Subtle Population Genetic Structure in Yelloweye Rockfish (*Sebastes ruberrimus*) Is Consistent with a Major Oceanographic Division in British Columbia, Canada

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Abstract

The boundaries between oceanographic domains often function as dispersal barriers for many temperate marine species with a dispersive pelagic larval phase. Yelloweye rockfish (Sebastes ruberrimus, YR) are widely distributed across the northeastern Pacific Ocean, inhabiting coastal rocky reefs from the Aleutian Islands in Alaska through southern California. This species exhibits an extended pelagic larval duration and has the capacity for long distance larval transport. We assayed 2,862 YR individuals from 13 general areas in the northeast Pacific Ocean for allelic variation at nine microsatellite loci. Bayesian model-based clustering analyses grouped individuals from the Strait of Georgia (SG) into a distinct genetic cluster, while individuals from outer coastal water locations (OCLs) were partitioned equally across two genetic clusters, including the cluster associated with the SG fish. Pairwise F_{ST} values were consistently an order of magnitude higher for comparisons between the SG and OCLs than they were for all OCL-OCL comparisons (~0.016 vs. ~0.001). This same pattern was observed across two time points when individuals were binned into an "old" and "young" group according to birth year (old: \sim 0.020 vs. 0.0003; young: \sim 0.020 vs. \sim 0.004). Additionally, mean allelic richness was markedly lower within the SG compared to the OCLs (8.00 vs. 10.54-11.77). These results indicate that the Strait of Georgia "deep-basin" estuary oceanographic domain acts as a dispersal barrier from the outer coastal waters via the Juan de Fuca Strait. Alternatively, selection against maladapted dispersers across this oceanographic transition may underlie the observed genetic differentiation between the Georgia basin and the outer coastal waters, and further work is needed to confirm the SG-OCL divide acts as a barrier to larval dispersal.

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Introduction

Population replenishment for a large number of marine populations depends upon the input of exogenously derived individuals [1–4], which can occur via an extended pelagic larval phase [5,6]. The cumulative effects of larval trajectory, planktonic survival, delivery to settlement habitat and postsettlement performance result in variable contributions in local larval input from outside sources [7,8]. Mismatches between the spatial scales of exploited management units and of population replenishment may precipitate overexploitation of targeted species [9–11]. Thus, effective management of harvested species requires an understanding of the pattern of larval dispersal [12]. Furthermore, the utility of marine protected area (MPA) networks increases with their ability to enhance net larval production and export into exploitable areas while remaining self-sufficient via larval input from upstream MPAs [13–15].

Oceanography is an important driver of population structure in temperate reef fishes, as ocean currents largely dictate larval trajectories, planktonic survival and delivery to settlement habitat [7,16]. Offshore advection of larvae or propagules may lead to long distance larval transport [17,18]. On the other hand, there are a number of retention mechanisms that limit dispersal distances [18–22]. These include fronts associated with upwelling [23] or the intersection of independent current systems [24], gyres that form around seamounts or other complex bathymetric features [25–27], and eddies [28,29], including those formed by currents moving around rugged coastlines [30,31]. In lieu of directly measuring dispersal distances, which is often difficult or intractable [32], population genetic structure can inform indirect estimates of the scale of dispersal [33–36]. Genetic structure that occurs across oceanographic features suggests that these features may function as long-standing barriers to dispersal [37–39], influencing the spatial scale of ecological and evolutionary processes.

Rockfishes (*Sebastes* spp.) constitute a diverse group of nearshore fishes that are distributed in temperate waters around the world [40]. The bulk of the diversity (\sim 65 species) occurs in the northeast Pacific Ocean, where rockfishes are found in every habitat type from the intertidal waters to depths greater than



Figure 1. Map of sampling locations in Northeast Pacific Ocean. Sampling locations and location codes of yelloweye rockfish (*Sebastes ruberrimus*), assayed at nine microsatellite loci. doi:10.1371/journal.pone.0071083.g001

1,500 meters [40]. This group shares many life-history attributes, such as slow maturation rate and highly variable juvenile recruitment that contribute to their low productivity and make them susceptible to overfishing [41]. Rockfishes are characterized by large populations, high fecundities, and pelagically dispersed larvae with the potential for widespread gene flow. Population genetic studies of rockfishes have shown that despite an extended pelagic larval duration (PLD), they often exhibit population structure over regional scales, potentially explained by the absence of settlement habitat [42] or concordant oceanographic divisions [43-49]. For example, Point Conception in southern California is associated with a genetic break in blue rockfish (S. mystinus) [47] and vermillion rockfish (S. miniatus) [50], while both copper rockfish (S. caurinus) and brown rockfish (S. auriculatus) from the Puget Sound basin exhibit significant genetic divergence from populations located along the outer coast [45,48,51].

In this study, we used microsatellite genetic markers to assess the population genetic structure of yelloweye rockfish (*S. ruberrimus;* YR) in the northeast Pacific Ocean. We predicted that population structure would occur across oceanographic features that likely

function as barriers to dispersal, resulting in three genetically differentiated subdivisions. These predicted genetic subdivisions include (i) the separation of the Georgia basin from the outer coast via the Juan de Fuca Strait and (ii) isolation of the Bowie Seamount location. We also predicted an isolation by distance signal would be detected across the outer coast locations, which span over 1,500 km of coastline. We detected subtle genetic structure, partitioning the Strait of Georgia population from a panmictic outer coast population. No signal of isolation by distance was detected across the outer coast sampling locations.

Materials and Methods

Study area and study species

Water circulation patterns in the northeast Pacific Ocean are largely structured by the eastward flowing Subarctic Current, which divides into the southern flowing California Current and northern flowing Alaska Current well offshore of Vancouver Island, British Columbia [52]. In British Columbia, the nearshore waters of the outer coast are dominated by the directional flow of

 Table 1. Sample location specific descriptive statistics.

Sample Location	Code	N	Mean A _R H _O		H _E	F _{IS}	P-value	
Southeast Alaska	AK	85	11.13	0.711	0.722	0.015	0.2038	
Brooks Bay	BB	50	10.59	0.723	0.731	0.012	0.2812	
Barber Point	BP	345	11.14	0.711	0.728	0.024	0.0034	
Bowie Seamount	BS	779	11.10	0.701	0.726	0.034	0.0004	
Calvert Island	CI	87	11.15	0.701	0.725	0.033	0.0308	
Cape St. James	CJ	327	10.99	0.692	0.723	0.043	0.0004	
Esperanza	ES	46	10.99	0.725	0.730	0.007	0.3735	
Oregon	OR	75	10.54	0.684	0.714	0.043	0.0150	
Strait of Georgia	SG	123	8.00	0.630	0.631	0.003	0.4406	
Tasu	TA	231	11.32	0.743	0.741	-0.002	0.5795	
Top Knot	ΤK	167	11.21	0.716	0.728	0.017	0.0718	
Triangle	TR	187	11.77	0.689	0.728	0.054	0.0004	
Washington	WA	81	11.12	0.707	0.723	0.024	0.0868	

Sample size (N), mean allelic richness (A_R), observed heterozygosity (H_O), expected heterozygosity (H_E), coefficient of inbreeding (F_{1S}) and associated Pvalues are shown for each sample location. P-values in **bold** denote significant heterozygote deficiencies at the 5% level after correcting for multiple comparisons (critical value: 0.00043).

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Table 2. Pairwise F_{ST} values.

the Davidson and Vancouver Island coastal currents, which dominate in winter and summer, respectively [53]. Circulation patterns within the inshore waters of the Georgia basin are typified by tidal currents and, in the summer, estuarine currents caused by freshwater input from the Fraser River [52,54]. Tidal forces cause intense vertical mixing of brackish surface waters and deep water from the Juan de Fuca Strait along sills located throughout the San Juan Archipelago and southern Georgia basin [54]. Through the vertical forcing of deep water and subsequent mixing that occurs with surface waters, these sills may act as physical barriers to larval dispersal [55,56]. Yelloweye rockfish are one of the larger and longest-lived species of rockfishes, reaching lengths greater than 80 cm and often living longer than 100 years [40]. Slow maturation rate, highly variable juvenile recruitment success, and long lifespans contribute to YR's susceptibility to overfishing [41,57], and conservation concerns in Canada and the United States have precipitated formal status assessments by The Committee on the Status of Endangered Wildlife in Canada (COSEWIC) and the National Marine Fisheries Service. This species is currently listed as "Special Concern" and "Threatened" in British Columbia and Puget Sound, respectively [58,59].

Sample collection

Yelloweye rockfish tissue samples were collected from 13 general areas in the coastal waters of British Columbia (including Bowie Seamount, approximately 200 km west of Haida Gwaii), southeast Alaska, Washington and Oregon (Figure 1, Table 1). Fish were sampled during Fisheries and Oceans Canada inshore rockfish stock assessment longline surveys, and opportunistically from commercial fishery vessels from 1998-2006. The majority of sample locations are comprised of individuals that were collected during a single survey. Strait of Georgia individuals, however, were collected over two years of sampling. Surveys were not restricted to a single season, and samples were collected over all seasons. Although we defined each sample area as a single location, they comprised samples collected from a range of individual sites within the general area of each of the 13 sample areas. Scientific collection permits and animal care approval were not required as tissue samples were taken from individuals already sampled as part of the DFO Inshore Rockfish Research Program, and opportunistically from commercial fishery vessels. Tissue samples were taken either onboard the fishing vessel or dockside and stored in 95% ethanol for genetic analyses. Sagittal otoliths were taken from a subset of individuals and locations for ageing.

Laboratory methods and scoring

Total genomic DNA was extracted with Qiagen Dneasy extraction kits (Qiagen, Valencia, California). Nine microsatellite

	AK	BB	BP	BS	сі	CJ	ES	OR	SG	ТА	тк	TR	WA
AK	-												
BB	0.0002	-											
BP	-0.0001	0.0019	-										
BS	0.0004	0.0003	0.0005	-									
CI	0.0033	0.0012	0.0015	0.0012	-								
CJ	0.0005	0	0.0002	-0.0001	0.0007	-							
ES	0.0006	0.0002	-0.0002	-0.0011	0.0010	-0.0018	-						
OR	0.0036	0.0012	0.0016	0.0006	-0.0004	0.0012	0.0013	-					
SG	0.0163	0.0189	0.0156	0.0165	0.0175	0.0142	0.018	0.0193	-				
ТА	0.0003	0.0002	0.0011	0.0002	0.0018	0.0009	-0.0001	0.0018	0.0193	-			
тк	0.0009	0.0006	0.0005	0.0004	0.0006	0.0003	0.0014	0.0007	0.0138	0.0001	-		
TR	0.0015	0.0004	0.0008	0.0003	0.0013	0.0006	0.0007	0.0016	0.0168	0.0011	0.0001	-	
WA	-0.0005	-0.0023	0.0010	-0.0001	0.0010	-0.0003	0.0002	0.0018	0.0164	0.0009	0.0008	0	-

Pairwise F_{ST} values for all sample locations are shown. Values in **bold** type are significant after correcting for multiple comparisons using the Bonferroni correction (adjusted critical value: 0.000641). All pairwise comparisons with the Strait of Georgia (SG) sample location are significant, and amongst the outer coast location comparisons, only the Barber Point (BP) – Tasu (TA) pairwise comparison is significant.

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loci (Table S1 in File S1) were amplified using polymerase chain reaction (PCR). Typical PCR cycling conditions included an initial denaturation at 94° C for 2 min, followed by ~30 cycles of 94° for 1 min, 46° – 60° for 30 sec, a 72° extension for 1 min, and a final extension at 72° for 10 min. Annealing times and temperatures were adjusted to optimize specific locus amplification. Forward PCR primers were fluorescently labeled and fragment sizing was conducted on the ABI 377 automated DNA sequencer platform (Applied Biosystems (ABI), Foster City, California). Fragments were sized using the GeneScan-500 size standard and allele scoring was performed with GENEMAPPER version 3.7 (ABI).

Data analysis

Deviations from Hardy-Weinberg equilibrium were assessed for each locus-population combination with GENEPOP version 3.1 [60]. Estimates of the exact P-values were obtained using the Markov chain method (1000 batches, 5000 iterations per batch). Genotypic linkage disequilibrium for all combinations of locus pairs within sample locations was calculated in GENEPOP (100 batches, 1000 iterations per batch). Measures of heterozygosity and allelic richness were calculated in FSTAT version 2.9.3.2 [61]. The assumption of selective neutrality was assessed with the selection detection workbench LOSITAN [62], which implements the F_{ST} outlier approach of FDIST [63]. Tests for isolation-bydistance (IBD) were performed with Mantel tests using the *ade4* package [64] in the R environment [65]. Pairwise F_{ST} values were linearized (F_{ST} /1- F_{ST}) according to Rousset [66].

Population structure was assessed with estimates of the summary statistic, F_{ST} , as well as with the genetic clustering program STRUCTURE version 2.3 [67-69]. Pairwise FST values were estimated with θ [70] using the permutation approach implemented in FSTAT. To test for temporal stability of pairwise FST values, 420 individuals (which have accompanying age data) from eight sample locations were binned into a "young" or "old" age group according to birth year. Individuals born before and after 1980 were binned into the "old" and "young" group, respectively. The year 1980 was chosen as the cutoff as it is the median birth year, which also maximizes sample size similarity between the sample locations for each age group (Table S2 in File S1). STRUCTURE was used to determine the number of distinct genetic clusters (K) among the sample locations, and to estimate individual assignment probabilities for each fish to each resolved cluster. The number of putative genetic clusters assessed ranged from 1 to 7. Each run consisted of a 500,000 step burn-in plus an additional 1,000,000 steps, and 20 iterations were run for each Kvalue. Due to low overall genetic structuring (global $F_{ST} = 0.002$), the number of genetic clusters was evaluated under both an admixture model including a location prior [69] in addition to an admixture model without a location prior. We did not utilize the method of Evanno et al. [71], which calculates the second order rate of change (ΔK) as an estimator of K, due to the limited success of this method observed when overall differentiation is low [72].

Results

Genotyping and scoring

A total of 2862 individuals from 13 general sample locations (sample sizes ranged from 46 to 779) were assayed at nine microsatellite loci. Only individuals with allele scores from a minimum of seven loci were retained for analysis. The majority of individuals (66.8%) had genotype scores across all nine loci, 628 (21.9%) contained missing data for one locus, and 323 (11.3%) individuals had missing data for two loci. Missing data, therefore, constitutes only 4.9% of the total dataset. There was no evidence

for selection acting at any of the microsatellite loci using LOSITAN (for all loci P>0.24; data not shown).

Within population variation

The number of alleles across all populations ranged from 10 (Sal3) to 39 (Sme3) with an average of 18.9 alleles per locus. Expected heterozygosity ranged from 0.488 (Sru9) to 0.877 (Sal1) and averaged 0.702 across loci and sample locations (Table S3 in File S1). A relatively high proportion of samples (24 out of 117 comparisons) were found to be out of Hardy-Weinberg equilibrium (HWE) with a critical value of 0.05. After using the Bonferroni correction for multiple comparisons, however, the number of departures from HWE dropped to eight (critical value: 0.0004). Samples exhibiting a departure from HWE were distributed across sample locations and loci, and, therefore are unlikely to affect our results because they do not point to consistently anomalous loci or localities. Only 39 (out of 468) locus-locus within sample location comparisons exhibited significant deviations from linkage equilibrium (P<0.05). These 39 departures were not concentrated on a particular locus pair or within specific sample locations, and are also unlikely to affect our results.

Among population variation

The amount of genetic variation attributable to differences between sample locations was low (global $F_{ST} = 0.002$, P < 0.01). Pairwise F_{ST} values range from less than 0 to 0.0193, and were consistently an order of magnitude higher for all comparisons between the Strait of Georgia (SG) sample location and outer coastal water locations (OCLs) than for comparisons between OCLs (~0.016 vs. ~0.001; Table 2). This same pattern was observed when comparisons were restricted to both the "old" and "young" groups (mean value for old SG-OCL vs. old OCL-OCL: ~0.0203 vs. ~0.0003; mean value for young SG-OCL vs. young OCL-OCL: 0.0204 vs. 0.0041; Table S4 and Table S5 in File S1). Furthermore, mean (across loci and locations) allelic richness was markedly lower in the SG location compared to the OCLs (8.00 vs. 10.54–11.77, average for OCLs: 11.10; Table 1).

Two admixture models, one without a location prior and one with a location prior were evaluated using STRUCTURE. The results based on the model without the location prior failed to detect more than a single genetic cluster (highest support for a K of 1) among our samples. By contrast, results from the model including a location prior had equal support for a K of 1 and a K of 2 (Table S6 in File S1), yet SG individuals exhibited high q-values (minimum is 0.87, mean is 0.93), while all OCL individuals exhibited approximately equal admixture of both genetic clusters (q-value approximately equal to 0.5; Figure 2).

No significant association between genetic distance and geographic distance was detected with the Mantel tests (Figure 3). Two IBD analyses were conducted, one including all pairwise sample location comparisons (Mantel P=0.250) and one with the Strait of Georgia and Bowie Seamount sample locations removed (Mantel P=0.268).

Discussion

Population genetic structure in yelloweye rockfish

We detected subtle population genetic structure that separates a putative Strait of Georgia population from a panmictic outer coast population. The results presented here constitute a re-analysis of YR population genetic structure originally investigated by Yamanaka et al. [73] and Yamanaka et al. [74]. Yamanaka et al. [73] present genetic data from several OCLs (including fish from



Figure 2. STRUCTURE analysis. STRUCTURE outputs for the admixture model without a location prior (*a*) and the admixture model with the location prior (*b*). The genome of each individual fish is represented by a thin vertical line as assayed by nine microsatellite markers, where each shade of grey represents a unique genetic cluster, and the proportion of each genetic cluster that contributes to an individual's genome is illustrated by the relative amount of each shade within each vertical line. Under the admixture model without a location prior, all individuals exhibit roughly equal admixture between the two genetic clusters (q is approximately 0.5). Under the admixture model with a location prior, however, the Strait of Georgia individuals exhibit a q-value close to 1.0, while the outer coast location individuals still exhibit equal admixture of both clusters. doi:10.1371/journal.pone.0071083.g002

Bowie Seamount), and failed to reject the hypothesis that all individuals were derived from a single, panmictic population. Yamanaka et al. [74] extended the sampling to include fish from the SG, and, using tree-based analyses, identified a putative SG-OCL genetic break with moderate bootstrap support. Our analyses support this putative population boundary, and provide some support for the temporal stability of this population structure. While increasing the number of sample locations within the Strait of Georgia, as well as the number of temporal comparisons would bolster the robustness of the population structure we observed, similar results were found in a separate analysis conducted with a different molecular marker type, Amplified Fragment Length Polymorphisms [75] and with results for other rockfishes [45,48,51].

Natural selection as a mechanism promoting genetic isolation of the Georgia basin

One interpretation of the genetic structure observed between the SG and the OCLs is that the oceanographic transition between these areas functions as a physical barrier to larval dispersal. Alternatively, dispersal may occur across this oceanographic divide and natural selection against maladapted dispersers reduces connectivity and promotes genetic differentiation [8]. The Georgia basin is a unique environment, and likely presents individuals with varying selective forces [52]. Selection, however, is more likely to differentiate a subset of functionally important loci [45], while the genetic differentiation we observed was due to the combined effects of all (putatively neutral) loci, rather than being driven by a single locus with a large effect. This suggests that a genome-wide process, such as reduced gene flow, underlies the observed genetic structure. If selection were driving this pattern, we would expect it to be operating on many loci spread across the genome in order to affect all nine of the microsatellite markers via hitchhiking, a scenario less consistent with our data.

Lack of genetic structure in the outer coastal waters

Bowie Seamount (58.23°N 135.74°W) is approximately 200 kilometers west of Haida Gwaii and rises to within 30 meters of the surface from depths over 3,000 meters creating rockfish habitat that is isolated from coastal areas [52]. Despite these putative barriers to larval exchange, we found no evidence for genetic differentiation of the Bowie Seamount sample from other outer coast sample locations. Episodic recruitment of larvae from coastal populations to Bowie Seamount may be driven by Haida eddies [76-79], mesoscale vortices that form along the west coast of Haida Gwaii and move westward into the Gulf of Alaska and may persist for several years [76,77]. While the PLD of YR is unknown, the PLDs of other rockfishes ranges from one to several months, but may be up to one year [40]. Given the average velocity of a Haida eddy, pelagic larval durations on the longer side may allow for sufficient transport time to Bowie Seamount, which may help explain the apparent genetic homogeneity we observed in this study.

The apparent lack of genetic structure or a significant IBD relationship between the OCLs should not be taken, on its own, as evidence supporting high connectivity between these locations.



Figure 3. Isolation-by-distance analysis. The relationship between genetic distance ($F_{ST}/1-F_{ST}$) and geographic distance is shown for all pairwise locations, excluding the Strait of Georgia and Bowie Seamount sample locations. No significant correlation was detected with a Mantel test (P>0.25).

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High effective population sizes (Ne), commonly observed in marine fishes, may resist the effects of drift and contribute to low F_{ST} values [80]. Using estimates of population genetic structure to infer contemporary patterns of gene flow (and migration rates) may be problematic, as these estimates reflect both historical and contemporary patterns of gene flow, [81]. Over evolutionary time, one migrant per generation can be sufficient to homogenize neutral allele frequencies [82,83]. Thus, rare historical dispersal events may be sufficient to confound contemporary patterns of gene flow, but not reflect ecologically relevant influx of larvae. Moreover, consistent dispersal following a stepping-stone pattern over long time scales may lead to genetic homogenization.

Management and conservation of yelloweye rockfish

Conservation concerns surrounding Pacific rockfishes (*Sebastes* spp.) in the late 1990's [41] precipitated a number of changes to the management of inshore rockfish in British Columbia. In 2002, Fisheries and Oceans Canada (DFO) introduced a Rockfish Conservation Strategy to address these concerns and slow apparent population declines. Four specific measures were implemented: (i) account for all rockfish catch, (ii) decrease fishing mortality, (iii) establish areas closed to fishing, and (iv) improve stock assessment and monitoring (for details please see [53]). Concurrent with the implementation of the Rockfish Conservation Strategy, YR were reviewed by COSEWIC [58] to determine if available information warranted a protected status. Based on the genetic and demographic differences identified by Yamanaka et al. [74] and supported by this study, two Designatable Units (DUs) of YR were delineated and assessed by COSEWIC. Both

the inside and outside DUs (SG and outer coastal waters, respectively) are listed as "Special Concern", owing mainly to the life-history characteristics (e.g., long-life spans, slow maturation rate, and highly variable juvenile recruitment) that make YR susceptible to overfishing and have the potential to slip into "Threatened" status [58].

As part of the Rockfish Conservation Strategy, areas suitable for protection were identified through a series of consultations and habitat modelling [57]. The first rockfish conservation areas (RCAs) were designated in 2002, and at present there are 164 RCAs in the coastal waters of British Columbia (Figure S1 in File S1). The RCAs constitute approximately 20% and 30% of outer coast and Georgia basin rockfish habitat, respectively. While the implementation of the RCAs constitutes a significant achievement in marine conservation, much work remains to understand how these protected areas affect demographic and ecological processes within the greater region.

Understanding larval movement is a critical component in understanding the regional effect of marine protected areas (MPAs) [84,85]. Larvae originating within protected areas may either recruit locally, enhancing populations within the MPA or spill over MPA boundaries, enhancing outside fished areas. The extent to which MPAs are self-sufficient (i.e. self-recruiting) depends upon the size of the MPA in relation to the dispersal distances of locally produced larvae. This is an important consideration for the RCAs, as they were not delineated to purposefully function as a network connected by larval dispersal. Furthermore, there is considerable variation in the size and spacing of RCAs between those located in inside and outside waters. Future studies will need to assess the number of individuals residing in outside and inside RCAs to understand how larval output from RCAs may vary between outside and inside areas, and what effect that entails for the role of upstream RCAs acting as larval sources for downstream RCAs.

Conclusion

Our analyses support a population boundary between the inshore waters of the Strait of Georgia and the outer coastal waters that coincides with the transition between two oceanographic domains via the Juan de Fuca Strait. Several important implications can be drawn from this result. Principally, the genetic data suggest that dispersal is restricted regionally by major oceanographic features. This scale of dispersal seems to match the scale at which YR stocks are managed, although more detailed studies are needed to fully elucidate the complex metapopulation dynamics observed. If dispersal between the Strait of Georgia and the outer coastal waters is indeed rare or demographically insignificant, it is unlikely that outside areas will function as substantial larval sources for the inshore population.

Supporting Information

File S1 Supporting figures and tables. Figure S1. Rockfish Conservation Areas. The distribution of rockfish conservation areas (RCAs) in British Columbia (figure reproduced with permission from Yamanaka & Logan [57]). Information about the RCAs can be found on the DFO website: http://www.pac.dfo-mpo.gc.ca/fm-gp/maps-cartes/rca-acs/index-eng.htm. Table S1. Primer information. The forward (F) and reverse (R) primer sequences, PCR annealing temperatures (T_A), Genbank accession number (GB AC #), and reference for each microsatellite locus are shown below. Table S2. Temporal population structure comparisons. The sample size for the "old" and "young" datasets (N_{old}, N_{young}), as well as the pairwise F_{ST} value for each within

location, "old" vs. "young" comparison is shown below. None of the pairwise F_{ST} values is statistically significant. Table S3. Individual locus descriptive statistics. Mean number of alleles (N_A), total number of alleles (N_T), observed heterozygosity (H_O), expected heterozygosity (H_E), inbreeding coefficient (F_{IS}), theta, and standard error of theta (S.E.) for each locus are shown. Table S4. "Old" pairwise FST values. Pairwise FST values for the "old" dataset are shown below. Statistically significant values are shown in **bold**. Table S5. "Young" pairwise FST values. Pairwise FST values for the "young" dataset are shown below. Statistically significant values are shown in **bold**. Table S6. STRUCTURE mean log likelihood results. Mean log likelihood of each K value (LnP(K)) and standard deviation (S.D.) are shown for both models evaluated in STRUCTURE: an admixture model without a location prior, and an admixture model with a location prior. The most likely value of *K* is shown in **bold** under each model. (DOC)

References

- Roughgarden J, Gaines SD, Possingham H (1988) Recruitment dynamics in complex life cycles. Science 241: 1460–1466.
- Caley MJ, Carr MH, Hixon MA, Hughes TP, Jones GP, et al. (1996) Recruitment and the local dynamics of open marine populations. Annu Rev Ecol Syst 27: 477–500.
- Bode M, Bode L, Armsworth PR (2006) Larval dispersal reveals regional sources and sinks in the Great Barrier Reef. MEPS 308: 17–25.
- Cowen RK, Sponaugle S (2009) Larval dispersal and marine population connectivity. Annu Rev Mar Sci 1: 443–466.
- Thorson G (1950) Reproductive and larval ecology of marine bottom invertebrates. Biol Rev 25: 1–45.
- Scheltema RS (1986) On dispersal and planktonic larvae of benthic invertebrates: an eclectic review. Bull Mar Sci 39: 290–322.
- Gunderson DR, Vetter RD (2006) Temperate rocky reef fishes. In Kritzer JP, Sale PF, editors. Marine Metapopulations. San Francisco: Academic Press. 69– 117.
- Marshall DJ, Monro K, Bode M, Keough MJ, Swearer S (2010) Phenotypeenvironment mismatches reduce connectivity in the sea. Ecol Lett 13: 128–140.
- Ihssen PE, Booke HE, Casselman JM, McGlade JM, Payne NR, et al. (1981) Stock identification: materials and methods. CJFAS 38: 1838–1855.
- Carvalho GR, Hauser L (1994) Molecular genetics and the stock concept in fisheries. Rev Fish Biol Fisher 4: 326–350.
- Hilborn R, Orensanz JM, Parma AM (2005) Institutions, incentives and the future of fisheries. Philos Trans R Soc Lond B 360: 47–57.
- White C, Costello C (2011) Matching spatial property rights fisheries with scales of fish dispersal. Ecol Appl 21(2): 350–362.
- Botsford LW, Hastings A, Gaines SD (2001) Dependence of sustainability on the configuration of marine reserves and larval dispersal distance. Ecol Lett 4: 144– 150.
- Shanks AL, Grantham BA, Carr MH (2003) Propagule dispersal distance and the size and spacing of marine reserves. Ecol Appl 13(1) Supplement: S159– S169.
- Harrison HB, Williamson DH, Evans RD, Almany GR, Thorrold SR, et al. (2012) Larval export from marine reserves and the recruitment benefit for fish and fisheries. Curr Biol 22(11): 1023–1028.
- Largier JL (2003) Considerations in estimating larval dispersal distances from oceanographic data. Ecol Appl 13(1) Supplement: S71–S89.
- Gaines SD, Roughgarden J (1985) Larval settlement rate: a leading determinant of structure of an ecological community of the marine intertidal zone. Proc Natl Acad Sci USA 82(11): 3707–3711.
- Shanks AL (2009) Pelagic larval duration and dispersal distance revisited. Biol Bull 216(3): 373–385.
- Lobel PS, Robinson AR (1986) Transport and entrapment of fish larvae by ocean mesoscale eddies and currents in Hawaiian waters. Deep-Sea Res 33(4): 483–500.
- Cowen RK, Castro LR (1994) Relation of coral-reef fish larval distributions to island scale circulation around Barbados, West Indies. Bull Mar Sci 54: 228– 244.
- Cowen RK, Kamazima LMM, Sponaugle S, Paris CB, Olson DB (2000) Connectivity of marine populations: open or closed? Science 287: 857–859.
- McManus MA, Woodson CB (2012) Plankton distribution and ocean dispersal. J Exp Biol 215: 1008–1016.
- Morgan SG, Fisher JL, Miller SH, McAfee ST, Largier JL (2009) Nearshore larval retention in a region of strong upwelling and recruitment limitation. Ecology 90(12): 3489–3502.
 Feng M, Slawinski D, Beckley LE, Keesing JK (2010) Retention and dispersal of

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- Dower JF, Perry RI (2001) High abundance of larval rockfish over Cobb Seamount, an isolated seamount in the Northeast Pacific. Fish Oceanogr 10(3): 268–274.
- Brewin PE, Stocks KI, Haidvogel DB, Condit C, Gupta A (2009) Effects of oceanographic retention on decapod and gastropod community diversity on seamounts. MEPS 383: 225–237.
- Hayward TL, Mantyla AW (1990) Physical, chemical and biological structure of a coastal eddy near Cape Mendocino. J Mar Res 48(4): 825–850.
- Sabatés A, Salat J, Raya V, Emelianov M (2012) Role of mesoscale eddies in shaping the spatial distribution of the coexisting *Engraulis encrasicolus* and *Sardinella aurita* larvae in the northwestern Mediterranean. J Mar Res 111: 108–119.
- Wolanski E (1994) Physical Oceanographic Processes of the Great Barrier Reef. Boca RatonFL: CRC Press. 194 p.
- Andutta FP, Kingsford MJ, Wolanski E (2012) 'Sticky water' enables the retention of larvae in a reef mosaic. Estuar Coast Shelf S 101: 54–63.
- 32. Levin LA (2006) Recent progress in understanding larval dispersal: new directions and digressions. Integr Comp Biol 46(3): 282–297.
- Hellberg ME, Burton RS, Neigel JE, Palumbi SR (2002) Genetic assessment of connectivity among marine populations. Bull Mar Sci 70(1) Supplement: 273– 290.
- Purcell JFH, Cowen RK, Hughes CR, Williams DA (2006) Weak genetic structure indicates strong dispersal limits: a tale of two coral reef fish. Proc R Soc B 273: 1483–1490.
- Weersing K, Toonen RJ (2009) Population genetics, larval dispersal, and connectivity in marine systems. MEPS 393: 1–12.
- Pinsky ML, Montes HR, Palumbi SR (2010) Using isolation by distance and effective density to estimate dispersal scales in anemone fish. Evolution 64(9): 2688–2700.
- Avise JC (2000) Phylogeography: the history and formation of species. CambridgeMA: Harvard University Press. 447 p.
- Dawson MN (2001) Phylogeography in coastal marine animals: a solution from California? J Biogeogr 28: 723–736.
- Pelc RA, Warner RR, Gaines SD (2009) Geographical patterns of genetic structure in marine species with contrasting life histories. J Biogeogr 36: 1881– 1890.
- Love MS, Yaklovich M, Thorsteinson L (2002) The Rockfishes of the Northeast Pacific. BerkeleyCA: University of California Press, 404 p.
- Parker SJ, Berkeley SA, Golden JT, Gunderson DR, Heifetz J, et al. (2000) Management of Pacific Rockfish. Fisheries 25(3): 22–29.
- Johansson ML, Banks MA, Glunt KD, Hassel-Finnegan HM, Buonaccorsi VP (2008) Influence of habitat discontinuity, geographical distance, and oceanography on fine-scale population genetic structure of copper rockfish (*Sebastes caurinus*). Mol Ecol 17: 3051–3061.
- Rocha-Olivares A, Vetter RD (1999) Effects of oceanographic circulation on the gene flow, genetic structure, and phylogeography of the rosethorn rockfish (Sebastes helvomaculatus) CJFAS 56: 803–813.
- Withler RE, Beacham TD, Schulze AD, Richards LJ, Miller KM (2001) Coexisting populations of Pacific ocean perch, *Sebastes alutus*, in Queen Charlotte Sound, British Columbia. Mar Biol 139: 1–12.
- Buonaccorsi VP, Kimbrell CA, Lynn EA, Vetter RD (2002) Population structure of copper rockfish (*Sebastes caurinus*) reflects postglacial colonization and contemporary patterns of gene flow. CJFAS 59: 1374–1384.
- Buonaccorsi VP, Westerman M, Stannard J, Kimbrell C, Vetter RD (2004) Molecular genetic structure suggests limited larval dispersal in grass rockfish, *Sebastes rastrelliger*. Mar Biol 145: 779–788.

geography. Mar Freshwater Res 61: 1259-1267.

- Cope JM (2004) Population genetics and phylogeography of the blue rockfish (Sebastes mystimus) from Washington to California. CJFAS 61: 332–342.
- Buonaccorsi VP, Kimbrell CA, Lynn EA, Vetter RD (2005) Limited realized dispersal and introgressive hybridization influence genetic structure and conservation strategies for brown rockfish, *Sebastes auriculatus*. Conser Genet 6: 697–713.
- Burford MO (2009) Demographic history, geographical distribution and reproductive isolation of distinct lineages of blue rockfish (*Sebastes mystinus*), a marine fish with a high dispersal potential. J Evolution Biol 22(7): 1471–1486.
- Hyde JR, Vetter RD (2009) Population genetic structure in the redefined vermillion rockfish (*Sebastes miniatus*) indicates limited larval dispersal and reveals natural management units. CJFAS 66: 1569–1581.
- Seeb LW (1998) Gene flow and introgression within and among three species of rockfishes, Sebastes auriculatus, S. caurinus, and S. maliger. J Hered 89(5): 393–403.
- Thomson RE (1981) Oceanography of the British Columbia Coast. Canadian Special Publications of Fisheries and Aquatic Sciences 56: 291 p.
- Freeland HJ, Crawfrod WR, Thomson RE (1984) Currents along the pacific coast of Canada. Atmos Ocean 22(2): 151–172.
- Waldichuck M (1957) Physical oceanography of the Strait of Georgia, British Columbia. Journal of Fisheries Research Board, Canada 14(3): 321–486.
- Sköld M, Wing SR, Mladenov PV (2003) Genetic subdivision of a sea star with high dispersal capability in relation to physical barriers in a fjordic seascape. MEPS 250: 163–174.
- Perrin C, Wing SR, Roy MS (2004) Effects of hydrographic barriers on population genetic structure of the sea star *Coscinasterias muricata* (Echinodermata, Asteroidea) in the New Zealand fiords. Mol Ecol 13: 2183–2195.
- Yamanaka KL, Logan G (2010) Developing British Columbia's inshore rockfish conservation strategy. Mar Coast Fish: 28–46.
- COSEWIC (2008) COSEWIC Assessment and Status Report on the Yelloweye Rockfish Sebastes ruberrimus in Canada. Committee on the Status of Endangered Wildlife in Canada. Available: http://www.sararegistry.gc.ca/species/ speciesDetails_e.cfm?sid = 1020. Accessed 8 September 2012.
- 59. Federal Register (2010) vol. 75 no. 8, 22276–22290.
- Raymond M, Rousset F (1995) GENEPOP (version 1.2): population genetics software for exact tests and ecumenicism. J Hered 86: 248–249.
- Goudet J (2001) FSTAT, version 2.9.3, A program to estimate and test gene diversities and fixation indices. *Lausanne University*, Lausanne, Switzerland.
- Antao T, Lopes A, Lopes RJ, Beja-Pereira A, Luikart G (2008) LOSITAN: A workbench to detect molecular adaptation based on a F_{ST}-outlier method. BMC bioinformatics. doi: 10.1186/1471-2105-9-323.
- Beaumont MA, Nichols RA (1996) Evaluating loci for use in the genetic analysis of population structure. Proc R Soc B 363: 1619–1626.
- 64. Dray S, Dufour AB (2007) The ade4 package: implementing the duality diagram for ecologists. J Stat Softw 22(4): 1–20.
- R Core Team. (2012) R: A language and environment for statistical computing. Vienna, Austria. Available: http://www.R-project.org.
- Rousset F (1997) Genetic differentiation and estimation of gene flow from Fstatistics under isolation by distance. Genetics 145(4): 1219–1228.

- Pritchard JK, Stephens M, Donnelly P (2000) Inference of population structure using multilocus genotype data. Genetics 155: 945–959.
- Falush D, Stephens M, Pritchard JK (2007) Inference of population structure using multilocus genotype data: dominant markers and null alleles. Mol Ecol 7(4): 574–578.
- Hubisz MJ, Falush D, Stephens M, Pritchard JK (2009) Inferring weak population structure with the assistance of sample group information. Mol Ecol Res 9: 1322–1332.
- Weir BS, Cockerham CC (1984) Estimating F-statistics for the analysis of population structure. Evolution 38(6): 1358–1370.
- Evanno G, Regnaut S, Goudet J (2005) Detecting the number of clusters of individuals using the software STRUCTURE: a simulation study. Mol Ecol 14: 2611–2620.
- Waples RS, Gaggiotti O (2006) What is a population? An empirical evaluation of some genetic methods for identifying the number of gene pools and their degree of connectivity. Mol Ecol 15: 1419–1439.
- Yamanaka K, Withler RE, Miller KM (2000) Structure of Yelloweye Rockfish (Sebastes ruberrimus) Populations in British Columbia. Canadian Stock Assessment Secretariat Research Document 172.
- 74. Yamanaka KL, Lacko LC, Withler RE, Grandin C, Lockhead JK, et al. (2006) A review of yelloweye rockfish (*Sebastes ruberrimus*) along the Pacific coast of Canada: biology, distribution and abundance trends. Canadian Science Advisory Secretariat Research Document 76.
- Siegle MR (2011) Population structure in yelloweye rockfish (Sebastes ruberrimus) driven by limited dispersal and selection. MSc Thesis. The University of British Columbia: Canada.
- Crawford WR (2002) Physical characteristics of Haida Eddies. J Oceanogr 58: 703–713.
- Yelland D, Crawford WR (2005) Currents in Haida eddies. Deep-Sea Res PT II 52: 875–892.
- Dower J, Freeland H, Juniper K (1992) A strong biological response to oceanic flow past Cobb Seamount. Deep-Sea Res 39: 1139–1145.
- Mullineaux LS, Mills SW (1997) A test of the larval retention hypothesis in seamount-generated flows. Deep-Sea Res PT I 44(5): 745–770.
- Allendorf FW, Phelps SR (1981) Use of allelic frequencies to describe population structure. CJFAS 38: 1507–1514.
- Conover DO, Clarke LM, Munch SB, Wagner GN (2006) Spatial and temporal scales of adaptive divergence in marine fishes and the implications for conservation. J Fish Biol 69: 21–47.
- Slatkin M (1987) Gene flow and the geographic structure of natural populations. Science 236: 787–792.
- Waples RS (1998) Separating the wheat from the chaff: patterns of genetic differentiation in high gene flow species. J Hered 89: 438–450.
- Hastings A, Botsford LW (2003) Comparing designs of marine reserves for fisheries and for biodiversity. Ecol Appl 13(1) Supplement: S65–S70.
- Palumbi SR (2003) Population genetics, demographic connectivity, and the design of marine reserves. Ecol Appl 13(1) Supplement: S146–S158.