

Biomass Yield Efficiency of the Marine Anammox Bacterium, “*Candidatus Scalindua* sp.,” is Affected by Salinity

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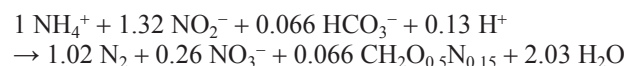
The growth rate and biomass yield efficiency of anaerobic ammonium oxidation (anammox) bacteria are markedly lower than those of most other autotrophic bacteria. Among the anammox bacterial genera, the growth rate and biomass yield of the marine anammox bacterium “*Candidatus Scalindua* sp.” is still lower than those of other anammox bacteria enriched from freshwater environments. The activity and growth of marine anammox bacteria are generally considered to be affected by the presence of salinity and organic compounds. Therefore, in the present study, the effects of salinity and volatile fatty acids (VFAs) on the anammox activity, inorganic carbon uptake, and biomass yield efficiency of “*Ca. Scalindua* sp.” enriched from the marine sediments of Hiroshima Bay, Japan, were investigated in batch experiments. Differences in VFA concentrations (0–10 mM) were observed under varying salinities (0.5%–4%). Anammox activity was high at 0.5%–3.5% salinity, but was 30% lower at 4% salinity. In addition, carbon uptake was higher at 1.5%–3.5% salinity. The results of the present study clearly demonstrated that the biomass yield efficiency of the marine anammox bacterium “*Ca. Scalindua* sp.” was significantly affected by salinity. On the other hand, the presence of VFAs up to 10 mM did not affect anammox activity, carbon uptake, or biomass yield efficiency.

Key words: anaerobic ammonium oxidation (anammox), “*Candidatus Scalindua*”, biomass yield efficiency, salinity, volatile fatty acids

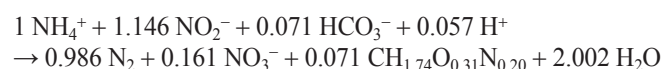
Anaerobic ammonium oxidation (anammox) is a microbiological process in which ammonium is oxidized to dinitrogen gas under anoxic conditions with nitrite as the electron acceptor (20, 31, 34). This process is mediated by anammox bacteria belonging to the phylum *Planctomycetes* (31). Previous studies demonstrated the ubiquitous distribution of anammox bacteria in artificial and natural ecosystems such as wastewater treatment plants, freshwater and marine sediments, and soils (2, 7, 8, 19, 24, 28, 29, 39). A minimum of five genera of candidate anammox bacteria have been tentatively proposed in *Planctomycetes*; *i.e.*, “*Candidatus Brocadia*,” “*Ca. Kuenenia*,” “*Ca. Scalindua*,” “*Ca. Anammoxoglobus*,” and “*Ca. Jettenia*” (9). “*Ca. Scalindua*” is primarily found in marine environments (16, 37), and 16S rRNA gene sequences have revealed that this genus contains taxonomically diverse members, even though only a few have been successfully grown in enrichment cultures to date (12, 14, 21, 36). Detailed, but partial information regarding some of the physiological characteristics of “*Ca. Brocadia*,” “*Ca. Kuenenia*,” and “*Ca. Scalindua*” are currently available (4, 6, 25, 26, 32), even though they have not yet been isolated in pure cultures.

Anammox bacteria are slow growing bacteria even under optimal growth conditions in laboratories with reported doubling times of 11 d for “*Ca. Brocadia anammoxidans*” (30), 8.3–11 d for “*Ca. Kuenenia stuttgartiensis*” (35), 7 d for “*Ca. Brocadia sinica*” (25), 14.4 d for “*Ca. Scalindua* sp.” (4), 3.6–5.4 d for anammox bacteria enriched from activated sludge and estimated by quantitative PCR (33), and 6–10 d for anammox bacteria enriched from a freshwater lake sediment and estimated by ²⁹N₂ measurements (38). These low

growth rates have been supported by the following stoichiometry of the anammox process calculated by considering inorganic carbon fixation to cells (anabolism), as originally reported by Strous *et al.* (30):



Using a membrane bioreactor, Lotti *et al.* (17) recently reported the following recalculated stoichiometry under no mass transfer limitations and no growth of non-anammox bacteria:



Among the anammox genera, the growth rate of the marine genera “*Ca. Scalindua*” is markedly lower than that of the freshwater genera. This may be attributed to a lower inorganic carbon fixation (uptake) rate, leading to a lower biomass yield efficiency because the maximum specific ammonium consumption rate (65 μmol NH₄⁺ [g protein]⁻¹ min⁻¹) of “*Ca. Scalindua* sp.” (4) is similar to that of freshwater anammox bacteria (25, 30, 31). In addition, the biomass yield efficiency of “*Ca. Scalindua* sp.” was determined to be 0.030 at 28°C and 3.5% salinity (4). Therefore, the differences observed in growth characteristics between freshwater and marine anammox bacteria raise questions regarding the influence of environmental factors.

We hypothesized that differences in the environmental concentrations of salt and certain organic compounds may affect the inorganic carbon uptake, biomass yield efficiency, and anammox activity of marine anammox bacteria. Therefore,

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the present study was conducted to determine the anammox activity and inorganic carbon uptake of the marine anammox bacterium, “*Ca. Scalindua* sp.” using batch experiments under different salinity and volatile fatty acid (VFA) concentrations. VFAs are also oxidized by both marine and freshwater anammox bacteria with nitrate as the electron acceptor (10, 11, 36). This metabolic variety may relate to a survival strategy of anammox bacteria in nutrient-limiting environments.

Materials and Methods

Biomass samples

Biomass samples were obtained from an upflow column reactor inoculated with an anammox biofilm, as previously described (14). The inoculant anammox biofilm contained two phylogenetically distinct “*Ca. Scalindua* sp.” operated at 20°C (14); therefore, the reactor was operated at 28°C to dominantly proliferate one species, as previously described (3, 4). The reactor volume was 920 mL and the hydraulic retention time was set to 4 h. The following slightly modified version of the synthetic marine nutrient medium reported by Kindaichi *et al.* (14) was fed into the reactor: 28 g L⁻¹ of an artificial sea salt, SEALIFE (Marine Tech, Tokyo, Japan), 70 mg-N L⁻¹ of (NH₄)₂SO₄, 70 mg-N L⁻¹ of NaNO₂, 500 mg L⁻¹ of KHCO₃, 27 mg L⁻¹ of KH₂PO₄, 300 mg L⁻¹ of MgSO₄·7H₂O, 180 mg L⁻¹ of CaCl₂·2H₂O, and 1 mL of trace element solutions I and II, as described by van de Graaf *et al.* (34). The medium was flushed with N₂ gas for 1 h to maintain an anaerobic condition (<0.5 mg L⁻¹ of dissolved oxygen) before the addition of nutrients.

Phylogenetic analysis

Total DNA was extracted from the upflow column reactor using the Fast DNA spin kit for soil (MP Biomedicals, Irvine, CA, USA) according to the manufacturer’s instructions. To construct the clone library, 16S rRNA gene fragments were amplified using the *Planctomycetales*-specific primer set of Pla46f (22) and 1390r (40). The PCR conditions were as follows: 4 min of initial denaturation at 94°C, followed by 30 cycles of 45 s at 94°C, 50 s at 58°C, and 3 min at 72°C. The final extension was performed for 10 min at 72°C. PCR products were confirmed using a 1% (w/v) agarose gel and were purified using the QIAquick PCR Purification Kit (Qiagen, Hilden, Germany). The purified PCR products were ligated into a pCR-XL-TOPO vector and transformed into One Shot *Escherichia coli* cells following the manufacturer’s instructions (TOPO XL PCR cloning kit; Invitrogen, Carlsbad, CA, USA). Cloned 16S rRNA genes were randomly selected and the clone library was constructed. The 16S rRNA genes were sequenced at the Dragon Genomics Center (Takara Bio, Otsu, Japan). Sequences with ≥97% homology were grouped into operational taxonomic units (OTUs) using the ARB neighbor-joining (NJ) method with similarity corrections performed using the ARB software (18). A phylogenetic tree was

constructed using the ARB NJ (Felsenstein correction), maximum parsimony (Phylip DNAPARS), and maximum likelihood (RAxML) methods, which were implemented by the ARB software with a database SSU Ref NR 111 dataset (27). A bootstrap resampling analysis for 1,000 replicates was conducted to estimate the confidence of the tree topologies.

Fluorescence in situ hybridization

Biomass samples were collected from the upflow column reactor and fixed in 4% paraformaldehyde. FISH was subsequently conducted according to the procedure described by Okabe *et al.* (23). The 16S rRNA-targeted oligonucleotide probes used in this study were EUBmix, composed of EUB338 (1), EUB338II, and EUB338III (5), which are specific for most bacteria; Amx368 (29), which is specific for all anammox bacteria; and Sca1129a, specific for uncultured bacterium husup-a2, and Sca1129b, specific for the uncultured bacterium husup-a7, which distinguish two phylogenetically different “*Ca. Scalindua* sp.” in the inoculum used in this study (14). The probes were labeled with Cy3 or Alexa Flour488 at the 5' end. A LSM5 PASCAL model confocal laser-scanning microscope, equipped with an Ar ion laser (488 nm) and HeNe laser (543 nm; Carl Zeiss, Oberkochen, Germany), was used for microscopy. The average surface area fraction was determined from at least 24 representative laser scanning microscopy projection images using the LSM5 PASCAL software provided by Carl Zeiss (13).

Batch experiments

Batch experiments were performed to investigate the influence of salinity (0.5%–4%) and VFAs (0–10 mM) on the anammox activity, inorganic carbon uptake, biomass yield efficiency, and VFA oxidation rates of “*Ca. Scalindua* sp.” as summarized in Table 1. It should be noted that the batch experiment under 0% salinity conditions was not performed because of the lack of anammox activity observed in a previous study (4). The biomass from the upflow column reactor was washed twice in the synthetic marine nutrient medium without ammonium and nitrite and was then suspended in the synthetic marine nutrient medium containing 7 mM ammonium and nitrite at the volatile suspended solids (VSS) concentration of 1 mg VSS mL⁻¹. A biomass suspension (4 mL) was dispensed into 7-mL serum vials and sealed with butyl rubber stoppers. The headspace was replaced by repeatedly vacuuming and purging with helium gas (>99.99995%). The vials were then statically incubated at 28°C for 1 d to determine the ammonium oxidation rate and inorganic carbon uptake and for 6 h to determine the VFA oxidation rate as described by Kartal *et al.* (10). All batch experiments were conducted in triplicate. Formate, acetate, and propionate were selected as organic compounds that can be oxidized by anammox bacteria as electron donors with nitrate as the electron acceptor, as previously reported (36). VFAs were added in the salt form (*i.e.*, sodium formate, sodium acetate, and sodium propionate) to prevent pH decreases by the addition of fatty acids. The pH was adjusted to 7.5 using 1 M NaOH or 1 M HCl after the addition of VFAs. [¹⁴C]

Table 1. Batch experimental conditions

Experiment*	NH ₄ ⁺ (mM)	NO ₂ ⁻ (mM)	NO ₃ ⁻ (mM)	Salinity (%)	VFA (mM)	[¹⁴ C]bicarbonate (μCi vial ⁻¹)	Incubation time (h)
Salinity	7	7	0	0.5–4		0	24
	7	7	0	0.5–4		20	24
VFA	7	7	0	2.8	Formate, 0–10	0	24
	7	7	0	2.8	Formate, 0–10	20	24
	7	7	0	2.8	Acetate, 0–10	0	24
	7	7	0	2.8	Acetate, 0–10	20	24
	7	7	0	2.8	Propionate, 0–10	0	24
	7	7	0	2.8	Propionate, 0–10	20	24
VFA oxidation rate	7	7	4	2.8	Formate, 3	0	6
	7	7	4	2.8	Acetate, 3	0	6
	7	7	4	2.8	Propionate, 3	0	6

* All experiments were conducted in triplicate at 28°C with 1 mg-VSS mL⁻¹.

bicarbonate was added to vials at a final concentration of 20 μCi (740 kBq vial⁻¹). Anammox activity and inorganic carbon uptake were expressed as a specific ammonium consumption and [¹⁴C] bicarbonate uptake rate, respectively. The biomass yield efficiency was calculated by dividing the specific [¹⁴C]bicarbonate uptake rate by the specific ammonium consumption rate.

Analytical methods

Ammonium concentrations were determined using Nessler's method with a UV-visible spectrophotometer (DR-2800; Hach, Loveland, CO, USA). Nitrite and nitrate concentrations were determined using ion chromatography (HPLC 20A; Shimadzu, Kyoto, Japan) with a Shodex Asahipak NH2P-50 4D anion column (Showa Denko, Tokyo, Japan) and UV-VIS detector (SPD-20A; Shimadzu) following the filtration of samples through 0.2- μm pore size membranes (Advantec, Tokyo, Japan), as previously described (14). The concentration of VFAs was determined using HPLC (HPLC 20A; Shimadzu) with a Shim-packSCR-102H column (Shimadzu) and conductivity detector (CDD-10A vp; Shimadzu) following the filtration of samples through 0.2- μm pore size membranes (Advantec). [¹⁴C]bicarbonate uptake was confirmed by liquid scintillation counting. The biomass was collected, washed three times with phosphate buffered saline, and mixed with the scintillation cocktail (Clear-sol I; Nacalai Tesque, Kyoto, Japan). Radioactivity was subsequently determined using an LSC-5100 liquid scintillation counter (Hitachi-Aloka Medical, Tokyo, Japan).

Statistical analysis

The average nitrogen stoichiometric ratios for consumed NO_2^- and consumed NH_4^+ ($\Delta\text{NO}_2^-/\Delta\text{NH}_4^+$) and produced NO_3^- and consumed NH_4^+ ($\Delta\text{NO}_3^-/\Delta\text{NH}_4^+$) in batch experiments for VFAs were compared using the unpaired Welch's *t*-test in Microsoft Excel. A *P* value of less than 0.05 was considered significant.

Nucleotide sequence accession number

The sequence data of the partial 16S rRNA gene obtained from the upflow column reactor was deposited in the GenBank/EMBL/DDBJ databases under accession number AB900163.

Results and Discussion

Dominant anammox species

To determine the dominant anammox species in the upflow column reactor, which had ammonium and nitrite removal efficiencies of 82.6% and 98.0%, respectively, a phylogenetic analysis on the basis of the 16S rRNA sequence was performed (Fig. 1). A total of 47 clones were randomly selected from the clone library and grouped into an OTU (AMX_B02) on the basis of $\geq 97\%$ sequence identity. The sequence similarity between AMX_B02 and "*Ca. Scalindua* sp. SH" and the clone husup-a7 was 99.9%, indicating that a single species of "*Ca. Scalindua* sp." was successfully enriched at 28°C, even though the inoculum contained two distinct species of anammox bacteria belonging to the "*Ca. Scalindua*" group (93.1% similarity to the clone husup-a2) (14). The dominance of a single species was further confirmed by a FISH analysis, which revealed that most anammox cells hybridized with the Amx368 (29) and Sca1129b (14) probes, whereas no cells hybridized with the Sca1129a probe (14). These results supported the findings of a previous study in which Sca1129b probe-hybridized cells dominated a reactor operated at 30°C (3). Anammox bacteria that were defined by Amx368 accounted for 73.5% \pm 7.4% of all bacteria identified by the EUBmix probes. In the inoculum, Amx368 probe- and

Sca1129a and Sca1129b probe-positive cells accounted for 60.8%, 17.3%, and 39.9% of all bacteria, respectively (14). The higher abundance of anammox bacteria in the present study may be related to the optimum growth temperature of "*Ca. Scalindua* sp." because they have been shown to have higher anammox activities at 25°C–28°C than that at 30°C (4).

Effects of salinity

The effects of 0.5%–4% salinity on the anammox activity of and carbon uptake by anammox bacteria were investigated in batch experiments. After the 1-d incubation, the simultaneous consumption of NH_4^+ and NO_2^- was observed under all conditions. Higher anammox activities were observed at 0.5%–3.5% salinity, whereas anammox activity was lower at 4% salinity (approximately 30% of the highest activity at 1.5% salinity; Fig. 2A). Higher carbon uptake activity was also observed at 1.5%–3.5% salinity (Fig. 2B). The highest activity and carbon uptake values occurred at different salinity concentrations, *i.e.*, the highest activity was at 1.5%, but

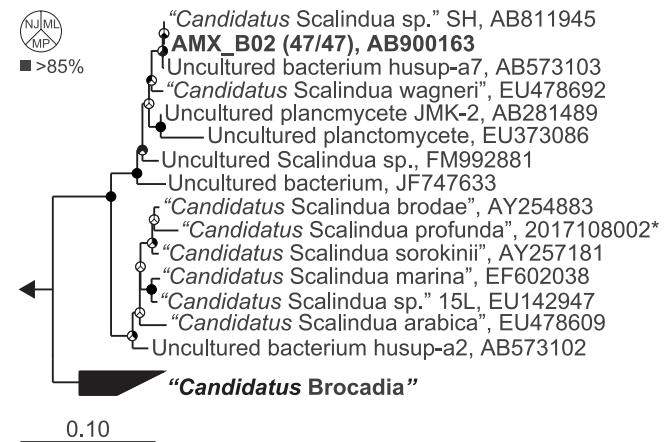


Fig. 1. Maximum-likelihood (ML) tree of the "*Ca. Scalindua*" genus based on 16S rRNA gene sequences.

The GenBank/EMBL/DDBJ accession numbers are indicated. The numbers in parentheses indicate the frequencies of identical clones analyzed. The scale bar represents the number of nucleotide changes per sequence position. Pie charts at the nodes represent the confidence of branch topology, and bootstrap values (1,000 replicates) greater than 85% are shaded black. The pie charts depict the neighbor-joining (NJ) method in the upper-left sector, ML method in the upper-right sector, and maximum parsimony (MP) method in the bottom sector. The asterisk represents the taxon ID in the Integrated Microbial Genomes with Microbiome Samples (IGM). The *Planctomyces brasiliensis* sequence (CP002546) served as the outgroup to root the tree.

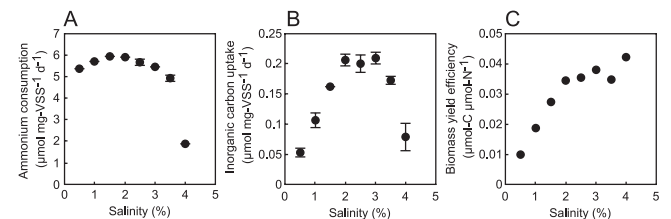


Fig. 2. Effects of salinity on the specific ammonium consumption rate (A) and specific inorganic carbon uptake rate (B). The values in (A) and (B) are the mean standard deviations of the results of independent triplicate experiments. Biomass yield efficiency (C) is calculated by dividing the average specific inorganic carbon uptake rate by the average specific ammonium consumption rate.

the highest carbon uptake was at 3% salinity, suggesting that the optimal conditions of “*Ca. Scalindua* sp.” differed with regard to anammox activity (catabolism) and carbon uptake (anabolism). Biomass yield efficiency was calculated to be 0.010–0.042 $\mu\text{mol-C } \mu\text{mol-N}^{-1}$ under 0.5%–4.0% salinity conditions (Fig. 2C). The maximum growth rate at 3% salinity was calculated to be 0.00021 h^{-1} ; however, this value was markedly lower than the previous reported value (0.0020 h^{-1}) (4) due to the batch incubation. It should be noted that the maximum value of 0.042 $\mu\text{mol-C } \mu\text{mol-N}^{-1}$ at 4% salinity was not appropriate because anammox activity and carbon uptake were both low, and, thus, apparent biomass yield efficiency was high. Therefore, the average biomass yield efficiencies of 0.036 (in the range of 0.035–0.038) at 2–3% salinity, at which carbon uptake was high, were similar to previous findings: 0.07 for “*Ca. Brocadia anammoxidans*” (32), 0.063 for “*Ca. Brocadia sinica*” (25), and 0.030 for “*Ca. Scalindua* sp. SH” (4). The wide range of biomass yield efficiencies (in the range of 0.010–0.038) under $\leq 3.5\%$ salinity conditions determined in this study indicated that the biomass yield efficiency of “*Ca. Scalindua* sp.” was strongly influenced by salinity. Under low salinity conditions ($\leq 1\%$), biomass yield efficiencies (range, 0.010–0.019) were markedly lower than the previously reported value of 0.030 at 28°C and 3.5% salinity (4). In addition, a previous study showed that the biomass yield efficiency of “*Ca. Scalindua* sp. SH” was not affected by temperature differences (28°C and 37°C) (4). These results indicated that an appropriate salinity range (2%–3%) may be a key environmental factor for the growth of “*Ca. Scalindua* sp.” in estuary environments, which was originally collected as the inoculum for an anammox enrichment culture (14, 15). The biomass yield efficiencies of freshwater anammox bacteria also need to be estimated under various salinity conditions in future studies for comparisons between marine and freshwater species.

Effects of volatile fatty acids

To investigate the effects of VFAs on anammox activity and inorganic carbon uptake, formate, acetate, and propionate, all of which are known to be oxidized by anammox bacteria with nitrate as the electron acceptor (10, 11, 36), were selected and tested (Fig. 3). After the 1-d incubation, the simultaneous consumption of NH_4^+ and NO_2^- was observed under all conditions. The average nitrogen stoichiometric ratios for $\Delta\text{NO}_2^-/\Delta\text{NH}_4^+$ and $\Delta\text{NO}_3^-/\Delta\text{NH}_4^+$ at 0 mM of formate, acetate, and propionate were 1.44 and 0.17, 1.20 and 0.03, and 1.34 and 0.08, respectively (Tables S1–S3). No significant differences were observed in these ratios ($P > 0.05$) among the three VFAs. The average nitrogen stoichiometric ratios for $\Delta\text{NO}_2^-/\Delta\text{NH}_4^+$ and $\Delta\text{NO}_3^-/\Delta\text{NH}_4^+$ at 1–10 mM were then compared with the ratio at 0 mM for each VFA. The stoichiometric ratio was not significantly different ($P > 0.05$) in any VFA experiment (Tables S1–S3). These results indicated that microbial processes other than anammox, such as VFA oxidation with nitrate and denitrification by co-existing bacteria in the biomass, were negligible. The average stoichiometric ratios in VFA experiments obtained in this study were similar to the previously reported ratios of marine anammox enrichment cultures, and were as follows: 1.22 and 0.22 (36), 1.28 and 0.24 (14), 1.21 and 0.15 (15), and 1.13

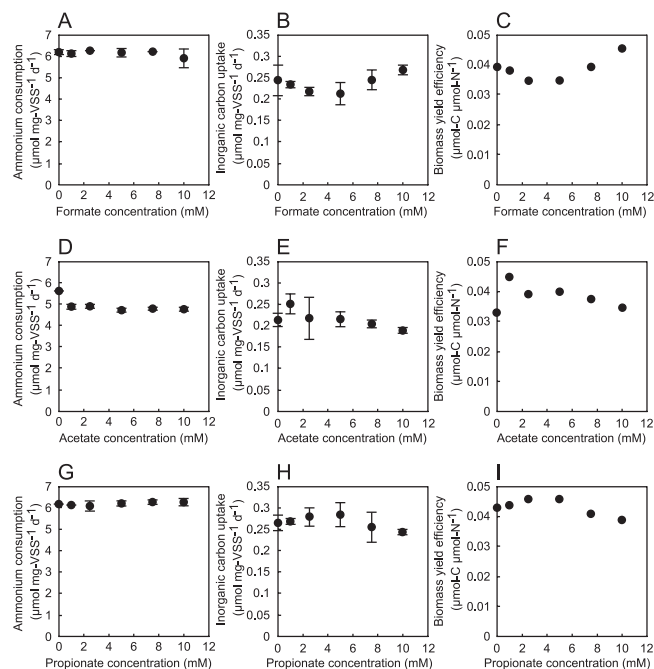


Fig. 3. Effects of formate (A–C), acetate (D–F), and propionate (G–I) on the specific ammonium consumption rate (A, D, G) and specific inorganic carbon uptake rate (B, E, H). The values in (A, D, G) and (B, E, H) are the means \pm SD of the results of independent triplicate experiments. Biomass yield efficiencies (C, F, I) were calculated by dividing the average specific inorganic carbon uptake rates by the average specific ammonium consumption rates.

and 0.18 (26). In the case of formate and propionate, ammonium consumption did not change with increasing concentrations relative to ammonium consumption at 0 mM (Fig. 3A and G). On the other hand, ammonium consumption slightly decreased with the addition of acetate (Fig. 3D), although this decrease was not dose-dependent. No significant decrease in inorganic carbon uptake was observed, even though average values fluctuated. The average carbon uptake observed with the addition of formate, acetate, or propionate was in the range of 20% of the uptake at 0 mM (Fig. 3B, E, and H). Based on ammonium consumption and carbon uptake results, average biomass yield efficiencies under the addition of formate, acetate, or propionate were calculated to be 0.039, 0.038, or 0.043 $\mu\text{mol-C } \mu\text{mol-N}^{-1}$, respectively (Fig. 3C, F, and I). These values were similar to the value of 0.030 under the 28°C and 2.8% salinity conditions reported by Awata *et al.* (4), but were 50% lower than the value reported for freshwater anammox bacteria (25, 32). One possible explanation for this lower yield efficiency is that “*Ca. Scalindua* spp.” require more energy for osmotic pressure adjustments or cell maintenance when salinity conditions are not optimal. However, future studies need to be conducted in order to confirm this phenomenon.

VFA oxidation rates

Anammox bacteria are recognized as consumers of VFAs with nitrate as the electron acceptor (10, 11, 36). In the present study, the oxidation rates of VFAs of “*Ca. Scalindua* sp.” were similar to those of other anammox bacteria previously reported (Table 2). The VFA oxidation rates of “*Ca.*

Table 2. Volatile fatty acid (VFA) oxidation rates of different anaerobic ammonium oxidation (anammox) species

Species	Oxidation rate of VFA ($\mu\text{mol g-protein}^{-1} \text{ min}^{-1}$)			Reference
	Formate	Acetate	Propionate	
" <i>Candidatus Brocadia anammoxidans</i> "	6.5±0.6	0.57±0.05	0.12±0.01	(10)
" <i>Ca. Brocadia fulgida</i> "	7.6±0.6	0.95±0.04	0.31±0.007	(11)
" <i>Ca. Kuenenia stuttgartiensis</i> "	5.8±0.6	0.31±0.03	0.12±0.01	(10)
" <i>Ca. Anammoxoglobus propionicus</i> "	6.7±0.6	0.79±0.07	0.64±0.05	(10)
" <i>Ca. Scalindua</i> spp."	7	0.7	0.3	(36)
" <i>Ca. Scalindua</i> sp."	5.2±0.1	0.78±0.19	0.36±0.05	This study

Scalindua sp." in the present study and "*Ca. Scalindua* spp." (36) were here than those of other freshwater anammox species, except for the acetate and propionate-consuming anammox bacteria "*Ca. Brocadia fulgida*" (11) and "*Ca. Anammoxoglobus propionicus*" (10). The higher VFA oxidation rates and tolerance to salinity of "*Ca. Scalindua*" species indicated that they had undergone niche adaptations to marine environments with lower ammonium and nitrite concentrations.

Conclusion

The anammox activity, inorganic carbon uptake, and biomass yield efficiency of "*Ca. Scalindua* sp." were determined by batch experiments under different salinities and VFA concentrations. We clearly demonstrated that the presence of VFAs did not affect these factors, whereas salinity significantly affected the carbon uptake of "*Ca. Scalindua* sp." These results provide important new insights into the activity and growth characteristics of "*Ca. Scalindua* sp." and indicate that salinity is one of the crucial factors regarding the nitrogen cycle in marine environments, especially at sites in which salinity fluctuates such as estuarine environments.

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References

- Amann, R.I., B.J. Binder, R.J. Olson, S.W. Chisholm, R. Deverux, and D. Stahl. 1990. Combination of 16S rRNA-targeted oligonucleotide probes with flow-cytometry for analyzing mixed microbial populations. *Appl. Environ. Microbiol.* 56:1919–1925.
- Amano, T., I. Yoshinaga, T. Yamagishi, C.V. Thuoc, P.T. Thu, S. Udea, K. Kato, Y. Sako, and Y. Suwa. 2011. Contribution of anammox bacteria to benthic nitrogen cycling in a mangrove forest and shrimp ponds, Haiphong, Vietnam. *Microbes Environ.* 26:1–6.
- Awata, T., K. Tanabe, T. Kindaichi, N. Ozaki, and A. Ohashi. 2012. Influence of temperature and salinity on microbial structure of marine anammox bacteria. *Water Sci. Technol.* 66:958–964.
- Awata, T., M. Oshiki, T. Kindaichi, N. Ozaki, A. Ohashi, and S. Okabe. 2013. Physiological characterization of an anaerobic ammonium-oxidizing bacterium belonging to the "*Candidatus Scalindua*" group. *Appl. Environ. Microbiol.* 79:4145–4148.
- Daims, H., R. Brühl, R. Amann, K.H. Schleifer, and M. Wagner. 1999. The domain-specific probe EUB338 is insufficient for the detection of all bacteria: development and evaluation of a more comprehensive probe set. *Syst. Appl. Microbiol.* 22:434–444.
- Dapena-Mora, A., I. Fernandez, J.L. Campos, A. Mosquera-Corral, R. Mendez, and M.S.M. Jetten. 2007. Evaluation of activity and inhibition effects on anammox process by batch based on the nitrogen gas production. *Enzyme Microb. Technol.* 40:859–865.
- Hu, B.L., L.D. Shen, X.Y. Xu, and P. Zheng. 2011. Anaerobic ammonium oxidation (anammox) in different natural ecosystems. *Biochem. Soc. Trans.* 39:174–177.
- Ishii, S., S. Ikeda, K. Minamisawa, and K. Senoo. 2011. Nitrogen cycling in rice paddy environments: Past achievements and future challenges. *Microbes Environ.* 26:282–292.
- Jetten, M.S.M., H.J.M. Op den Camp, J.G. Kuenen, and M. Strous. 2009. Family I. "*Candidatus Brocadiaceae*" fam. nov., p. 918–925. In N.R. Krieg, J.T. Staley, D.R. Brown, B.P. Hedlund, B.J. Paster, N.L. Ward, W. Ludwig, and W.B. Whitman (ed.), *Bergey's Manual of Systematic Bacteriology: the Bacteroidetes, Spirochaetes, Tenericutes (Mollicutes), Acidobacteria, Fibrobacteres, Fusobacteria, Dictyoglomi, Gemmatimonadetes, Lentisphaerae, Verrucomicrobia, Chlamydiae, and Planctomycetes*, 2nd ed, vol. 4. Springer-Verlag, New York.
- Kartal, B., J. Rattray, L.A. van Niftrik, *et al.* 2007. *Candidatus "Anammoxoglobus propionicus"* a new propionate oxidizing species of ammonium oxidizing bacteria. *Syst. Appl. Microbiol.* 30:39–49.
- Kartal, B., L.A. van Niftrik, J. Rattray, J. van de Vossenberg, M.C. Schmid, J.S. Damste, M.S.M. Jetten, and M. Strous. 2008. *Candidatus "Brocadia fulgida"*: an autofluorescent anaerobic ammonium oxidizing bacterium. *FEMS Microbiol. Ecol.* 63:46–55.
- Kawagoshi, Y., Y. Nakamura, H. Kawashima, K. Fujisaki, K. Furukawa, and A. Fujimoto. 2010. Enrichment of marine anammox bacteria from seawater-related samples and bacterial community study. *Water Sci. Technol.* 61:119–126.
- Kindaichi, T., T. Ito, H. Satoh, and S. Okabe. 2004. Ecophysiology interaction between nitrifying bacteria and heterotrophic bacteria in autotrophic nitrifying biofilms as determined by microautoradiography-fluorescence in situ hybridization. *Appl. Environ. Microbiol.* 70:1641–1650.
- Kindaichi, T., T. Awata, Y. Suzuki, K. Tanabe, M. Hatamoto, N. Ozaki, and A. Ohashi. 2011. Enrichment using an up-flow column reactor and community structure of marine anammox bacteria from coastal sediment. *Microbes Environ.* 26:67–73.
- Kindaichi, T., T. Awata, K. Tanabe, N. Ozaki, and A. Ohashi. 2011. Enrichment of marine anammox bacteria in Hiroshima Bay sediments. *Water Sci. Technol.* 63:964–969.
- Kuyppers, M.M.M., A.O. Sliemers, G. Lavik, M. Schmid, B.B. Jørgensen, J.G. Kuenen, J.S.S. Damsté, M. Strous, and M.S.M. Jetten. 2003. Anaerobic ammonium oxidation by Anammox bacteria in the Black Sea. *Nature* 422:608–611.
- Lotti, T., R. Kleerebezem, C. Lubello, and M.C.M. van Loosdrecht. 2014. Physiological and kinetic characterization of a suspended cell anammox culture. *Water Res.* 60:1–14.
- Ludwig, W., O. Strunk, R. Westram, *et al.* 2004. ARB: A software environment for sequence data. *Nucleic Acids Res.* 32:1363–1371.
- Meng, L., H. Cao, Y.-G. Hong, and J.-D. Gu. 2011. Seasonal dynamics of anammox bacteria in estuarial sediment of the Mai Po Nature Reserve revealed by analyzing the 16S rRNA and hydrazine oxidoreductase (*hzo*) genes. *Microbes Environ.* 26:15–22.
- Mulder, A., A.A. van de Graaf, L.A. Robertson, and J.G. Kuenen. 1995. Anaerobic ammonium oxidation discovered in a denitrifying fluidized-bed reactor. *FEMS Microbiol. Ecol.* 16:177–183.
- Nakajima, J., M. Sakka, T. Kimura, K. Furukawa, and K. Sakka. 2008. Enrichment of anammox bacteria from marine environment for the construction of bioremediation reactor. *Appl. Microbiol. Biotechnol.* 77:1159–1166.

22. Neef, A., R. Amann, H. Schlesner, and K.H. Schleifer. 1998. Monitoring a widespread bacterial group: in situ detection of planctomycetes with 16S rRNA-targeted probes. *Microbiology* 144:3257–3266.
23. Okabe, S., H. Satoh, and Y. Watanabe. 1999. In situ analysis of nitrifying biofilms as determined by *in situ* hybridization and the use of microelectrodes. *Appl. Environ. Microbiol.* 65:3182–3191.
24. Op den Camp, H.J.M., B. Kartal, D. Guven, *et al.* 2006. Global impact and application of the anaerobic ammonium-oxidizing (anammox) bacteria. *Biochem. Soc. Trans.* 34:1811–1816.
25. Oshiki, M., M. Shimokawa, N. Fujii, H. Satoh, and S. Okabe. 2011. Physiological characteristics of the anaerobic ammonium-oxidizing bacterium '*Candidatus* Brocadia sinica'. *Microbiology* 157:1706–1713.
26. Oshiki, M., T. Awata, T. Kindaichi, H. Satoh, and S. Okabe. 2013. Cultivation of planktonic anaerobic ammonium oxidation (anammox) bacteria using membrane bioreactor. *Microbes Environ.* 28:436–443.
27. Pruesse, E., C. Quast, K. Knittel, B. Fuchs, W. Ludwig, J. Peplies, and F.O. Glöckner. 2007. SILVA: a comprehensive online resource for quality checked and aligned ribosomal RNA sequence data compatible with ARB. *Nucleic Acids Res.* 35:7188–7196.
28. Sato, Y., H. Ohta, T. Yamagishi, *et al.* 2012. Detection of anammox activity and 16S rRNA genes in ravine paddy field soil. *Microbes Environ.* 27:316–319.
29. Schmid, M.C., B. Mass, A. Dapena, *et al.* 2005. Biomarkers for in situ detection of anaerobic ammonium-oxidizing (anammox) bacteria. *Appl. Environ. Microbiol.* 71:1677–1684.
30. Strous, M., J.J. Heijnen, J.G. Kuenen, and M.S.M. Jetten. 1998. The sequencing batch reactor as a powerful tool for the study of slowly growing anaerobic ammonium-oxidizing microorganisms. *Appl. Microbiol. Biotechnol.* 50:589–596.
31. Strous, M., J.A. Fuerst, E.H. Kramer, S. Logemann, G. Muyzer, K.T. van de Pas-Schoonen, R. Webb, J.G. Kuenen, and M.S.M. Jetten. 1999. Missing lithotroph identified as new planctomycete. *Nature* 400:446–449.
32. Strous, M., J.G. Kuenen, and M.S.M. Jetten. 1999. Key physiology of anaerobic ammonium oxidation. *Appl. Environ. Microbiol.* 65:3248–3250.
33. Tsushima, I., T. Kindaichi, and S. Okabe. 2007. Quantification of anaerobic ammonium-oxidizing bacteria in enrichment cultures by real-time PCR. *Water Res.* 41:785–794.
34. van de Graaf, A.A., P. de Bruijn, L.A. Robertson, M.S.M. Jetten, and J.G. Kuenen. 1996. Autotrophic growth of anaerobic ammonium-oxidizing micro-organisms in a fluidized bed reactor. *Microbiology* 142:2187–2196.
35. van de Vossenberg, J., J.E. Rattray, W. Geerts, B. Kartal, L. van Niftrik, E.G. van Donselaar, J.S.S. Damsté, M. Strous, and M.S.M. Jetten. 2008. Enrichment and characterization of marine anammox bacteria associated with global nitrogen gas production. *Environ. Microbiol.* 10:3120–3129.
36. van der Star, W.R.L., A.I. Miclea, U.G.J.M. van Dongen, G. Muyzer, C. Picoreanu, and M.C.M. van Loosdrecht. 2008. The membrane bioreactor: a novel tool to grow anammox bacteria as free cells. *Biotechnol. Bioeng.* 101:286–294.
37. Woebken, D., P. Lam, M.M.M. Kuypers, S.W.A. Naqvi, B. Kartal, M. Strous, M.S.M. Jetten, B.M. Fuchs, and R. Amann. 2008. A microdiversity study of anammox bacteria a novel *Candidatus* *Scalindua* phylotype in marine oxygen minimum zones. *Environ. Microbiol.* 10:3106–3119.
38. Yasuda, T., M. Waki, I. Yoshinaga, T. Amano, K. Suzuki, Y. Tanaka, T. Yamagishi, and Y. Suwa. 2011. Evidence of exponential growth of an anammox population in an anaerobic batch culture. *Microbes Environ.* 26:266–269.
39. Yoshinaga, I., T. Amano, T. Yamagishi, K. Okada, S. Ueda, Y. Sako, and Y. Suwa. 2011. Distribution and diversity of anaerobic ammonium oxidation (Anammox) bacteria in the sediment of a eutrophic freshwater lake, Lake Kitaura, Japan. *Microbes Environ.* 26:189–197.
40. Zheng, D., E.W. Alm, D.A. Stahl, and L. Raskin. 1996. Characterization of universal small-subunit rRNA hybridization probes for quantitative molecular microbial ecology studies. *Appl. Environ. Microbiol.* 62:4504–4513.