

## Research



**Cite this article:** Kapsenberg L, Miglioli A, Bitter MC, Tambutté E, Dumollard R, Gattuso J-P. 2018 Ocean pH fluctuations affect mussel larvae at key developmental transitions.

*Proc. R. Soc. B* **285**: 20182381.

<http://dx.doi.org/10.1098/rspb.2018.2381>

Received: 24 October 2018

Accepted: 27 November 2018

**Subject Category:**

Development and physiology

**Subject Areas:**

developmental biology

**Keywords:**

ocean acidification, pH fluctuations, mollusc, development, shell field, trochophore

**Author for correspondence:**

L. Kapsenberg

e-mail: [kapsenberg@icm.csic.es](mailto:kapsenberg@icm.csic.es)

<sup>†</sup>Present address: Department of Marine Biology and Oceanography, CSIC Institute of Marine Sciences, Barcelona, Spain.

Electronic supplementary material is available online at <https://dx.doi.org/10.6084/m9.figshare.c.4325828>.

# Ocean pH fluctuations affect mussel larvae at key developmental transitions

L. Kapsenberg<sup>1,†</sup>, A. Miglioli<sup>1,3</sup>, M. C. Bitter<sup>4</sup>, E. Tambutté<sup>5</sup>, R. Dumollard<sup>2</sup> and J.-P. Gattuso<sup>1,6</sup>

<sup>1</sup>Laboratoire d'Océanographie de Villefranche, and <sup>2</sup>Laboratoire de Biologie du Développement de Villefranche, Sorbonne Université, CNRS, 181 chemin du Lazaret, 06230 Villefranche-sur-mer, France

<sup>3</sup>Dipartimento di Scienze della Terra, dell'Ambiente e della Vita, DISTAV, Università di Genova, Genova, Italy

<sup>4</sup>Department of Ecology and Evolution, University of Chicago, Chicago, IL, USA

<sup>5</sup>Marine Biology Department, Centre Scientifique de Monaco, 8 Quai Antoine 1er, MC98000, Monaco, Monaco

<sup>6</sup>Institute for Sustainable Development and International Relations, Sciences Po, 27 rue Saint Guillaume, 75007 Paris, France

LK, 0000-0002-7361-9061; MCB, 0000-0001-7607-2375

Coastal marine ecosystems experience dynamic fluctuations in seawater carbonate chemistry. The importance of this variation in the context of ocean acidification requires knowing what aspect of variability biological processes respond to. We conducted four experiments (ranging from 3 to 22 days) with different variability regimes (pH<sub>T</sub> 7.4–8.1) assessing the impact of diel fluctuations in carbonate chemistry on the early development of the mussel *Mytilus galloprovincialis*. Larval shell growth was consistently correlated to mean exposures, regardless of variability regimes, indicating that calcification responds instantaneously to seawater chemistry. Larval development was impacted by timing of exposure, revealing sensitivity of two developmental processes: development of the shell field, and transition from the first to the second larval shell. Fluorescent staining revealed developmental delay of the shell field at low pH, and abnormal development thereof was correlated with hinge defects in D-veligers. This study shows, for the first time, that ocean acidification affects larval soft-tissue development, independent from calcification. Multiple developmental processes additively underpin the teratogenic effect of ocean acidification on bivalve larvae. These results explain why trochophores are the most sensitive life-history stage in marine bivalves and suggest that short-term variability in carbonate chemistry can impact early larval development.

## 1. Introduction

Coastal marine ecosystems experience dynamic spatio-temporal variability in seawater inorganic carbonate chemistry [1]. Short-term variability occurs on top of baseline changes associated with ocean acidification. Ocean acidification causes a decrease in mean seawater pH and aragonite saturation state ( $\Omega_a$ ) via absorption of anthropogenic carbon dioxide (CO<sub>2</sub>) emissions [2]. As ocean acidification progresses, the acidity of naturally occurring low pH events is enhanced [3]. How the interplay of acidification and coastal pH variability will impact marine organisms requires understanding what aspect of pH variability influences biological processes that occur over similar time frames [4,5].

Fluctuations in coastal carbonate chemistry often occur on a diel period. Diel fluctuations result from daytime photosynthetic removal of CO<sub>2</sub> by phytoplankton and benthic photoautotrophs followed by night-time respiration of the whole community [6]. Depending on habitat characteristics, seagrasses, macroalgae and kelp forests can induce diel pH fluctuations that span a few

tenths to a full pH unit [5–8]. These short-term pH fluctuations are of similar magnitude as anthropogenic ocean acidification (–0.4 units pH by end of the century) [9], which is known to affect a myriad of biological processes [10]. In addition to baseline changes in carbonate chemistry, ocean acidification is expected to increase the magnitude of diel fluctuations [5].

The biological impact of fluctuating carbonate chemistry in the context of ocean acidification is understudied [11]. A few recent studies that address this issue show non-generalizable responses across calcifying taxa and species [11,12]. For molluscs, one the most studied taxa, growth appears largely indifferent to variability [13–16] (but see [17]) despite the fact that exposure to fluctuating conditions may be energetically costly [18]. Understanding impacts of variability requires knowing which parameter of carbonate chemistry drives a biological response and the reaction norm thereof. Given recent advances in these criteria for molluscs, here we focus on the early development of bivalve larvae.

The early development of bivalves is completed within 2–3 days after fertilization and is extremely sensitive to carbonate chemistry [19–22]. Specifically, abnormal development and reduced growth of D-veliger larvae occur under conditions of low  $\Omega_a$  or low substrate inhibitor ratio ( $SIR = [\text{HCO}_3^-]/[\text{H}^+]$ ) [23,24], which are tightly coupled in manipulations using  $\text{CO}_2$ . Previous work suggests that the first 24 h of development are insensitive to  $\text{CO}_2$  acidification and exposure thereafter drives abnormal development [21]. Subsequent shell growth is highly dependent on  $\Omega_a$  or SIR [23,24], due to the fact that larvae have limited control over carbonate chemistry at the site of calcification [25]. These findings suggest that early development of bivalves is comprised of process-specific sensitivities, and short-term fluctuations in carbonate chemistry typical of shallow coastal waters may be important to larval development. Despite its environmental relevance, only a few studies have investigated the impact of variable carbonate chemistry on early development of marine bivalves [13,14], with inconsistent outcomes across species [17]. For example, Frieder *et al.* [17] found that semi-diel variability enhanced larval growth in *Mytilus galloprovincialis*, but not in *M. californianus*. As the aforementioned studies do not control for timing of fluctuations, it remains unclear how fluctuating carbonate chemistry affects specific developmental processes in bivalve larvae.

The aim of this study was to identify what aspects of diel fluctuations in carbonate chemistry (e.g. timing, magnitude) are important to the early growth and development of the mussel *M. galloprovincialis*. Global aquaculture of *M. galloprovincialis* largely depends on natural recruitment for seed supply [26], so natural variability in seawater chemistry in the context of ocean acidification is highly relevant to the persistence of this industry. The first experiment (Exp. 1) tests the hypothesis that the magnitude of variability influences larval growth and development. Results from Exp. 1 informed the design of Exp. 2 and Exp. 3, which test the hypothesis that development is influenced by the timing of variable exposures. A fourth experiment (Exp. 4) was conducted to specifically test the hypothesis that  $\text{CO}_2$ -acidified seawater alters development of the trochophore shell field. For two of the four experiments, larvae from unique parental crosses were cultured in isolation to explore the biological variation of the response across parental pairs (Exp. 2 and Exp. 4).

## 2. Material and methods

For reading simplicity and based on methods of manipulation, we describe treatments in terms of pH, whereby low pH represents the full suite of chemical changes brought on by  $\text{CO}_2$ -acidification under stable temperature and salinity (electronic supplementary material, table S1). Low pH treatments were chosen to reach levels of aragonite undersaturation known to induce abnormal development and reduced larval growth [24].

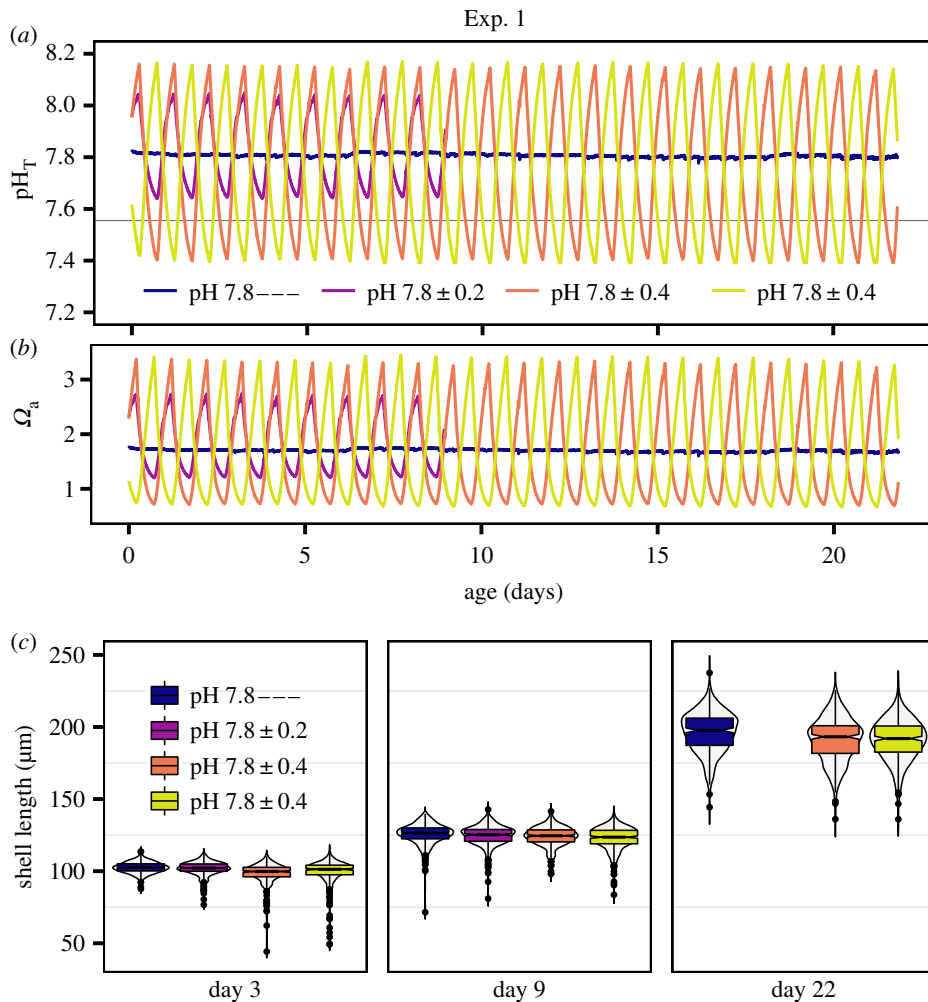
### (a) Experimental design

In the context of ocean acidification in dynamic coastal zones, a range of variable pH treatments was used ( $\text{pH}_T$  7.4–8.1). Exp. 1 assessed the impact of stable compared to variable pH treatments with the same mean pH (figure 1; electronic supplementary material, table S1). Four treatments were set up: control treatment of stable  $\text{pH}_T$  7.8 (pH 7.8—), and three variability treatments with a mean  $\text{pH}_T$  7.8 and a diel range of either 0.4 (pH  $7.8 \pm 0.2$ ), 0.8 (pH  $7.8 \pm 0.4$ ), or 0.8 offset by 12 h (pH  $7.8 \pm 0.4$ ), in order to control for the timing of variability exposures. Embryos from three unique families, using three males and nine females (1 male  $\times$  3 females, replicated three times with unique individuals), were cultured separately in the four treatments (12 cultures total,  $N = 3$  biological replicates). Larvae were sampled on day 3 (size and morphology), 9 (size) and 22 (size). Treatment pH  $7.8 \pm 0.2$  was discontinued after day 9 for technical reasons.

Variable pH treatments in Exp. 2 and Exp. 3 (figure 2a; electronic supplementary material, table S1) were designed to assess the effect of pH exposure at the start of shell morphogenesis, using a single diel pH fluctuation down from  $\text{pH}_T$  8.1 (pH  $8.1^-$ ) or up from  $\text{pH}_T$  7.4 (pH  $7.4^+$ ) centred around the start of calcification, approximately 30 hours post-fertilization (hpf). Stable  $\text{pH}_T$  exposures were used as controls (pH 8.1— and pH 7.4—) and larvae were sampled on day 3 (size and morphology). In Exp. 2, embryos from three unique pairs were cultured separately (12 cultures in total,  $n = 3$  biological replicates per treatment). In Exp. 3, embryos from nine unique pairs were pooled and then distributed across three replicate cultures (total of 12 cultures,  $n = 3$  technical replicates). In Exp. 4 (figure 2a), embryos from five unique pairs were cultured separately in  $\text{pH}_T$  8.1 and  $\text{pH}_T$  7.4 (10 cultures in total,  $n = 5$  biological replicates per treatment) and sampled at two time points: 35 hpf (fluorescent staining) and 68 hpf on day 3 (size, morphology, scanning electron microscopy [SEM]).

### (b) Larval cultures

Gravid *M. galloprovincialis* were collected from two sites during local spawning seasons. The first group was collected from a dock in Thau Lagoon (43.415°N, 3.688°E), a shallow semi-enclosed bay in Sète, France on 11 October 2016 (Exp. 1). The second group was collected from shallow buoy lines in the Bay of Villefranche-sur-Mer (43.682°N, 7.319°E), France on 9 February 2017 (Exp. 2–4). During the spawning season, pH variability in Thau Lagoon ranged from  $\text{pH}_T$  7.80 to 8.10 (L.K., 2016, unpublished data) and from  $\text{pH}_T$  8.10 to 8.15 in the Bay of Villefranche [27]. Adult mussels were kept in a single temperature-controlled flow-through sea table (approx. 16°C) and fed three times per week a mixture of Instant Algae (Reed Mariculture, Iso 1800 and Shellfish Diet 1800) until spawning. Experiments were conducted from January through March 2017 (see electronic supplementary material). Spawning was induced by cleaning the mussels of epibionts and warming seawater to 28°C. Sperm was kept on ice and eggs were kept at approximately 16°C until fertilization. Test fertilizations were performed to ensure gamete compatibility prior to batch fertilization. Fertilization was performed simultaneously across biological replicates, except for Exp. 4, which required 30 min



**Figure 1.** D-veliger shell size correlates to mean exposures: Exp. 1. (a) pH time series of Exp. 1, grey references  $\Omega_a = 1$ , hpf is hours post-fertilization. (b)  $\Omega_a$  time series of Exp. 1. (c) Shell size of larvae in Exp. 1 on days 3, 9 and 22. Box plots denote median, median confidence interval (notch width), quartiles and outliers (dots). Violin plots show the size distribution of the observations ( $n \approx 300$ ).

staggered fertilization by pair to facilitate sampling at precisely 35 hpf. Successful fertilization was determined by greater than 90% presence of polar bodies, after which embryos were transferred to culture vessels at a density of 14 embryos  $\text{ml}^{-1}$ . All experiments were run at 14.3°C (near February habitat conditions [27]) to maintain equal developmental rates. In Exp. 1, larvae were fed  $1 \times 10^8$  live cells of *Tisochrysis lutea* (CCAP 92714) once or twice per day after day 3.

### (c) Experimental system and seawater chemistry

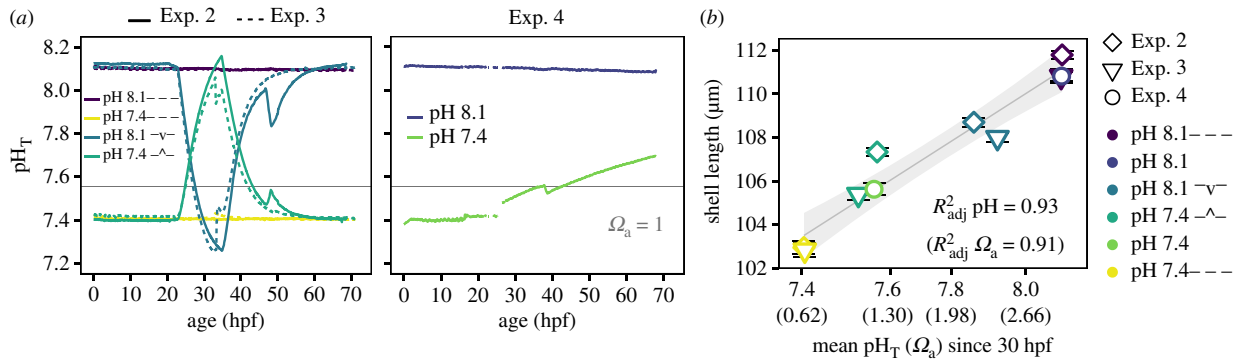
Larvae were cultured in a temperature-controlled pH variability system described in Kapsenberg *et al.* [28]. Four header tanks (35 l) were supplied with seawater pumped from 5 m depth in the Bay Villefranche, filtered to 0.35  $\mu\text{m}$  and UV-sterilized (FSW). Header tank pH was manipulated via the addition of pure  $\text{CO}_2$  gas using a glass pH electrode feedback system (IKS Aquastar) and constant aeration with  $\text{CO}_2$ -free air. Treatment water was pumped from header tanks to culture buckets ( $n = 3$  per header tank, see electronic supplementary material for control of header tank effects) using irrigation drippers ( $21 \text{ h}^{-1}$ ). Variable pH treatments switched header tank pH every 12 h, resulting in smooth pH oscillations in cultures as documented by Honeywell Durafet III pH sensors (for performance quality see electronic supplementary material). Exp. 4 consisted of static cultures (sourced from two independent header tanks of  $\text{pH}_T$  8.1 and  $\text{pH}_T$  7.4) due to the use of calcein dye. Water samples were collected from each treatment header tank for salinity (Mettler Toledo SevenEasy Conductivity) and total alkalinity ( $A_T$ ) every

2–3 days for Exp. 1, every 2 days for Exp. 2, every day during Exp. 3, and once at the start of Exp. 4. Samples for  $A_T$  were run in duplicate using open cell titration (Metrohm Titrando 888) [29]. Accuracy of  $A_T$  measurements ranged between  $-7$  and  $+5 \mu\text{mol kg}^{-1}$  as compared to certified reference material (A. Dickson, Scripps Institution of Oceanography). Precision of  $A_T$  measurements was  $2 \mu\text{mol kg}^{-1}$  (mean standard deviation of all duplicate measurements,  $n = 68$ ).  $\Omega_a$  and  $p\text{CO}_2$  were calculated from  $\text{pH}_T$  using mean temperature, and header tank salinity and  $A_T$  per treatment, per experiment (electronic supplementary material, table S1). Carbonate system calculations were performed in RStudio (version 1.0.143) using the seacarb R package [30], with constants  $K_1$  and  $K_2$ ,  $K_f$ , and  $K_s$  from Lueker *et al.* [31], Perez *et al.* [32] and Dickson [33], respectively.

### (d) Shell size, morphology and SEM

Shell size was determined as the maximum shell length parallel to the hinge, using bright field microscopy and IMAGEJ software ( $n \geq 100$  culture $^{-1}$ ). On day 3, larval morphology was scored ( $n \geq 100$  culture $^{-1}$ ) according to His *et al.* [34]: normal (D-shaped shell with a straight hinge), trochophore (undeveloped larvae, often shell-less), abnormal hinge (concave D-shaped shell), protruding mantle (D-shaped shell with protruding tissue or velum), and combined abnormal hinge and protruding mantle (D-shaped shell with both abnormalities). For SEM imaging, D-veligers were preserved in 100% EtOH. Prior to imaging (JEOL 6010LV), shells were cleaned (rinses of 2 min tap water, 5 min R.O.  $\text{H}_2\text{O}$  with 1% bleach, 2 min tap water, 5 s





**Figure 2.** D-veliger shell size correlates to mean exposures: Exp. 2–4. (a) pH time series of Exp. 2–4, grey references  $\Omega_a = 1$ , hpf is hours post-fertilization. (b) Mean shell size ( $n \approx 300$ ,  $\pm$  s.e.) of D-veligers from Exp. 2–4 (symbols) on day 3 plotted against the treatment-respective mean pH exposure since the start of calcification (30 hpf) to sampling time (68–72 hpf). Mechanistically, as calcification is driven by  $\Omega_a$  (or SIR) and not pH, the corresponding  $\Omega_a$  data is shown in parentheses ( $x$ -axis,  $R^2_{adj}$ ). Colours indicate pH treatments in (a). Linear regression for pH is in grey ( $\pm$  s.e.).

ddH<sub>2</sub>O, temporary storage in 100% EtOH), dried at 42°C, and sputter coated with gold.

### (e) Fluorescent staining

To visualize the first CaCO<sub>3</sub> precipitation, calcein (0.001 M, final concentration) was added to static cultures of Pair 2 and 4 at 25 hpf in Exp. 4 [35]. Pilot experiments using calcein dye showed that the first calcification occurred during the early trochophore stage at 30 hpf in 14.3°C. Immediately after calcein addition, motorized paddles were turned on in all culture vessels, resulting in slight CO<sub>2</sub> off-gassing in pH<sub>T</sub> 7.4. Calcein culture pH was checked daily using a glass pH electrode which was compared against a calibrated Durafet (addition of calcein caused a 0.03 unit pH<sub>T</sub> increase in the high pH treatment and 0.10 unit pH<sub>T</sub> decrease in the low pH treatment).

To visualize the organic matrix of trochophores in Exp. 4, larvae from all pairs were sampled and stained live with calcofluor (Calcofluor White M2R, #F3543 from Sigma-Aldrich) at 35 hpf. Calcofluor is a fluorochrome that binds to chitin (and cellulose) and has been used to study chitin in adult abalone [36]. Its precision for identifying chitin from other matrix molecules has been debated [37] and the results presented here are restricted to assessing the shape and extent of the organic matrix in general. Concentrated larvae were stained by calcofluor in FSW for 5 min (final exposure of 1:50000 w/v by diluting a stock solution of 1:100 w/v in DMSO, stored at -20°C), washed three times with FSW, fixed with a drop of 4% paraformaldehyde, and immediately imaged on a confocal microscope (Leica SP8, electronic supplementary material). Images were 3D rendered and composite images were rotated for each larva to measure the area of one valve stained by calcein and calcofluor via manual drawing using IMAGEJ software ( $n = 13$ –42 culture<sup>-1</sup>).

### (f) Statistical analyses

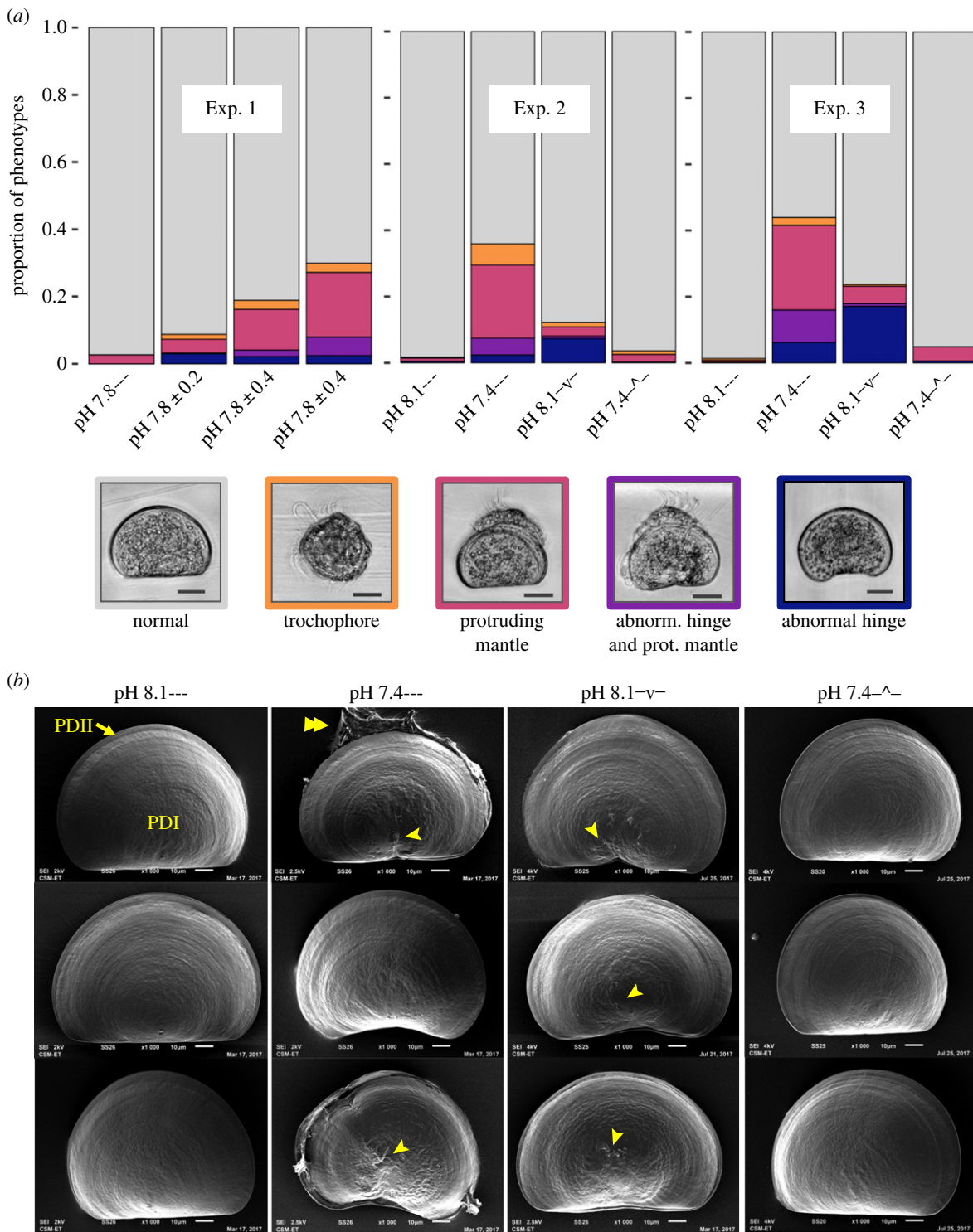
Data analysis was performed in RSTUDIO (v. 1.0.143) [38]. Residuals of larval shell size data were not normally distributed, violating assumptions required for ANOVAs. For Exp. 1, a two-factor permutation analysis of variance was used with 1000 permutations in RVAideMemoire R package [39]. Treatment and age were fixed factors (interaction was not significant and removed from the final model) and family was used as a blocking factor. Size data from the first 100 larval measurements were used and data from treatment pH 7.8  $\pm$  0.2, stopped after day 9, was excluded. For Exp. 2–4, a linear regression was used to assess the impact of mean pH exposure on mean larval size during the shell growing period (30–68 hpf). Exp. 2–4 had different biological design (isolating versus pooling larvae from unique parental pairs), so a linear regression was performed on

overall mean larval size ( $n \approx 300$  treatment<sup>-1</sup> experiment<sup>-1</sup>). Cultures with calcein dye in Exp. 4 were excluded as these cultures did not contain a Durafet by which to calculate the mean pH exposure. For Exp. 1 and Exp. 2 (family and pair replication), proportions of larvae with normal development was analysed using a generalized linear mixed effects model with treatment as a fixed effect and family or pair as a random effect, using the lmer R package [40]. Significance of the fixed effect was tested against the null model using a likelihood ratio test. For Exp. 3 (bucket replication), treatment effect was assessed using a one-way ANOVA. Analyses were repeated for Exp. 2 and Exp. 3 for proportion of larvae with an abnormal hinge (irrespective of a protruding mantle). All model residuals exhibited a normal distribution (Shapiro–Wilk normality test) and equal variance (Levene’s test). Least-squares means were used for *post hoc* pairwise contrasts of treatment effects for Exp. 2 and Exp. 3, using a Bonferroni correction for six comparisons in the lsmeans R package [41].

## 3. Results

### (a) Shell size responds to mean conditions

The impact of variable carbonate chemistry on shell growth was analysed by comparing shell length of larvae reared under treatments of either constant pH<sub>T</sub> 7.8 or mean pH<sub>T</sub> 7.8 with a total diel range of 0.4 or 0.8 units pH<sub>T</sub> in Exp. 1 (figure 1). Shell length was not significantly affected by a 0.8 unit pH<sub>T</sub> diel variation over the course of a three-week exposure (figure 1c; treatment effect:  $F_{1,2} 3.33$ ,  $p = 0.093$ ; shell size increased over time,  $F_{1,2} 1712.68$ ,  $p < 0.001$ ), which suggests that shell length is a function of mean exposures. This was empirically tested and verified in Exp. 2–4, using a combination of stable and variable treatments spanning a range of mean exposures (figure 2). Shell length was significantly correlated to mean exposures during the shell growing period, 30 hpf until the sampling time on day 3 (for pH:  $F_{1,8} 119.6$ ,  $p < 0.0001$ ; for  $\Omega_a$ :  $F_{1,8} 92.57$ ,  $p < 0.0001$ ). Mean pH<sub>T</sub> ( $\Omega_a$ ) explained 93% (91%) of the variance in average shell size of larvae across treatments, respectively (figure 2b). This correlation was independent of developmental effects (abnormal larvae are smaller than normal D-veligers in low pH treatments; electronic supplementary material, figure S1 and table S2) as the correlation was maintained for larvae in the top 10th percentile size class (for  $\Omega_a$ ,  $F_{1,8} 32.27$ ,  $p < 0.001$ ,  $R^2_{adj} = 0.78$ ).



**Figure 3.** Variable pH alters development of mussel larvae. (a) Proportion of larval phenotypes on day 3 in Exp. 1, Exp. 2 and Exp. 3. Treatment labels match those in figures 1a and 2a. Colour coding follows the border colour of representative phenotype images (scale bar = 30  $\mu\text{m}$ ). Per experiment, proportions of specific phenotypes were calculated from summed observations across three cultures per treatment ( $n \approx 300$ ; see text for design of biological replication of cultures). (b) SEM images of D-veliger shells observed on day 3 in Exp. 3 (scale bar = 10  $\mu\text{m}$ , brightness and contrast was adjusted per image). Examples of three larvae were chosen to highlight phenotypes associated with treatment effects. Prodissoconch I (PDI) and II (PDII) is indicated for a larva from pH 8.1— (top). Examples from pH 7.4— and pH 8.1<sup>-v-</sup> show larvae with abnormal hinges and scarring at the centre of PDI (arrowhead). Larva with a protruding mantle is indicated by a double arrowhead.

### (b) Larval development depends on timing of exposure to acidified seawater

In Exp. 1, abnormal development increased with the magnitude of the diel pH range (figure 3a;  $p$ -value < 0.007 for all pairwise comparisons, electronic supplementary material, table S3). The increase in abnormal development across treatments with diel cycles offset by 12 h (pH 7.8  $\pm$  0.4 versus pH

7.8  $\pm$  0.4;  $p = 0.0013$ ) suggests that the timing of low pH conditions was an important factor for developmental outcomes. This was empirically tested and verified in Exp. 2 and Exp. 3 (electronic supplementary material, tables S4 and S5). In both Exp. 2 and Exp. 3, the most striking result was that  $\geq 95\%$  of larvae developed normally in both pH 8.1— and pH 7.4—<sup>-^</sup> (figure 3a), whereas larvae in pH 8.1<sup>-v-</sup> exhibited 76–88% normal development, mostly due to the presence of abnormal

hinge phenotypes. In pH 7.4—, only 56–64% of larvae exhibited normal development due to frequent observations of larvae with protruding mantles and abnormal hinges. Larvae with both an abnormal hinge and protruding mantle were most prevalent in pH 7.4— (5.1% in Exp. 2, and 9.8% in Exp. 3), wherein 60–70% of larvae with a hinge abnormality also exhibited a protruding mantle. The combination of an abnormal hinge and a protruding mantle was observed in less than 1% of larvae in pH 8.1<sup>-v-</sup> (due to few protruding mantles), and never in pH 8.1— and pH 7.4-^-(due to absence of abnormal hinges). The temporal partitioning of unfavourable seawater chemistry exposure and associated abnormal phenotypes suggests that abnormalities in the hinge and mantle are additive. This additive effect was also apparent in the shell size of these phenotypes (electronic supplementary material, figure S1): larvae with an abnormal hinge, protruding mantle or both are 1, 4 and 6% smaller, respectively, than normal D-veligers from the same low pH treatment (electronic supplementary material, table S2).

### (c) Abnormal development of the shell field is correlated to hinge abnormalities

In both Exp. 2 and Exp. 3, the proportion of larvae with an abnormal hinge (irrespective of a protruding mantle) was significantly greater in treatments that experienced low pH conditions around 30 hpf (7–8% in Exp. 2, 16–18% in Exp. 3, in pH 7.4— and 8.1<sup>-v-</sup>) compared to those that experienced high pH at this time (less than 1% in pH 8.1— and pH<sub>T</sub> 7.4-^-(pairwise comparisons  $p$ -values  $\leq 0.004$ ; electronic supplementary material, tables S6 and S7) and equal between treatments which experienced the same pH at that time ( $p$ -values = 1.0; electronic supplementary material, tables S6 and S7). Therefore, hinge abnormalities were induced by exposure to unfavourable low pH conditions around 30 hpf (27–35 hpf), independent of prior or later exposures. Larvae exposed to unfavourable carbonate chemistry during this period frequently exhibited irregular texture and scarring in the centre of the first larval shell (PDI, prodissoconch I) (e.g. figure 3b; in pH 7.4— and pH 8.1<sup>-v-</sup> in Exp. 3). In contrast, D-veliger shells of larvae reared in pH 8.1— and pH 7.4-^- were indistinguishable (figure 3b).

The developmental process underpinning the abnormal shell development was investigated in Exp. 4. In pH<sub>T</sub> 7.4 at 35 hpf, the organic matrix of some larvae exhibited an indentation along the hinge line (figure 4a). In contrast, in pH<sub>T</sub> 8.1, this chitinous isthmus was straight and well developed. The proportion of larvae with an indented matrix at 35 hpf was highly correlated to the proportion of D-veligers with an abnormal hinge at 68 hpf (figure 4b;  $F_{1,8} = 47.89$ ,  $p = 0.0001$ ,  $R^2_{\text{adj}} = 0.84$ ). Larvae from Pair 1 exhibited greater than 99% normal development in pH<sub>T</sub> 7.4 at both trochophore (35 hpf) and D-veliger stage (68 hpf), with normal flat shell matrices in both pH<sub>T</sub> 8.1 and pH<sub>T</sub> 7.4 at 35 hpf (figure 4c). In contrast, 55% of larvae from Pair 2 exhibited matrix indentations in pH<sub>T</sub> 7.4 ( $n = 12$ , of 22 larvae), and 35% of D-veligers exhibited hinge abnormalities by day 3 ( $n = 42$ , of 121 larvae). While some variation in the correlation between phenotypes at 35 and 68 hpf may be related to sample size, these data, along with results from Exp. 2 and 3, provide strong evidence that matrix

indentations persists throughout development and cause hinge abnormalities in D-veligers.

Matrix indentations occurred in the presence and absence of CaCO<sub>3</sub> (electronic supplementary material, figure S2), suggesting that abnormal shell field development is a phenotype that is fixed prior to calcification. At 35 hpf, only 87% of pH<sub>T</sub> 7.4 larvae had started calcification ( $n = 40$  out of 46). Shell abnormalities in pH<sub>T</sub> 7.4 were visible as keyhole-shaped indentations near the hinge, indicating no calcification in that region by 35 hpf (figure 4a). In pH<sub>T</sub> 8.1, this abnormality was not observed and all larvae exhibited substantial shell growth ( $n = 58$ ). Both the organic matrix and larval shell were larger for larvae reared in pH<sub>T</sub> 8.1 compared with those in pH<sub>T</sub> 7.4 at 35 hpf (electronic supplementary material, figure S3), regardless of the normal or abnormal developmental trajectory of the larval cohort (visually apparent in figure 4a,c).

### (d) Protruding mantle indicates sensitivity of the PDI-II transition

The protruding mantle phenotype occurred during the early D-veliger stage around 40 hpf. In Exp. 2 and 3, larvae with protruding mantles (irrespective of an abnormal hinge) were associated with low pH exposure, with approximately 30% observed in pH 7.4— and  $\leq 1\%$  in pH 8.1— (figure 3a). In contrast, treatments pH 7.4-^- and pH 8.1<sup>-v-</sup> induced a low occurrence of protruding mantle (respectively, 2.3 versus 3.5% in Exp. 2; 4.3 versus 6.1% in Exp. 3) indicating that this phenotype was induced by low pH at a time when pH in both treatments were similar. There are only two periods where this occurs, near 27 and 40 hpf. The most relevant time-point is 40 hpf, as embryos are resistant to CO<sub>2</sub> acidification during the first day of development [21], and developmental processes around 40 hpf relate to the shell–mantle interface at the transition from the first (PDI) to the second larval shell (PDII).

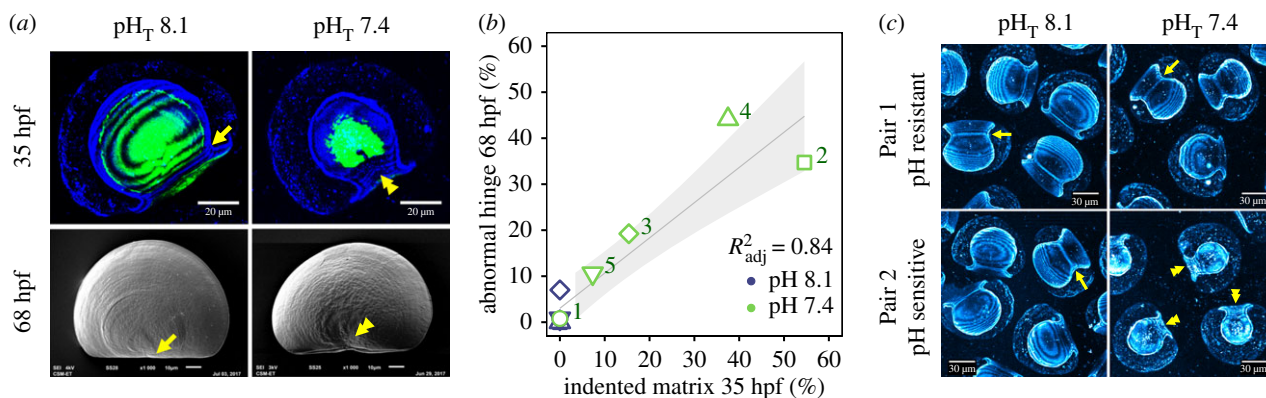
## 4. Discussion

The aim of this study was to identify which aspect of mussel larval development is influenced by temporal variability of carbonate chemistry. By using a range of variability treatments, we show that shell growth responds to mean exposures while development depends on time-sensitive exposures (Exp. 1–4), and these processes additively contribute to overall D-veliger phenotypes. Based on timing of unfavourable conditions, sensitive developmental processes were identified and linked to specific abnormal phenotypes (Exp. 2–3), with mechanistic evidence at the tissue layer (Exp. 4). Electronic supplementary material, figure S4 details the developmental timeline and windows of sensitivity.

### (a) Mussel larval growth and development

Larval shell growth was driven by mean exposures regardless of variability regimes (figures 1c and 2b). This reflects the instantaneous nature of  $\Omega_a$ -dependent precipitation kinetics in bivalve larvae [42], whereby the benefit of high  $\Omega_a$  is neutralized by the negative effect of low  $\Omega_a$ . This has also been observed in other mollusc larvae, including abalone, hard clams, oysters and scallops [13–15], although oddly not for *M. galloprovincialis* from California [17]. Frieder





**Figure 4.** Relationship between organic matrix morphology and D-veliger shell phenotype in Exp. 4. (a) Images of larvae in pH<sub>T</sub> 8.1 (straight hinge, arrow) and pH<sub>T</sub> 7.4 (indented matrix and hinge, double arrowhead). Composite confocal images show larvae at 35 hours post-fertilization (hpf; organic matrix stained in blue, CaCO<sub>3</sub> stained in green, scale bar = 20 μm) and SEM show larvae at 68 hpf (scale bar = 10 μm). Larvae are of unique individuals from Pair 2 and were selected to best visualize the correlation shown in (b). (b) Percentage of larvae with an indented organic matrix at 35 hpf ( $n = 13–42$ ) correlated with percentage of D-veliger larvae exhibiting an abnormal hinge at 68 hpf ( $n = 109–155$ ), across five unique male–female pairs (symbol, numbered for pH<sub>T</sub> 7.4). Linear regression line is grey ( $\pm$  s.e.). (c) Trochophore larvae (Pair 1 and Pair 2) reared in pH<sub>T</sub> 8.1 and pH<sub>T</sub> 7.4 and stained with calcofluor (cyan) at 35 hpf. Colour was adjusted equally across images for clarity of the hinge line (scale bar = 30 μm). Arrows denote normal organic matrices and hinge lines (arrow) and indented matrix (double arrowhead). Ring-like pattern in staining is an artefact of confocal imaging larvae through multiple slices.

*et al.* [17] found that *M. galloprovincialis* larvae increased growth in low pH treatments when semi-diurnal fluctuations were included. Culture temperatures and larval growth rates in control treatments in their study and ours are comparable (approx. 125 μm; day 8 in  $\Omega_a$  1.9 at 15.7°C versus day 9 in  $\Omega_a$  1.7 at 14.4°C in our study). As Frieder *et al.* [17] conducted a single experiment with few replicate measurements, further experiments will be necessary to identify the source of these contrasting observations.

In terms of development, we identified two processes that, when they occur during conditions of low pH (low  $\Omega_a$  and SIR), give rise to abnormal D-veligers: (i) formation of the shell field prior to calcification around 30 hpf, and (ii) velum retraction around 40 hpf prior to the transition from PDI to PDII.

Exposure to low pH conditions around 30 hpf produced D-veligers with an abnormal hinge (figure 3, Exp. 2 and 3). The major developmental process occurring at this time is the formation of the shell field during the mid-trochophore stage (electronic supplementary material, figure S4). This process is initiated in the early trochophore stage by the invagination of a group of ectoderm cells, which create a pore. This brings together a rosette of outer surface cells that secrete what appears to eventually become the periostracum [43]. The pore closes and deeper invaginated cells then evaginate back to the surface epithelium where, under control conditions, they create a flat region that expands via mitotic division, under the expanding periostracum [43–45]. During this process, shell field cells exude a chitin-based organic matrix wherein calcification takes place [43,46,47]. In low pH treatments, the organic matrix often exhibited a central indentation at the site of the shell field invagination (figure 4). At 35 hpf, the lack of calcification in this area and scarring in the centre of PDI on D-veligers suggests a treatment effect on cells associated with the shell field evagination (figure 3b). Larvae in pH 7.4– $\wedge$ – calcified smooth shells, despite being in unfavourable carbonate chemistry from approximately 40 hpf onward. This indicates that the effect of unfavourable carbonate chemistry at the start of calcification (approx. 30 hpf), which generated abnormal

shell texture, is different from the effect of low  $\Omega_a$  that drives the rate of linear, but smooth, extension of the shells after this period [24]. Similar distortions in PDI have previously been observed in oyster larvae of *Ostrea edulis* and linked to the process of the shell field evagination, although this was not in the context of environmental conditions [48]. Hinge abnormalities may result from abnormal or incomplete restructuring of the ectoderm during development of the shell field, probably prior to calcification. The resulting abnormal trochophore body shape alters the calcification blueprint, whereby the shell simply takes on the shape of the cellular landscape over which the organic matrix is exuded, thereby producing D-veligers with an indented hinge.

Within low pH treatments, the abnormal hinge phenotype is only 1% smaller than normal D-veligers (electronic supplementary material, figure S1). In Exp. 1, abnormal hinge larvae were still present on day 9 (L.K. 2017, personal observation). By day 22, however, curvature of the shell masked the angle of the shell hinge, so there is no evidence to infer the effect of the abnormal hinge phenotype on larval fitness.

Regardless of normal or abnormal shell field development, the organic matrix and calcified area at 35 hpf was consistently smaller in low pH treatments, which we interpret as developmental delay. Developmental delay likely occurs from exposures during the early trochophore stage (electronic supplementary material, figure S4). The energetic cost of building the protein-rich organic matrix is much greater than the cost of external calcification [49,50]. In oyster larvae, low pH causes a decrease in the protein deposition efficiency [51]. Developmental delay of the shell field may thus stem from protein production issues associated with building the organic matrix. Consequently, this could also delay the onset of calcification.

Exposure to low pH conditions near 40 hpf was linked to increased frequency of D-veligers with protruding mantle tissue or velum (irrespective of an abnormal hinge). At this time, the extension of the calcified PDI shell has caught up with the leading edge of the expanding organic matrix, which

then both fully cover the larval body (L.K. 2017, personal observation). Around 47 hpf, larvae gain the ability to retract their velum and close their shell (a behavioural response to ethanol exposure; L.K. 2016, personal observation). Shell closure marks the transition from PDI growth to concentric growth lines distinctive of PDII (electronic supplementary material, figure S5) [48]. Protruding tissue is evidence of either abnormal tissue development or inability to retract the velum. It is likely that these larvae cannot progress in development to PDII growth. Such a developmental arrest is evident in shell size. D-veligers with protruding tissue tend to be 4–6% smaller than normal D-veligers (electronic supplementary material, figure S1). Previous experiments showed that protruding mantle phenotypes of *M. galloprovincialis* lacked PDII growth, even after a 5-day exposure to  $\Omega_a$  0.49 [21], which suggests that this phenotype is terminal.

From this study, it is clear that CO<sub>2</sub>-acidified seawater impacts development of the larval body in multiple ways, independent from impacts on calcification (electronic supplementary material, figure S4). This study raises the question as to how seawater CO<sub>2</sub>-acidification disrupts tissue development in mussel larvae and highlights the need to identify mechanisms of ocean acidification impacts at a cellular level.

### (b) Environmental and global change context

Our results suggest that developmental success of mussel larvae will be unpredictable in habitats with high variability in carbonate chemistry. Cues that determine the moment of a spawning event in the field are not well understood. The mussel population in Villefranche appears to spawn out during storms, and mussels have spawned in a bucket on the boat, which suggests that the final trigger may be physical and unrelated to time of day. Temperature-dependent developmental rates could also influence when a given larval cohort will be sensitive to carbonate chemistry. As shell growth depends on mean conditions, areas with high mean  $\Omega_a$ , regardless of variability regimes, may significantly benefit normally-developed D-veliger larvae, if larvae are retained in this body of water throughout their pelagic phase.

Normal larval development in pH<sub>T</sub> 7.4 ranged from 30% to 99% across unique parental pairs (electronic supplementary material, figures S6 and S7). Such biological variation is common among bivalves and may facilitate adaptation to global change [52,53]. Consequently, as ocean acidification progresses, dormant genotypes previously unselected for in natural populations may be favoured via new natural selection pressures operating on early development [54].

Although we did not test for this, evidence for adaptation can be found via population comparisons between sites with different pH regimes [55]. Overall, future recruitment of *M. galloprovincialis* is likely to increase in variance as emerging factors such as local carbonate chemistry variability, timing of spawning, parental effects, and other co-occurring global change stressors gain importance in determining successful development of a larval cohort.

## 5. Conclusion

By controlling the timing of low pH events, this study partitions the effect of ocean acidification on mussel larval growth from that on development. Larval growth is a function of mean exposures. This extends previous research on the  $\Omega_a$ -dependency of calcification in bivalve larvae [23,24] by demonstrating that calcification responds instantaneously to changes in seawater chemistry. Independent from calcification, abnormal development was driven by sensitivity to low pH conditions during specific soft-tissue developmental processes: (i) formation of the shell field and (ii) transition from PDI to PDII. This is the first study documenting ocean acidification sensitivity in soft tissues of bivalve larvae and how formation thereof determines D-veliger morphology. These additive and short-lived processes explain why trochophores are the most sensitive life-history stage in marine bivalves.

**Ethics.** Number of adult mussels used and duration of stress was minimized.

**Data accessibility.** Data are archived on BCO-DMO, <https://www.bco-dmo.org/project/720349>.

**Authors' contributions.** L.K. designed the research with input from A.M., M.C.B. and J.-P.G. L.K., A.M., M.C.B. and E.T. performed research. R.D. and J.-P.G. contributed resources. L.K. analysed the data and wrote the paper with contributions from all authors.

**Competing interests.** The authors declare they have no competing interests.

**Funding.** This research was funded by the US National Science Foundation (NSF; OCE-1521597 to L.K.). L.K. was also supported by the European Commission Horizon 2020 Marie Skłodowska-Curie Action (No. 747637). M.C.B. was supported by US Department of Education (Grant No. 200A150101) and NSF Graduate Research Fellowship (No. 1000198423). A.M. was supported by an Erasmus+ traineeship scholarship (University of Genova) and by the Observatoire Océanologique de Villefranche.

**Acknowledgements.** The authors thank technical support at the Laboratoire d'Océanographie de Villefranche: Samir Alliouane, Laurent Gilletta, Régis Lasbleiz, Guillaume de Liège and David Luquet. We are grateful for training from Janet Chenevert and Mebarek Temagoult, and advice from Frédéric Gazeau and Daniel K. Okamoto.

## References

- Hofmann GE *et al.* 2011 High-frequency dynamics of ocean pH: a multi-ecosystem comparison. *PLoS ONE* **6**, e28983. (doi:10.1371/journal.pone.0028983)
- Rhein M *et al.* 2013 Observations: ocean. In *Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change* (eds TF Stocker *et al.*), pp. 255–315. Cambridge, UK: Cambridge University Press.
- Feely RA, Sabine CL, Hernandez-Ayon JM, Ianson D, Hales B. 2008 Evidence for upwelling of corrosive 'acidified' water onto the continental shelf. *Science* **320**, 1490–1492. (doi:10.1126/science.1155676)
- Waldbusser GG, Salisbury JE. 2014 Ocean acidification in the coastal zone from an organism's perspective: multiple system parameters, frequency domains, and habitats. *Annu. Rev. Mar. Sci.* **6**, 221–247. (doi:10.1146/annurev-marine-121211-172238)
- Pacella SR, Brown CA, Waldbusser GG, Labiosa RG, Hales B. 2018 Seagrass habitat metabolism increases short-term extremes and long-term offset of CO<sub>2</sub> under future ocean acidification. *Proc. Natl Acad. Sci. USA* **115**, 3870–3875. (doi:10.1073/pnas.1703445115)
- Kapsenberg L, Hofmann GE. 2016 Ocean pH time-series and drivers of variability along the northern Channel Islands, California, USA. *Limnol. Oceanogr.* **61**, 953–968. (doi:10.1002/lno.10264)



7. Baumann H, Wallace RB, Tagliaferri T, Gobler CJ. 2015 Large natural pH, CO<sub>2</sub> and O<sub>2</sub> fluctuations in a temperate tidal salt marsh on diel, seasonal, and interannual time scales. *Estuaries Coast.* **38**, 220–231. (doi:10.1007/s12237-014-9800-y)
8. Wahl M *et al.* 2018 Macroalgae may mitigate ocean acidification effects on mussel calcification by increasing pH and its fluctuations. *Limnol. Oceanogr.* **63**, 3–21. (doi:10.1002/lno.10608)
9. Gattuso J-P *et al.* 2015 Contrasting futures for ocean and society from different anthropogenic CO<sub>2</sub> emissions scenarios. *Science* **349**, aac4722. (doi:10.1126/science.aac4722)
10. Kroeker KJ *et al.* 2013 Impacts of ocean acidification on marine organisms: quantifying sensitivities and interaction with warming. *Glob. Change Biol.* **19**, 1884–1896. (doi:10.1111/gcb.12179)
11. Rivest EB, Comeau S, Cornwall CE. 2017 The role of natural variability in shaping the response of coral reef organisms to climate change. *Curr. Clim. Change Rep.* **3**, 271–281. (doi:10.1007/s40641-017-0082-x)
12. Cornwall CE *et al.* 2018 Resistance of corals and coralline algae to ocean acidification: physiological control of calcification under natural pH variability. *Proc. R. Soc. B* **285**, 20181168. (doi:10.1098/rspb.2018.1168)
13. Clark HR, Gobler CJ. 2016 Diurnal fluctuations in CO<sub>2</sub> and dissolved oxygen concentrations do not provide a refuge from hypoxia and acidification for early-life-stage bivalves. *Mar. Ecol. Prog. Ser.* **558**, 1–14. (doi:10.3354/meps11852)
14. Gobler CJ, Clark HR, Griffith AW, Lusty MW. 2017 Diurnal fluctuations in acidification and hypoxia reduce growth and survival of larval and juvenile bay scallops (*Argopecten irradians*) and hard clams (*Mercenaria mercenaria*). *Front. Mar. Sci.* **3**, 282. (doi:10.3389/fmars.2016.00282)
15. Onitsuka T *et al.* 2018 Effects of ocean acidification with pCO<sub>2</sub> diurnal fluctuations on survival and larval shell formation of Ezo abalone, *Haliotis discus hannai*. *Mar. Environ. Res.* **134**, 28–36. (doi:10.1016/j.marenvres.2017.12.015)
16. Keppel AG, Breitbart DL, Wikfors GH, Burrell RB, Clark VM. 2015 Effects of co-varying diel-cycling hypoxia and pH on disease susceptibility in the eastern oyster *Crassostrea virginica*. *Mar. Ecol. Prog. Ser.* **538**, 169–183. (doi:10.3354/meps11479)
17. Frieder CA, Gonzalez JP, Bockmon EE, Navarro MO, Levin LA. 2014 Can variable pH and low oxygen moderate ocean acidification outcomes for mussel larvae? *Glob. Change Biol.* **20**, 754–764. (doi:10.1111/gcb.12485)
18. Mangan S, Urbina MA, Findlay HS, Wilson RW, Lewis C. 2017 Fluctuating seawater pH/pCO<sub>2</sub> regimes are more energetically expensive than static pH/pCO<sub>2</sub> levels in the mussel *Mytilus edulis*. *Proc. R. Soc. B* **284**, 20171642. (doi:10.1098/rspb.2017.1642)
19. Gazeau F *et al.* 2013 Impacts of ocean acidification on marine shelled molluscs. *Mar. Biol.* **160**, 2207–2245. (doi:10.1007/s00227-013-2219-3)
20. Gaylord B *et al.* 2011 Functional impacts of ocean acidification in an ecologically critical foundation species. *J. Exp. Biol.* **214**, 2586–2594. (doi:10.1242/jeb.055939)
21. Kurihara H, Asai T, Kato S, Ishimatsu A. 2008 Effects of elevated pCO<sub>2</sub> on early development in the mussel *Mytilus galloprovincialis*. *Aquat. Biol.* **4**, 225–233. (doi:10.3354/ab00109)
22. Talmage SC, Gobler CJ. 2010 Effects of past, present, and future ocean carbon dioxide concentrations on the growth and survival of larval shellfish. *Proc. Natl Acad. Sci. USA* **107**, 17 246–17 251. (doi:10.1073/pnas.0913804107)
23. Thomsen J, Haynert K, Wegner KM, Melzner F. 2015 Impact of seawater carbonate chemistry on the calcification of marine bivalves. *Biogeosciences* **12**, 4209–4220. (doi:10.5194/bg-12-4209-2015)
24. Waldbusser GG *et al.* 2015 Saturation-state sensitivity of marine bivalve larvae to ocean acidification. *Nat. Clim. Change* **5**, 273–280. (doi:10.1038/nclimate2479)
25. Ramesh K, Hu MY, Thomsen J, Bleich M, Melzner F. 2017 Mussel larvae modify calcifying fluid carbonate chemistry to promote calcification. *Nat. Commun.* **8**, 1709. (doi:10.1038/s41467-017-01806-8)
26. FAO. 2004 *Mytilus galloprovincialis*. See [www.fao.org/fishery/culturedspecies/Mytilus-galloprovincialis/en](http://www.fao.org/fishery/culturedspecies/Mytilus-galloprovincialis/en).
27. Kapsenberg L, Alliouane S, Gazeau F, Mousseau L, Gattuso JP. 2017 Coastal ocean acidification and increasing total alkalinity in the northwestern Mediterranean Sea. *Ocean Sci.* **13**, 411–426. (doi:10.5194/os-13-411-2017)
28. Kapsenberg L *et al.* 2017 Advancing ocean acidification biology using Durafet pH electrodes. *Front. Mar. Sci.* **4**, 321. (doi:10.3389/fmars.2017.00321)
29. Dickson AG, Sabine CL, Christian JR. 2007 *Guide to best practices for ocean CO<sub>2</sub> measurements*. PICES Special Publication 3. Sidney, Canada: North Pacific Marine Science Organization.
30. Gattuso J-P, Epitalon J-M, Lavigne H, Orr J. 2017 seacarb: Seawater Carbonate Chemistry. R package version 3.2.2. See <https://cran.r-project.org/package=seacarb>.
31. Lueker TJ, Dickson AG, Keeling CD. 2000 Ocean pCO<sub>2</sub> calculated from dissolved inorganic carbon, alkalinity, and equations for K<sub>1</sub> and K<sub>2</sub>: validation based on laboratory measurements of CO<sub>2</sub> in gas and seawater at equilibrium. *Mar. Chem.* **70**, 105–119. (doi:10.1016/S0304-4203(00)00022-0)
32. Perez FF, Fraga F. 1987 The pH measurements in seawater on the NBS scale. *Mar. Chem.* **21**, 315–327. (doi:10.1016/0304-4203(87)90054-5)
33. Dickson AG. 1990 Standard potential of the reaction: AgCl(s) + 12 H<sub>2</sub>(g) = Ag(s) + HCl(aq), and the standard acidity constant of the ion HSO<sub>4</sub><sup>-</sup> in synthetic sea water from 273.15 to 318.15 K. *J. Chem. Thermodyn.* **22**, 113–127. (doi:10.1016/0021-9614(90)90074-Z)
34. His E, Seaman MNL, Beiras R. 1997 A simplification the bivalve embryogenesis and larval development bioassay method for water quality assessment. *Water Res.* **31**, 351–355. (doi:10.1016/S0043-1354(96)00244-8)
35. Moran AL, Marko PB. 2005 A simple technique for physical marking of larvae of marine bivalves. *J. Shellfish Res.* **24**, 567–571. (doi:10.2983/0730-8000(2005)24[567:ASTFPM]2.0.CO;2)
36. Bezares J, Asaro RJ, Hawley M. 2008 Macromolecular structure of the organic framework of nacre in *Haliotis rufescens*: implications for growth and mechanical behavior. *J. Struct. Biol.* **163**, 61–75. (doi:10.1016/j.jsb.2008.04.009)
37. Furuhashi T, Schwarzingler C, Miksik I, Smrz M, Beran A. 2009 Molluscan shell evolution with review of shell calcification hypothesis. *Comp. Biochem. Physiol. B Biochem. Mol. Biol.* **154**, 351–371. (doi:10.1016/j.cbpb.2009.07.011)
38. R Core Team. 2016 *R: a language and environment for statistical computing*. Vienna, Austria: R Foundation for Statistical Computing. See <https://www.r-project.org/>.
39. Hervé M. 2017 RVAideMemoire: testing and plotting procedures for biostatistics. R package version 0.9-68. See <https://cran.r-project.org/web/packages/RVAideMemoire/>.
40. Bates D, Maechler M, Bolker B, Walker S. 2015 Fitting linear mixed-effects models using lme4. *J. Stat. Softw.* **67**, 1–48. (doi:10.18637/jss.v067.i01)
41. Lenth RV. 2016 Least-squares means: the R package lsmeans. *J. Stat. Softw.* **69**, 1–33. (doi:10.18637/jss.v069.i01)
42. Waldbusser GG *et al.* 2013 A developmental and energetic basis linking larval oyster shell formation to acidification sensitivity. *Geophys. Res. Lett.* **40**, 2171–2176. (doi:10.1002/grl.50449)
43. Kniprath E. 1981 Ontogeny of the molluscan shell field: a review. *Zool. Scr.* **10**, 61–79. (doi:10.1111/j.1463-6409.1981.tb00485.x)
44. Kniprath E. 1980 Larval development of the shell and the shell gland in *Mytilus* (Bivalvia). *Wilhelm Roux. Arch. Dev. Biol.* **188**, 201–204. (doi:10.1007/BF00849049)
45. Hayakaze E, Tanabe K. 1999 Early larval shell development in Mytilid bivalve *Mytilus galloprovincialis*. *Venus Jpn. J. Malacol.* **58**, 119–127.
46. Weiss IM, Schönitzer V. 2006 The distribution of chitin in larval shells of the bivalve mollusk *Mytilus galloprovincialis*. *J. Struct. Biol.* **153**, 264–277. (doi:10.1016/j.jsb.2005.11.006)
47. Falini G, Albeck S, Weiner S, Addadi L. 1996 Control of aragonite or calcite polymorphism by mollusk shell macromolecules. *Science* **271**, 67–69. (doi:10.1126/science.271.5245.67)
48. Waller TR. 1981 Functional morphology and development of veliger larvae of the European oyster, *Ostrea edulis* Linné. *Smithson. Contrib. Zool.* **328**, 1–70. (doi:10.5479/si.00810282.328)
49. Palmer AR. 1983 Relative cost of producing skeletal organic matrix versus calcification: evidence from marine gastropods. *Mar. Biol.* **75**, 287–292. (doi:10.1007/BF00406014)
50. Palmer AR. 1992 Calcification in marine molluscs: how costly is it? *Proc. Natl Acad. Sci. USA* **89**, 1379–1382. (doi:10.1073/pnas.89.4.1379)
51. Frieder CA, Applebaum SL, Pan TCF, Manahan DT. 2018 Shifting balance of protein synthesis and

- degradation sets a threshold for larval growth under environmental stress. *Biol. Bull.* **234**, 45–57. (doi:10.1086/696830)
52. Sunday JM, Crim RN, Harley CDG, Hart MW. 2011 Quantifying rates of evolutionary adaptation in response to ocean acidification. *PLoS ONE* **6**, e22881. (doi:10.1371/journal.pone.0022881)
53. Vihtakari M, Havenhand J, Renaud PE, Hendriks IE. 2016 Variable individual- and population- level responses to ocean acidification. *Front. Mar. Sci.* **3**, 51. (doi:10.3389/fmars.2016.00051)
54. Gibson G, Dworkin I. 2004 Uncovering cryptic genetic variation. *Nat. Rev. Genet.* **5**, 681. (doi:10.1038/nrg1426)
55. Thomsen J *et al.* 2017 Naturally acidified habitat selects for ocean acidification-tolerant mussels. *Sci. Adv.* **3**, e1602411. (doi:10.1126/sciadv.1602411)