

NH₂OH Disproportionation Mediated by Anaerobic Ammonium-oxidizing (Anammox) Bacteria

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Anammox bacteria produce N_2 gas by oxidizing NH_4^+ with NO_2^- , and hydroxylamine (NH_2OH) is a potential intermediate of the anammox process. N_2 gas production occurs when anammox bacteria are incubated with NH_2OH only, indicating their capacity for NH_2OH disproportionation with NH_2OH serving as both the electron donor and acceptor. Limited information is currently available on NH_2OH disproportionation by anammox bacteria; therefore, the stoichiometry of anammox bacterial NH_2OH disproportionation was examined in the present study using ¹⁵N-tracing techniques. The anammox bacteria, *Brocadia sinica, Jettenia caeni*, and *Scalindua* sp. were incubated with the addition of ¹⁵NH₂OH, and the production of ¹⁵N-labeled nitrogenous compounds was assessed. The anammox bacteria tested performed NH_2OH disproportionation and produced ¹⁵⁻¹⁵N₂ gas and NH_4^+ as reaction products. The addition of acetylene, an inhibitor of the anammox process, reduced the activity of NH_2OH disproportionation, but not completely. The growth of *B. sinica* by NH_2OH disproportionation (-240.3 kJ mol NH_2OH^{-1} under standard conditions) was also tested in 3 up-flow column anammox reactors fed with 1) 0.7 mM NH_2OH only, 2) 0.7 mM NH_2OH and 0.5 mM NH_4^+ , and 3) 0.7 mM NH_2OH and 0.5 mM NO_2^- . NH_2OH consumption activities were markedly reduced after 7 d of operation, indicating that *B. sinica* was unable to maintain its activity or biomass by NH_2OH disproportionation.

Key words: anammox bacteria, hydroxylamine (NH₂OH), NH₂OH disproportionation, ¹⁵N-tracing technique, up-flow column reactor

Anaerobic ammonium-oxidizing (anammox) bacteria were initially discovered in a denitrifying bioreactor in 1995 (Mulder *et al.*, 1995), and are now recognized as the main players in the global nitrogen cycle (Kuypers *et al.*, 2003; 2005; Amano *et al.*, 2007; Yoshinaga *et al.*, 2011). In the anammox process, NH_4^+ is oxidized to N_2 gas using NO_2^- as an electron acceptor, and the stoichiometry of the anammox process has been described as follows (Lotti *et al.*, 2014):

1 NH₄⁺+1.146 NO₂⁻+0.071 HCO₃⁻+0.057 H⁺

 \rightarrow 0.986 N₂+0.161 NO₃⁻⁺0.071 CH_{1.74}O_{0.31}N_{0.20}+2.002 H₂O eq. 1 As an intermediate, hydrazine (N₂H₄) is synthesized from NH₄⁺ and NO or NH₂OH by hydrazine synthase, and its biosynthesis appears to be unique to metabolism by anammox bacteria (Strous *et al.*, 2006; Kartal *et al.*, 2011; Oshiki *et al.*, 2016a). Anammox bacteria are monophyletically affiliated into the bacterial order *Brocadiales* in the phylum *Planctomycetota* (Strous *et al.*, 1999), and five candidate genera have been proposed: *Candidatus Brocadia*, *Kuenenia*, *Jettenia*, *Anammoxoglobus*, and *Scalindua* (Egli *et al.*, 2003; Kuypers *et al.*, 2003; Quan *et al.*, 2008). These anammox bacteria have been detected in various natural and man-made ecosystems and are significantly involved in

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Hydroxylamine (NH₂OH), a well-known intermediate of the aerobic NH₃ oxidation reaction (Madigan *et al.*, 2019), is also a potential intermediate of the anammox process (van de Graaf et al., 1997). The anammox bacterium, Brocadia sinica reduced NO₂⁻ to NH₂OH (Oshiki et al., 2016a) potentially by using reductive hydroxylamine dehydrogenase (rHao) (Ferousi et al., 2021), and synthesized N₂H₄ from the NH₂OH and NH₄⁺ using hydrazine synthase (Oshiki et al., 2016a). Kuenenia stuttgartiensis reduced NO_2^- to NO (Kartal et al., 2011), and NO was further reduced to NH_2OH by hydrazine synthase to synthesize N_2H_4 (Dietl et al., 2015). Apart from the anammox process, the nitrogen transformation reaction, NH₂OH disproportionation, has been reported in anammox bacteria (van der Star et al., 2008). A disproportionation reaction is a reaction in which a chemical compound serves as both an electron donor and accepter, and the disproportionation reaction of inorganic sulfur (Finster, 2008) is an example of a microbial disproportionation reaction. In NH₂OH disproportionation, NH₂OH is converted to N₂ gas and NH₄⁺ using the following stoichiometry (Pacheco et al., 2011):

 $3 \text{ NH}_2\text{OH}+\text{H}^+ \rightarrow \text{NH}_4+\text{N}_2+3 \text{ H}_2\text{O}$ eq. 2

NH₂OH disproportionation involves the following 2 reactions, N_2H_4 production and consumption (Soler-Jofra *et al.*, 2020):

 $NH_4^+ + NH_2OH \rightarrow N_2H_4 + H_2O + H^+$ eq. 3

 $2 \text{ NH}_2\text{OH} + \text{N}_2\text{H}_4 + 2 \text{ H}^+ \rightarrow 2 \text{ NH}_4^+ + \text{N}_2 + 2 \text{ H}_2\text{O}$ eq. 4

Although NH₂OH disproportionation by K. stuttgartiensis (van der Star *et al.*, 2008; Soler-Jofra *et al.*, 2020) and B. sinica (Oshiki *et al.*, 2016a) has been described, further studies are required to obtain a more detailed understanding of anammox bacterial NH₂OH disproportionation for the

following reasons. Although the kinetics of NH2OH disproportionation have been investigated using K. stuttgartiensis (van der Star et al., 2008; Soler-Jofra et al., 2020), the amounts of N₂ gas produced in NH₂OH disproportionation (see eq. 2) were not measured in previous studies and the stoichiometry of NH₂OH disproportionation was not established. Furthermore, we previously examined NH₂OH disproportionation by B. sinica; however, we only reported the occurrence of NH2OH disproportionation (Oshiki et al., 2016a) and did not investigate the effects of NH₂OH concentrations on NH₂OH disproportionation by repeating batch incubations. In addition, although eq. 2 yields -240.3 kJ mol NH₂OH⁻¹ of free energy under standard conditions (Soler-Jofra et al., 2020), the growth of anammox bacteria with NH₂OH disproportionation has not yet been examined.

Therefore, the present study investigated anammox bacterial NH₂OH disproportionation. The phylogenetically different anammox bacteria, B. sinica, Jettenia caeni, and Scalindua sp. were incubated with ¹⁵N-labeled NH₂OH, and the stoichiometry of NH₂OH disproportionation was carefully assessed based on measurements of ¹⁵⁻¹⁵N₂ and NH₄⁺ concentrations. NH₂OH disproportionation was also analyzed under acetylene inhibition conditions. Acetylene is a strong inhibitor of aerobic NH₃ oxidation, N₂O reduction to N_2 (*i.e.*, denitrification), and the anammox process (Jensen et al., 2007); however, its effects on NH₂OH disproportionation currently remain unknown. B. sinica was cultured in up-flow column reactors with the addition of NH₂OH to establish whether it grows on NH₂OH, and the activity of NH₂OH consumption and the abundance of the anammox bacterial 16S rRNA gene were evaluated.

Materials and Methods

Anammox bacterial cultures

Planktonic cells of B. sinica, J. caeni, and Scalindua sp. were cultivated in membrane bioreactors (MBRs) equipped with a hollow fiber membrane module (pore size of 0.1 µm, polyethylene) as previously described (Oshiki et al., 2013; Zhang and Okabe, 2020). Culture media fed into MBRs contained KH_2PO_4 (24.4 mg L-1), MgSO₄·7H₂O (60 mg L-1), CaCl₂ (51 mg L-1), yeast extract (Becton, Dickinson and Company) (1.0 mg L-1), and 0.5 mL of trace element solutions I and II (van de Graaf et al., 1996). The artificial sea salt SEALIFE (Marine Tech) (Kindaichi et al., 2011) was supplemented into media for "Ca. Scalindua sp." at a final concentration of 28 g L⁻¹. Equimolar amounts of $NH_4(SO_4)_2$ and NaNO₂ were supplemented into media at 10 mM for B. sinica and Scalindua sp. and 5 mM for J. caeni, and the nitrogen loading rates of the MBRs for B. sinica, J. caeni, and Scalindua sp. were set at 0.55, 0.18, and 0.45 kg N m⁻³ d⁻¹, respectively. MBRs were operated at 37°C for B. sinica and at 25°C for J. caeni and Scalindua sp. pH was not controlled in MBRs, but was in the range of pH 7.6-8.0. Anammox bacterial cells accounted for more than 90% of the total biomass in MBRs as measured by a fluorescence insitu hybridization (FISH) analysis using the oligonucleotide probes AMX820 (Schmid et al., 2001) and EUBmix composed of equimolar EUB338, EUB338II, and EUB338III (Daims et al., 1999). Anammox bacterial species were routinely checked based on the partial anammox bacterial 16S rRNA gene sequence using Sanger sequencing (Oshiki et al., 2011).

Batch incubations of anammox bacteria

Standard anaerobic techniques were employed in an anaerobic chamber (Coy Laboratories Products) in which the concentration of oxygen was maintained at <1 ppm. Culture media and stock solutions were prepared by purging N_2 gas for >30 min, and then repeatedly vacuuming and purging He gas. The ¹⁵N enrichment of ¹⁵NH₂OH·HCl (Cambridge Isotope Laboratories) was >98%.

Anammox bacterial cells collected from MBRs were centrifuged at 13,420×g at 20°C for 10 min, washed, and then resuspended in the above culture media without NH_4^+ and NO_2^- at concentrations of 0.5 mg protein mL⁻¹. Twenty-five milliliters of the cell suspension was dispensed into 70-mL serum glass vials (Nichiden-Rika glass), and the headspace was replaced with He gas (>99.99995%) after sealing with butyl rubber stoppers and aluminium caps. Vials were incubated after the addition of ¹⁵NH₂OH (final concentration of 1.0 to 10 mM) and acetylene (30 µM) (Jensen et al., 2007) in the dark at 37°C for B. sinica and at 25°C for J. caeni and Scalindua sp.. Liquid samples were collected using a 1-mL plastic disposable syringe, immediately filtered using a 0.2-µm cellulose acetate filter, and subjected to measurements of NH₄⁺, NO₂⁻, NO₃⁻, and NH₂OH concentrations. Gas samples were collected using a gas-tight glass syringe and immediately injected into a gas chromatograph to assess $^{14-15}N_2$ and $^{15-15}N_2$ concentrations.

Up-flow column reactors fed with NH₂OH

Three 255-mL up-flow column reactors were operated at 37°C in the dark with the continuous feeding of the above culture media containing 1) 0.7 mM ¹⁴NH₂OH, 2) 0.7 mM ¹⁴NH₂OH and 0.5 mM $^{14}NH_4^+$, or 3) 0.7 mM $^{14}NH_2OH$ and 0.5 mM $^{14}NO_2^-$. B. sinica cells immobilized on polyvinyl alcohol (PVA)-sodium alginate (SA) (6 and 2% [w/v], respectively) beads were inoculated into the column reactors at a packing ratio of 50% (v/v). The gel immobilization of B. sinica cells in PVA-SA gel beads was performed as previously described (Ali et al., 2015). Briefly, the planktonic cells of B. sinica collected from the above MBR were resuspended in culture media, and mixed with an equal volume of PVA-alginate (FUJIFILM Wako) solution (12 and 4% [w/v], respectively). The gel solution was dropped using a disposable 50-mL plastic syringe (Terumo) into a 4% (w/v) CaCl₂ solution to form gel beads (diameter of ca. 2 mm). After an overnight incubation at 20°C, gel beads were washed with fresh inorganic medium. The gel beads obtained were inoculated into the up-flow column reactors.

Chemical analysis

 $\rm NH_4^+, \rm NO_2^-,$ and $\rm NO_3^-$ concentrations were measured using the ion chromatograph IC-2010 equipped with the TSKgel SuperIC-Anion HS or TSKgel SuperIC-Cation HS column (Tosoh). $\rm NH_2OH$ concentrations were measured colorimetrically (Frear and Burrell, 1955). Briefly, liquid samples were mixed with 0.48% (w/v) trichloroacetic acid, 0.2% (w/v) 8-hydroxyquinoline, and 0.2 M $\rm Na_2CO_3$, heated at 100°C for 1 min, and absorbance was then measured at a wavelength of 705 nm using the spectrophotometer V-630bio (Jasco). $\rm N_2H_4$ concentrations were measured colorimetrically using *p*-dimethyl-aminobenzaldehyde (Watt and Chrisp, 1952). Briefly, liquid samples were mixed with 0.12 M *p*-dimethyl-aminobenzaldehyde, and absorbance was measured at a wavelength of 460 nm. $\rm ^{14-15}N_2$ and $\rm ^{15-15}N_2$ concentrations were measured by gas chro-

¹⁴⁻¹⁵N₂ and ¹⁵⁻¹⁵N₂ concentrations were measured by gas chromatography mass spectrometry (GC/MS) (Isobe *et al.*, 2011a; 2011b). Fifty microliters of the headspace gas was collected using a 100-μL gas-tight glass syringe and immediately injected into the gas chromatograph GCMS-QP 2010 SE (Shimadzu) equipped with a fused silica capillary column (Agilent Technologies). Peaks at *m/z*=29 and 30 corresponding to ¹⁴⁻¹⁵N₂ and ¹⁵⁻¹⁵N₂ were monitored, and concentrations were calculated using a standard curve prepared using ¹⁵⁻¹⁵N₂ gas (Cambridge Isotope Laboratories). The ¹⁴⁻¹⁵N₂ and ¹⁵⁻¹⁵N₂ concentrations of ¹⁵NH₂OH were calculated by considering the ¹⁵N enrichment of ¹⁵NH₂OH·HCl (¹⁵N, 98%) and the natural abundance of $^{14\text{-}15}N_2$ and $^{15\text{-}15}N_2$ in atmospheric N_2 gas contaminated at the injection of the gas sample.

qPCR assay

The copy numbers of the anammox bacterial 16S rRNA gene were measured using a qPCR assay. Genomic DNA was extracted from gel beads collected from the up-flow column reactors using the FastDNA SPIN kit (Qiagen) according to the instruction manual supplied by the manufacturer. The qPCR assay was conducted using the ABI7500 fast Real-Time PCR System (Thermo Fisher Scientific) and Premix Ex Taq (Probe qPCR) (TakaraBio) under previously described thermal conditions (Zhang and Okabe, 2017a). The oligonucleotide primers and TaqMan probe used for the PCR amplification of the *B. sinica* 16S rRNA gene were BRS95F, BRS170R, and BRS130P. Standard curves (10¹ to 10⁶ copies μ L⁻¹) were prepared using a dilution series of plasmid DNAs containing a partial *B. sinica* 16S rRNA gene sequence.

Results

*NH*₂*OH disproportionation by phylogenetically different anammox bacteria*

Anammox bacterial cells of *B. sinica*, *J. caeni*, and *Scalindua sp.* were incubated with the addition of 1.0 to 2.5 mM ¹⁵NH₂OH. As shown in Fig. 1a, c, and e, ¹⁵NH₂OH consumption occurred in the culture concurrently with the production of ¹⁵⁻¹⁵N₂ gas and NH₄⁺. N₂H₄ was also produced and markedly increased when NH₂OH concentrations decreased below 0.1 mM (*i.e.*, a 60-min incubation for *B. sinica* and *Scalindua* sp. and a 120-min incubation for *J. caeni*). ¹⁴⁻¹⁵N₂, NO₂⁻, and NO₃⁻ were not detected during any incubations, and ¹⁵NH₂OH consumption was negligible in vials without anammox bacterial cells (*i.e.*, abiotic control incubation).

Batch incubations were repeated with the addition of 30 μ M acetylene. Anammox bacterial cells consumed ¹⁵NH₂OH (Fig. 1b, d, and f), whereas consumption rates were >3-fold lower than those without acetylene. Although ¹⁵⁻¹⁵N₂ gas and NH₄⁺ were produced during the incubation, N₂H₄ was not produced in any anammox bacterial cultures.

The above batch incubation of *B. sinica* was repeated with an increase in the initial NH₂OH concentration (from 5 to 10 mM) with/without 30 μ M acetylene. As shown in Fig. 2, ¹⁵NH₂OH consumption and the concomitant production of ¹⁵⁻¹⁵N₂ and NH₄⁺ occurred, similar to the batch incubation with the addition of 2.5 mM ¹⁵NH₂OH (*i.e.*, Fig. 1a and b). It is important to note that N₂H₄ production only occurred when NH₂OH concentrations decreased below 1.5 mM (Fig. 2a after 7.5 h of the incubation), and N₂H₄ production was not observed in Fig. 2c and e.

The stoichiometry and nitrogen mass balance of the above batch incubations are shown in Table 1. The theoretical values for $\Delta^{15\cdot15}N_2/\Delta^{15}NH_2OH$ and $\Delta^{15}NH_4^{+}/\Delta^{15}NH_2OH$ were 0.33 and 0.33, respectively when NH₂OH disproportionation occurred according to eq. 2. The values for $\Delta^{15\cdot15}N_2/\Delta^{15}NH_2OH$ and $\Delta^{15}NH_4^{+}/\Delta^{15}NH_2OH$ obtained from batch incubations with the addition of 2.5 or 5 mM ¹⁵NH₂OH were in the range of 0.19–0.45 and 0.17–0.36, respectively. These values were generally close to the theoretical values and those for $\Delta^{15}NH_4^{+}/\Delta^{15}NH_2OH$ in the *K. stuttgartiensis* culture (0.19–0.48), except for the following batch incu-

bations: *B. sinica* with 2.5 mM ¹⁵NH₂OH without acetylene (0.2 for Δ^{15} NH₄⁺/ Δ^{15} NH₂OH), *B. sinica* with 5 mM ¹⁵NH₂OH with acetylene (0.19 for $\Delta^{15-15}N_2/\Delta^{15}$ NH₂OH), and *Scalindua* sp. with 2.5 mM ¹⁵NH₂OH with and without acetylene (0.17 or 0.21 for Δ^{15} NH₄⁺/ Δ^{15} NH₂OH). Batch incubations of *B. sinica* with 7.5 or 10 mM NH₂OH generally resulted in lower values for $\Delta^{15-15}N_2/\Delta^{15}$ NH₂OH and Δ^{15} NH₄⁺/ Δ^{15} NH₂OH, which were in the ranges of 0.14–0.19 and 0.09–0.19, respectively.

Cultivation of B. sinica with NH_2OH in up-flow column reactors

NH₂OH consumption and the abundance of the anammox bacterial 16S rRNA gene were examined in the 1) NH₂OH-, 2) NH₂OH and NH₄⁺-, and 3) NH₂OH and NO₂⁻-feeding up-flow column reactors. In all operated reactors, NH₂OH consumption markedly decreased after 7 d of operation, and halted after 15 d of operation (Fig. 3). NH₄⁺ concentrations in the effluents of the NH₂OH- and NH₂OH and NO₂⁻-feeding reactors were in the same range (Fig. 3a and c, respectively), and the consumption of NH₄⁺ did not occur in the NH₂OH and NH₄⁺-feeding reactor (Fig. 3b). The copy numbers of the anammox bacterial 16S rRNA gene decreased over time, and its abundance after 14 d of operation was an order of magnitude less than that after 1 d of operation. Copy numbers were not measured in the reactor fed with NH₂OH and NO₂⁻ (Fig. 3c).

Discussion

NH₂OH disproportionation was examined using phylogenetically different anammox bacteria, and the present study clearly indicated that NH₂OH disproportionation is a common nitrogen transformation process of anammox bacteria (Table 1). Although previous studies (van der Star et al., 2008; Soler-Jofra et al., 2020) reported anammox bacterial NH₂OH disproportionation, they did not examine the amount of N₂ gas produced. The present study performed sophisticated ¹⁵NH₂OH-tracing batch incubations, and revealed that anammox bacteria yielded both N2 gas and NH₄⁺ from NH₂OH disproportionation. ¹⁵NH₂OH-tracing batch incubations enabled the stoichiometry of ¹⁵⁻¹⁵N₂ and ¹⁵NH₄⁺ production to ¹⁵NH₂OH consumption to be examined (Table 1), and the results obtained indicated that an increase in the initial concentration of NH2OH resulted in a decrease in $\Delta^{15-15}N_2/\Delta^{15}NH_2OH$ and $\Delta^{15}NH_4^+/\Delta^{15}NH_2OH$ for *B. sinica*. Furthermore, the addition of acetylene resulted in a decrease and increase in $\Delta^{15-15}N_2/\Delta^{15}NH_2OH$ and $\Delta^{15}NH_4/\Delta^{15}NH_2OH$, respectively. These variations in stoichiometry suggested that ¹⁵N₂ and/or ¹⁵NH₄⁺ were produced by multiple nitrogen transformation pathways other than NH₂OH disproportionation. Anammox bacterial hydroxylamine dehydrogenase oxidizes NH₂OH to NO (Maalcke et al., 2014), and N₂ gas may be produced using the NO formed by the anammox process (the coupling of NH₄⁺ and NO) (Kartal *et al.*, 2013). Although the reduction of NH₂OH to NH₄⁺ by anammox bacterial cells has not yet been demonstrated, anammox bacteria are capable of dissimilatory nitrite reduction to ammonium (Kartal et al., 2007).

N₂ and NH₄⁺-forming NH₂OH disproportionation has not

12

9

6

з

0

12

9

6

3

180

N₂H₄ (µmol [N vial]⁻¹)

240

N₂H₄ (µmol [N vial]⁻¹)



b) Brocadia sinica with acetylene

-↔ 60

d) Jettenia caeni with acetylene

60

f) Scalindua sp. with acetylene

-¹⁵NH₂OH

15-15N2

120

(min)

120

(min)

-15NH2OH

15-15N2

180

50

40

30

20

10

0

50

40

30

20

10

0

C 0

0

NH2OH, NH4⁺, ¹⁵⁻¹⁵N2 (µmol [N vial]⁻¹)



c) Jettenia caeni







Fig. 1. $^{15}NH_2OH$ -tracing batch incubations to demonstrate anammox bacterial NH₂OH disproportionation. Twenty-five milliliters of *Brocadia* sinica (panel a and b), *Jettenia caeni* (panel c and d), and *Scalindua* sp. (panel e and f) cultures were incubated with the addition of 2.5 mM $^{15}NH_2OH$ in 70-mL glass vials. Incubations were repeated with the addition of 30 μ M acetylene (panel b, d, and f). All incubations were performed in triplicate, and symbols and error bars represent mean values and the range of standard deviations, respectively. The standard deviations of data points are often within the symbols.

a) Brocadia sinica 5 mM NH₂OH



c) Brocadia sinica 7.5 mM NH₂OH









d) Brocadia sinica 7.5 mM NH2OH with acetylene



e) Brocadia sinica 10 mM NH2OH

f) Brocadia sinica 10 mM NH2OH with acetylene



Fig. 2. Effects of the initial NH₂OH concentration on NH₂OH disproportionation by Brocadia sinica. Twenty-five milliliters of the Brocadia sinica culture was incubated with the addition of 5 mM (panel a and b), 7.5 mM (panel c and d), and 10 mM 15 NH₂OH (panel e and f) in 70-mL glass vials. Incubations were repeated with the addition of 30 μ M acetylene (panel b, d, and f). All incubations were performed in triplicate, and symbols and error bars represent mean values and the range of standard deviations, respectively. The standard deviations of data points are often within the symbols.

Table 1. Stoichiometry of NH₂OH disproportionation mediated by anammox bacteria. The values obtained in the present study were mean values from triplicate biological replicates, and values were calculated from the initial and final concentrations during batch incubations. The N-mass balance (%) was calculated by dividing the total amounts of ¹⁵NH₂OH, ¹⁵NH₄⁺, and ¹⁵⁻¹⁵N₂ at the end of the incubation (µmol [N vial]⁻¹) by the initial amounts of ¹⁵NH₂OH (µmol [N vial]⁻¹).

| Species | NH ₂ OH (mM) | Acetylene* | N-mass balance | $\Delta^{15-15}N_2/\Delta NH_2OH$ (mol/mol) | $\Delta^{15}NH_4^+/\Delta NH_2OH$ (mol/mol) | References |
|--------------------|-------------------------|------------|----------------|---|---|---|
| B. sinica | 2.5 | w/o | 91% | 0.35 | 0.20 | This study |
| | 5 | w/o | 85% | 0.28 | 0.27 | |
| | 7.5 | w/o | 66% | 0.19 | 0.12 | |
| | 10 | w/o | 72% | 0.16 | 0.09 | |
| | 2 | + | 91% | 0.27 | 0.35 | |
| | 5 | + | 81% | 0.19 | 0.30 | |
| | 7.5 | + | 76% | 0.16 | 0.19 | |
| | 10 | + | 77% | 0.14 | 0.15 | |
| J. caeni | 1 | w/o | 83% | 0.26 | 0.25 | |
| | | + | 91% | 0.26 | 0.36 | |
| Scalindua sp. | 2 | w/o | 82% | 0.29 | 0.17 | |
| | 2.5 | + | 111% | 0.45 | 0.21 | |
| K. stuttgartiensis | 1.6–10 | w/o | NA | ND | 0.19-0.48 | van der Star <i>et al.</i> , 2008 Soler-Jofra <i>et al.</i> , 2020 |
| B. flugida | 4 | w/o | NA | ND | 0.25 | van der Star et al., 2008 |

*; w/o and +: without and with the addition of 30 µM acetylene, respectively, NA: not available, ND: not determined.



Fig. 3. NH₂OH consumption by *Brocadia sinica* in continuous up-flow column reactors. Up-flow column reactors were inoculated with PVA-SA gel beads immobilizing *B. sinica* cells and were operated with the continuous feeding of NH₂OH (panel a), NH₂OH and NH₄⁺ (panel b), and NH₂OH and NO₂⁻ (panel c).

yet been reported for microorganisms other than anammox bacteria. It requires the formation of a nitrogen-nitrogen bond for N_2 gas, and only nitric oxide reductase and hydrazine synthase catalyze this reaction (Dietl *et al.*, 2015). This may be the reason why N_2 and NH_4^+ -forming NH_2OH disproportionation has not yet been detected in microbial cultures other than anammox bacteria. Apart from N_2 and NH_4^+ -forming NH_2OH disproportionation, 1) NH_4^+ and NO_- , 2) NH_4^+ and NO_2^- -, and 3) N_2O and NH_4^+ -forming NH_2OH disproportionation are thermodynamically favorable (Pacheco *et al.*, 2011); however, limited information is currently available on their activities or involvement in the nitrogen transformation process in natural and man-made ecosystems.

The addition of acetylene did not induce the accumulation of N₂H₄, but reduced NH₂OH consumption (Fig. 1b, d, and f), which indicated that acetylene inhibited the enzymes involved in both NH2OH consumption and N2H4 production reactions. B. sinica hydrazine synthase utilizes NH₂OH as a substrate for N_2H_4 synthesis (Oshiki *et al.*, 2016a); therefore, hydrazine synthase may be the enzyme that is inhibited by acetylene. The inhibition of anammox bacterial activity by acetylene has been reported (Jensen et al., 2007); however, the underlying mechanism(s) remain unknown. Apart from hydrazine synthase, acetylene inhibited the NO₂ reduction reaction of anammox bacteria (Kartal et al., 2011; Oshiki et al., 2016a), indicating that acetylene suppresses multiple nitrogen transformation reactions of anammox bacteria. Although the acetylene inhibition of copper-containing metalloproteins, such as ammonia monooxygenase and nitrous oxide reductase (Amo and Nos, respectively) has been described (Ensign et al., 1993; Gilch et al., 2009), the involvement of copper-containing metalloproteins in anammox bacterial metabolism, particularly N₂H₄ production, remains unclear. The binding site of acetylene to anammox bacterial enzymes is of interest for obtaining a more detailed understanding of the acetylene inhibition of anammox bacteria. The addition of 30 µM acetylene did not completely inhibit the production of ${}^{15-15}N_2$ gas or NH_4^+ (Fig. 1 and 2). This result suggests the incomplete inhibition of anammox bacterial hydrazine synthase with the addition of 30 µM acetylene (Jensen et al., 2007) and/or the production of ${}^{15-15}N_2$ gas and NH_4^+ , but not through N_2H_4 . The corresponding mechanism has not yet been elucidated, and a further understanding of acetylene inhibition by anammox bacteria will provide novel insights.

Apart from the above acetylene inhibition, high NH₂OH concentrations (>0.1 mM NH₂OH in Fig 1a, c, and e, and >1.5 mM NH₂OH in Fig. 2a) did not induce the accumulation of N_2H_4 . This behavior cannot simply be explained by the inhibition of anammox bacterial hydrazine dehydrogenase involved in N₂H₄ oxidation to N₂. The N₂H₄ oxidation activities of purified anammox bacterial hydrazine dehydrogenase were inhibited in the presence of 2.4 to 7.9 µM NH₂OH (Shimamura et al., 2007; Maalcke et al., 2016), and this inhibition provides an explanation for the accumulation of N₂H₄ during batch incubations, but not for the lack of its accumulation at high NH₂OH concentrations. On the other hand, anammox bacterial N₂H₄ accumulation may be explained by the balance between N₂H₄ production and consumption reactions (eq. 3 and 4, respectively), as reported in previous studies (van der Star et al., 2008; Soler-Jofra et al., 2020). N₂H₄ production and consumption reactions require 1 and 2 moles of NH₂OH, respectively, and higher NH₂OH concentrations increase N₂H₄ consumption rates more than N₂H₄ production rates; therefore, N₂H₄ does not accumulate at high NH₂OH concentrations. The affinity constants of eq. 3 and 4 for NH₂OH need to be examined in more detail in order to clarify the N₂H₄ accumulation behavior of anammox bacterial cells.

The results of the up-flow column reactor experiments revealed that B. sinica did not proliferate with NH2OH disproportion (Fig. 3), and this is the first experimental evidence to show that an anammox culture cannot be maintained with NH₂OH as the sole energy source. A 1 log reduction in the B. sinica 16S rRNA gene copy number clearly indicated that B. sinica did not proliferate in the operated reactors. B. sinica cells may have been compromised during reactor operation due to the high toxicity and mutagenesis of NH₂OH, which resulted in a decrease in the B. sinica 16S rRNA gene copy number in PVA-SA gel beads. We previously reported a log reduction in the anammox bacterial 16S rRNA gene copy number in PVA-SA gel beads under unfavorable cultivation conditions (Zhang et al., 2017b). It is important to note that B. sinica cells preferentially performed NH₂OH disproportionation over anammox. Anammox using NO2- or NH2OH did not occur in the NH₂OH and NO₂⁻- or NH₂OH and NH₄⁺-feeding reactors (Fig. 3c and b, respectively), which indicated that the anammox activities of B. sinica were inhibited in these reactors by NH₂OH. Therefore, low NH₂OH concentrations need to be maintained in order to avoid the inhibition of anammox activities and achieve the stable performance of nitrogen removal. Although the inhibitory effects of NH₂OH on nitrifying bacteria have been investigated (Kindaichi et al., 2004; Soler-Jofra et al., 2021), the IC₅₀ concentration of NH₂OH for anammox bacteria has not yet been systematically examined. In the present study, the sudden deterioration of NH₂OH consumption occurred after 7 d of operation; however, the concentration of NH2OH in the influent was markedly lower than those in batch incubations in which the marked deterioration of NH2OH consumption did not occur (i.e., 0.7 and 2.5-10 mM, respectively). Therefore, it was not possible to approximate the IC₅₀ concentration of NH₂OH inhibition from short-term batch incubations, and further studies are warranted to assess the IC₅₀ concentration of the long-term inhibition of NH₂OH. NH₂OH may accumulate at the submicromolar range in natural aquatic environments (Fukumori et al., 2003; Bikbulatova et al., 2007) and at the submillimolar range in nitrifying cultures (Soler-Jofra et al., 2021 and references therein); however, the impact of NH₂OH on nitrogen transformation reactions remains unclear. Further studies are needed to examine NH₂OH transformation, including anammox bacterial NH₂OH disproportionation, in natural and man-made ecosystems.

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References

Ali, M., Oshiki, M., Rathnayake, L., Ishii, S., Satoh, H., and Okabe, S. (2015) Rapid and successful start-up of anammox process by immobilizing the minimal quantity of biomass in PVA-SA gel beads. *Water Res* **79**: 147–157.

- Amano, T., Yoshinaga, I., Okada, K., Yamagishi, T., Ueda, S., Obuchi, A., *et al.* (2007) Detection of anammox activity and diversity of anammox bacteria-related 16S rRNA genes in coastal marine sediment in Japan. *Microbes Environ* 22: 232–242.
- Bikbulatova, E.M., Stepanova, I.E., and Bikbulatov, E.S. (2007) Concentration and localization of hydroxylamine in the reservoirs and lakes in the territory of European Russia. *Water Resour* 34: 554–562.
- Daims, H., Brühl, A., Amann, R., Schleifer, K.H., and Wagner, M. (1999) The domain-specific probe EUB338 is insufficient for the detection of all bacteria. Development and evaluation of a more comprehensive probe set. Syst Appl Microbiol 22: 434–444.
- Dietl, A., Ferousi, C., Maalcke, W.J., Menzel, A., Vries, S. de, Keltjens, J.T., et al. (2015) The inner workings of the hydrazine synthase multiprotein complex. *Nature* 527: 394–397.
- Egli, K., Bosshard, F., Werlen, C., Lais, P., Siegrist, H., Zehnder, A.J.B., and van der Meer, J.R. (2003) Microbial composition and structure of a rotating biological contactor biofilm treating ammonium-rich wastewater without organic carbon. *Microb Ecol* **45**: 419–432.
- Ensign, S.A., Hyman, M.R., and Arp, D.J. (1993) In vitro activation of ammonia monooxygenase from *Nitrosomonas europaea* by copper. J Bacteriol 175: 1971–1980.
- Ferousi, C., Schmitz, R.A., Maalcke, W.J., Lindhoud, S., Versantvoort, W., Jetten, M.S.M., *et al.* (2021) Characterization of a nitritereducing octaheme hydroxylamine oxidoreductase that lacks the tyrosine cross-link. *J Biol Chem* 296: 100476.
- Finster, K. (2008) Microbiological disproportionation of inorganic sulfur compounds. *J Sulfur Chem* **29**: 281–292.
- Frear, D.S., and Burrell, R.C. (1955) Spectrophotometric method for determining hydroxylamine reductase activity in higher plants. *Anal Chem* 27: 1664–1665.
- Fukumori, R., Senga, Y., Okumura, M., Fujinaga, K., and Seike, Y. (2003) Pretreatment and pre concentration methods for hydroxyl amine in environmental water samples using solid-phase extraction, followed by a spectrophotometric determination. *Bunseki Kagaku* 52: 747–753 (in Japanese with an English abstract).
- Gilch, S., Vogel, M., Lorenz, M.W., Meyer, O., and Schmidt, I. (2009) Interaction of the mechanism-based inactivator acetylene with ammonia monooxygenase of *Nitrosomonas europaea*. *Microbiology* 155: 279–284.
- Isobe, K., Koba, K., Ueda, S., Senoo, K., Harayama, S., and Suwa, Y. (2011a) A simple and rapid GC/MS method for the simultaneous determination of gaseous metabolites. *J Microbiol Methods* 84: 46–51.
- Isobe, K., Suwa, Y., Ikutani, J., Kuroiwa, M., Makita, T., Takebayashi, Y., et al. (2011b) Analytical techniques for quantifying ¹⁵N/¹⁴N of nitrate, nitrite, total dissolved nitrogen and ammonium in environmental samples using a gas chromatograph equipped with a quadrupole mass spectrometer. *Microbes Environ* 26: 46–53.
- Jensen, M.M., Thamdrup, B., and Dalsgaard, T. (2007) Effects of specific inhibitors on anammox and denitrification in marine sediments. *Appl Environ Microbiol* 73: 3151–3158.
- Kartal, B., Kuypers, M.M.M., Lavik, G., Schalk, J., Op den Camp, H.J.M., Jetten, M.S.M., and Strous, M. (2007) Anammox bacteria disguised as denitrifiers: nitrate reduction to dinitrogen gas via nitrite and ammonium. *Environ Microbiol* 9: 635–642.
- Kartal, B., Maalcke, W.J., de Almeida, N.M., Cirpus, I., Gloerich, J., Geerts, W., et al. (2011) Molecular mechanism of anaerobic ammonium oxidation. *Nature* 479: 127–130.
- Kartal, B., de Almeida, N.M., Maalcke, W.J., Op den Camp, H.J.M., Jetten, M.S.M., and Keltjens, J.T. (2013) How to make a living from anaerobic ammonium oxidation. *FEMS Microbiol Rev* 37: 428–461.
- Kindaichi, T., Okabe, S., Satoh, H., and Watanabe, Y. (2004) Effects of hydroxylamine on microbial community structure and function of autotrophic nitrifying biofilms determined by in situ hybridization and the use of microelectrodes. *Water Sci Technol* **49**: 61–68.
- Kindaichi, T., Awata, T., Suzuki, Y., Tanabe, K., Hatamoto, M., Ozaki, N., and Ohashi, A. (2011) Enrichment using an up-flow column reactor and community structure of marine anammox bacteria from coastal sediment. *Microbes Environ* 26: 67–73.
- Kuypers, M.M.M., Sliekers, A.O., Lavik, G., Schmid, M., Jørgensen, B.B., Kuenen, J.G., *et al.* (2003) Anaerobic ammonium oxidation by anammox bacteria in the Black Sea. *Nature* **422**: 608–611.

- Kuypers, M.M.M., Lavik, G., Woebken, D., Schmid, M., Fuchs, B.M., Amann, R., et al. (2005) Massive nitrogen loss from the Benguela upwelling system through anaerobic ammonium oxidation. Proc Natl Acad Sci U S A 102: 6478–6483.
- Lotti, T., Kleerebezem, R., Lubello, C., and van Loosdrecht, M.C.M. (2014) Physiological and kinetic characterization of a suspended cell anammox culture. *Water Res* 60: 1–14.
- Maalcke, W.J., Dietl, A., Marritt, S.J., Butt, J.N., Jetten, M.S.M., Keltjens, J.T., *et al.* (2014) Structural basis of biological NO generation by octaheme oxidoreductases. *J Biol Chem* 289: 1228–1242.
- Maalcke, W.J., Reimann, J., de Vries, S., Butt, J.N., Dietl, A., Kip, N., et al. (2016) Characterization of anammox hydrazine dehydrogenase, a key N₂-producing enzyme in the global nitrogen cycle. J Biol Chem 291: 17077–17092.
- Madigan, M.T., Bender, K.S., Buckley, D.H., Sattley, W.M., and Stahl, D.A. (2019) Brock Biology of Microorganisms, 15th edn. London: Pearson Education.
- Mulder, A., van de Graaf, A.A., Robertson, L.A., and Kuenen, J.G. (1995) Anaerobic ammonium oxidation discovered in a denitrifying fluidized bed reactor. *FEMS Microbiol Ecol* 16: 177–183.
- Oshiki, M., Shimokawa, M., Fujii, N., Satoh, H., and Okabe, S. (2011) Physiological characteristics of the anaerobic ammoniumoxidizing bacterium "*Candidatus* Brocadia sinica." *Microbiology* 157: 1706–1713.
- Oshiki, M., Awata, T., Kindaichi, T., Satoh, H., and Okabe, S. (2013) Cultivation of planktonic anaerobic ammonium oxidation (anammox) bacteria by using membrane bioreactor. *Microbes Environ* **28**: 436–443.
- Oshiki, M., Ali, M., Shinyako-Hata, K., Satoh, H., and Okabe, S. (2016a) Hydroxylamine-dependent anaerobic ammonium oxidation (anammox) by "*Candidatus* Brocadia sinica." *Environ Microbiol* **18**: 3133–3143.
- Oshiki, M., Satoh, H., and Okabe, S. (2016b) Ecology and physiology of anaerobic ammonium oxidizing (anammox) bacteria. *Environ Microbiol* **18**: 2784–2796.
- Pacheco, A.A., McGarry, J., Kostera, J., and Corona, A. (2011) Techniques for investigating hydroxylamine disproportionation by hydroxylamine oxidoreductases. *Methods Enzymol* 486: 447–463.
- Quan, Z.X., Rhee, S.K., Zuo, J.E., Yang, Y., Bae, J.W., Park, J.R., et al. (2008) Diversity of ammonium-oxidizing bacteria in a granular sludge anaerobic ammonium-oxidizing (anammox) reactor. *Environ Microbiol* 10: 3130–3139.
- Schmid, M., Schmitz-Esser, S., Jetten, M., and Wagner, M. (2001) 16S-23S rDNA intergenic spacer and 23S rDNA of anaerobic ammonium-oxidizing bacteria: Implications for phylogeny and in situ detection. *Environ Microbiol* 3: 450–459.
- Shimamura, M., Nishiyama, T., Shigetomo, H., Toyomoto, T., Kawahara, Y., Furukawa, K., and Fujii, T. (2007) Isolation of a multiheme protein with features of a hydrazine-oxidizing enzyme from an anaerobic ammonium-oxidizing enrichment culture. *Appl Environ Microbiol* 73: 1065–1072.
- Soler-Jofra, A., Laurenia, M., Warmerdama, M., Pérezb, J., and van Loosdrecht, M.C.M. (2020) Hydroxylamine metabolism of *Ca*. Kuenenia stuttgartiensis. *Water Res* 184: 116188.
- Soler-Jofra, A., Pérez, J., and van Loosdrecht, M.C.M. (2021) Hydroxylamine and the nitrogen cycle: A review. *Water Res* 190: 116723.
- Strous, M., Fuerst, J., Kramer, E., Logemann, S., Muyzer, G., van de Pas-Schoonen, K., *et al.* (1999) Missing lithotroph identified as new planctomycete. *Nature* **400**: 446–449.
- Strous, M., Pelletier, E., Mangenot, S., Rattei, T., Lehner, A., Taylor, M., et al. (2006) Deciphering the evolution and metabolism of an anammox bacterium from a community genome. *Nature* 440: 790–794.
- van de Graaf, A.A., de Bruijn, P., Robertson, L.A., Jetten, M.S.M., and Kuenen, J.G. (1996) Autotrophic growth of anaerobic ammoniumoxidizing micro-organisms in a fluidized bed reactor. *Microbiology* 142: 2187–2196.
- van de Graaf, A.A., de Bruijn, P., Robertson, L.A., Jetten, M.S.M., and Kuenen, J.G. (1997) Metabolic pathway of anaerobic ammonium oxidation on the basis of ¹⁵N studies in a fluidized bed reactor. *Microbiology* 143: 2415–2421.

- van der Star, W.R.L., van de Graaf, M.J., Kartal, B., Picioreanu, C., Jetten, M.S.M., and van Loosdrecht, M.C.M. (2008) Response of anaerobic ammonium-oxidizing bacteria to hydroxylamine. *Appl Environ Microbiol* 74: 4417–4426.
- Watt, G.W., and Chrisp, J.D. (1952) A spectrophotometric method for the determination of hydrazine. Anal Chem 24: 2006–2008.
- Yoshinaga, I., Amano, T., Yamagishi, T., Okada, K., Ueda, S., Sako, Y., and Suwa, Y. (2011) Distribution and diversity of anaerobic ammonium oxidation (anammox) bacteria in the sediment of a eutrophic freshwater lake, lake Kitaura, Japan. *Microbes Environ* 26: 189–197.
- Zhang, L., and Okabe, S. (2017a) Rapid cultivation of free-living planktonic anammox cells. *Water Res* **127**: 204–210.
- Zhang, L., Narita, Y., Gao, L., Ali, M., Oshiki, M., and Okabe, S. (2017b) Maximum specific growth rate of anammox bacteria revisited. *Water Res* 116: 296–303.
- Zhang, L., and Okabe, S. (2020) Ecological niche differentiation among anammox bacteria. *Water Res* 171: 115468.