

## Article

# High Removal Efficiency of Diatomite-Based X Zeolite for Cu<sup>2+</sup> and Zn<sup>2+</sup>

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**Abstract:** Diatomite-based X zeolite was obtained and its crystallinity, morphology, and interface properties were investigated by XRD, BET, SEM, EDS, and XRF. The obtained X zeolite possessed a unique meso-microporous structure and showed good ion exchange properties for Cu<sup>2+</sup> and Zn<sup>2+</sup>. The pseudo-second-order model and Langmuir isotherm model can best describe the adsorption kinetics and isotherms of Cu<sup>2+</sup> and Zn<sup>2+</sup>, respectively. The maximal adsorption capacities of X zeolite for Cu<sup>2+</sup> and Zn<sup>2+</sup> were 146 and 195 mg/g at 323 K, respectively. Meanwhile, the adsorption process for Cu<sup>2+</sup> and Zn<sup>2+</sup> were chemical adsorption and ion exchange, respectively. Furthermore, the adsorption data turned out to be an endothermic and spontaneous process. Compared with other reported materials, the adsorption capacity of X zeolite synthesized from diatomite was among the highest. Therefore, it could be a promising adsorbent for the disposal of wastewater that contains metal ions.

**Keywords:** X zeolite; Cu<sup>2+</sup> and Zn<sup>2+</sup>; adsorption thermodynamic; adsorption kinetics



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## 1. Introduction

Nowadays, Water pollution has turned out to be more and more serious in the world and especially in developing countries. Heavy metal pollution [1] usually occurred from many industries, such as electroplating, metal mining operations, battery manufacturing, tanneries, fertilizer industries, and leather industries. Furthermore, the US Environmental Protection Agency designated heavy metals as “priority pollutants”, because they are unbiodegradable and tend to accumulate in living organisms [2]. Among these heavy metals, Cu<sup>2+</sup> and Zn<sup>2+</sup> are special, because they played an important role in living organisms and act as micronutrients with trace amounts. However, intake of excess Cu<sup>2+</sup> might cause vomiting, shock, liver and kidney failure, or even death. Excessive intake of Zn<sup>2+</sup> can also cause serious health problems, such as diarrhea, fever, or even acute gastrointestinal disturbances accompanied by nausea. Therefore, we must regulate the maximum allowed concentrations of Cu<sup>2+</sup> and Zn<sup>2+</sup> before releasing them into surface waters. To meet the strict environmental regulations and the rising demands for water quality, the safe disposal of heavy metal pollution has become more and more important and urgent.

Currently, the disposal methods of heavy metals from the aqueous system are usually chemical precipitation [3], coagulation [4–7], membrane separation [8], photocatalytic degradation [9], extraction [10], chemical oxidation, and reduction [11], biological treatment [12] and adsorption technologies [13,14]. Among all these methods, adsorption is highly effective and economical. In addition, the great advantages of adsorption over others are the low generation of residues, the possibility of large-scale processing, easy metal recovery, and the possibility of the reuse of the adsorbents [15]. Therefore, different

types of materials have been developed as adsorbents for the removal of heavy metals, such as modified natural minerals [16,17], activated carbon [18], polymer [19,20], modified metal oxides [21–23] and zeolite [24,25].

Zeolites are well-known adsorbents as they can easily absorb and exchange metal ions with positive charges in their framework because they possess the net negative charges generated from the isomorphic substitution of  $\text{Si}^{4+}$  by  $\text{Al}^{3+}$  due to the compensation by alkali and alkaline-earth metal ions. Meanwhile, the greatest advantage of zeolites over other adsorbents is that they also show ion exchange selectivity. Because they have a three-dimensional framework constructed by tetrahedral units and then linked by shared oxygen atoms to create cavities and channels of different dimensions to control ion selectivity in zeolite structures. In addition, they are highly porous aluminosilicates with different cavity structures that offer large internal and external surface areas, which are beneficial for adsorption [26,27].

Here, we synthesized X zeolite from diatomite. And systematically characterized their physicochemical properties via various techniques. Then we evaluated the adsorption performance of the obtained sample via the removal of  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$ . To the best of our knowledge, the removal of  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  by X zeolite synthesized from diatomite has not been investigated. This work demonstrates the potential performance of X zeolite for the removal of  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  in industrial effluent. Moreover, we systematically studied the adsorption kinetics, isotherms, thermodynamics, and the adsorption mechanisms of X zeolite for  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$ .

## 2. Experiments

### 2.1. Materials

Diatomite (Linjiang, China), copper(II) chloride dehydrate ( $\text{CuCl}_2 \cdot 2\text{H}_2\text{O}$ ), sodium hydroxide (NaOH), zinc sulfate heptahydrate ( $\text{ZnSO}_4 \cdot 7\text{H}_2\text{O}$ ), aluminum hydroxide ( $\text{Al}(\text{OH})_3$ ), hydrochloric acid (HCl). The whole reagents were analytical reagent grade, and distilled water was used in all the experiments.

### 2.2. Batch Adsorption Experiments

We synthesized X zeolite as reported in our previous work [26,28]. And then we investigated its adsorption behaviors via bath experiments. First, forty milligrams of X zeolite was added into 40 mL of  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  solution of different initial concentrations and different temperatures. Then the concentration of  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  solution was measured at adsorption equilibrium and the amount of the adsorbed  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  was determined as follows.

$$q_e = \frac{v(c_i - c_e)}{m \times 1000} \quad (1)$$

where  $q_e$  (mg/g) is the adsorption capacity at equilibrium,  $v$  (mL) is the volume of aqueous solution,  $m$  (g) is the dosage of the adsorbent,  $c_i$  (mg/L) is the initial concentration of  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  solution,  $c_e$  (mg/L) is the equilibrium concentration of  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  solution.

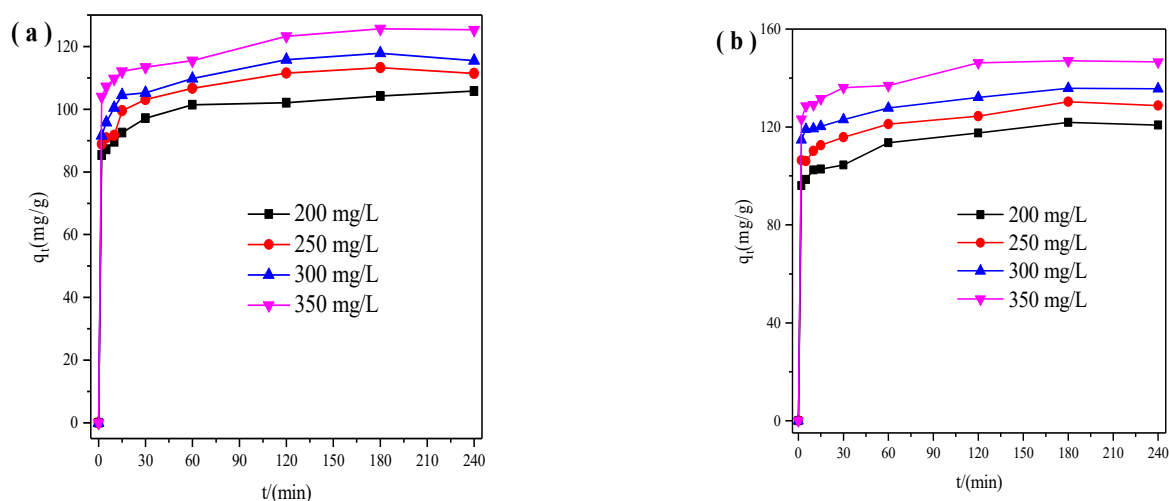
Similarly, the effects of initial solution pH on the adsorption of  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  were studied by adjusting the pH value. First, we studied the adsorption kinetics and added 40 mg of X zeolite into 40 mL of  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  solutions with an initial concentration of 200 mg/L, 250 mg/L, 300 mg/L, and 350 mg/L in order to determine the minimum time needed for equilibrium. Furthermore, we measured the concentrations of  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  solution at fixed intervals. Then we studied the thermodynamic properties and added 40 mg of X zeolite into 40 mL of  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  solution with different initial concentrations at temperatures of 303, 313, and 323 K. All adsorption experiments here were carried out using three replicates and the average values were displayed.

### 3. Results and Discussion

#### 3.1. Adsorption Experiments

##### 3.1.1. Kinetic Models

First, we studied the adsorption behavior of X zeolite for  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  and then got the relationship between time and adsorption capacity (Figure 1). The adsorption capacity of X zeolite for  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  grew fast within 15 min, and then slowly reached up to adsorption equilibrium within 2 h. The easily accessible sorption sites and high surface area of the adsorbent contributed to the rapid adsorption. Meanwhile, the equilibrium adsorption capacities of X zeolite for  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  went up at higher concentrations. The equilibrium adsorption capacities for  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  increased from 104 and 122 mg/g to 126 and 147 mg/g when  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  concentration increased from 200 to 350 mg/L, respectively. The adsorption capacities of  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  increased with their initial concentration raising, due to the higher driving force produced by the concentration gradient [29].



**Figure 1.** Relationship between time and adsorption capacity of X zeolite for  $\text{Cu}^{2+}$  (a) and  $\text{Zn}^{2+}$  (b) with different concentrations.

Then we utilized the pseudo-first and pseudo-second-order model to evaluate the adsorption mechanism of X zeolite for  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  during the adsorption process. The pseudo-first-order model was shown in the following form [30]:

$$\ln(q_e - q_t) = \ln q_e - k_1 t \quad (2)$$

where  $k_1$  ( $\text{min}^{-1}$ ) is the rate constant of pseudo-first-order reaction,  $q_e$  (mg/g) is the equilibrium adsorption capacity of X zeolite for  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$ ,  $q_t$  (mg/g) is the adsorption capacity of X zeolite for  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  at any time  $t$ . From the linear plot of  $\ln(q_e - q_t)$  versus  $t$  (Figure 2), both  $q_e$  and  $k_1$  can be obtained.

The pseudo-second order model was displayed in the following format [31]:

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e} \quad (3)$$

$$h = k_2 q_e^2 \quad (4)$$

where  $k_2$  ( $\text{min}^{-1}$ ) is the rate constant of pseudo-second-order reaction and  $h$  [mg/(g·min)] is the initial sorption rate. The rate constants can be determined via the linear plot  $t/q_t$  against  $t$  (Figure 3).

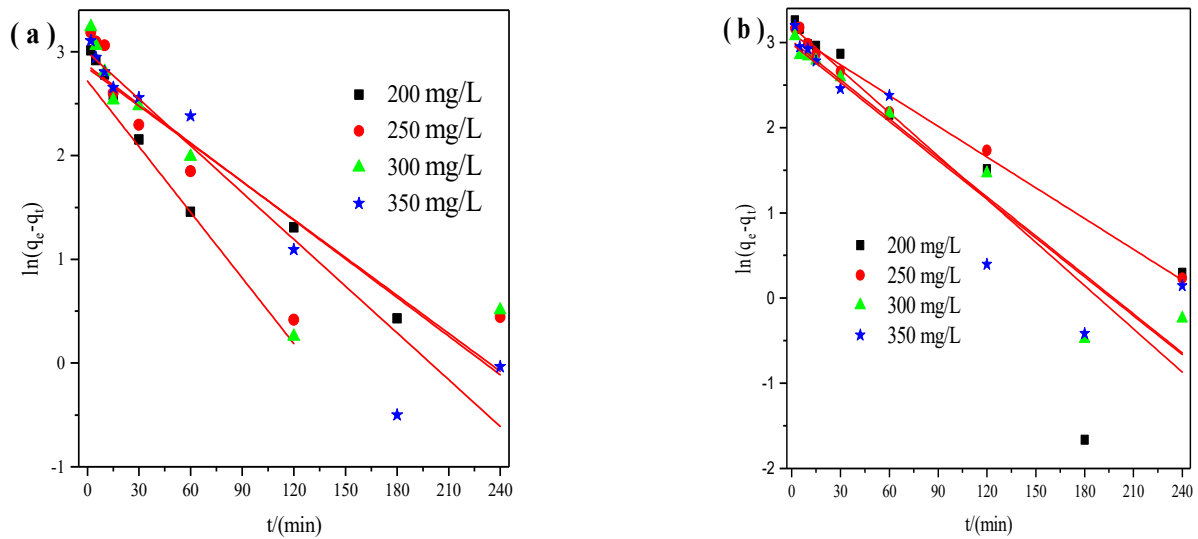


Figure 2. Pseudo–first order model of X zeolite for Cu<sup>2+</sup> (a) and Zn<sup>2+</sup> (b).

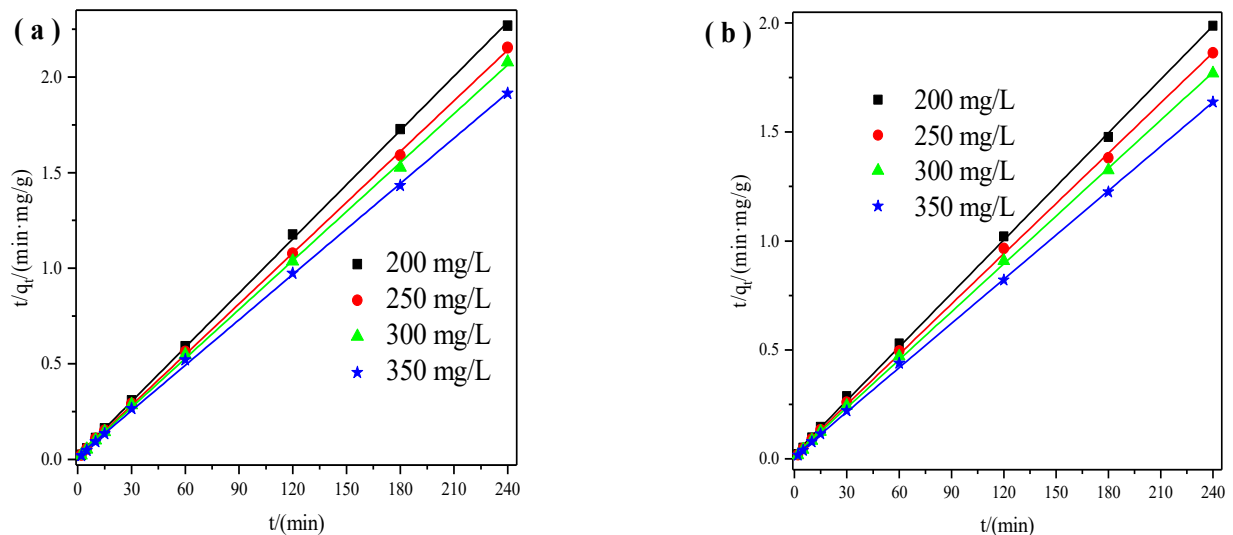


Figure 3. Pseudo–second order model of X zeolite for Cu<sup>2+</sup> (a) and Zn<sup>2+</sup> (b).

Tables 1 and 2 summarized the calculated parameters of the pseudo-first and pseudo-second-order model. The correlation coefficients ( $R^2$ ) of the pseudo-second-order model ranged from 0.9995 to 0.9997, and the pseudo-second-order model ranged from 0.7519 to 0.9867. The correlation coefficients suggested that the adsorption kinetics can best fit the pseudo-second-order model, which indicates that chemical sorption played an important effect during its adsorption process. It may be owing to the exchange of electrons between X zeolite and Cu<sup>2+</sup> or Zn<sup>2+</sup> metal ions [32]. Meanwhile, it found that the values of initial sorption rate ( $h$ ) increased at a higher initial concentration, which indicated that a higher driving force produced by the concentration gradient can promote the adsorption process [22]. Furthermore, the maximum adsorption capacities for Cu<sup>2+</sup> and Zn<sup>2+</sup> could reach 126 and 148 mg/g, respectively.

**Table 1.** Parameters of the pseudo–second order model.

Samples	Cu				Zn			
	$k_2$ (g/(mg·min))	$h$ (mg/(g·min))	$q_e$ (mg/g)	$R^2$	$k_2$ (g/(mg·min))	$h$ (mg/(g·min))	$q_e$ (mg/g)	$R^2$
200	0.00475	53.08	106	0.9997	0.00312	46.51	122	0.9995
250	0.00473	60.39	113	0.9998	0.00330	55.80	130	0.9995
300	0.00471	64.58	117	0.9996	0.00369	68.68	136	0.9997
350	0.00371	59.14	126	0.9996	0.00390	85.09	148	0.9997

**Table 2.** Parameters of pseudo–first order model.

Samples	Cu			Zn		
	$k_1$ (min <sup>-1</sup> )	$q_e$ (mg/g)	$R^2$	$k_1$ (min <sup>-1</sup> )	$q_e$ (mg/g)	$R^2$
200	0.02109	15	0.8617	0.01691	24	0.7519
250	0.01241	18	0.7860	0.01204	22	0.9867
300	0.01217	17	0.7526	0.01526	20	0.9368
350	0.01502	20	0.9160	0.01521	20	0.8668

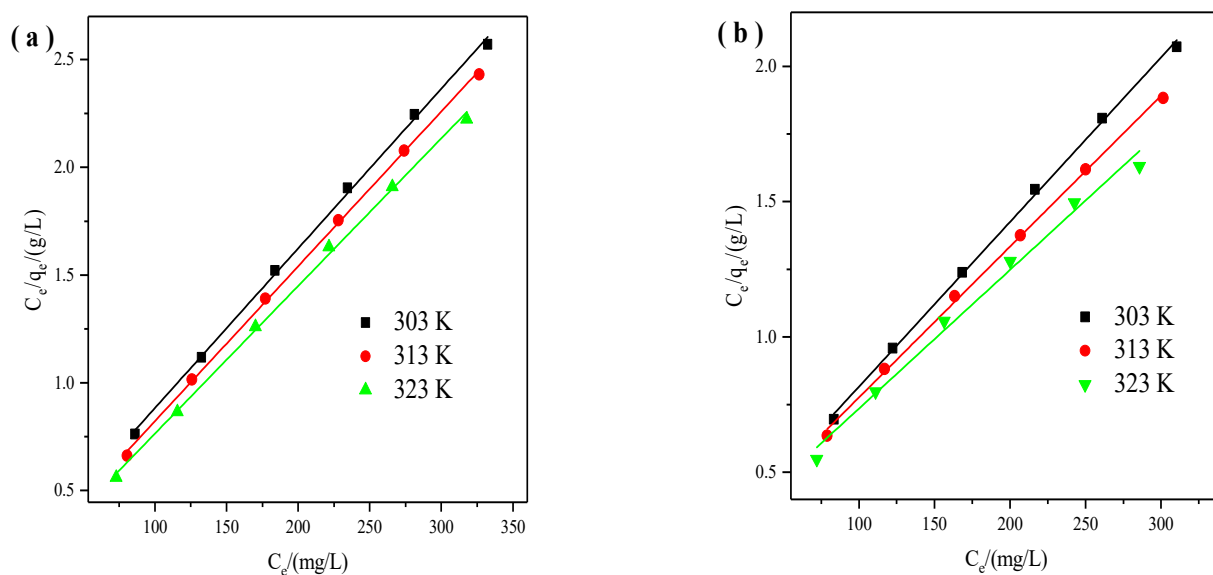
3.1.2. Isotherm Models

To get a deep understanding of the surface property of X zeolite and maximum adsorption capacity, three classic physical isotherm models [32] were utilized to describe the adsorption mechanism of Cu<sup>2+</sup> or Zn<sup>2+</sup> by X zeolite.

We used the Langmuir isotherm model to describe the relationship between Cu<sup>2+</sup> or Zn<sup>2+</sup> solution and the surface of X zeolite. And the Langmuir isotherm model was expressed in the following format [33]:

$$\frac{C_e}{q_e} = \frac{1}{q_m} C_e + \frac{1}{q_m K_L} \tag{5}$$

where  $K_L$  (L/mg) is the Langmuir constant related to the adsorption energy,  $q_m$  (mg/g) is the maximum adsorption capacity of X zeolite. From the linear plot of  $C_e/q_e$  against  $C_e$  (Figure 4), Both  $q_m$  and  $K_L$  can be determined.

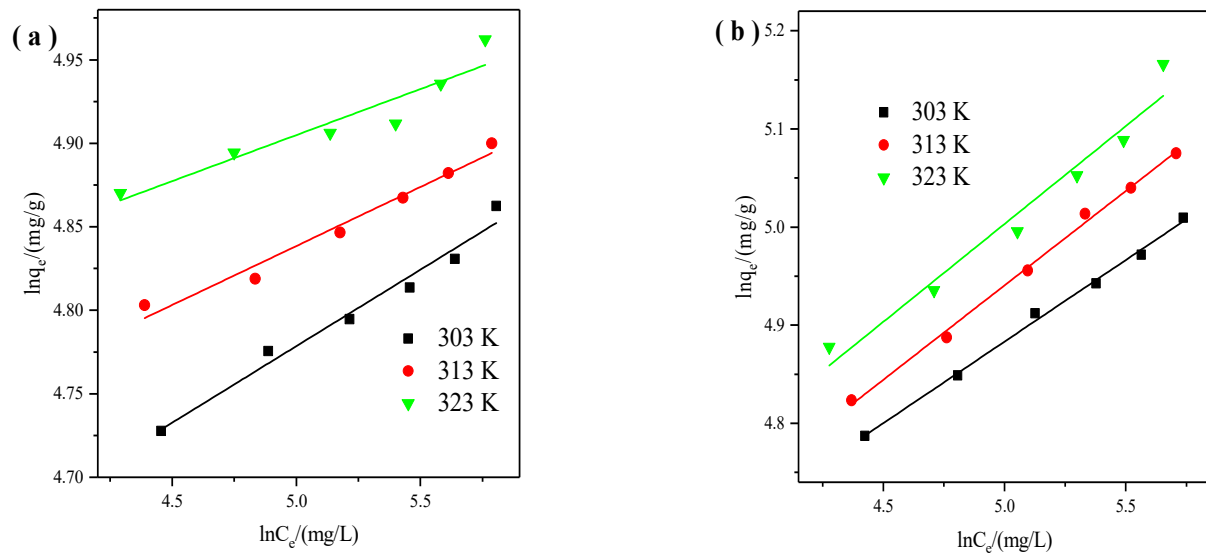


**Figure 4.** Langmuir isotherms of X zeolite for Cu<sup>2+</sup> (a) and Zn<sup>2+</sup> (b).

Meanwhile, We used the Freundlich isotherm model to describe the heterogeneous systems [31]. The equation is described in the following format:

$$\ln q_e = \frac{1}{n} \ln C_e + \ln K_F \quad (6)$$

where  $K_F$  is related to the adsorption capacity,  $n$  is related to the adsorption intensity. From data correlation (Figure 5), both  $n$  and  $K_F$  can be determined.



**Figure 5.** Freundlich isotherms of X zeolite for  $\text{Cu}^{2+}$  (a) and  $\text{Zn}^{2+}$  (b).

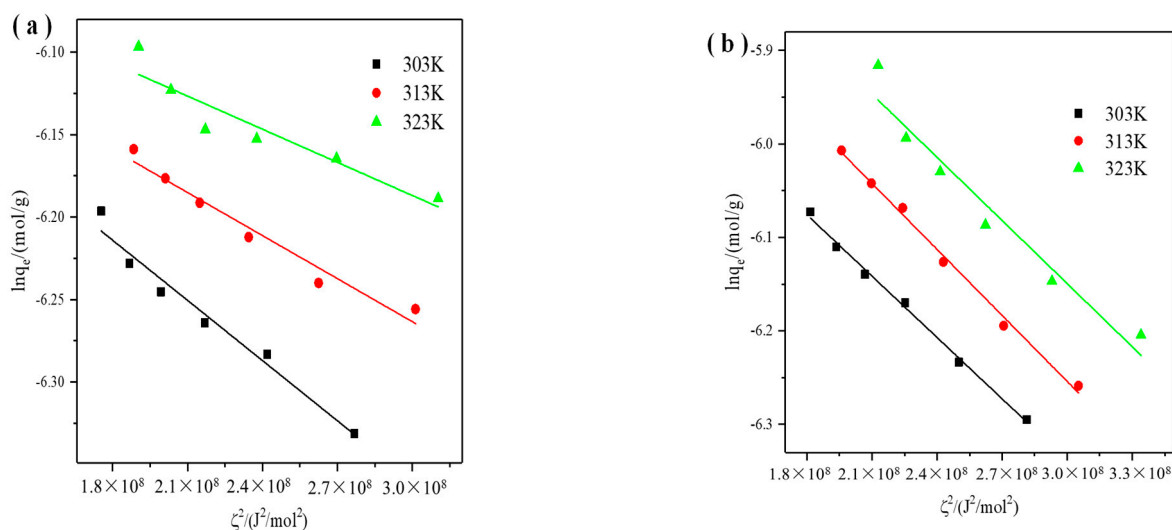
The Dubinin-Radushkevich (D–R) isotherm model is generally used to estimate the adsorption mechanism (physical or chemical adsorption) of microporous adsorbents, which was based on the adsorption potential theory. The adsorption potential theory reflected the change in the Gibbs free energy of an adsorbent after adsorbing a unit molar mass of adsorbate. The equation is expressed as follows:

$$\ln q_e = \ln q_m - \beta \zeta^2 \quad (7)$$

$$\zeta = RT \ln \left( 1 + \frac{1}{C_e} \right) \quad (8)$$

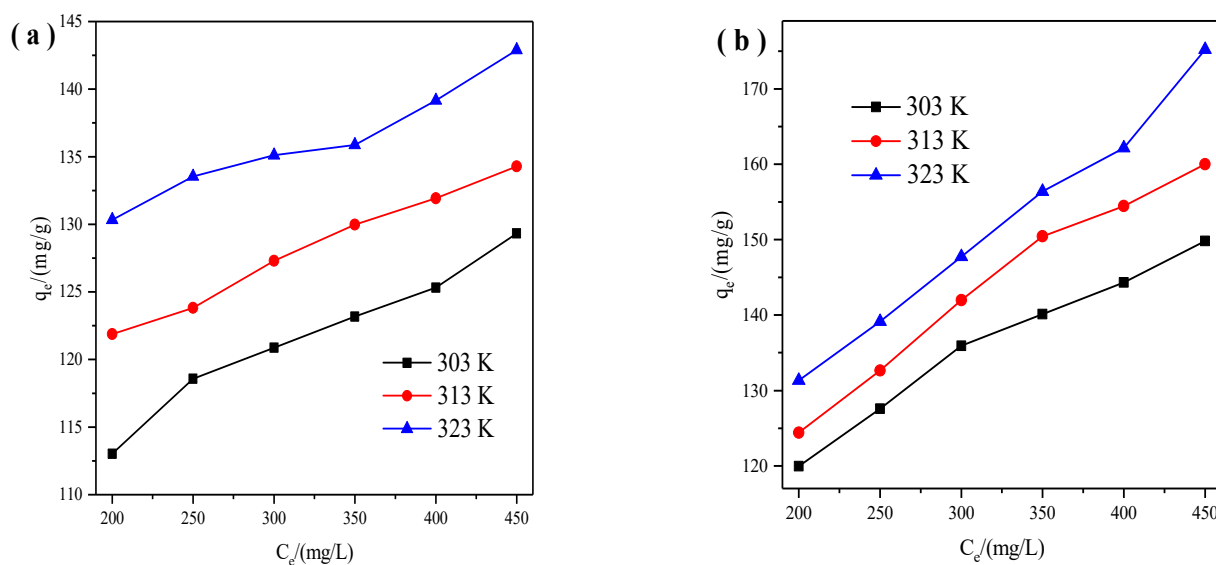
$$E = \frac{1}{(2\beta)^{\frac{1}{2}}} \quad (9)$$

where  $q_e$  (mol/g) is the quantity of adsorbate adsorbed by the adsorbent,  $q_{\max}$  (mol/g) is the quantity of maximum adsorption capacity of the adsorbent (mol/g).  $R$  (8.314/(mol·K)) is the universal gas constant,  $T$  (K) is the absolute solution temperature.  $\beta$  ( $\text{mol}^2/\text{J}^2$ ) and  $\zeta$  (J/mol) are the constants related to adsorption energy and Polanyi adsorption potential, respectively.  $C_e$  is the equilibrium concentration (mol/L). From the linear plot of  $\ln q_e$  against  $\zeta^2$  (Figure 6), both  $q_m$  and  $\beta$  can be determined.  $E$  is the mean free energy of adsorption, which is applied to distinguish the adsorption type. When the magnitude of  $E$  is below 8 KJ/mol, the adsorption process is physical adsorption. When the magnitude of  $E$  is between 8 and 16 KJ/mol, the adsorption process is ion exchange. When the magnitude of  $E$  is above 16 KJ/mol, the adsorption process is chemical adsorption.



**Figure 6.** D-R isotherms of X zeolite for Cu<sup>2+</sup> (a) and Zn<sup>2+</sup> (b).

We handled the experiments at 303 K, 313 K, and 323 K, and the results are displayed in Figure 7. It can be found that the adsorption capacities of X zeolite for Cu<sup>2+</sup> and Zn<sup>2+</sup> reached as the temperature went up, which indicated the endothermic process.



**Figure 7.** Adsorption isotherms of X zeolite for Cu<sup>2+</sup> (a) and Zn<sup>2+</sup> (b).

Tables 3–5 summarized the calculated parameters of Langmuir, Freundlich, and D-R isotherms. All the correlation coefficient values ( $R^2$ ) of the Langmuir, Freundlich, and D-R models are high. However, the Langmuir model correlation coefficient ( $R^2$ ) values of Cu<sup>2+</sup> and Zn<sup>2+</sup> were a little higher demonstrating that the adsorption data followed well with the Langmuir isotherm. These results indicated that Cu<sup>2+</sup> and Zn<sup>2+</sup> were unevenly adsorbed in the form of monolayer coverage. The  $K_L$  constant of the Langmuir parameters demonstrated the binding affinity between X zeolite and Cu<sup>2+</sup> or Zn<sup>2+</sup> [34]. The  $K_L$  values of Cu<sup>2+</sup> range from 0.0521 to 0.0865, and the  $K_L$  values range from 0.0230 to 0.0293. The  $K_L$  values suggested that X zeolite possesses stronger adsorption of Cu<sup>2+</sup> than those of Zn<sup>2+</sup>. The  $n$  values of Freundlich isotherms for Cu<sup>2+</sup> and Zn<sup>2+</sup> were all more than 1 at 303, 313, 323 K demonstrating the facile adsorption between X zeolite and Cu<sup>2+</sup> or Zn<sup>2+</sup> [35]. Meanwhile, higher adsorption capacities were obtained at higher temperatures suggesting the endothermic nature. In addition, the theoretical maximum adsorption capacities for

$\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  by the D-R isotherm model are higher. It may be due to that there are still many active adsorption sites in the micropore when the surface adsorption of X zeolite reaches saturation. And the internal diffusion rate of heavy metals is slow, which decreased the binding chance between heavy metals and adsorption sites in X zeolites, thus resulting in a smaller adsorption capacity. The values of E for  $\text{Cu}^{2+}$  are above 16 kJ/mol suggesting that the adsorption process is chemical adsorption. The values of E for  $\text{Zn}^{2+}$  are between 8 and 16 kJ/mol demonstrating that the adsorption process is ion exchange. The values of E for  $\text{Cu}^{2+}$  range from 20.27 to 27.32 kJ/mol, and the values of E for  $\text{Zn}^{2+}$  range from 14.58 to 15.13 kJ/mol. The values of E suggested that X zeolite possesses stronger adsorption of  $\text{Cu}^{2+}$  than those of  $\text{Zn}^{2+}$ , which is consistent with the characteristics of  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$ . Therefore, the prepared X zeolite probably possessed selective adsorption towards  $\text{Cu}^{2+}$  [27]. Moreover, the maximum adsorption capacities of X zeolite for  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  can reach 146 and 195 mg/g, respectively.

**Table 3.** Parameters of Langmuir and Freundlich isotherms for the adsorption of  $\text{Cu}^{2+}$ .

T/(K)	Langmuir			Freundlich		
	$q_m$ (mg/g)	$K_L$ (L/mg)	$R^2$	$K_F$ ( $\text{mg}^{1-(1/n)}\text{L}^{1/n}\text{g}^{-1}$ )	$n$	$R^2$
303	135	0.0521	0.9985	75.24	10.92	0.9708
313	139	0.0691	0.9993	88.69	14.15	0.9683
323	146	0.0865	0.9982	102.44	18.14	0.8707

**Table 4.** Parameters of Langmuir and Freundlich isotherms for the adsorption of  $\text{Zn}^{2+}$ .

T/(K)	Langmuir			Freundlich		
	$q_m$ (mg/g)	$K_L$ (L/mg)	$R^2$	$K_F$ ( $\text{mg}^{1-(1/n)}\text{L}^{1/n}\text{g}^{-1}$ )	$n$	$R^2$
303	164	0.0293	0.9985	57.53	6.02	0.9956
313	179	0.0255	0.9984	53.44	5.20	0.9956
323	195	0.0230	0.9881	54.92	5.01	0.9530

**Table 5.** Parameters of D-R isotherms for the adsorption of  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$ .

T/(K)	$\text{Cu}^{2+}$				$\text{Zn}^{2+}$			
	$q_m$ (mg/g)	$\beta$ ( $\text{mol}^2/\text{J}^2$ )	E (KJ/mol)	$R^2$	$q_m$ (mg/g)	$\beta$ ( $\text{mol}^2/\text{J}^2$ )	E (KJ/mol)	$R^2$
303	158	$1.22 \times 10^{-9}$	20.27	0.9654	221	$2.19 \times 10^{-9}$	15.13	0.9947
313	157	$0.87 \times 10^{-9}$	23.97	0.9541	253	$2.35 \times 10^{-9}$	14.58	0.9933
323	160	$0.67 \times 10^{-9}$	27.32	0.8510	273	$2.25 \times 10^{-9}$	14.91	0.9373

In addition, we compared the adsorption capacities of X zeolite for  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  with some other reported materials. As shown in Tables 6 and 7, the prepared X zeolite possessed higher adsorption capacities for  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  than those of other reported adsorbents. Meanwhile, the present synthetic method of X zeolite is environmentally friendly and perhaps low-cost compared with other methods. Therefore, X zeolite prepared from diatomite could be a candidate for removing  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  from wastewater.



**Table 6.** Comparison of adsorption capacities of various adsorbents for Cu<sup>2+</sup>.

Adsorbent	Synthetic Method	Q <sub>m</sub> (mg/g)	Refs.
Zeolite from fly ash	Hydrothermal method	64	[36]
A zeolite from coal fly ash	Hydrothermal and impregnation methods	50	[37]
Hematite	Co-precipitation method	84	[38]
α-MnO <sub>2</sub>	Precipitation method	83	[39]
Fe <sub>3</sub> O <sub>4</sub> @APS@AA-co-CA MNPs	Chemical co-precipitation methods	127	[21]
X zeolite	Hydrothermal method	146	The paper

**Table 7.** Comparison of adsorption capacities of various adsorbents for Zn<sup>2+</sup>.

Adsorbent	Synthetic Method	Q <sub>m</sub> (mg/g)	Refs.
A zeolite	Hydrothermal and calcination methods	80	[24]
X zeolite	Hydrothermal and calcination methods	68	[24]
A zeolite from coal fly ash	Hydrothermal and impregnation methods	31	[37]
Hydrous manganese dioxide	Precipitation method	57	[40]
Fe <sub>3</sub> O <sub>4</sub> @APS@AA-co-CA MNPs	Chemical co-precipitation methods	43	[21]
X zeolite	Hydrothermal method	195	The paper

### 3.1.3. Adsorption thermodynamics

Generally, we can get more detailed data of internal energy within the adsorption process through the study of thermodynamics. And the thermodynamic parameters for the adsorption of Cu<sup>2+</sup> and Zn<sup>2+</sup> can be described in the following format [41].

$$\Delta G^0 = \Delta H^0 - T\Delta S^0 \quad (10)$$

$$\ln K_d = \frac{\Delta S^0}{R} - \frac{\Delta H^0}{RT} \quad (11)$$

$$K_d = \frac{q_e}{C_e} \quad (12)$$

where  $\Delta G^0$  (kJ/mol) is the standard free energy,  $\Delta S^0$  (J/K) is the standard entropy,  $\Delta H^0$  (kJ/mol) is the standard enthalpy,  $R$  (8.314/(molK)) is the universal gas constant,  $T$  (K) is the absolute solution temperature,  $K_d$  is the distribution coefficient. Finally, we can get the  $\Delta G^0$ ,  $\Delta H^0$ , and  $\Delta S^0$ , and they are presented in Figure 8.

The calculated thermodynamic parameters of Cu<sup>2+</sup> and Zn<sup>2+</sup> were listed in Tables 8 and 9, respectively. The values of  $\Delta H^0$  for Cu<sup>2+</sup> and Zn<sup>2+</sup> range from 5.88 to 12.52 KJ/mol, and the values of  $\Delta S^0$  for Cu<sup>2+</sup> and Zn<sup>2+</sup> range from 68.92 to 100.92 J/mol. They demonstrated the endothermic nature and good affinity of X zeolite for Cu<sup>2+</sup> and Zn<sup>2+</sup> metal ions, respectively. And positive  $\Delta S^0$  values of Cu<sup>2+</sup> and Zn<sup>2+</sup> suggested the increased randomness between the interface of X zeolite and Cu<sup>2+</sup> or Zn<sup>2+</sup>. However,  $\Delta S^0$  for Cu<sup>2+</sup> and Zn<sup>2+</sup> went down as their initial concentration went up, which is consistent with the previous literature [42]. The values of  $\Delta G^0$  for Cu<sup>2+</sup> and Zn<sup>2+</sup> at three temperatures are all negative indicating the feasible and spontaneous nature. Moreover,  $\Delta G^0$  for Cu<sup>2+</sup> and Zn<sup>2+</sup> all went down with temperature went up indicating the spontaneous nature with high affinity [43]. The  $\Delta G^0$  for Cu<sup>2+</sup> and Zn<sup>2+</sup> went up as their initial concentration increased, which is also in accordance with the previous work [32].

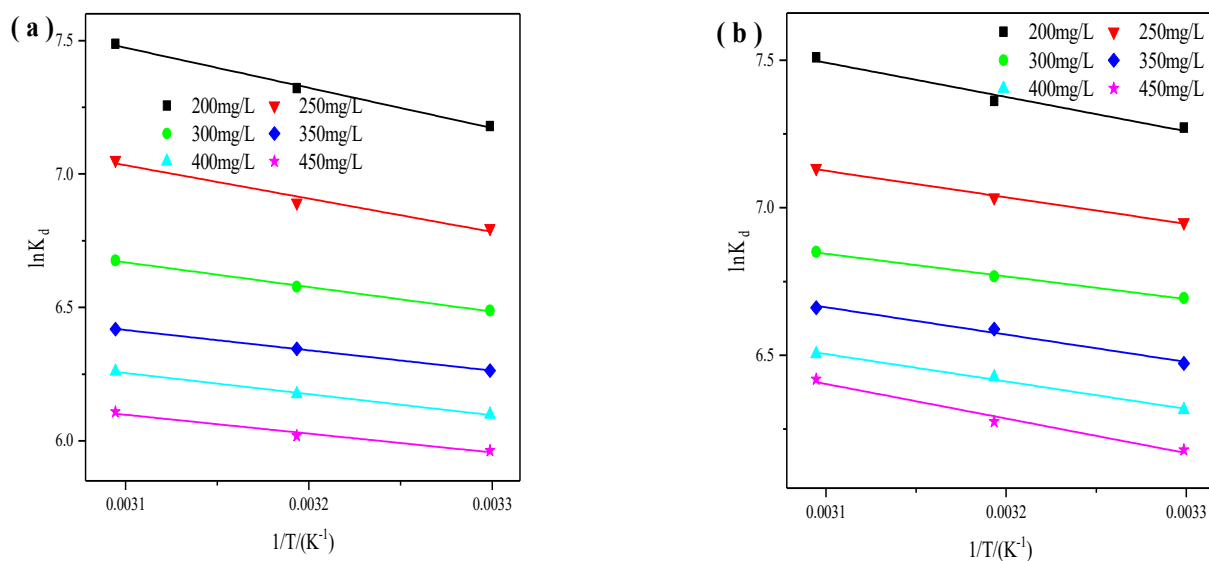


Figure 8. Plot of  $\ln K_d$  versus  $1/T$  at various initial concentrations of  $Cu^{2+}$  (a) and  $Zn^{2+}$  (b).

Table 8. Thermodynamic parameters for  $Cu^{2+}$  adsorption.

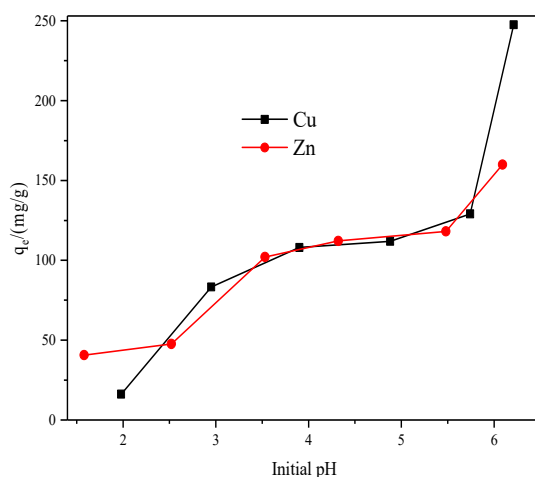
$C_0$ (mg/L)	$\Delta H^0$ (KJ/mol)	$\Delta S^0$ (J/mol)	$\Delta G^0$ (KJ/mol)		
			303 K	313 K	323 K
200	12.52	100.92	−18.07	−19.08	−20.08
250	10.38	90.64	−17.09	−17.99	−18.90
300	7.68	79.26	−16.33	−17.13	−17.92
350	6.33	72.97	−15.78	−16.51	−17.24
400	6.58	72.39	−15.36	−16.08	−16.80
450	5.88	68.92	−15.01	−15.69	−16.38

Table 9. Thermodynamic parameters for  $Zn^{2+}$  adsorption.

$C_0$ (mg/L)	$\Delta H^0$ (KJ/mol)	$\Delta S^0$ (J/mol)	$\Delta G^0$ (KJ/mol)		
			303 K	313 K	323 K
200	9.66	92.22	−18.29	−19.21	−20.13
250	7.47	82.38	−17.50	−18.32	−19.14
300	6.38	76.69	−16.85	−17.62	−18.39
350	7.70	79.26	−16.32	−17.11	−17.90
400	7.72	78.02	−15.92	−16.70	−17.48
450	9.74	83.43	−15.54	−16.37	−17.21

### 3.1.4. Effect of Initial pH

Because the solution pH is one of the critical factors that directly influence the solubility of metals, the effect of initial pH on the adsorption capacity of X zeolite was studied in the pH range of 1.0–6.0. Generally, the increase in pH results in a decrease in solubility of metals due to the formation of metal hydroxides of low solubility. Therefore, making metals soluble in solution is critical for adsorption. As shown in Figure 9, the adsorption capacity of X zeolite increased as the pH value increased. According to the experimental data, the initial pH values of  $Cu(II)$  and  $Zn(II)$  solutions were 4.88 and 5.48, respectively with the concentration of 250 ppm. Some blue and white flocs appeared with increasing the pH value of  $Cu(II)$  and  $Zn(II)$  solution due to hydrolysis and the higher adsorption capacity at high pH value was due to the hydroxide precipitation [44]. However, the lower adsorption capacity at lower solution pH was owing to the excessive  $H^+$ , which will compete with the heavy metals for adsorption sites.

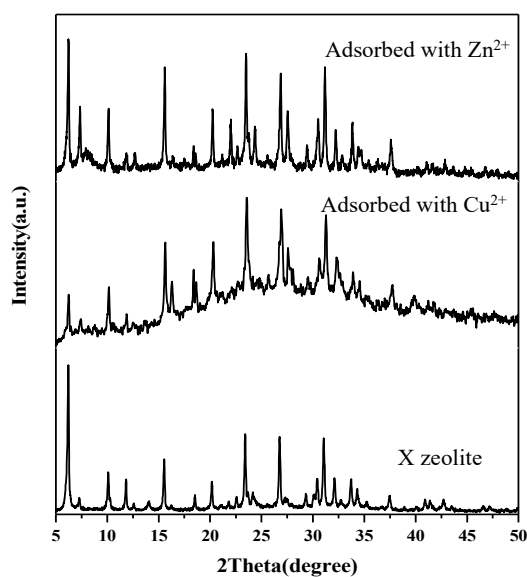


**Figure 9.** Effect of initial pH on the adsorption capacity of X zeolite for  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$ .

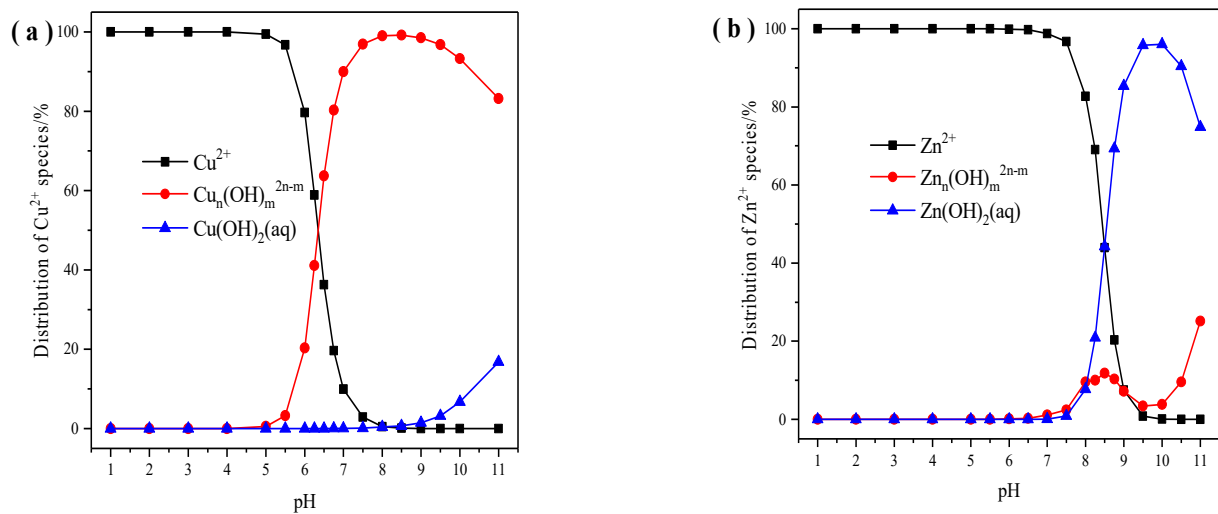
### 3.2. Characterization of X Zeolite

#### 3.2.1. XRD

XRD patterns of X zeolite and samples after adsorption with  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  are shown in Figure 10. After adsorption of  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$ , we can still observe the mineral structure of X zeolite, and the intensity of X zeolite at  $2\theta = 6.10^\circ, 9.97^\circ, 15.39^\circ, 23.24^\circ, 26.58^\circ,$  and  $30.86^\circ$  decreased in intensity and samples after adsorption with  $\text{Cu}^{2+}$  appears to show some amorphous background possibly due to Cu hydroxide precipitation. Moreover, the software Visual MINTEQ was applied to get the state distribution of  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  as pH values range from 1–11. When the pH of the  $\text{Cu}^{2+}$  solution is lower than 5.0, the copper existed in the  $\text{Cu}^{2+}$  state (Figure 11a). The pH of the  $\text{Cu}^{2+}$  solution became 5.28 after adsorption with  $\text{Cu}^{2+}$ . When the pH of the  $\text{Zn}^{2+}$  solution is lower than 7.5, the zinc existed in the  $\text{Zn}^{2+}$  state (Figure 11b). The pH of the  $\text{Zn}^{2+}$  solution became 6.25 after adsorption with  $\text{Zn}^{2+}$ . Based on the above analysis, the state of copper and zinc after adsorption of  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  are  $\text{Zn}^{2+}$ ,  $\text{Cu}^{2+}$ , and a small amount of Cu hydroxide precipitation, respectively. The pH analysis is in agreement with the XRD data.



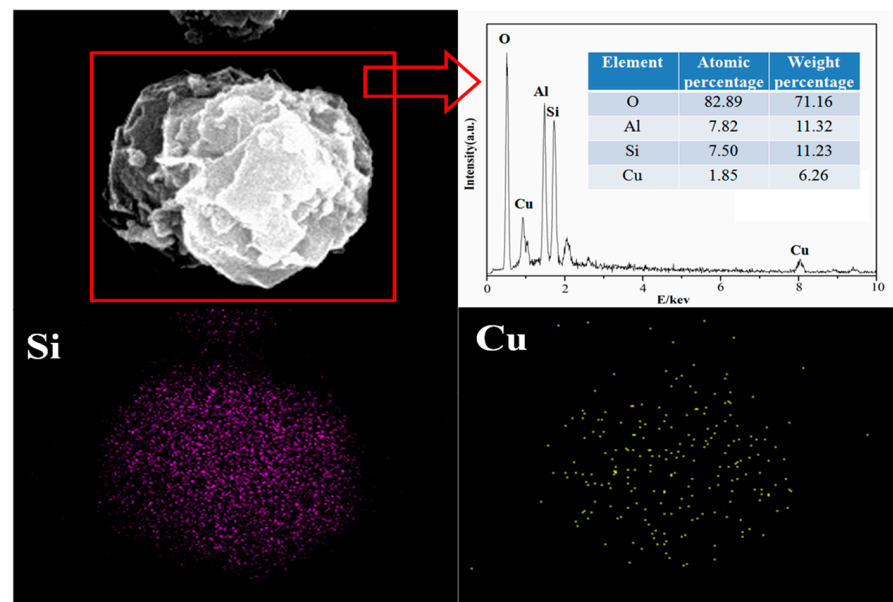
**Figure 10.** XRD patterns of X zeolite and samples after adsorption with  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$ .



**Figure 11.** The state distribution of  $\text{Cu}^{2+}$  (a) and  $\text{Zn}^{2+}$  (b) at different pH values.

### 3.2.2. SEM

The morphology of X zeolite and samples after adsorption with  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  are shown in Figures 12 and 13, respectively. The SEM images of X zeolite are an aggregation of small particles in spherical shape according to our previous work [45]. The morphology of X zeolite was almost unchanged with  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$  dispersed well on X zeolite after adsorption of  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$ , respectively. The compositions of X zeolite and samples after adsorption of  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  are presented in Table 10. After adsorption, the composition of sodium greatly decreased. These chemical analyses suggest that Na ions were exchanged with  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  by the ion-exchange process.



**Figure 12.** SEM images and EDS analysis of X zeolite adsorbed with  $\text{Cu}^{2+}$ .

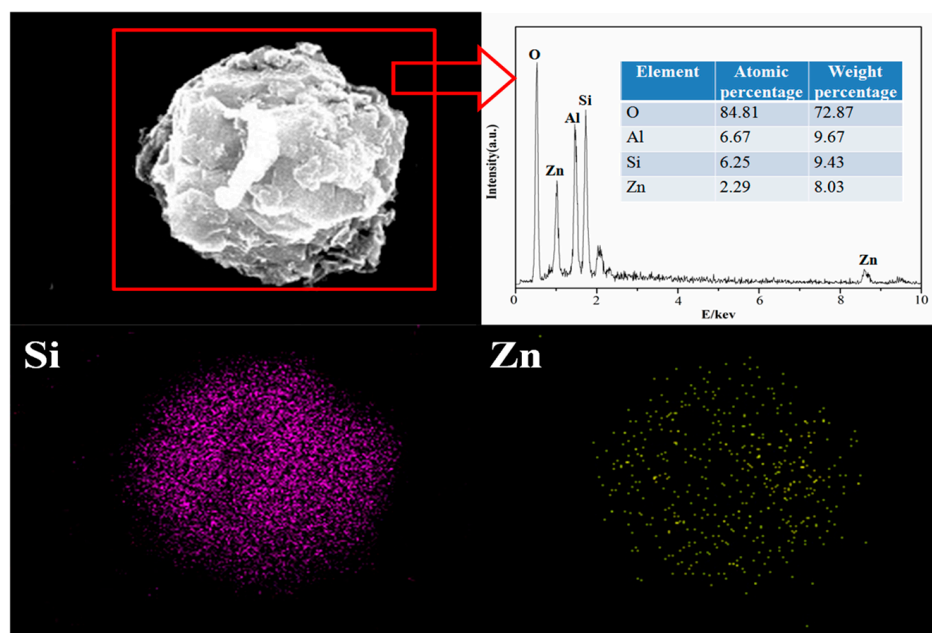


Figure 13. SEM images and EDS analysis of X zeolite adsorbed with  $Zn^{2+}$ .

Table 10. The composition of X zeolite and samples of X zeolite after adsorption with  $Cu^{2+}$  or  $Zn^{2+}$ .

Samples	Si	Al	Cu	Zn	Na	O	$n_{(Si/Al)}$
X zeolite	21.41	17.07	/	/	14.6	45.33	1.21
Sample adsorbed with $Cu^{2+}$	19.72	16.58	16.14	/	2.40	42.51	1.15
Sample adsorbed with $Zn^{2+}$	18.53	16.01	/	16.69	4.56	42.29	1.12

### 3.2.3. BET

According to our previous work [27], the obtained X zeolite from diatomite possessed a unique meso-microporous structure with a high BET surface of  $453 \text{ m}^2/\text{g}$ , which can improve the mass diffusion and transport of  $Cu^{2+}$  and  $Zn^{2+}$  by utilizing its large cavity volumes and mesopore channels. And then it will increase the effective meso-microporous contact area, which is beneficial for the adsorption process [46]. The adsorption mechanism of X zeolite for  $Cu^{2+}$  and  $Zn^{2+}$  can be described as follows: first, metal ions diffused to the surface of X zeolite through its liquid membrane; then metal ions diffused from the surface to the interior of X zeolite; finally, metal ions exchanged with cations at the active sites in the X zeolite.

## 4. Conclusions

The X zeolite was synthesized from diatomite displayed a unique meso-microporous structure. The Langmuir isotherm model can best describe the isotherms of  $Cu^{2+}$  and  $Zn^{2+}$  with the maximum adsorption capacities of 146 and 195 mg/g, respectively. The pseudo-second-order model can best describe the kinetic of  $Cu^{2+}$  and  $Zn^{2+}$ . The adsorption process for  $Cu^{2+}$  and  $Zn^{2+}$  is chemical adsorption and ion exchange, respectively. Therefore, the prepared X zeolite probably possessed selective adsorption towards  $Cu^{2+}$ . Moreover, the Langmuir model can best describe the isotherm of  $Cu^{2+}$  and  $Zn^{2+}$  indicating that  $Cu^{2+}$  and  $Zn^{2+}$  were unevenly adsorbed in the form of monolayer coverage. In addition, the adsorption processes of  $Cu^{2+}$  and  $Zn^{2+}$  are both endothermic and spontaneous. The excellent adsorption performance of X zeolite could be attributed to the meso-microporous structure, which improved its mass diffusion and transport of  $Cu^{2+}$  and  $Zn^{2+}$  by its large cavity volumes and mesopore channels with effective meso-microporous contact area.

Therefore, the X zeolite is a promising and potentially cost-effective adsorbent for the efficient disposal of wastewater that contains more metal ions.

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## References

1. Fu, F.; Wang, Q. Removal of heavy metal ions from wastewaters: A review. *J. Environ. Manag.* **2011**, *92*, 407–418. [\[CrossRef\]](#)
2. Zhang, Y.; Zheng, R.; Zhao, J.; Zhang, Y.; Wong, P.K.; Ma, F. Biosorption of zinc from aqueous solution using chemically treated rice husk. *BioMed Res. Int.* **2013**, *2013*, 365163. [\[CrossRef\]](#)
3. Taty-Costodes, V.C.; Fauduet, H.; Porte, C.; Delacroix, A. Removal of Cd(II) and Pb(II) ions, from aqueous solutions, by adsorption onto sawdust of *Pinus sylvestris*. *J. Hazard. Mater.* **2003**, *105*, 121–142. [\[CrossRef\]](#)
4. Kulik, N.; Panova, Y.; Trapido, M. The Fenton Chemistry and Its Combination with Coagulation for Treatment of Dye Solutions. *Sep. Sci. Technol.* **2007**, *42*, 1521–1534. [\[CrossRef\]](#)
5. López-Maldonado, E.A.; Oropeza-Guzmán, M.T. Nejayote biopolyelectrolytes multifunctionality (glucurono feruloylated arabinoxylans) in the separation of hazardous metal ions from industrial wastewater. *Chem. Eng. J.* **2021**, *423*, 130210. [\[CrossRef\]](#)
6. Xiao, X.; Yu, Y.; Sun, Y.; Zheng, X.; Chen, A. Heavy metal removal from aqueous solutions by chitosan-based magnetic composite flocculants. *J. Environ. Sci.* **2021**, *108*, 22–32. [\[CrossRef\]](#) [\[PubMed\]](#)
7. Macczak, P.; Kaczmarek, H.; Ziegler-Borowska, M. Recent Achievements in Polymer Bio-Based Flocculants for Water Treatment. *Materials* **2020**, *13*, 3951. [\[CrossRef\]](#) [\[PubMed\]](#)
8. Zhang, L.; Zhao, Y.-H.; Bai, R. Development of a multifunctional membrane for chromatic warning and enhanced adsorptive removal of heavy metal ions: Application to cadmium. *J. Membr. Sci.* **2011**, *379*, 69–79. [\[CrossRef\]](#)
9. Sun, Z.; Zheng, L.; Zheng, S.; Frost, R.L. Preparation and characterization of TiO<sub>2</sub>/acid leached serpentinite tailings composites and their photocatalytic reduction of chromium(VI). *J. Colloid Interface Sci.* **2013**, *404*, 102–109. [\[CrossRef\]](#) [\[PubMed\]](#)
10. Luo, J.; Li, J.; Yang, Z.; Liu, X. Removal of chromium(III) from aqueous waste solution by predispersed solvent extraction. *Trans. Nonferrous Met. Soc. China* **2013**, *23*, 524–529. [\[CrossRef\]](#)
11. Ölmez, T.; Kabdaşlı, I.; Tünay, O. The effect of the textile industry dye bath additive EDTMPA on colour removal characteristics by ozone oxidation. *Water Sci. Technol.* **2007**, *55*, 145. [\[CrossRef\]](#)
12. Allègre, C.; Moulin, P.; Maisseu, M.; Charbit, F. Treatment and reuse of reactive dyeing effluents. *J. Membr. Sci.* **2006**, *269*, 15–34. [\[CrossRef\]](#)
13. Feng, Y.; Gong, J.-L.; Zeng, G.-M.; Niu, Q.-Y.; Zhang, H.-Y.; Niu, C.-G.; Deng, J.-H.; Yan, M. Adsorption of Cd (II) and Zn (II) from aqueous solutions using magnetic hydroxyapatite nanoparticles as adsorbents. *Chem. Eng. J.* **2010**, *162*, 487–494. [\[CrossRef\]](#)
14. Banerjee, S.S.; Chen, D.H. Fast removal of copper ions by gum arabic modified magnetic nano-adsorbent. *J. Hazard. Mater.* **2007**, *147*, 792–799. [\[CrossRef\]](#)
15. Zhang, S.; Zhang, Y.; Ding, J.; Zhang, Z.; Gao, C.; Halimi, M.; Demey, H.; Yang, Z.; Yang, W. High phosphate removal using La(OH)<sub>3</sub> loaded chitosan based composites and mechanistic study. *J. Environ. Sci.* **2021**, *106*, 105–115. [\[CrossRef\]](#) [\[PubMed\]](#)
16. Ivanets, A.I.; Kitikova, N.V.; Shashkova, I.L.; Oleksienko, O.V.; Levchuk, I.; Sillanpää, M. Removal of Zn<sup>2+</sup>, Fe<sup>2+</sup>, Cu<sup>2+</sup>, Pb<sup>2+</sup>, Cd<sup>2+</sup>, Ni<sup>2+</sup> and Co<sup>2+</sup> ions from aqueous solutions using modified phosphate dolomite. *J. Environ. Chem. Eng.* **2014**, *2*, 981–987. [\[CrossRef\]](#)
17. Abu-Zurayk, R.A.; al Bakain, R.Z.; Hamadneh, I.; Al-Dujaili, A.H. Adsorption of Pb(II), Cr(III) and Cr(VI) from aqueous solution by surfactant-modified diatomaceous earth: Equilibrium, kinetic and thermodynamic modeling studies. *Int. J. Miner. Process.* **2015**, *140*, 79–87. [\[CrossRef\]](#)
18. Terdkiatburana, T.; Wang, S.; Tadó, M.O. Competition and complexation of heavy metal ions and humic acid on zeolitic MCM-22 and activated carbon. *Chem. Eng. J.* **2008**, *139*, 437–444. [\[CrossRef\]](#)

19. Tao, P.; Shao, M.; Song, C.; Wu, S.; Cheng, M.; Cui, Z. Preparation of porous and hollow Mn<sub>2</sub>O<sub>3</sub> microspheres and their adsorption studies on heavy metal ions from aqueous solutions. *J. Ind. Eng. Chem.* **2014**, *20*, 3128–3133. [[CrossRef](#)]
20. Wang, W.; Ding, Z.; Cai, M.; Jian, H.; Zeng, Z.; Li, F.; Liu, J.P. Synthesis and high-efficiency methylene blue adsorption of magnetic PAA/MnFe<sub>2</sub>O<sub>4</sub> nanocomposites. *Appl. Surf. Sci.* **2015**, *346*, 348–353. [[CrossRef](#)]
21. Ge, F.; Li, M.M.; Ye, H.; Zhao, B.X. Effective removal of heavy metal ions Cd<sup>2+</sup>, Zn<sup>2+</sup>, Pb<sup>2+</sup>, Cu<sup>2+</sup> from aqueous solution by polymer-modified magnetic nanoparticles. *J. Hazard. Mater.* **2012**, *211–212*, 366–372. [[CrossRef](#)] [[PubMed](#)]
22. Han, R.; Zou, W.; Zhang, Z.; Shi, J.; Yang, J. Removal of copper(II) and lead(II) from aqueous solution by manganese oxide coated sand I. Characterization and kinetic study. *J. Hazard. Mater.* **2006**, *137*, 384–395. [[CrossRef](#)]
23. Hua, M.; Zhang, S.; Pan, B.; Zhang, W.; Lv, L.; Zhang, Q. Heavy metal removal from water/wastewater by nanosized metal oxides: A review. *J. Hazard. Mater.* **2012**, *211–212*, 317–331. [[CrossRef](#)] [[PubMed](#)]
24. Nibou, D.; Mekatel, H.; Amokrane, S.; Barkat, M.; Trari, M. Adsorption of Zn<sup>2+</sup> ions onto NaA and NaX zeolites: Kinetic, equilibrium and thermodynamic studies. *J. Hazard. Mater.* **2010**, *173*, 637–646. [[CrossRef](#)] [[PubMed](#)]
25. Zanin, E.; Scapinello, J.; de Oliveira, M.; Rambo, C.L.; Franscescon, F.; Freitas, L.; de Mello, J.M.M.; Fiori, M.A.; Oliveira, J.V.; Magro, J.D. Adsorption of heavy metals from wastewater graphic industry using clinoptilolite zeolite as adsorbent. *Process Saf. Environ. Prot.* **2017**, *105*, 194–200. [[CrossRef](#)]
26. Yao, G.; Lei, J.; Zhang, X.; Sun, Z.; Zheng, S.; Komarneni, S. Mechanism of zeolite X crystallization from diatomite. *Mater. Res. Bull.* **2018**, *107*, 132–138. [[CrossRef](#)]
27. Yao, G.; Zhang, X.; Sun, Z.; Zheng, S. High adsorption selectivity of zeolite X in the binary ionic system of Cu(II) and Zn(II). *J. Porous Mater.* **2019**, *26*, 1197–1207. [[CrossRef](#)]
28. Yao, G.; Lei, J.; Zhang, X.; Sun, Z.; Zheng, S. One-Step Hydrothermal Synthesis of Zeolite X Powder from Natural Low-Grade Diatomite. *Materials* **2018**, *11*, 906. [[CrossRef](#)]
29. Zou, W.; Han, R.; Chen, Z.; Jinghua, Z.; Shi, J. Kinetic study of adsorption of Cu(II) and Pb(II) from aqueous solutions using manganese oxide coated zeolite in batch mode. *Colloids Surf. A Physicochem. Eng. Asp.* **2006**, *279*, 238–246. [[CrossRef](#)]
30. Park, Y.; Sun, Z.; Ayoko, G.A.; Frost, R.L. Bisphenol A sorption by organo-montmorillonite: Implications for the removal of organic contaminants from water. *Chemosphere* **2014**, *107*, 249–256. [[CrossRef](#)]
31. Park, Y.; Ayoko, G.A.; Horvath, E.; Kurdi, R.; Kristof, J.; Frost, R.L. Structural characterisation and environmental application of organoclays for the removal of phenolic compounds. *J. Colloid Interface Sci.* **2013**, *393*, 319–334. [[CrossRef](#)] [[PubMed](#)]
32. Sun, Z.; Yao, G.; Liu, M.; Zheng, S. In situ synthesis of magnetic MnFe<sub>2</sub>O<sub>4</sub>/diatomite nanocomposite adsorbent and its efficient removal of cationic dyes. *J. Taiwan Inst. Chem. Eng.* **2017**, *71*, 501–509. [[CrossRef](#)]
33. Zheng, S.; Sun, Z.; Park, Y.; Ayoko, G.A.; Frost, R.L. Removal of bisphenol A from wastewater by Ca-montmorillonite modified with selected surfactants. *Chem. Eng. J.* **2013**, *234*, 416–422. [[CrossRef](#)]
34. Ates, A. Role of modification of natural zeolite in removal of manganese from aqueous solutions. *Powder Technol.* **2014**, *264*, 86–95. [[CrossRef](#)]
35. Cui, L.; Hu, L.; Guo, X.; Zhang, Y.; Wang, Y.; Wei, Q.; Du, B. Kinetic, isotherm and thermodynamic investigations of Cu<sup>2+</sup> adsorption onto magnesium hydroxyapatite/ferroferrous oxide nano-composites with easy magnetic separation assistance. *J. Mol. Liq.* **2014**, *198*, 157–163. [[CrossRef](#)]
36. Wang, S.; Soudi, M.; Li, L.; Zhu, Z.H. Coal ash conversion into effective adsorbents for removal of heavy metals and dyes from wastewater. *J. Hazard. Mater.* **2006**, *133*, 243–251. [[CrossRef](#)]
37. Hui, K.S.; Chao, C.Y.; Kot, S.C. Removal of mixed heavy metal ions in wastewater by zeolite 4A and residual products from recycled coal fly ash. *J. Hazard. Mater.* **2005**, *127*, 89–101. [[CrossRef](#)]
38. Chen, Y.H.; Li, F.A. Kinetic study on removal of copper(II) using goethite and hematite nano-photocatalysts. *J. Colloid Interface Sci.* **2010**, *347*, 277–281. [[CrossRef](#)] [[PubMed](#)]
39. Koivula, R.; Pakarinen, J.; Sivenius, M.; Sirola, K.; Harjula, R.; Paatero, E. Use of hydrometallurgical wastewater as a precursor for the synthesis of cryptomelane-type manganese dioxide ion exchange material. *Sep. Purif. Technol.* **2009**, *70*, 53–57. [[CrossRef](#)]
40. Su, Q.; Pan, B.; Wan, S.; Zhang, W.; Lv, L. Use of hydrous manganese dioxide as a potential sorbent for selective removal of lead, cadmium, and zinc ions from water. *J. Colloid Interface Sci.* **2010**, *349*, 607–612. [[CrossRef](#)]
41. Zhou, J.; Wang, Y.; Wang, J.; Qiao, W.; Long, D.; Ling, L. Effective removal of hexavalent chromium from aqueous solutions by adsorption on mesoporous carbon microspheres. *J. Colloid Interface Sci.* **2016**, *462*, 200–207. [[CrossRef](#)]
42. Al-Qodah, Z.; Lafi, W.K.; Al-Anber, Z.; Al-Shannag, M.; Harahsheh, A. Adsorption of methylene blue by acid and heat treated diatomaceous silica. *Desalination* **2007**, *217*, 212–224. [[CrossRef](#)]
43. Ballav, N.; Choi, H.J.; Mishra, S.B.; Maity, A. Polypyrrole-coated halloysite nanotube clay nanocomposite: Synthesis, characterization and Cr(VI) adsorption behavior. *Appl. Clay Sci.* **2014**, *102*, 60–70. [[CrossRef](#)]
44. Kocaoba, S.; Orhan, Y.; Akyüz, T. Kinetics and equilibrium studies of heavy metal ions removal by use of natural zeolite. *Desalination* **2007**, *214*, 1–10. [[CrossRef](#)]
45. Yao, G.; Lei, J.; Zhang, W.; Yu, C.; Sun, Z.; Zheng, S.; Komarneni, S. Antimicrobial activity of X zeolite exchanged with Cu(2+) and Zn(2+) on Escherichia coli and Staphylococcus aureus. *Environ. Sci. Pollut. Res. Int.* **2019**, *26*, 2782–2793. [[CrossRef](#)] [[PubMed](#)]
46. Yang, J.; Xu, W.; He, C.; Huang, Y.; Zhang, Z.; Wang, Y.; Hu, L.; Xia, D.; Shu, D. One-step synthesis of silicon carbide foams supported hierarchical porous sludge-derived activated carbon as efficient odor gas adsorbent. *J. Hazard. Mater.* **2017**, *344*, 33–41. [[CrossRef](#)] [[PubMed](#)]