The International Vertebrate Pet Trade Network and Insights from US Imports of Exotic Pets

JAMES S. SINCLAIR®, OLIVER C. STRINGHAM, BRADLEY UDELL, NICHOLAS E. MANDRAK, BRIAN LEUNG, CHRISTINA M. ROMAGOSA, AND JULIE L. LOCKWOOD

The international trade in exotic vertebrate pets provides key social and economic benefits but also drives associated ecological, ethical, and human health impacts. However, despite its clear importance, we currently lack a full understanding of the structure of the pet trade, hampering efforts to optimize its benefits while mitigating its negative effects. In the present article, we represent and review the structure of the pet trade as a network composed of different market actors (nodes) and trade flows (links). We identify key data gaps in this network that, if filled, would enable network analyses to pinpoint targets for management. As a case study of how data-informed networks can realize this goal, we quantified spatial and temporal patterns in pets imported to the United States. Our framework and case study illustrate how network approaches can help to inform and manage the effects of the growing demand for exotic pets.

Keywords: pet trade, biodiversity, disease, invasion, animal welfare

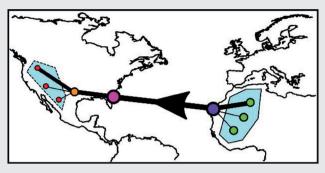
he legal, international transport of undomesticated vertebrates as pets (hereafter, *pet trade*) is widespread and growing (Bush et al. 2014, Lockwood et al. 2019), involving billions of dollars in trade and millions of animals each year (Karesh et al. 2007, Bush et al. 2014). The pet trade is therefore a key economic resource, particularly for local traders that depend on it for their income (e.g., Raghavan et al. 2013, Robinson et al. 2018). Pets also provide important social and human health benefits (McNicholas et al. 2005), even being considered as family members within some cultures (Lockwood et al. 2019). However, the pet trade is also responsible for the widespread translocation of biota and associated pathogens across the globe, which can have a variety of negative impacts, including species harvested to near extinction (Nijman et al. 2012, Poole and Shepherd 2017), animal abuse (Baker et al. 2013, Warwick 2014, Elwin et al. 2020), biological invasions (Kraus 2008, Lockwood et al. 2019), and the introduction of new zoonotic diseases that harm native species and human health (Chomel et al. 2007, Smith et al. 2017, O'Hanlon et al. 2018). Retaining the positive benefits of the pet trade while addressing its associated negative impacts therefore requires an understanding of the full structure of, and flow of animals through, the pet trade commodity chain.

Despite the broad scope and clear importance of the pet trade on a variety of economic, cultural, ecological, ethical, and human health systems, a comprehensive understanding and approach to mitigating the negative effects of this trade remains elusive (Rhyne et al. 2017, Smith et al. 2017, Scheffers et al. 2019). This implementation gap persists because, outside of select taxonomic groups, there is scarce empirical information on where wild animals are initially harvested to supply the pet market or the precise volume traded between regions. We also lack critical information on the full spectrum of species traded in these markets, the dynamics of traded species over space and time, and how imported pets are geographically distributed in the countries into which they are sold. Data that are readily available often come from wildlife export and import records (e.g., Rhyne et al. 2012, Bush et al. 2014, Reino et al. 2017), which frequently address trade only at the coarse scale of whole countries, are not specific to the pet trade (e.g., include food and product animals), or misidentify the species involved (Rhyne et al. 2012). Furthermore, available pet trade data that are specific to finer-scale activities are typically highly species specific (e.g., Rabemananjara et al. 2008, Nijman et al. 2012, Martin et al. 2018). Taken together, the limited empirical understanding of the pet trade hampers our ability to extrapolate to the broader trade commodity chain, which consists of a variety of species traded across a complex web of actors spanning multiple spatial and temporal scales. Without this more comprehensive view of the pet trade, the suite of policy and enforcement solutions that can mitigate its negative effects remain opaque and therefore difficult to realize in practice.

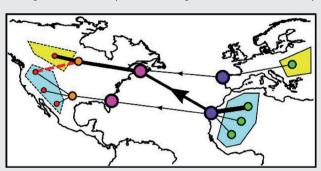
BioScience 71: 977–990. © The Author(s) 2021. Published by Oxford University Press on behalf of the American Institute of Biological Sciences. This is an Open Access article distributed under the terms of the Creative Commons Attribution License (http://creativecommons.org/licenses/by/4.0/), which permits unrestricted reuse, distribution, and reproduction in any medium, provided the original work is properly cited doi:10.1093/biosci/biab056 Advance Access publication 9 June 2021

Box 1. Insights from pet trade networks.

Pre- and postimport networks can be used to improve estimates of source harvest and the risks associated with pet introduction, which we illustrate using a hypothetical and simplified example of a trade network (right figure). In this example, pets are initially collected across a network of geographically dispersed harvesting locations (green nodes) and consolidated at an export location (purple node). Following arrival at the port of entry (pink node), pets are then redistributed into successively more geographically dispersed areas (orange then red nodes). On the basis of the nodes and link strengths (represented using line thickness) in this network, we can determine that the greatest harvest intensity occurs in the topmost green node, that the purple and pink nodes act as key transport hubs, and that



most pets are eventually sold in the topmost red node, which may therefore be at higher risk of disease transmission or invasion.



Spatial information on the pet network can also be used to estimate disease or invasion risk through environmental matching. Using another simplified example network (see the left figure), animals for the trade may be primarily harvested in one environment (blue) but sold in a different environment (yellow), creating an environmental mismatch. Therefore, successful disease transmission or nonnative species establishment may be less likely because pets are primarily sold into an unsuitable or unfamiliar environment. Conversely, the smaller trade routes between similar environments (i.e., blue to blue or yellow to yellow) or secondary transport from an unsuitable to a suitable region (e.g., blue region imports transported across the red dashed line) may be more important targets for management efforts.

The pet trade is a spatial network

We posit that addressing empirical data gaps across the global pet trade and mitigating its negative effects can be better accomplished by representing the trade as a spatially explicit network, with each of the actors and their trade interactions treated as respective nodes and links in a network model (Reino et al. 2017, Vall-llosera and Cassey 2017, Frost et al. 2019). In this trade network, pets originate in a source region and are transported from one linked node to the next along the commodity chain until their eventual point of retail sale and ultimate use. Actors in the trade serve as the nodes because they have a physical location in which pets are received, stored, and transported through the network. This approach provides a suitable framework for the pet trade by summarizing the principal market actors involved (the nodes), helping to quantify the spatial patterns and trade flows of pets (the links), standardizing terminology, and targeting data gathering efforts to key network components. Wellknown network-level and node-level statistics can also be used to identify trade hubs to prioritize for information gathering or management intervention, and to identify trading partners that are highly influential in structuring the trade network (Reino et al. 2017, Vall-llosera and Cassey 2017, Sánchez-Mercado et al. 2020). Such network approaches are powerful tools to map the topology of the pet trade and to focus management to the areas that best minimize its negative impacts (box 1).

To develop the network structure of the pet trade, we first systematically gathered published literature on international pet trade commodity chains to build a comprehensive, composite representation of the different types of actors and trade connections (see supplemental file S1 for the methodology). Using this composite representation as an information framework, we then conducted a second review designed to evaluate gaps in empirical data related to each type of node and link (see supplemental file S1). Finally, we used a case study of vertebrates imported to the continental United States to illustrate the value of a network approach. The United States is a principal driver of the global pet trade (Kraus 2008, Bush et al. 2014, Ribeiro et al. 2019) and our analyses show how some of the best available data on this trade fits into our framework and how a better understanding of spatial and temporal trade patterns can reveal key targets for management efforts. We also use this case study to infer potential spatial patterns of where traded individuals were likely harvested and where they are imported within the United States. These initial inflows and final outflows of the trade network are poorly understood but are crucial for evaluating the extinction risk of vertebrates feeding the pet trade and its role in introducing nonnative animals and their associated pathogens.

The pet trade commodity chain is a complex, multistage process

Our systematic review of the wildlife trade literature provided a general structure of the pet trade commodity chain by identifying 11 broad types of actors or node categories that harbor or transport pets for international trade (see figure 1). Below, we review each of these different node categories and the links between them.

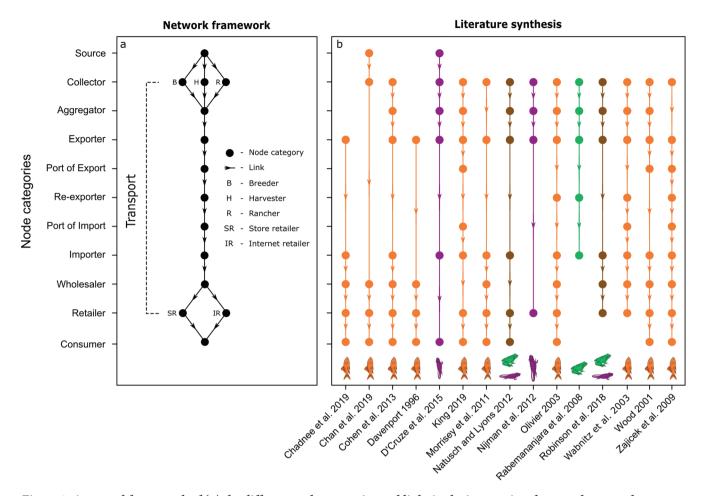


Figure 1. A general framework of (a) the different node categories and links in the international pet trade network, which was informed by (b) our review of publications that describe the general actors and connections involved in the legal trade of vertebrates as pets. The publications we reviewed were focused on the trade in amphibians (green), fishes (orange), reptiles (purple), or amphibians and reptiles (brown). We found no studies on birds or mammals that met our review criteria (see supplemental file S1). Our network framework outlines that pet transport (panel a, dashed line) is a directional, multistage process with pets moved from one category of nodes to the next in the chain as they are collected, transported, and eventually purchased. We also identified different subcategories of nodes from our review, such as pets collected for breeding versus ranching or store versus online retailers. However, different subcategories likely exist for all stages. Not all international trade will involve all node categories and some or many nodes can be skipped, such as pets sold directly from exporters to consumers.

Source and collector. Animals for the pet trade are first collected from their native habitats (D'Cruze et al. 2015, Chan et al. 2019). Although many animals are bred in captivity, the first point of entry of a species into the pet trade can always be traced back to a wild population. There have also been some reported instances of captive-breeding facilities using wild-caught animals to maintain captive population viability (Warwick 2014, Tensen 2016) or illegally reporting wild-caught animals as being bred in captivity (Nijman et al. 2012, Warwick 2014).

Animals collected from the wild tend to flow into three different subcategories of collector nodes, which are commonly termed *harvesters*, *breeders*, and *ranchers* (e.g., Morrisey et al. 2011, Natusch and Lyons 2012, Harrington et al. 2020). Harvesters do not raise or breed collected animals prior to

transfer through the commodity chain. Ranchers raise wild-caught juveniles until they can be sold as adults. Breeders collect animals to start or maintain their breeding stock, with most bred offspring sold into the market.

Past research suggests that the geographic locations for collector nodes are likely determined by the native range of the species and by which parts of their geographical range are the easiest for collectors to access, such as those closest to access roads (Ceballos and Fitzgerald 2004, Gravuer et al. 2008). The strength of the links between the source and collector node categories—that is, how many animals are collected from the wild—is also influenced by customer demand and the resulting exporter demand (Robinson et al. 2018), by population density and the seasonal trends of the targeted species

(Ceballos and Fitzgerald 2004, Rabemananjara et al. 2008), by the regional legality of wild harvest for the pet trade (Nijman et al. 2012), by individual-level variation in collector skill (Wood 2001, Ceballos and Fitzgerald 2004), and by animal traits (e.g., behavior; Chapple et al. 2012).

Aggregator. Once collected from the wild, animals can be transferred through intermediary aggregators prior to export (Natusch and Lyons 2012, Robinson et al. 2018). Aggregators are individuals or groups that obtain captured animals from local collectors, often from multiple locations, and consolidate these animals for eventual transfer to international exporting individuals or companies (exporters; figure 1). Likely geographic locations for aggregator nodes include areas in which target species are common—and therefore in which collection probability is higher—and areas in which transportation costs are lowest, such as regions with access roads and proximity to major cities (Rabemananjara et al. 2008). Some of the studies we reviewed also suggested that the link strength between different collectors and aggregators may be primarily influenced by the size of the exporters' business because aggregators typically aim to fill specific orders provided by exporters (Davenport 1996, Robinson et al. 2018).

Exporter. Exporters consolidate animals from aggregators or obtain animals directly from collectors (Carpenter et al. 2005, Rosa et al. 2011). Some studies indicated that exporter nodes for the legal trade may tend to be in both the source region of collected animals and close to transport hubs, such as international airports. For example, most Indonesian exporters that specialize in trading geckos are located on Java, where the animals are primarily harvested (Nijman et al. 2012), which improves accessibility, reduces transport mortality, and is proximate to major ports for ease of international air transport. The number of animals that ultimately flow into each exporter node is determined by a combination of factors, including the size of business (Wood 2001), the quantity and mortality of animals provided by collectors or aggregators (Wood 2001, Robinson et al. 2018), the demand for pets from customers and import partners (Rabemananjara et al. 2008), and legal restrictions placed on the exporters by the exporting or recipient country (Natusch and Lyons 2012, Robinson et al. 2018).

Ports of export, reexport, and import. Once prepared for shipment by the exporter, animals are transferred to the ports of export within the source country (King 2019). Export typically occurs via air transport but direct ground transport from the exporter to the port of import can occur between adjacent countries (e.g., Canada and the United States; Smith et al. 2017). During transport, animals can be moved to an intermediate location and reexported. At this location, multiple shipments can be consolidated and sometimes mixed with animals derived from other geographical locations, then divided into new shipments and reexported to their intended ports of import (Olivier 2003).

Importer. Following arrival at the port of import within the importing country and passage through customs inspection (if any), animals are dispersed from the port of import to consolidating locations owned by the importers (Olivier 2003, King 2019). Here, the animals are unpacked, examined, and acclimated to their new holding facilities. Importers may tend to be geographically located near international airports (Zajicek et al. 2009), but these locations can be difficult to pinpoint because registered business addresses do not necessarily reflect where the animals are being housed. The strength of the links between a given port and importer operation were not discussed in any of the studies we reviewed. However, we infer that, like exporters, these links may be primarily controlled by the size of the company and the severity of legal restrictions and enforcement, which would be imposed by the importing country and the port of import.

Sale (wholesaler, retailer, and consumer). After import, the pet trade commodity chain concludes when animals are eventually sold and housed as pets with a consumer. Sale can occur directly from importers to consumers, or animals can pass through the domestic-breeding market and through various stages of domestic transport as they are transferred from importer to wholesalers, to brick-andmortar and internet retailers, and finally to consumers (Morrisey et al. 2011, Chan et al. 2019). Wholesalers may tend to be located near ports of import (Zajicek et al. 2009) because they purchase large quantities of pets from earlier actors in the commodity chain (often importers) for resale to retailers or consumers. However, the location of most retail stores and pet owners is likely more geographically dispersed and may be clustered in urban centers (Ceballos and Fitzgerald 2004).

Large data gaps exist for most nodes and links in the trade network

Our second review, aimed at identifying data gaps, highlighted the sheer quantity of empirical information missing on each network component in the framework detailed above (see box 2), which limits our ability to understand and manage the pet trade (Cohen et al. 2013, Rhyne et al. 2017, Smith et al. 2017). Data availability varied widely for each network node category and their respective links. Empirical information was only broadly available for two nodes: the likely region from which animals were collected and the locations of pet retailers. Even focusing on these two parts of the network is still problematic for drawing broader conclusions about where animals came from and where they are transported. The greater data availability on pet source regions, for example, was primarily due to publicly available biodiversity databases. However, distribution databases were not created to track the pet trade and so cannot be used to infer the number of individuals collected, whether a pet was wild caught or captive bred, nor mortality associated with collection. Such knowledge gaps are becoming increasingly

Box 2. Pet trade data availability.

We found that data availability varied widely for each network component in the international pet trade. Empirical information was generally available for the earliest node categories and links, and then became progressively less available when pets are transferred through a multitude of intermediaries. Empirical information increased for the export and import node categories and links owing to legal requirements related to the transport of legally protected species, such as the Convention on International Trade in Endangered Species of Wild Fauna and Flora (CITES). After import into a country, data availability declined dramatically again as animals were transferred through multiple intermediaries prior to sale as pets. Our conclusions of data availability for each node and link in the international pet trade network are provided in the table below (see supplemental file S1 for methods). Common data sources, proxies, or methods reported by relevant publications, and example publications, are respectively listed under the "Example sources or proxies" and "Example publications" headings. We limited the example publications to a maximum of three and so the number of references does not reflect differences in data availability (see supplemental file S1 for full references). Interviews were commonly used to obtain data across every component, and so this method is not listed as an example data source to limit repetition.

Node category	Name	Data needs	Example sources or proxies	Example publications
1	Source	Location or distribution ^a	IUCN, FishBase, GBIF, field surveys of species sold by retailers	Andreone et al. 2006, Nijman et al. 2017, Scheffers et al. 2019
		Abundance ^b	Government records, stock assessments, productivity	Shaw et al. 2011, Buchanan et al. 2019, Ngo et al. 2019
2a	Harvester	Location ^b	Government permits, aggregator or exporter records, field surveys	González 2003, Lyons and Natusch 2011, Raghavan et al. 2013
		Volume collected ^b	Government permits, collector records, back calculation from exporter or aggregator orders	González 2003, Lyons and Natusch 2011, Raghavan et al. 2013
		Mortality or escape ^c	Harvester records, simulated harvesting experiment	González 2003, Militz et al. 2016, Raghavan et al. 2018
2b, 2c	Breeder or rancher	Location ^c	Government permits or census, aggregator or exporter records	Lyons and Natusch 2011
		Volume traded ^b	Government permits or census, breeder records, back calculation from exporter or aggregator orders	Van Wilgen et al. 2010, Lankau et al. 2017, Vall-llosera and Cassey 2017
		Mortality or escape ^c	Breeder records, mortality recorded in zoos or introduction programs	Vagelli 2004, Fischer 2012
3	Aggregator	Location ^c	Government permits, exporter records	Rabemananjara et al. 2008, Robinson et al. 2018
		Volume traded ^c	Aggregator records, back calculation from exporter orders	Shuman et al. 2004, Bušina et al. 2018
		Mortality or escape ^c	Aggregator records	Schmidt and Kunzmann 2005
4	Exporter	Location ^c	Government permits, customs records (e.g., LEMIS), internet records	Rabemananjara et al. 2008, Raghavan et al. 2013, Robinson et al. 2018
		Volume traded ^b	Exporter records	Martin-Smith et al. 2006, Nijman et al. 2012, Raghavan et al. 2013
		Mortality or escape ^c	Exporter records, investigation	Rubec and Cruz 2005, Schmidt and Kunzmann 2005
5	Port of Export	Location ^b	Transportation least-cost network, international trade networks, custom records, survey social media	Moreau and Coomes 2007, Rhyne et al. 2012, Raghavan et al. 2013
		Volume traded ^b	Customs records, port records	Moreau and Coomes 2007, Rhyne et al. 2012, Raghavan et al. 2013
		Mortality or escaped	Port inspection records	
6	Reexporter	Location ^c	Airline travel records, cargo tracking codes, CITES	Martin 2018, Martin et al. 2018
		Volume traded ^c	Exporter or reexporter records	Affre et al. 2005, Nijman and Shepherd 2010, Martin 2018
		Mortality or escape ^d	Back calculate from volume exported minus volume received	

Continued.					
Node category	Name	Data needs	Example sources or proxies	Example publications	
7	Port of import	Location ^b	Transportation least-cost network, international trade networks, customs records, survey social media	Chapman et al. 1997, Vall-llosera and Cassey 2017, Eskew et al. 2020	
		Volume traded ^b	Customs records, port records	Chapman et al. 1997, Vall-llosera and Cassey 2017, Eskew et al. 2020	
		Mortality or escape ^c	Port inspection or quarantine records	Iñigo-Elias and Ramos 1991	
8	Importer	Location ^d	Government permits, customs records, online records of company locations, LEMIS		
		Volume traded ^c	Importer records	Biondo 2018	
		Mortality or escape ^c	Importer records, official inspection	Steinmetz et al. 1998, Faruk et al. 2012	
9	Wholesaler	Location ^d	Government permits, online records of company locations, transportation least-cost network		
		Volume traded ^c	Importer shipping or wholesaler order records, governmental investigation	Ashley et al. 2014	
		Mortality or escape ^c	Governmental investigation	Ashley et al. 2014	
10a	Store retailer	Location ^a	Online records of company locations, US records of pet and pet supply stores, field surveys	Strecker et al. 2011, Su et al. 2014 Nijman et al. 2017	
		Volume traded ^c	Wholesaler shipping or retailer order records, field surveys, estimate from advertisement	Strecker et al. 2011, Nijman et al. 2017	
		Mortality or escape ^c	Retailer records, field surveys	Faruk et al. 2012, Ashley et al. 2014	
10b	Internet retailer	Location ^c	Return address on received shipments, online advertisements	Olden et al. 2021	
		Volume traded ^c	Retailer records, survey trade activity on auction sites or social media	Kikillus et al. 2012, Martin et al. 2018, Olden et al. 2021	
		Mortality or escape ^d	Survey retailers, records of refunded shipments		
11	Consumer	Location ^d	Pet ownership statistics, population centers		
		Volume traded ^c	Retailer sale records, American Veterinary Medical Association, US census of pet ownership	Gertzen et al. 2008, Strecker et al. 2011	
		Mortality or escape ^c	Field surveys, citizen science	Ceballos-Mago et al. 2010, Robinson et al. 2015, Vall-llosera and Cassey 2017	

urgent to fill as the pet and wildlife trade affects an evergrowing proportion of global biodiversity (Di Minin et al. 2019, Scheffers et al. 2019, Marshall et al. 2020).

Regarding retail locations, the greater information availability for this node category was due to publicly available, online, and governmental information on brick-and-mortar stores, such as from census records or social media (e.g., Strecker et al. 2011, Stringham et al. 2020). This type of data could be used as a proxy to estimate how animals are distributed within a country if pet consumers follow similar patterns to other types of retail customers and tend to shop

close to where they live (Huff 1963). However, location records generally do not provide more proprietary information on store sales. Without sales information, retail location records can only be loosely linked to the likelihood of zoonotic disease transmission or nonnative establishment. These risks would be better informed by data on the types of animals being sold (e.g., amphibians, birds), the quantity sold, and where they are sold. A greater proportion of the wildlife trade is also shifting to online markets (Lavorgna 2014), which are not captured by data on brick-and-mortar retail locations.

Table 1. Top five exporting countries and importing ports in the US trade of pet amphibians, birds, fishes, mammals, and reptiles during 1999 through 2013.

Top five export or import	Amphibians	Birds	Fishes	Mammals	Reptiles
Average export richness	Madagascar (20.5) Tanzania (15.1) Suriname (12.8) Indonesia (12.7) Canada (12.2)	Canada (56.1) Belgium (23.3) Peru (18.7) Suriname (17.3)	Indonesia (17.0) Peru (11.1) Colombia (10.9) Thailand (10.1) Philippines (9.6)	Guyana (9.1) Canada (8.3) Czech Republic (6.8) Netherlands (6.4) Indonesia (3.5)	Indonesia (89.3) Germany (60.9) Tanzania (53.2) Madagascar (39.2) Guyana (31.5)
Percent of total export richness	62% (203 of 326)	Tanzania (14.8) 57% (375 of 652)	53% (155 of 290)	58% (86 of 147)	66% (745 of 1119)
Average export quantity	Hong Kong (709K) Singapore (414K) China (383K) Indonesia (188K) South Korea (97K)	Taiwan (29K) Tanzania (25K) Belgium (22K) Senegal (22K) Australia (15K)	Trinidad and Tobago (3.1M) Thailand (1.8M) China (1.4M) Hong Kong (1.3M) Malaysia (354K)	Netherlands (48K) Czech Republic (42K) Canada (8K) Peru (1K) Indonesia (1K)	Vietnam (302K) El Salvador (191K) Colombia (94K) Togo (86K) Indonesia (78K)
Percent of total import quantity	87% (2.7M of 3.1M)	62% (1.7M of 2.7M)	90% (12.0M of 13.3M)	97% (1.5M of 1.54M)	60% (11.2M of 18.7M)
Average import richness	Miami (84.1) Los Angeles (67.0) Dallas–Fort Worth (14.1) New York (13.5) Atlanta (5.7)	Los Angeles (103.2) Miami (64.9) New York (32.4) Buffalo (27.1) Detroit (13.5)	Houston (33.6) New York (29.8) Los Angeles (13.8) Atlanta (13.0) San Francisco (12.7)	Miami (16.5) Dallas–Fort Worth (14.1) Chicago (7.1) Los Angeles (5.7) Houston (3.8)	Miami (363.1) Los Angeles (227.3) Dallas–Fort Worth (92.1) New York (38.0) Atlanta (25.9)
Percent of total import richness	89% (289 of 326)	91% (594 of 652)	81% (235 of 290)	86% (127 of 147)	95% (1066 of 1119)
Average import quantity	Los Angeles (1.4M) New York (168K) Tampa (141K) Miami (116K) Chicago (106K)	Los Angeles (150K) New York (18K) Miami (6K) Pembina (5K) Buffalo (1K)	New York (3.9M) San Francisco (2.1M) Los Angeles (1.7M) Atlanta (338K) Tampa (181K)	Dallas–Fort Worth (52K) Chicago (16K) Atlanta (9K) Los Angeles (9K) New York (6K)	Miami (673K) Los Angeles (487K) Dallas–Fort Worth (45K) New York (11K) Baltimore (7K)
Percent of total import quantity	96% (3.0M of 3.1M)	99% (2.7M of 2.72M)	93% (12.4M of 13.3M)	91% (1.4M of 1.5M)	98% (18.3M of 18.7M)

Note: Trade richness is the average number of species exported or imported across years (species per year), and quantity is the average number of animals (individuals per year). The percentage of total trade richness (species) or quantity (individuals) represented by the top five countries or ports across all years combined is also provided for each animal clade. The quantity values are rounded to the nearest 1000, 10,000, or 100,000, depending on trade magnitude.

US trade data illustrates the value of a network approach to pet-trade impact mitigation

Although there is a disheartening prevalence of data gaps throughout the pet trade network, relevant insights can still be gained by collecting empirical data and conducting network analyses on key components. To illustrate this point, we analyzed over 230,000 shipments of live, wild-caught and captive-bred vertebrates—exotic amphibians, birds, freshwater and marine fishes, and reptiles—that were legally imported to the United States during 1999 through 2013. This data set was obtained from the US Fish and Wildlife Service's Law Enforcement Management Information System (LEMIS) and was curated to only include shipments that could be reliably attributed to the pet trade (see supplemental file S2 for the methods). The finalized data set included 38,485 amphibian, 11,210 bird, 5667 mammal, 38,974 freshwater and marine fish, and 89,720 reptile shipment records that together detailed the trade of over 187 million individual animals over 15 years. This pet trade data is not without its limitations as it is restricted to imports into a single country and some shipments had to be excluded owing to data entry errors (which are further described in Rhyne et al. 2012 and Eskew et al. 2020). However, through an analysis of spatial and temporal patterns in US pet exports and imports (methods detailed in supplemental file S3), we show that even limited data on the trade network can help to direct future research and management approaches.

Finding hotspots of biodiversity loss and economic dependence. Our US case study illustrated how empirical data on pet exports can inform efforts to improve the economic benefits of the trade, and mitigate its biodiversity impacts, by identifying trade hotspots—that is, regions of high biodiversity loss and economic dependence. We found that just a handful of countries accounted for most of the species richness (53%–66%; table 1) and quantity (60%–97%; table 1) of animal clades exported to the United States for sale as pets. A further subset of these countries both export and harbor a

Table 2. Agreement between the number of species exported from a given country to the United States and how many species in the pet trade naturally occur within that country for each animal clade.

Match between native and export					
richness	Amphibians	Birds	Mammals	Reptiles	
High export and high diversity (total species	Madagascar	Peru	Indonesia	Indonesia	
exported of total native species in trade)	(42 of 38)	(74 of 132)	(17 of 19)	(200 of 147)	
	Indonesia	Tanzania	Guyana	Tanzania	
	(41 of 29)	(73 of 128)	(16 of 32)	(144 of 113)	
	Suriname	Guinea	Ghana	China	
	(41 of 28)	(63 of 103)	(11 of 19)	(88 of 125)	
Low export and high diversity	Colombia	Columbia	Bolivia ^a	Mexico	
	(5 of 52)	(10 of 143)	(0 of 30)	(32 of 162)	
	Brazil ^a	Brazil	Brazil	Myanmar	
	(0 of 45)	(13 of 146)	(3 of 32)	(1 of 96)	
	Bolivia ^a	Kenya	Venezuela ^a	Brazil	
	(0 of 38)	(3 of 134)	(0 of 27)	(3 of 97)	
High export and low diversity	Germany	Canada	Canada	Germany	
	(43 of 8)	(186 of 31)	(39 of 13)	(380 of 8)	
	Canada	Belgium	Czech Republic	Canada	
	(42 of 14)	(78 of 40)	(34 of 11)	(138 of 23)	
	Netherlands	Netherlands	Netherlands	Netherlands	
	(29 of 5)	(73 of 38)	(25 of 10)	(70 of 4)	

Note: Countries with a high diversity of exports and a high diversity of native species in the pet trade (purple countries in figure 1) are potential trade hotspots. Countries with low export richness but a high richness of native species in the trade (pink countries in figure 1) are potential source regions of pets reexported, bred, or smuggled in other countries. Countries with much higher export richness than their native richness (blue countries in figure 1) are potentially prolific breeders or reexporters. The top three such countries are listed in each table subsection. *Countries that are never recorded as the country of origin for a given animal clade in the LEMIS database despite having a high diversity of native species present in the pet trade.

high diversity of traded species (see table 2 and figure 2). For example, 95% of the amphibians exported out of Madagascar and into the United States are native to Madagascar. Such countries may also have a higher percentage of their population living in relative poverty and who depend on the pet trade as their primary income (e.g., Madagascar or Indonesia; Rabemananjara et al. 2008, Ferse et al. 2012). These local collectors are also generally paid a small fraction of the retail value of traded pets (e.g., Rabemananjara et al. 2008, Natusch and Lyons 2012, Raghavan et al. 2013). It is this combination of high native biodiversity, economic dependence on animal collection, and poor compensation that may create a trade hotspot because harvest is tightly coupled to the growing demand for new pets (Romagosa 2014, Scheffers et al. 2019, Marshall et al. 2020), and there is little monetary incentive to harvest sustainably (Wood 2001, Natusch and Lyons 2012). We cannot determine from our data whether unsustainable harvest is occurring in each country nor which traded animals are sourced from captive breeding programs versus wild capture. However, our results show how export data can help to identify potential trade hotspots. These hotspots can then be confirmed in followup analyses and can be targeted for livelihood diversification and equitable benefit sharing (e.g., Ferse et al. 2012, Hinsley et al. 2018), and ultimately culturally appropriate

interventions to reduce the biodiversity impacts of the trade (e.g., Agu and Gore 2020).

Reducing disease transmission. A network approach can also inform mitigation efforts aimed at controlling wildlife-associated disease spillover. Knowing the location of the source nodes (figure 1) aids in the identification of which diseases are potentially being moved with their animal hosts (Springborn et al. 2015), as can information on wild capture versus captive breeding (Smith et al. 2017). Furthermore, high volume reexport and import nodes can consolidate closely related animals from multiple source regions under stressful transport conditions, which may aid the transmission of pathogens between individuals and species (Warwick 2014). Knowing the principal reexporters and ports of import can therefore greatly aid biosecurity enforcement aimed at preventing such transmissions through inspections and quarantines.

For example, our case study showed that four US airports—Dallas, Los Angeles, Miami, and New York—consistently received the highest richness (table 1 and figure 3) and quantity (table 1) of pet imports. Intensifying enforcement of biosecurity policy at these large animal import hubs could have a strong influence on mitigating disease introductions (e.g., van Roon et al. 2019). In addition, our network

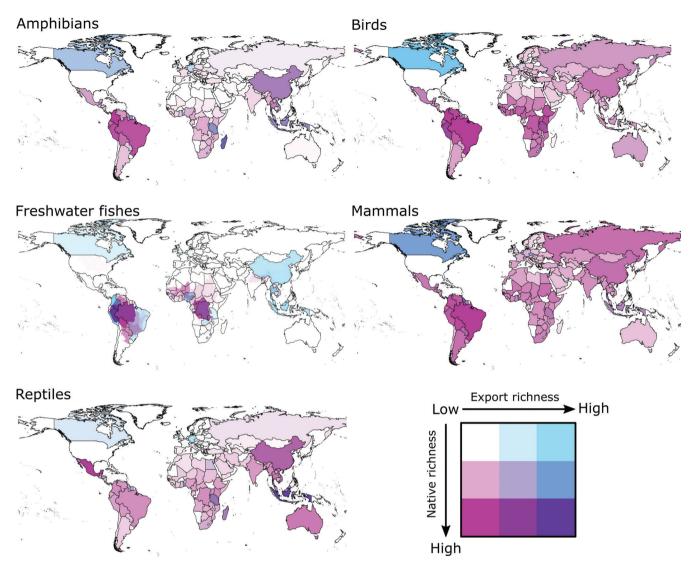


Figure 2. Average species richness of pet amphibians, birds, freshwater fishes, mammals, and reptiles exported from each country to the United States during 1999 through 2013 (blue; increasing values are illustrated using an increasing color gradient) overlaid on the total richness of species native to that country that are recorded in the US pet trade (pink). Richness of traded freshwater fishes is not available by country and is instead provided across global watersheds. Marine fishes are excluded from this analysis because export values are country specific and cannot be overlaid on ocean distributions. The methods for producing these maps and links to online, interactive versions are provided in supplemental file S3.

analyses of incoming and outgoing trade connections indicated that just a few US ports were responsible for importing pets from a very high number of source countries, and a few countries spatially dispersed their animal exports to a very high number of US ports. These trade patterns were evidenced by strongly right-skewed distributions of trade connections (figure 4). Our network results also showed that the US trade in amphibian and reptile pets has spatially consolidated over time, whereas the trade in fishes and mammals has become more spatially dispersed (on the basis of the data available during 1999 through 2013; figure 5). Exporters and animal clades that are creating broader spatial distribution,

such as Canada, which exported birds to an average of 13 different US ports across years, could be important targets for increased enforcement because greater distribution can facilitate more widespread disease introduction (Shirley and Rushton 2005). Conversely, ports and animal clades that are spatially consolidating can help to simplify biosecurity to a small number of target ports that warrant higher staffing and monitoring by inspectors (Floerl et al. 2009, Silk et al. 2017).

Preventing nonnative introductions and revealing regions of high invasion risk. Efforts to prevent nonnative species

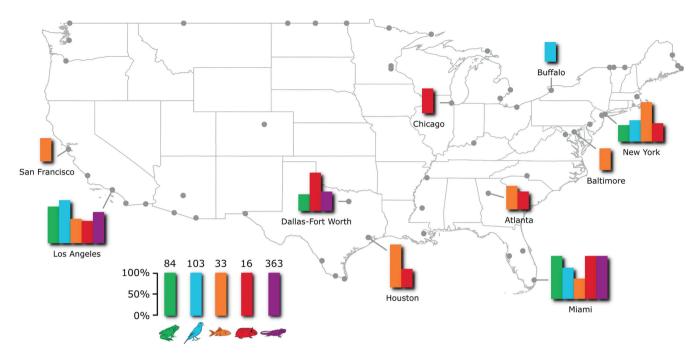


Figure 3. Average species richness of amphibians (green), birds (blue), fishes (orange), mammals (red), and reptiles (purple) imported to each US port of entry (air and land ports) for sale as pets during 1999 through 2013. There are strong spatial patterns in where different animal clades tend to be imported into the United States such that only a handful of ports—primarily the Dallas–Fort Worth, Los Angeles, Miami, and New York airports—import the majority of traded species (these same four ports also import the greatest quantities; see supplemental file S3). Bars are scaled relative to the maximum import richness for each animal clade, which is listed at the top of each bar in the legend. Ports whose imports constitute less than 5% of maximum import richness, and are therefore not principal ports of entry for imported pets, are plotted as grey points.

introductions via the pet trade can also benefit from our network approach. Identifying the principal countries of export (e.g., Canada, Indonesia, or Madagascar in the US trade; table 1) and ports of import can provide key targets for mitigation efforts aiming to prevent new species introductions (Reaser et al. 2008). In addition, information on nonnative pet origins (i.e., the source nodes), and distribution within recipient countries from importer to consumer (figure 1), can be used to inform predictions of environmental matching between native and potential nonnative distributions (Howeth et al. 2016). For instance, some commonly traded amphibian, bird, and reptile species that we found in the US trade tend to be native to warmer climates, such as the western dwarf clawed frog (Hymenochirus curtipes), the savannah monitor (Varanus exanthematicus), and the estrildid finches (family Estrildidae). These same animals also tended to be imported to the warmer regions of the United States (e.g., Miami and Los Angeles). We do not know whether they were subsequently sold as pets in these warmer regions, but a high environmental match between the source and import regions suggests a higher risk of successful establishment.

Finally, although data are scarce, a better understanding of the later nodes in the commodity chain (i.e., retailers and consumers) is critical for predicting pet escape or release, which are major pathways driving the establishment of new nonnative animals worldwide (Hulme et al. 2008, Stringham and Lockwood 2018). Although our US case study did not extend to these later nodes, the differences we found in animal import patterns provides an example of potential differences in spatial distribution that could be informative of establishment risk. We found that the majority of amphibians, birds, and reptiles were imported via Miami and Los Angeles, the majority of freshwater and marine fishes were imported via New York and Dallas, and the majority of mammals were imported to Miami and Dallas (figure 3). These spatial import patterns may result in some regions receiving more introductions of particular animal clades, such as more reptiles being introduced to the southeastern and southwestern United States, which can translate to a higher invasion risk (Lockwood et al. 2005).

Identifying potential breeders, reexporters, or sources of illegally traded pets. Even network approaches using incomplete trade data, such as the LEMIS shipment records that are not designed to track the pet trade, can inform management efforts by helping to locate key network components, such

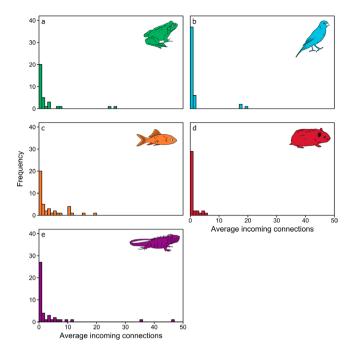


Figure 4. Frequency distributions of incoming trade connections for pet (a) amphibians, (b) birds, (c) fishes, (d) mammals, and (e) reptiles imported to each US port of entry. The incoming connections for each port are calculated as the number of different countries a port receives shipments from for each animal clade averaged from 1999 through 2013. These right-skewed distributions indicate that most ports tend to import pets from just a single country, whereas a small number of ports (usually the Los Angeles, Miami, or New York airports) import and therefore consolidate pets from many different countries. Histograms of outgoing connections are similarly right skewed and are provided in supplementary materials S3.

as prolific reexporters, breeders, or possibly routes of illegal wildlife trade. In our US case study, for example, we found a variety of mismatches in which countries were exporting more species than they possessed as native residents (table 2). Such mismatches could indicate data deficiencies where the reported country of origin in export and import records may not reflect the actual source region of traded animals, providing insight into potentially important nodes and trade links. For instance, Canada, Germany, and the Netherlands were often listed as the region of origin of a diverse array of animals not native to any of these countries, which suggests that these countries are likely home to a variety of captive breeding programs or are prolific reexporters. This information can subsequently inform management efforts depending on the source of the traded animals and the conditions of captivity. If these countries are breeding different species for sale as pets then they may be low priority management targets because well-managed breeding facilities can result in lower biodiversity and disease impacts (Smith et al. 2009).

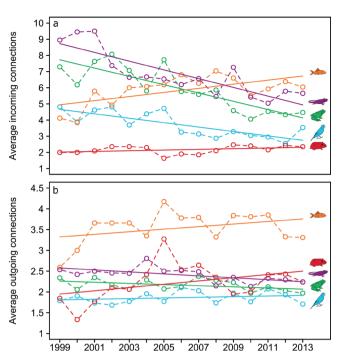


Figure 5. Temporal changes in the average number of (a) incoming and (b) outgoing trade connections during 1999 through 2013 for each animal clade. Both incoming and outgoing connections have declined over time for amphibians (green) and reptiles (purple), indicating a spatial consolidation of trade, whereas these metrics have increased in fishes (orange) and mammals (red). For birds (blue), incoming connections have decreased over time, whereas outgoing connections have increased. Best-fit lines are plotted for each animal clade on the basis of their linear relationship with time predicted from ANCOVA models (see supplementary materials S3).

However, poorly managed programs may be high priority targets owing to stressful and crowded conditions that can increase disease susceptibility and transmission (Warwick 2014). Alternatively, if the exporting countries are prolific reexporters then they are potentially facilitating the harvest of animals and consolidation of wildlife diseases from a variety of source regions. These types of countries may therefore be important targets for enforcement of inspection and quarantine regulations and for determining the origins of their shipments.

Other mismatches occurred in our US case study in which some countries harbored a high diversity of traded species but were rarely listed as the countries of origin in any shipment records, such as Brazil, Columbia, and Bolivia (table 2). It is possible that, although a variety of traded species naturally occur in these regions, traded animals could be mostly sourced from neighboring countries in which they also naturally occur or are being sourced from captive-breeding programs. Alternatively, countries that harbor a

high diversity of traded species but are not themselves listed as the origin of these shipments could be the unreported source regions for many traded pets being shipped or smuggled to intermediate countries prior to import into the United States (Nóbrega Alves et al. 2013, Patoka et al. 2015). This form of reexporting may appear legal if significant time is spent in the layover country or if breeding facilities are established, effectively laundering the initial wild source region of these animals. Such hidden source countries may therefore be unrecognized trade hotspots or overlooked actors in the illegal trade of wild-caught pets, which can be revealed through network approaches and by filling in data gaps in the pet trade network.

Conclusions

The vertebrate pet trade is a poorly understood but increasingly important driver of the translocation of millions of animals around the world, presenting serious conservation and public health concerns. However, this same trade offers myriad positive economic and social benefits ranging from supporting local economies to providing emotional companionship. Here, we show that developing conceptual and empirical network representations of the pet trade allows targeted interventions that mitigate its negative impacts and could improve its positive economic benefits. Our literature synthesis revealed that the full pet trade network is a primarily directional, multistage process, and empirical data are sorely lacking for most stages. The complexity and data gaps of the trade network stem from animals transferring through many different entities and locations on their journey from their source habitats to a consumer. Encouragingly, our case study demonstrates that even networks sparsely informed by empirical data can provide important insights into the potential spatial and temporal patterns of the pet trade, such as inferring the initial geographic origins and eventual distribution of purchased animals. These insights provide a better understanding of the pet trade, which is necessary to improve the management of its impacts, to identify the people that depend on it financially, and to ensure it is conducted sustainably so that it can continue to provide its beneficial services.

Acknowledgments

We would like to thank Marie-Josée Létourneau for producing the animal artwork used in some of our figures and for agreeing to its publication. This work was supported by the National Socio-Environmental Synthesis Center under funding received from the US National Science Foundation (grant no. DBI-1052875) and with project funding from the Centre for Invasive Species Solutions (grant no. PO1-I-002: "Understanding and intervening in illegal trade in nonnative species").

Supplemental material

Supplemental data are available at *BIOSCI* online.

References cited

- Agu HU, Gore ML. 2020. Women in wildlife trafficking in Africa: A synthesis of literature. Global Ecology and Conservation 23: e01166.
- Baker SE, Cain R, van Kesteren F, Zommers ZA, D'Cruze N, Macdonald DW. 2013. Rough trade: Animal welfare in the global wildlife trade. BioScience 63: 928-938.
- Bush ER, Baker SE, Macdonald DW. 2014. Global trade in exotic pets 2006-2012. Conservation Biology 28: 663-676.
- Carpenter AI, Robson O, Rowcliffe JM, Watkinson AR. 2005. The impacts of international and national governance changes on a traded resource: A case study of Madagascar and its chameleon trade. Biological Conservation 123: 279-287.
- Ceballos CP, Fitzgerald LA. 2004. The trade in native and exotic turtles in Texas. Wildlife Society Bulletin 32: 881-891.
- Chapple DG, Simmonds SM, Wong BBM. 2012. Can behavioral and personality traits influence the success of unintentional species introductions? Trends in Ecology and Evolution 27: 57-62.
- Chan FT, et al. 2019. Leaving the fish bowl: The ornamental trade as a global vector for freshwater fish invasions. Aquatic Ecosystem Health and Management 22: 417-439.
- Chomel BB, Belotto A, Meslin FX. 2007. Wildlife, exotic pets, and emerging zoonoses. Emerging Infectious Diseases 13: 6-11.
- Cohen FPA, Valenti WC, Calado R. 2013. Traceability issues in the trade of marine ornamental species. Reviews in Fisheries Science 21:
- D'Cruze N, Singh B, Morrison T, Schmidt-Burbach J, Macdonald DW, Mookerjee A. 2015. A star attraction: The illegal trade in Indian Star Tortoises. Nature Conservation 13: 1-19.
- Davenport KE. 1996. Characteristics of the current international trade in ornamental fish, with special reference to the European Union. Revue Scientifique et Technique de l'Office International des Epizoonties 15: 435-443.
- Di Minin E, et al. 2019. Identifying global centers of unsustainable commercial harvesting of species. Science Advances 5: eaau2879.
- Elwin A, Green J, D'Cruze N. 2020. On the record: An analysis of exotic pet licences in the UK. Animals 10: 2373.
- Eskew EA, White AM, Ross N, Smith KM, Smith KF, Rodríguez JP, Zambrana-Torrelio C, Karesh WB, Daszak P. 2020. United States wildlife and wildlife product imports from 2000-2014. Scientific Data 7: 22.
- Ferse SCA, Knittweis L, Krause G, Maddusila A, Glaser M. 2012. Livelihoods of ornamental coral fishermen in South Sulawesi/Indonesia: Implications for management. Coastal Management 40: 525–555.
- Floerl O, Inglis GJ, Dey K, Smith A. 2009. The importance of transport hubs in stepping-stone invasions. Journal of Applied Ecology 46:
- Frost CM, Allen WJ, Courchamp F, Jeschke JM, Saul W-C, Wardle DA. 2019. Using network theory to understand and predict biological invasions. Trends in Ecology and Evolution 34: 831-843.
- Gravuer K, Sullivan JJ, Williams PA, Duncan RP. 2008. Strong human association with plant invasion success for Trifolium introductions to New Zealand. Proceedings of the National Academy of Sciences of the United States of America 105: 6344-6349.
- Harrington LA, Green J, Muinde P, Macdonald DW, Auliya M, D'Cruze N. 2020. Snakes and ladders: A review of ball python production in West Africa for the global pet market. Nature Conservation 41: 1–24.
- Hinsley A, et al. 2018. A review of the trade in orchids and its implications for conservation. Botanical Journal of the Linnean Society 186:
- Howeth JG, et al. 2016. Predicting invasiveness of species in trade: Climate match, trophic guild and fecundity influence establishment and impact of non-native freshwater fishes. Diversity and Distributions 22: 148-160.
- Huff DL. 1963. A probabilistic analysis of shopping center trade areas. Land Economics 39: 81-90.
- Hulme PE, et al. 2008. Grasping at the routes of biological invasions: A framework for integrating pathways into policy. Journal of Applied Ecology 45: 403-414.

- Karesh WB, Cook RA, Gilbert M, Newcomb J. 2007. Implications of wildlife trade on the movement of avian influenza and other infectious diseases. Journal of Wildlife Diseases 43: 55–59.
- Kraus F. 2008. Alien Reptiles and Amphibians: A Scientific Compendium and Analysis. Springer.
- King TA. 2019. Wild caught ornamental fish: a perspective from the UK ornamental aquatic industry on the sustainability of aquatic organisms and livelihoods. Journal of Fish Biology 94: 925–936.
- Lavorgna A. 2014. Wildlife trafficking in the Internet age. Crime Science 3: 5.Lockwood JL, Cassey P, Blackburn TM. 2005. The role of propagule pressure in explaining species invasions. Trends in Ecology and Evolution 20: 223–228.
- Lockwood JL, et al. 2019. When pets become pests: The role of the exotic pet trade in producing invasive vertebrate animals. Frontiers in Ecology and the Environment 17: 323–330.
- Marshall BM, Strine C, Hughes AC. 2020. Thousands of reptile species threatened by under-regulated global trade. Nature Communications 11: 4738.
- Martin RO, Senni C, D'Cruze NC. 2018. Trade in wild-sourced African grey parrots: Insights via social media. Global Ecology and Conservation 15: e00429.
- McNicholas J, Gilbey A, Rennie A, Ahmedzai S, Dono J-A, Ormerod E. 2005. Pet ownership and human health: A brief review of evidence and issues. BMJ 331: 1252–1254.
- Morrisey D, Inglis G, Neil K, Bradley A, Fitridge I. 2011. Characterization of the marine aquarium trade and management of associated marine pests in Australia, a country with stringent import biosecurity regulation. Environmental Conservation 38: 89–100.
- Natusch DJD, Lyons JA. 2012. Exploited for pets: The harvest and trade of amphibians and reptiles from Indonesian New Guinea. Biodiversity and Conservation 21: 2899–2911.
- Nijman V, Shepherd CR. 2010. The role of Asia in the global trade in CITES II-listed poison arrow frogs: Hopping from Kazakhstan to Lebanon to Thailand and beyond. Biodiversity and Conservation 19: 1963–1970.
- Nijman V, Shepherd CR, Mumpuni, Sanders KL. 2012. Over-exploitation and illegal trade of reptiles in Indonesia. Herpetological Journal 22: 83–89
- Nóbrega Alves RR, de Farias Lima JR, Araujo HFP. 2013. The live bird trade in Brazil and its conservation implications: An overview. Bird Conservation International 23: 53–65.
- O'Hanlon SJ, et al. 2018. Recent Asian origin of chytrid fungi causing global amphibian declines. Science 360: 621–627.
- Olivier K. 2003. World trade in ornamental species. Pages 49–64 in Cato JC, Brown CL, eds. Marine Ornamental Species. Blackwell Publishing Company.
- Patoka J, Kalous L, Kopeck O. 2015. Imports of ornamental crayfish: The first decade from the Czech Republic's perspective. Knowledge and Management of Aquatic Ecosystems 416: 04.
- Poole CM, Shepherd CR. 2017. Shades of grey: The legal trade in CITESlisted birds in Singapore, notably the globally threatened African grey parrot *Psittacus erithacus*. Oryx 51: 411–417.
- Rabemananjara FCE, Rasoamampionona Raminosoa N, Ravoahangimalala Ramilijaona O, Andreone F, Bora P, Carpenter AI, Glaw F, Razafindrabe T, Vallan D, Vieites DR. 2008. Malagasy poison frogs in the pet trade: A survey of levels of exploitation of species in the genus *Mantella*. Amphibian and Reptile Conservation 5: 3–16.
- Raghavan R, Dahanukar N, Tlusty MF, Rhyne AL, Krishna Kumar K, Molur S, Rosser AM. 2013. Uncovering an obscure trade: Threatened freshwater fishes and the aquarium pet markets. Biological Conservation 164: 158–169
- Reaser JK, Meyerson LA, Von Holle B. 2008. Saving camels from straws: How propagule pressure-based prevention policies can reduce the risk of biological invasion. Biological Invasions 10: 1085–1098.
- Reino L, Figureueira R, Beja P, Araújo MB, Capinha C, Strubbe D. 2017. Networks of global bird invasion altered by regional trade ban. Science Advances 3: e1700783.

- Rhyne AL, Tlusty MF, Schofield PJ, Kaufman L, Morris JA Jr, Bruckner AW. 2012. Revealing the appetite of the marine aquarium fish trade: The volume and biodiversity of fish imported into the United States. PLOS ONE 7: e35808.
- Rhyne AL, Tlusty MF, Szczebak JT, Holmberg RJ. 2017. Expanding our understanding of the trade in marine aquarium animals. Peer J 5: e2949.
- Ribeiro J, et al. 2019. Trends in legal and illegal trade of wild birds: A global assessment based on expert knowledge. Biodiversity and Conservation 28: 3343–3369.
- Robinson JE, Fraser IM, St. John FAV, Randrianantoandro JC, Andriantsimanarilafy RR, Razafimanahaka JH, Griffiths RA, Roberts DL. 2018. Wildlife supply chains in Madagascar from local collection to global export. Biological Conservation 226: 144–152.
- Romagosa C. 2014. Patterns of live vertebrate importation into the United States: Analysis of an invasion pathway. Pages 115–146 in Keller RP, Cadotte MW, Sandiford G, eds. Invasive Species in a Globalized World: Ecological, Social, and Legal Perspectives on Policy. Chicago University Press.
- Rosa IL, Oliveira TPR, Osório FM, Moraes LE, Castro ALC, Barros GML, Alves RRN. 2011. Fisheries and trade of seahorses in Brazil: Historical perspective, current trends, and future directions. Biodiversity and Conservation 20: 1951–1971.
- Sánchez-Mercado A, Cardozo-Urdaneta A, Moran L, Ovalle L, Arvelo MÁ, Morales-Campos J, Coyle B, Braun MJ, Rodríguez-Clark KM. 2020. Social network analysis reveals specialized trade in an Endangered songbird. Animal Conservation 23: 132–144.
- Scheffers BR, Oliveira BF, Lamb I, Edwards DP. 2019. Global wildlife trade across the tree of life. Science 366: 71–76.
- Shirley MDF, Rushton SP. 2005. The impacts of network topology on disease spread. Ecological Complexity 2: 287–299.
- Silk MJ, Croft DP, Delahay RJ, Hodgson DJ, Boots M, Weber N, McDonald RA. 2017. Using social network measures in wildlife disease ecology, epidemiology, and management. BioScience 67: 245–257.
- Smith KF, Behrens M, Schloegel LM, Marano N, Burgiel S, Daszak P. 2009. Reducing the risks of the wildlife trade. Science 324: 594–595.
- Smith KM, et al. 2017. Summarizing US wildlife trade with an eye toward assessing the risk of infectious disease introduction. EcoHealth 14: 29–39.
- Springborn MR, Keller RP, Elwood S, Romagosa CM, Zambrana-Torrelio C, Daszak P. 2015. Integrating invasion and disease in the risk assessment of live bird trade. Diversity and Distributions 21: 101–110.
- Strecker AL, Campbell PM, Olden JD. 2011. The aquarium trade as an invasion pathway in the Pacific Northwest. Fisheries 36: 74–85.
- Stringham OC, Lockwood JL. 2018. Pet problems: Biological and economic factors that influence the release of alien reptiles and amphibians by pet owners. Journal of Applied Ecology 55: 2632–2640.
- Stringham OC, Toomes A, Kanishka AM, Mitchell L, Heinrich S, Ross JV, Cassey P. 2020. A guide to using the Internet to monitor and quantify the wildlife trade. Conservation Biology. doi:10.1111/cobi.13675
- Tensen L. 2016. Under what circumstances can wildlife farming benefit species conservation? Global Ecology and Conservation 6: 286–298.
- Vall-llosera M, Cassey P. 2017. 'Do you come from a land down under?' Characteristics of the international trade in Australian endemic parrots. Biological Conservation 207: 38–46.
- van Roon A, Maas M, Toale D, Tafro N, van der Giessen J. 2019. Live exotic animals legally and illegally imported via the main Dutch airport and considerations for public health. PLOS ONE 14: e0220122.
- Warwick C. 2014. The morality of the reptile "pet" trade. Journal of Animal Ethics 4: 74–94.
- Wood EM. 2001. Collection of Coral Reef Fish for Aquaria: Global Trade, Conservation Issues, and Management Strategies. Marine Conservation Society, Ross-on-Wye, United Kindgom.
- Zajicek P, Hardin S, Watson C. 2009. A Florida marine ornamental pathway risk analysis. Reviews in Fisheries Science 17: 156–169.

James Sinclair (james.sinclair270@gmail.com) is a postdoctoral researcher with the Aquatic Ecology Lab at The Ohio State University, in Columbus, Ohio, in the United States. Oliver Stringham is a postdoctoral researcher with the Invasion Science and Wildlife Ecology Lab at The University of Adelaide, in Adelaide, South Australia, in Australia. Bradley Udell is a former PhD student at the University of Florida, in Gainesville, and is currently an ecologist with the US Geological Survey's North American Bat Monitoring Program at the Fort Collins Science Center, in Fort Collins, Colorado, in the United States. Nicholas Mandrak is a professor at the University of Toronto Scarborough, in Toronto, Ontario, in Canada. Brian Leung is an assistant professor at McGill University, the UNESCO Chair for Dialogues on Sustainability, and the Director of the McGill-STRI Neotropical Environment Option, in Montreal, Quebec, in Canada. Christina Romagosa is an associate professor at the University of Florida, in Gainesville, Florida, in the United States. Julie Lockwood is a professor and chair of the Department of Ecology, Evolution, and Natural Resources at Rutgers University, in New Brunswick, New Jersey, in the United States.