1008–15.
 107. Phillips S, Wallis K, Lane A. Quantifying the impacts of bushfire on populations of wild koalas (*Phascolarctos cinereus*): insights from the 2019/20 fire season. *Ecol Manag Restor.* 2021;22(1):80–8.
 108. Powell, RA, and Boitani L (2012) Movements, home ranges, activity, and dispersal. In: *Carnivore ecology and conservation: a handbook of techniques, techniques in ecology and conservation* Oxford university press, London, United Kingdom. pp. 188–217. (Oxford University Press: London, United Kingdom)

109. Pyke, G (1983) Animal movements: an optimal foraging approach. In *The ecology of animal movement*. (Ed. IR Swingland and PJ Greenwood). (Clarendon Press Oxford).
110. QGIS Development Team (2021) QGIS Geographic Information System.
111. Qin Y, Xiao X, Wigner J-P, Ciais P, Canadell JG, Brandt M, Li X, Fan L, Wu X, Tang H. Large loss and rapid recovery of vegetation cover and aboveground biomass over forest areas in Australia during 2019–2020. *Remote Sens Environ*. 2022;278: 113087.
112. R Core Team (2020) R: A language and environment for statistical computing. In: R foundation for statistical computing, Vienna, Austria.
113. Radford Miller, SL (2012) Aspects of the ecology of the koala, *Phascolarctos cinereus*, in a tall coastal production forest in north eastern New South Wales. PhD Thesis, Southern Cross University.
114. Ramsay, S (1999) The ecology and dispersal patterns of juvenile koalas, *Phascolarctos cinereus*, in fragmented habitat. PhD Thesis, The University of Sydney.
115. Ream R, Harris R, Smith J, Boyd D. Movement patterns of a lone wolf, *Canis lupus*, in unoccupied wolf range, southeastern British Columbia. *Can Field-Nat*. 1985;99(2):234–9.
116. Rhodes JR, Wiegand T, McAlpine CA, Callaghan J, Lunney D, Bowen M, Possingham HP. Modeling species' distributions to improve conservation in semiurban landscapes: koala case study. *Conserv Biol*. 2006;20(2):449–59.
117. Rivrud IM, Loe LE, Myrsetrud A. How does local weather predict red deer home range size at different temporal scales? *J Anim Ecol*. 2010;79(6):1280–95.
118. Roff D. Dispersal in dipterans: its costs and consequences. *The J Animal Ecol*. 1977;46:443–56.
119. Russell-Smith J, Yates CP, Whitehead PJ, Smith R, Craig R, Allan GE, Thackway R, Frakes I, Cridland S, Meyer MC. Bushfires 'down under': patterns and implications of contemporary Australian landscape burning. *Int J Wildl Fire*. 2007;16(4):361–77.
120. Ryan MA, Whisson DA, Holland GJ, Arnould JP. Activity patterns of free-ranging koalas (*Phascolarctos cinereus*) revealed by accelerometry. *PLoS ONE*. 2013;8(11): e80366.
121. Saïd S, Gaillard JM, Widmer O, Débias F, Bourgoïn G, Delorme D, Roux C. What shapes intra-specific variation in home range size? A case study of female roe deer. *Oikos*. 2009;118(9):1299–306.
122. Santos JL, Hradsky BA, Keith DA, Rowe KC, Senior KL, Sitters H, Kelly LT. Beyond inappropriate fire regimes: a synthesis of fire-driven declines of threatened mammals in Australia. *Conserv Lett*. 2022;15(5): e12905.
123. Saunders D, Nguyen H, Cowen S, Magrath M, Marsh K, Bell S, Bobruk J. Radio-tracking wildlife with drones: a viewshed analysis quantifying survey coverage across diverse landscapes. *Wildl Res*. 2022;49(1):1–10.
124. Schoepf I, Schmohl G, König B, Pillay N, Schradin C. Manipulation of population density and food availability affects home range sizes of African striped mouse females. *Anim Behav*. 2015;99:53–60.
125. Schradin C, Pillay N. Female striped mice (*Rhabdomys pumilio*) change their home ranges in response to seasonal variation in food availability. *Behav Ecol*. 2006;17(3):452–8.
126. Schradin C, Schmohl G, Rödel HG, Schoepf I, Treffler SM, Brenner J, Bleeker M, Schubert M, König B, Pillay N. Female home range size is regulated by resource distribution and intraspecific competition: a long-term field study. *Anim Behav*. 2010;79(1):195–203.
127. Shaffer KE, Laudenslayer W Jr, Hedwall SJ. Fire and animal interactions. In: *Wagtendonk JW, Sugihara NG, Stephens SL, Thode AE, Shaffer KE, Fites-Kaufman JA, editors. Fire in California's ecosystems*. 2nd ed. California: University of California Press; 2018. p. 118–44.
128. Shaw RE, James AI, Tuft K, Legge S, Cary GJ, Peakall R, Banks SC. Unburnt habitat patches are critical for survival and in situ population recovery in a small mammal after fire. *J Appl Ecol*. 2021;58(6):1325–35.
129. Silva JA, Nielsen S, McLoughlin PD, Rodgers AR, Hague C, Boutin S. Comparison of pre-fire and post-fire space use reveals varied responses by woodland caribou (*Rangifer tarandus caribou*) in the Boreal Shield. *Can J Zool*. 2020;98(11):751–60.
130. Spencer WD. Home ranges and the value of spatial information. *J Mammal*. 2012;93(4):929–47.
131. State Government of NSW and Department of Planning and Environment (2020) Fire extent and severity mapping. Available at <https://datasets.seed.nsw.gov.au/dataset/fire-extent-and-severity-mapping-fesm>. (FESM v3).
132. State of NSW and Office of Environment and Heritage (2011) plant communities of the south eastern highlands and Australian Alps within the murrumbidgee catchment of new south wales. In: Version 1.1. a report to catchment action NSW. NSW office of environment and heritage; department of premier and cabinet, queanbeyan. report, office of environment and heritage NSW, Queanbeyan, NSW, Australia.
133. Sutherland EF, Dickman CR. Mechanisms of recovery after fire by rodents in the Australian environment: a review. *Wildl Res*. 1999;26(4):405–19.
134. Thompson MB, Medlin G, Hutchinson R, West N. Short-term effects of fuel reduction burning on populations of small terrestrial mammals. *Wildl Res*. 1989;16(2):117–29.
135. Tolhurst B, Grogan A, Hughes H, Scott D. Effects of temporary captivity on ranging behaviour in urban red foxes (*Vulpes vulpes*). *Appl Anim Behav Sci*. 2016;181:182–90.
136. Van Eeden, LM, Nimmo D, Mahony M, Herman K, Ehmke G, Driessen J, O'Connor J, Bino G, Taylor M, and Dickman C (2020) Impacts of the unprecedented 2019–2020 bushfires on Australian animals. Report.
137. Van Winkle W. Comparison of several probabilistic home-range models. *J Wildl Manag*. 1975;39(1):118–23.
138. Vernes K, Pope LC. Stability of nest range, home range and movement of the northern bettong (*Bettongia tropica*) following moderate-intensity fire in a tropical woodland, north-eastern Queensland. *Wildl Res*. 2001;28(2):141–50.
139. Vogelnest LW, Rupert, Veterinary considerations for the rescue, treatment, rehabilitation and release of wildlife. In 'medicine of Australian mammals'. Collingwood: CSIRO Publishing; 2008. p. 1–12.
140. Ward M, Tulloch A, Stewart R, Possingham HP, Legge S, Gallagher RV, Graham EM, Southwell D, Keith D, Dixon K. Restoring habitat for fire-impacted species' across degraded Australian landscapes. *Environ Res Lett*. 2022;17(8): 084036.
141. Ward M, Tulloch AI, Radford JQ, Williams BA, Reside AE, Macdonald SL, Mayfield HJ, Maron M, Possingham HP, Vine SJ. Impact of 2019–2020 mega-fires on Australian fauna habitat. *Nat Ecol Evol*. 2020;4(10):1321–6.
142. Watchorn DJ, Whisson DA. Quantifying the interactions between koalas in a high-density population during the breeding period. *Aust Mammal*. 2019;42(1):28–37.
143. Watson S, Taylor R, Nimmo D, Kelly L, Clarke M, Bennett A. The influence of unburnt patches and distance from refuges on post-fire bird communities. *Anim Conserv*. 2012;15(5):499–507.
144. Watson SJ, Taylor RS, Nimmo DG, Kelly LT, Haslem A, Clarke MF, Bennett AF. Effects of time since fire on birds: how informative are generalized fire response curves for conservation management? *Ecol Appl*. 2012;22(2):685–96.
145. Weigler BJ, Girjes AA, White NA, Kunst ND, Carrick FN, Lavin MF. Aspects of the epidemiology of Chlamydia psittaci infection in a population of koalas (*Phascolarctos cinereus*) in southeastern Queensland. *Aust J Wildl Dis*. 1988;24(2):282–91.
146. Whelan R, Rodgerson L, Dickman C, Sutherland E. Critical life cycles of plants and animals: developing a process-based understanding of population changes in fire-prone landscapes. In: Bradstock RA, Williams JE, Gill MA, editors. *Flammable Australia: the fire regimes and biodiversity of a continent*. Cambridge: Cambridge University Press; 2002. p. 94–124.
147. Whisson DA, Dixon V, Taylor ML, Melzer A. Failure to respond to food resource decline has catastrophic consequences for koalas in a high-density population in southern Australia. *PLoS ONE*. 2016;11(1): e0144348.
148. White NA. Ecology of the koala (*Phascolarctos cinereus*) in rural south-east Queensland. *Aust Wildl Res*. 1999;26(6):731–44.
149. Wintle BA, Legge S, Woinarski JC. After the megafires: What next for Australian wildlife? *Trends Ecol Evol*. 2020;35(9):753–7.
150. Witt RR, Beranek CT, Howell LG, Ryan SA, Clulow J, Jordan NR, Denholm B, Roff A. Real-time drone derived thermal imagery

- outperforms traditional survey methods for an arboreal forest mammal. *PLoS ONE*. 2020;15(11): e0242204.
151. Worton B. A review of models of home range for animal movement. *Ecol Model*. 1987;38(3–4):277–98.
 152. Worton BJ. Kernel methods for estimating the utilization distribution in home-range studies. *Ecology*. 1989;70(1):164–8.
 153. Worton BJ. Using Monte Carlo simulation to evaluate kernel-based home range estimators. *J Wildl Manag*. 1995;59(4):794–800.
 154. Yoder JM, Marschall EA, Swanson DA. The cost of dispersal: predation as a function of movement and site familiarity in ruffed grouse. *Behav Ecol*. 2004;15(3):469–76.
 155. Youngentob KN, Wood JT, Lindenmayer DB. The response of arboreal marsupials to landscape context over time: a large-scale fragmentation study revisited. *J Biogeogr*. 2013;40(11):2082–93.

Publisher's Note

Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.