

1 **Genomic Diversity Illuminates the Species History and Environmental**
2 **Adaptation of *Drosophila suzukii***

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9 **Abstract**

10 Biological invasions carry substantial practical and scientific importance, and represent
11 natural evolutionary experiments on contemporary timescales. Here, we investigated genomic
12 diversity, invasion history, and environmental adaptation of the crop pest *Drosophila suzukii* using
13 whole-genome sequencing data and environmental metadata for 29 population samples from its
14 native and invasive range. Our analysis of genome-wide variation revealed maximal diversity in
15 a population from eastern China, with other populations largely containing a subset of the
16 genetic diversity present in this sample, consistent with an ancestral or refugial species range
17 encompassing this region of Asia. Our analyses of genomic diversity and population history
18 recapitulated some previously suggested dynamics of genetic structure, invasion bottlenecks, and
19 admixture events. However, we suggest that clearer inferences of the geographic origins of
20 worldwide invasions will require denser geographic sampling of genetic variation, particularly in
21 Asia. Using a genome-environment association approach, we detected genetic signals of local
22 adaptation associated with nine distinct environmental factors related to altitude, wind speed,
23 precipitation, temperature, and human land use. We detected some unique functional signatures
24 for each environmental variable, such as a prevalence of cuticular genes associated with annual
25 precipitation. We also inferred biological commonalities in the adaptation to diverse selective
26 pressures, particularly in terms of the apparent contribution of nervous system evolution to
27 enriched processes (ranging from neuron development to circadian behavior) and/or top genes
28 associated with all nine environmental variables. Our findings provide insights into the invasion
29 history of *D. suzukii* and depict a finer-scale adaptive landscape underlying this species' invasion
30 success.

31

32 **Introduction**

33 One of the main goals of ecological and evolutionary genomics is to understand how
34 organisms evolve in response to novel environments. Biological invasions, while often
35 ecologically and economically damaging, represent unique opportunities to build our
36 understanding of local adaptation, as natural experiments that expose introduced species to new
37 biotic and abiotic factors on contemporary time scales (Lee 2002; Prentis et al. 2008; Colautti
38 and Lau 2016). Invasive species can exhibit rapid phenotypic and genetic changes during the
39 invasion process, driven by various evolutionary mechanisms such as selection, drift, mutation,
40 and gene flow (Colautti and Lau 2016; Hodgins et al. 2018). These changes can result in the
41 adaptive evolution of invasive populations to the novel environments they encounter (Colautti
42 and Barrett 2013). Although there have been emerging studies on the evolutionary biology of
43 invasive species in recent years, the source and nature of the genetic variation underlying such
44 adaptation are still not well characterized (Reznick et al. 2019; Welles and Dlugosch 2019).

45 *Drosophila suzukii*, also known as spotted wing *Drosophila*, is a promising model for studying
46 adaptive evolution during invasions. *Drosophila suzukii* is a highly polyphagous vinegar fly that
47 originated from Asia. It first expanded to Hawaii in 1980, and has recently invaded North
48 America and Europe, and subsequently Réunion Island (Indian Ocean) and South America since
49 the late 2000s (Asplen et al. 2015). *Drosophila suzukii* differs from other *Drosophila* species in its
50 unique ability to oviposit on both unripe and ripe fruits, using its serrated ovipositor to pierce the
51 skin of soft-skinned fruits. This has allowed it to exploit a novel ecological niche and avoid
52 competition with other vinegar flies that typically feed on overripe and rotting fruits (Cini et al.
53 2012; Atallah et al. 2014), causing severe economic losses to fruit crops (Knapp et al. 2021). It
54 also exhibits remarkable phenotypic plasticity and genetic diversity, which may facilitate its

55 adaptation to different climatic conditions and host plants (Gibert et al. 2019; Little et al. 2020;
56 Olazcuaga et al. 2020). To obtain a comprehensive evolutionary genetic understanding of the
57 invasion history of *D. suzukii*, we need to understand both the demography of this species
58 (including the history and structure of populations from Asia and introduced regions) and the
59 genetic basis and ecological drivers of adaptive evolutionary changes that have allowed this
60 species to occupy diverse worldwide environments.

61 Multiple genetic studies have investigated the demographic history of invasive
62 populations of *D. suzukii*. Adrion et al. (2014) analyzed six X-linked gene fragments from
63 populations in the continental US, Hawaii, Japan, and Spain. They detected differentiation
64 signals between European, Asian, and US populations, but not among US populations, and
65 suggested independent invasions into Europe and the continental US. Using data from 25
66 microsatellite markers, Fraimout et al. (2017) further characterized invasion bottlenecks and
67 inferred more comprehensive worldwide invasion scenarios of *D. suzukii*, using an approximate
68 Bayesian computation random forest (ABC-RF) approach. The resulting model, which was
69 constrained by the observed chronology of the species' invasion, indicated that some invading
70 populations had multiple genetic sources. A more recent population genomic study that focused
71 on samples from the US, as well as single sites in Brazil, Ireland, Italy, Japan, and Korea, used
72 autosomal single nucleotide polymorphisms (SNPs) to infer population structure and admixture
73 (Lewald et al. 2021). Their model mostly agrees with the above microsatellite analysis, except for
74 an admixture event from the western US back to Asia. Collectively, these studies suggest some
75 emerging consensus about the invasion history of *D. suzukii*, along with some lingering
76 uncertainty about admixture during this process. However, the limited study of Asian *D. suzukii*

77 samples has left the history of the species within this continent less clear, along with the precise
78 geographic origins of invading populations.

79 In contrast, the genetic basis of environmental adaptation in *D. suzukii* is still largely
80 unexplored. Among the few relevant studies is that of Olazcuaga et al. (2020), which used
81 genomic sequencing data from 22 worldwide populations to search for SNPs with greater
82 frequency differences between Asian (China, Japan) and non-Asian populations than are
83 observed at most loci due to founder event bottlenecks, in hopes of identifying genetic variants
84 that may underlie the invasion success of introduced populations. A subsequent study examining
85 this same data set also found a small number of transposon insertions with strong frequency
86 differences between American/European and Asian populations (Mérel et al. 2021). Another
87 study identified F_{ST} outliers among Hawaiian *D. suzukii* populations (Koch et al. 2020). Apart
88 from these limited comparisons, the genetic changes that may have helped *D. suzukii* to adapt to
89 diverse worldwide environments are entirely unknown.

90 With the increasing availability of genomic resources from non-model species, genotype-
91 environment association (GEA), also known as environmental association analysis (EAA), is
92 becoming a widely-used approach to understand the potential relationship between specific
93 environmental factors and adaptive genetic variation (Rellstab et al. 2015). GEA is also useful in
94 identifying subtle changes in allele frequencies that are difficult to detect with outlier tests based
95 on traditional population genomic approaches, especially when the number of studied
96 populations is relatively large, and there is high gene flow counteracting patterns of local
97 adaptation (Kawecki and Ebert 2004). The capability of GEA to identify adaptive genetic
98 changes and environmental drivers of local adaptation has been demonstrated with whole-
99 genome pool-seq data from *Drosophila melanogaster* (Bogaerts-Márquez et al. 2021). Therefore,

100 GEA could be helpful in understanding the genotype-environment relationships underlying the
101 invasion success of *D. sukuzii*.

102 In the present study, we perform population genomic analyses on whole-genome pool-seq
103 data of 29 population samples from native and invasive ranges to investigate the genomic
104 diversity, invasion history and environmental adaptation of *D. sukuzii*. We investigate the
105 geographic pattern of genetic diversity to obtain clues regarding the history of this species within
106 Asia. We investigate population structure and perform demographic inference, assessing
107 agreement with published models of invasion history. Finally, we test the association between
108 SNP frequencies and nine environmental variables across sampling locations, identifying both
109 specific and shared functional signatures of adaptation to these diverse selective pressures.

110

111 **Results**

112

113 **Genomic Diversity Reaches its Maximum in Eastern China**

114 We investigated the genomic polymorphism of 29 *D. sukuzii* populations derived from
115 Asia (n=8), Europe (n = 11), and North and South America (n = 10), and Asia (n = 8),
116 encompassing both newly reported and previously published samples (Figure 1A; Table S1;
117 Olazcuaga et al. 2020). Whole-genome sequences were obtained from 29 pooled samples
118 consisting of 50-212 female and male individuals. The depth of mapped reads after quality
119 control ranged from 23X to 66X among population samples, with an average of 45X (Table S1).

120 We first estimated nucleotide diversity (π_S) across synonymous SNPs to investigate the
121 effects of rapid invasions on neutral genetic diversity. The genome-wide π_S ranged from a

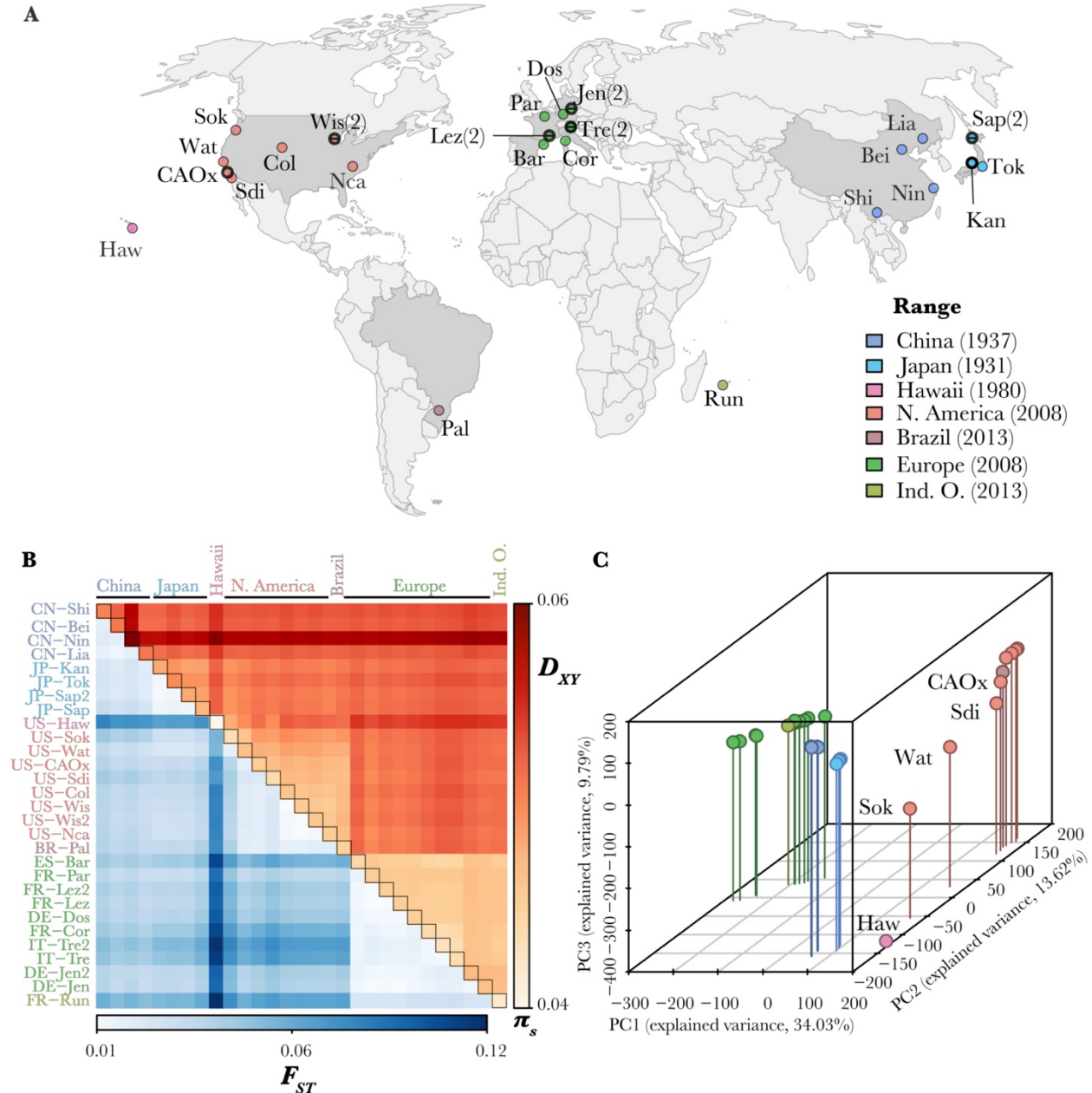
122 minimum of 0.041 (autosomes) and 0.019 (X chromosome) in Hawaii, US (US-Haw) to a
123 maximum of 0.059 (autosomes) and 0.045 (X chromosome) in Ningbo, China (CN-Nin). The
124 acute drop in π_S of introduced European and American populations relative to that of the native
125 Chinese and Japanese populations reflects previously reported founder event bottlenecks (Figure
126 1B, diagonal; Figure S1, S2). The observed patterns were also recapitulated with SNPs at all
127 types of sites (Figure S3). Lower π_S and more contrasting among-population π_S differences in X
128 chromosome reflect effectively prolonged bottlenecks due to their lower effective population size
129 (Pool & Nielsen, 2007; Figure S1). We also observed a greater loss of rare alleles in the invasive
130 populations as a typical consequence of bottlenecks (Figure S4). These founder events also
131 increased genetic differentiation as measured by F_{ST} , especially between continents (Figure 1B;
132 Figure S3), and particularly for the X chromosome (Figure S2).

133 Previous studies have hypothesized a relatively wide native range of *D. suzukii* in East and
134 Southeast Asia (Adrion et al. 2014; Fraimout et al. 2017). We examined genetic variation within
135 and among Asian populations to search for clues regarding the history of *D. suzukii* within its
136 native continent. The Ningbo population in eastern China had the highest π_S – 15.6 % greater
137 than any other population for the autosomes, and 30.6 % greater for the X chromosome (Figure
138 1B; Figure S1). Notably, D_{XY} values involving the Ningbo population did not meaningfully
139 exceed the magnitude of Ningbo's π , indicating that all other investigated locations primarily
140 have a subset of the genetic diversity present in Ningbo. These findings are consistent with the
141 other sampled populations descending in relatively recent population genetic time from a
142 Ningbo-like ancestor, and eastern China being within or closer to the ancestral range of the
143 species.

144

145 **Population Structure and Invasion History of *D. suzukii***

146 Principal component analysis (PCA) of allele frequencies was next performed to offer
147 insights into the population structure of *D. suzukii*. The top three principal components (PC 1-3)
148 explained 57.44% (autosomes, Figure 1C) and 72.35% (X chromosome, Figure S2B) of the
149 variance among populations. In both autosomes and X, three-dimensional PCA and matrices of
150 both D_{XY} and window F_{ST} recapitulated both continuous and hierarchical geographic structure.
151 These results together showed the expected clustering of populations into four distinct ranges
152 (East Asia, Hawaii, Americas, and Europe; Figure 1B, 1C; Figure S2). Much of the observed
153 population differentiation is most likely due to founder event bottlenecks and admixture during
154 worldwide expansion (Framout et al. 2017), whereas migration following population
155 establishment would need to be overwhelmingly high to have significant impacts given the very
156 brief time scale of the global invasion. The two more northerly populations from the western
157 US, Oregon (US-Sok) and central California (US-Wat), show an affinity with the Hawaiian
158 population, which aligns with the suggestion that these populations received a genetic
159 contribution from Hawaii in addition to East Asia (Framout et al. 2017), while populations from
160 southern California and the central and eastern US show less evidence of such admixture. These
161 findings are also reflected by the clustering patterns in F_{ST} -based neighbor-joining trees (Figure
162 S5).



163 Figure 1. *Drosophila suzukii* populations show maximal diversity in eastern China and continent-level genetic
 164 structure. (A) The geographic locations of the studied 29 natural populations are depicted as dots. In addition to the
 165 22 populations sampled by Olazcuaga et al. (2020), populations newly sampled at independent locations are circled
 166 in black. Populations newly sampled at nearby locations are circled and center-dashed in black, with the number of
 167 total population samples in brackets. The year of the first recorded occurrence in each geographic range (colored
 168 grey in the map) is given in brackets in the color legend. Further information about each sample is presented in
 169 Table S1. (B) Population differentiation in allele frequencies (F_{ST} ; lower triangle), between-population sequence

170 distances (D_{XY} ; higher triangle), and within-population nucleotide diversity (π_S ; diagonal) across autosomal
171 synonymous SNPs are displayed as a heatmap. Population names are colored by their geographic region. (C)
172 Autosomal genetic structure is shown by three-dimensional principal components analysis (PCA) based on allele
173 frequencies of the two most frequent alleles across all populations. Each dot represents a population. Labeled are
174 Hawaii and western coastal US populations. See the X chromosome version of (B) and (C) in Figure S2.

175

176 To provide a complementary understanding of the invasion history from genome-wide
177 information, we used TreeMix (Pickrell and Pritchard 2012) to build maximum-likelihood trees
178 with admixture graphs based on allele counts of SNPs from autosomes and X chromosome,
179 separately (Figure 2; Figure S7). We rooted the tree along the branch leading to the Ningbo
180 population (CN-Nin), since the above diversity analysis suggested that it could reflect the most
181 basal lineage among the investigated populations. We also sequentially added a range of
182 migration events (0 - 20) to trees to capture gene flow and parsimoniously improve the model fit,
183 and found that tree models with 11 mixture events explain most of the variance in relatedness
184 between populations (autosomes: 99.6%, X chromosome: 99.9%), while remaining interpretable
185 for both autosomes and X chromosome, with greater numbers of migration events resulting in
186 diminishing model improvements (Figure 2; Figure S6).

187 The topology of our TreeMix histories generally agreed with previous inferences (Adrion
188 et al. 2014; Fraimout et al. 2017; Lewald et al. 2021) and with our above results, including the
189 presence of clear genetic groupings from China, Japan, the United States, and Europe. The tree
190 structure largely agreed between our autosomal and X chromosome trees, and showed general
191 consistency across models where differing numbers of migration events were estimated. One
192 exception concerned the differential placement of the American clade. The autosomal tree
193 grouped the American and European lineages, which split from an Asian lineage closer to the

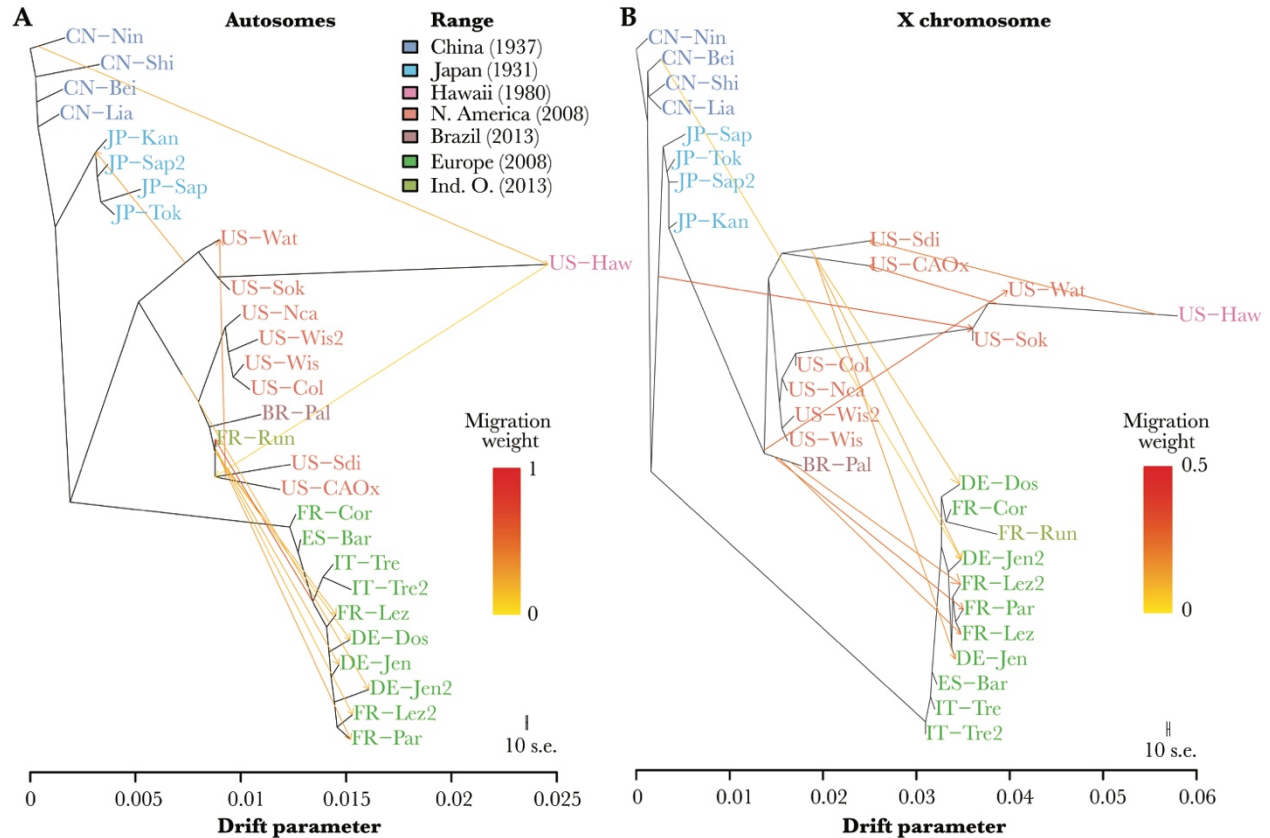
194 sampled Japan populations than to China. The X chromosome tree instead placed the
195 American group within the sampled Japan populations. Both results differ from the model of
196 Fraimout et al. (2017), which utilizes source populations from China for both continental
197 invasions, although each of our trees does include a specific migration event from China, and the
198 published model did entail ancestry from Japan to the mainland US via Hawaii.

199 The placement of the Hawaii population in both our autosomal and X chromosome trees
200 departed from the observed chronology of the species' expansion and from the model estimated
201 by Fraimout et al. (2017), which was conditioned on observed invasion dates. In both of our
202 trees, Hawaii was placed as an offshoot of the mainland US group, albeit with considerable drift
203 (Figure 2), even though *D. sukuzii* was observed in Hawaii about 30 years sooner than on the
204 mainland. Although our trees do entail admixture from Hawaii into southern California
205 populations, the previously proposed history of Hawaii also contributing higher levels of
206 admixture into the Oregon (US-Sok) and central California (US-Wat) populations (versus Hawaii
207 arising from a population similar to those two) is easier to reconcile with observed invasion
208 history. These results highlight the challenges in estimating population history from genetic data
209 alone, from a finite number of populations, and in the presence of substantial drift and admixture
210 along some lineages.

211 The autosomal and X chromosome trees also differed in their placement of the
212 population from Réunion (an Indian Ocean island that is part of France). The autosomal tree
213 placed Réunion within the American clade, but with substantial bidirectional migration between
214 this population and European samples. Whereas, the X chromosome tree placed Réunion
215 alongside a population from France, which is more consistent with the European origin estimated
216 by Fraimout et al. (2017).

217 The autosomal tree also indicated an admixture event from western US back to Japan
218 with an admixture proportion of 32% (Figure 2A), consistent with a previous inference of western
219 US/Asia admixture using a larger number of US populations (Figure 4 of Lewald et al. 2021).
220 Other instances of migration events from a more recent into an older population on a different
221 continent were also suggested from Réunion to Europe (from the autosomal tree discussed above)
222 and from Brazil to three separate French population samples based on the X chromosome tree
223 (Figure 2B). Whereas some “migration events” might actually reflect independent invasions into
224 the same continent, followed by secondary contact during the range expansion process, these
225 “reverse migrations” cannot be explained in the same way. It is not clear that neutral migration
226 between continents within just a few years should really be expected to displace any noticeable
227 portion of an established population’s gene pool. Alternatively, the precise accuracy of the
228 TreeMix models may be limited by incomplete population sampling, or by violations of the
229 neutral assumptions of the method, such as adaptive introgression, as was argued for gene flow
230 back into the African ancestral range of *D. melanogaster* (Svedberg et al. 2021).

231



232

233 Figure 2. Population topology and migration events of *D. suzukii* suggested by TreeMix. Maximum likelihood

234 graphs were built by allele counts at (A) autosomal and (B) X-chromosomal SNPs, allowing 11 migration events.

235 Horizontal branch lengths are proportional to the amount of genetic drift that has occurred on the branch. The

236 scale bar shows ten times the average standard error of the entries in the sample covariance matrix, reflecting the

237 precision of branch lengths estimates. Population names are colored by their geographic region. The year of the

238 first recorded occurrence in each geographic range is given in brackets in the color legend. Migration arrows are

239 colored according to their weight, which approximates admixture proportions. The residual fit from these graphs is

240 presented in Figure S7.

242 Figure 3. Chromosomal distribution of genetic polymorphism in *D. suzukii* informs the ordering and orientations of
243 contigs, as well as levels of centromeric and telomeric repression. Window nucleotide diversity (π_w) values are
244 displayed across (A) the X chromosome and major autosomal arms (B) 2L, (C) 2R, (D) 3L, (E) 3R. Chromosome 4 is
245 not shown as it only contains 12 windows. Each window is a continuous genomic region that includes 125,000
246 analyzed sites. Each dot represents the average π_w across populations within their geographic range as colored.
247 Only populations from major continental ranges are shown for chromosomal patterns to be clear. Within each
248 chromosome arm, contigs are ordered by length. Separate contigs are indicated by grey or white shading.

249

250 **Polymorphism, Divergence, and the Genomic Locations of *D. suzukii* Contigs**

251 Genome-wide patterns of polymorphism and divergence are useful to reveal various
252 aspects of the evolutionary history of a species. The *D. suzukii* reference assembly (Paris et al.
253 2020) is somewhat fragmented, with two or more contigs assigned to each major chromosome
254 arm, but with unknown orientation with respect to each other. We note that in other examined
255 *D. melanogaster* group species, crossing-over rate and window nucleotide diversity (π_w) are reduced
256 in centromere- and telomere-proximal regions (*e.g.*, True et al. 1996). When we examined the
257 chromosomal distribution of π_w (Figure 3), we noticed that certain arrangements of contigs
258 would result in the expected patterns of reduced diversity at the ends of each arm, and relatively
259 smooth shifts in the diversity of large windows. For instance, the second longest contig in
260 chromosome arm 3L is likely to the left of the longest contig in a reverse orientation, potentially
261 with the smaller high diversity contig in between, (Figure 3D), although either of the low diversity
262 contig end could be centromeric or telomeric. Similarly in chromosome arm 3R, although the
263 assembly was more fragmented and the centromere- or telomere-proximal regions consist of
264 multiple contigs, the second and third longest contigs appeared to contain the transitions between
265 the high diversity mid-arm region and the centromeric/telomeric zones (Figure 3E). Therefore,

266 genome-wide polymorphisms could serve as useful information to aid the ordering and orienting
267 of contig-level genome assemblies (Nakabayashi and Morishita 2020), although a combination of
268 both systematic neighbor-matching approaches and manual correction would be needed, and
269 integration with other forms of evidence (*e.g.*, comparative genomic or experimental data) would
270 be preferable.

271 Genomic polymorphism levels could also be indirectly influenced by centromeric and
272 telomeric repression of crossing over, leading to a stronger influence of natural selection reducing
273 linked variation (Maynard Smith & Haigh, 1974; Charlesworth et al. 1993). From the
274 distribution of window-based nucleotide diversity (π_w) across chromosome arms, we noticed that
275 centromere- and telomere-proximal regions of reduced diversity showed similarly narrow
276 declines in π_w (Figure 3). These patterns are more similar to those in *D. simulans* than to *D.*
277 *melanogaster* – which has broader centromeric regions of low crossing over (Langley et al. 2012).
278 These results suggest a relatively weaker suppression of crossing over in the centromere-proximal
279 regions in *D. suzukii*. It also appears that regions of low nucleotide diversity (which probably
280 coincide with regions of low recombination) cover a relatively small fraction of the genome,
281 which should limit the potentially biasing effects of natural selection on demographic inferences
282 and facilitate the localization of selective sweeps (Schridder et al. 2016; Lange and Pool 2018).

283 We next leveraged our large pooled sequencing data set to improve inferences about
284 which contigs map to the X chromosome. Out of a total of 546 contigs, 313 were previously
285 assigned to autosomes and X chromosome through either direct mapping or comparing a
286 female-to-male read depth ratio (Paris et al. 2020). We added to these annotations by
287 implementing an approach based on correlations in sequencing depth of coverage across
288 population samples that included varying numbers of females and males (see Materials and

289 Methods). Based on this analysis, we assigned 170 contigs as autosomal or X-linked. Our
290 classification of previously-assigned contigs was 96% consistent with past inferences, but we
291 corrected four previous assignments that were based on female-to-male read depth ratio, whereas
292 our method did not assign eight previously-assigned contigs.

293 To reveal levels of selective constraint, we estimated the divergence between *D. suzukii*
294 and its close relative *D. biarmipes* (Ometto et al. 2013; Suvorov et al. 2022; Figure S8). Compared
295 to the divergence of *D. melanogaster* from *D. simulans*, *D. suzukii* shows a higher genome-wide
296 divergence from *D. biarmipes*. The intronic and intergenic regions also have a higher divergence
297 relative to that of nonsynonymous sites, suggesting that a lower proportion of these non-coding
298 regions are constrained in *D. suzukii* and *D. biarmipes* than in *D. melanogaster* and *D. simulans* (Begun
299 et al. 2007; Lange and Pool 2018). Non-coding divergence is potentially inflated by repetitive
300 elements (i.e., repeatome), which represent about half of the genome expansion of *D. suzukii*
301 compared to the *D. melanogaster* genome (Paris et al. 2020). This expansion of the repeatome
302 could reflect a lower long-term effective population size (N_e) in the *D. suzukii* lineage. Given that
303 the repeatome is predominantly contributed by transposable elements (TEs), that hypothesis
304 aligns with the negative correlation between TE content and N_e found among 22 worldwide *D.*
305 *suzukii* populations also sampled in this study (Mérel et al. 2021). However, the greater π of *D.*
306 *suzukii* than *D. melanogaster* (Figure S1; Lack et al. 2016a; Lewald et al. 2021) could instead
307 indicate a greater N_e for *D. suzukii* within the past $4N_e$ generations.

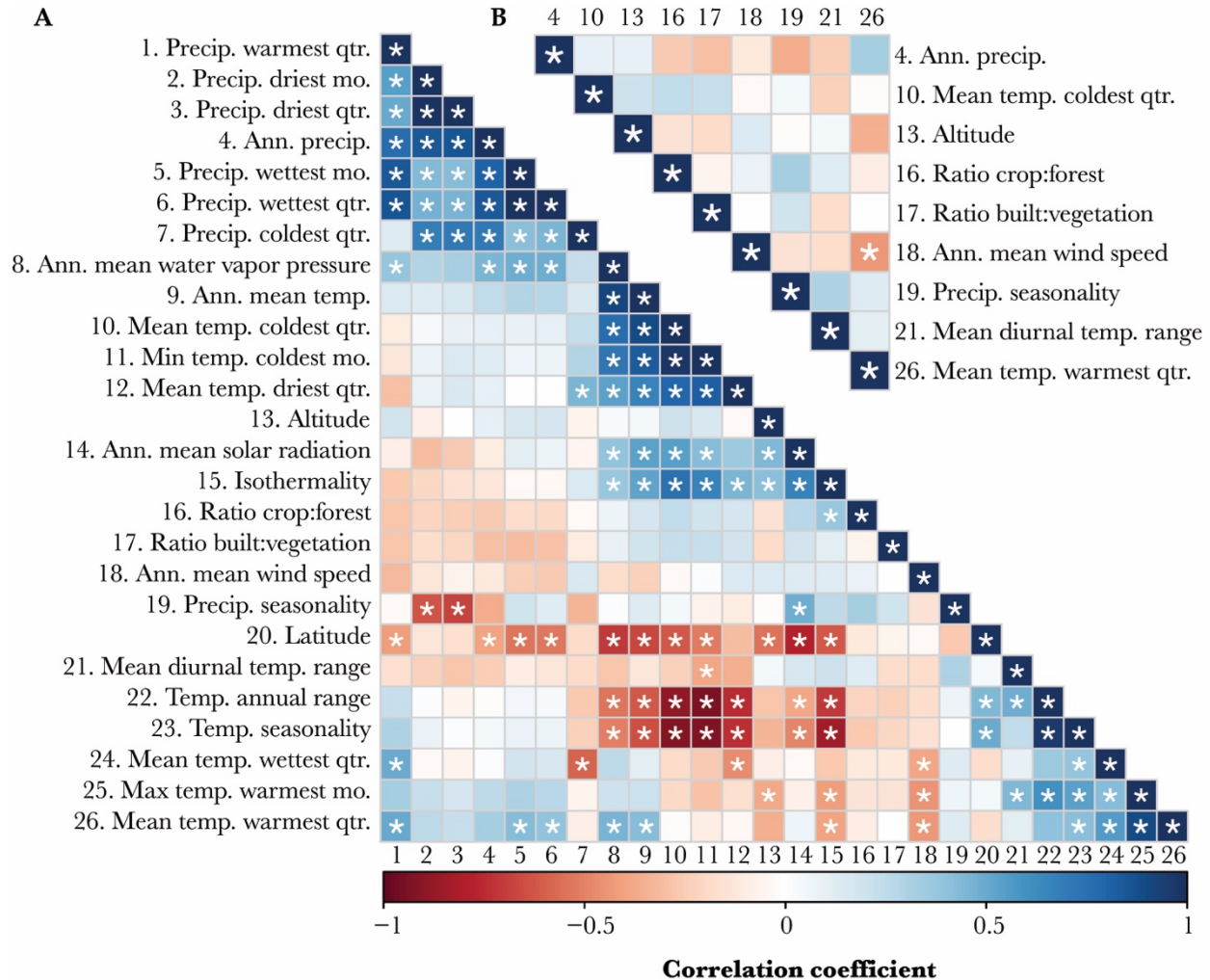
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309 **Environmental Adaptation Reveals the Genetic and Functional Basis of Invasion**

310 **Success**

311 The worldwide expansion of *D. suzukii* has exposed this species to selection pressures from
312 varying local environmental conditions (Olazcuaga et al. 2020; Mérel et al. 2021). To identify
313 environmental factors that have contributed to adaptive genetic differentiation at various levels
314 and loci under positive selection, we performed a whole-genome scan using GEA analysis
315 between environmental and genetic differentiation (see Materials and Methods). Our selection of
316 environmental variables for GEA started with a preliminary set of 26 candidate variables that are
317 potentially relevant in the adaptation process of *D. suzukii* (Figure 4A), based on data availability
318 and prior knowledge of *Drosophila* ecology (Kellermann et al. 2012; Hamby et al. 2016; Bogaerts-
319 Márquez et al. 2021). Among common strategies of performing GEA with a large set of relevant
320 environmental variables, univariate association with all environmental variables could increase
321 the number of statistical tests, thus increasing the difficulty of controlling rates of false discovery.
322 On the other hand, including multiple variables might cause a multicollinearity issue (Rellstab et
323 al. 2015). We opted to retain nine of the least correlated environmental variables for univariate
324 tests, representing altitude, wind speed, as well as multiple aspects of temperature, precipitation,
325 and human land usage for GEA analysis (Figure 4B). Although the temperature of the coldest
326 quarter had a significant negative correlation with wind speed, we kept both variables for GEA,
327 as cold stress and wind-related factors are known to be potential drivers of local adaptation in
328 *Drosophila* (Bogaerts-Márquez et al. 2021). As indicated by their coefficients of variation (CV s), the
329 selected environmental variables had moderate ($15\% \leq CV < 30\%$) to high ($CV \geq 30\%$)
330 variability across our sampling locations (Table S3).

331



332 Figure 4. Identification of least-correlated environmental variables for genotype-environment association (GEA)
 333 analysis in *D. sukuii*. (A) Pairwise correlations among a preliminary set of 26 environmental variables that are
 334 potentially impactful on *D. sukuii*. (B) A final set of nine of the most relevant and least correlated environmental
 335 variables that were chosen for GEA analysis. The Pearson correlation coefficients are colored from -1 (perfect
 336 negative correlation) to 1 (perfect positive correlation). Significance correlations ($p < 0.05$) are indicated by
 337 asterisks. See Table S3 for environmental values used to calculate correlation coefficients.

338

339 From 5,752,156 genome-wide SNPs with a minor allele frequency (MAF) higher than
 340 5%, we identified an average of 3,033 (SD = 823.4) unique candidate variants that were

341 significantly (genome-wide $q < 0.05$) associated with each of the nine candidate environmental
342 variables (Table S4). These variants corresponded to an average of 3,345 overlapping or
343 neighboring genes (Table S4), suggesting that selection pressures from the tested environmental
344 variables (or correlated factors) have meaningfully contributed to adaptive genetic variation in *D.*
345 *suzukii*, even on the brief time scale of its worldwide expansion. Among all tested environmental
346 factors, mean temperature of the coldest quarter was associated with the greatest number of
347 putatively adaptive variants (4,250 SNPs). Two precipitation-related variables, annual
348 precipitation (4,141 SNPs) and precipitation seasonality (3,389 SNPs), have the next largest loci
349 count. The ratio of built area to vegetation (i.e., crops and forests) and the ratio of crops to
350 forests were associated with the fewest genetic variants (2,369 and 1,608 SNPs, respectively).

351 To reveal the potential genetic and functional basis of invasion success under multiple
352 environmental challenges throughout the species' range, we first examined the functions of genes
353 linked to the top 10 environment-associated loci for each variable as ranked by association q -
354 value, and then by the g parameter estimating the sensitivity to environmental differentiation as
355 a tie-breaker. We found many of these genes have known functions that could facilitate
356 adaptation to the associated environmental factor (Table S5).

357 Among genes linked to altitude-associated loci, the second-ranked candidate *ab* is known
358 to control wing size in *Drosophila* (Simoes da Silva et al. 2019). Interestingly, wing size was found
359 to have increased in a highland Ethiopia *D. melanogaster* population, potentially assisting flight in
360 thin, cool air (Lack et al. 2016b). Ranked next to *ab* is *Gbs-70E*, which plays roles in glycogen
361 metabolism and the development of eggs inside the maternal ovary (Kerekes et al. 2014).
362 Another top gene, the lysine demethylase *Kdm2*, is upregulated in response to hypoxia (Batie et al.
363 2017).

364 With wind speed, the top first candidate *Ttc30* is an essential gene in the biogenesis of
365 sensory cilia, which are key to both chemosensory and mechanosensory functions in *Drosophila*
366 (Avidor-Reiss et al. 2004; Avidor-Reiss and Leroux 2015). Another top candidate, *Arr2*, is
367 involved in olfaction, hearing, and vision (Alloway and Dolph 1999; Elaine Merrill et al. 2005;
368 Senthilan et al. 2012). In light of the relevance of wind for insect flight, we also noted that a third
369 top candidate, *vn*, is a developmental gene named for its wing phenotype (Wang et al. 2000).

370 There is also some evidence for precipitation-related local adaptation. The top gene *mmv*
371 associated with precipitation seasonality (i.e., annual range of precipitation) was shown to
372 regulate chitin synthesis and cuticle production. Since precipitation is correlated with desiccation
373 resistance across the *Drosophila* phylogeny (Kellermann et al. 2012), *D. suzukii* may have
374 developed adaptive strategies of modifying chitin biosynthesis under conditions of desiccation
375 (Rezende et al. 2008; Clark et al. 2009), which was also implied in seasonal plasticity of natural
376 *Drosophila* populations (Shearer et al. 2016; Horváth et al. 2023). In addition, the gene *osy*
377 (*CG33970*) contributes to the formation of the outer cuticle layer and is expressed more highly in
378 *D. suzukii* than in *D. melanogaster* (Wang et al. 2020). Furthermore, two of the top genes associated
379 with annual precipitation (*Abd-B* and *bab1*) regulate cuticle pigmentation (Rogers et al. 2013),
380 which may or may not correlate with desiccation tolerance in *Drosophila* species (Wang et al.
381 2021). We also note that although environmental fitness effects on these testes-expressed genes
382 are not known, the same SNP near *CG17944* and *nxf4* was among the highest-scoring variants for
383 both annual precipitation and precipitation seasonality (variables that have a non-significantly
384 negative correlation between them; Figure 4).

385 Another important environmental barrier to invasion success is temperature. For the
386 mean temperature of the coldest quarter, a top gene was *Ac78C*, which has roles in circadian

387 regulation and taste (Ueno and Kidokoro 2008; Duvall and Taghert 2013). With the mean
388 temperature of the warmest quarter, the top genes *crp*, *Mrtf*, and *Ubx* help control the
389 development of trachea (Han et al. 2004; Guha and Kornberg 2005; Wong et al. 2015), which
390 may be important in limiting water loss in hot environments (Gibbs et al. 2003).

391 For the ratio of built to vegetated area, a different variant near the cuticle-related gene *osy*
392 (which was also indicated above for precipitation seasonality) was detected. Another top outlier
393 was the nervous system gene *trv*, which is involved in thermosensitivity (Honjo et al. 2016). For
394 the relative levels of crop and forest cover, the first-ranked variant was near *Mtk*, which encodes
395 an antifungal and antibacterial peptide (Levashina et al. 1995), and we note that mushrooms
396 (which are more available in forest) have been proposed as overwintering food sources for *D.*
397 *suzukii* (Wallingford et al. 2018), and the evolution of immune genes has been found to differ
398 strongly between mushroom-feeding and human commensal *Drosophila* species (Hill et al. 2019).
399 With regard to the differential light environments entailed by forest versus farm habitats, we note
400 that the next highest gene, *CadN2*, helps connect photoreceptor neurons to their targets (Prakash
401 et al. 2005).

402 Beyond the top candidate genes that have related functions to specific types of
403 environmental changes, we also found a wide range of nervous system genes associated with
404 multiple environmental factors. For instance, among the top five altitude-associated loci, three
405 have known functions in the nervous system of *Drosophila*, including the first-ranked gene *Cmpy*,
406 which enables proper growth control at neuromuscular junctions (James and Broihier 2011), *ab*,
407 which regulates dendritic complexity (Li et al. 2004; Sugimura et al. 2004), and *not*, which is
408 essential for stabilizing synaptic homeostasis within glia (Wang et al. 2020). Such genes were also
409 linked to top-10 loci associated with wind speed (*dpr6*), precipitation (*Msp300*, *Tusp*, *5-HT2A*),

410 temperature (*ATP6AP2*, *CG13579*, *D*), and land use variables (*Bsg*, *Mp*, *velo*). Different variants
411 associated with *scrt*, a regulator of neuronal cell fate, were among the top results for both mean
412 diurnal range and the ratio of crop to forest cover.

413 Next, we examined environment-specific adaptation on a more comprehensive basis
414 through a gene ontology (GO) enrichment analysis of the top 500 genes associated with each
415 environmental variable (Figure 5). As correlates of temperature in the coldest quarter, cAMP
416 metabolic process was the top enriched category, followed by two other related purine
417 metabolism groupings. We note that cAMP is important in circadian regulation (*e.g.*, Palacios-
418 Muñoz & Ewer, 2018), which is known to play an important role in *Drosophila* environmental
419 adaptation (*e.g.*, Helfrich-Förster et al. 2020), as also implicated by the presence of “entrainment
420 of circadian clock” on our top GO category list for altitude. More broadly, purine metabolism
421 was inferred as a strategy of cold acclimation in *D. suzukii* (Enriquez and Colinet 2019). For
422 diurnal temperature range, the top category was “regulation of growth”, and we note that some
423 drosophilids have evolved to have larger body sizes in more challenging thermal environments
424 (Gilchrist and Partridge 1999; Calboli et al. 2003; Lack, Yassin, et al. 2016).

425 With precipitation, we identified “chitin metabolic process” as the top GO term
426 associated with annual precipitation, as well as “chitin-binding” with precipitation seasonality.
427 Together with the chitin synthesis genes we described above for precipitation, adaptation to the
428 overall intensity and seasonal variation of precipitation by modifying cuticular chitin may be
429 implied. For crop to forest ratio, the category “antimicrobial humoral response” included the top
430 gene *Mtk* listed above.

431

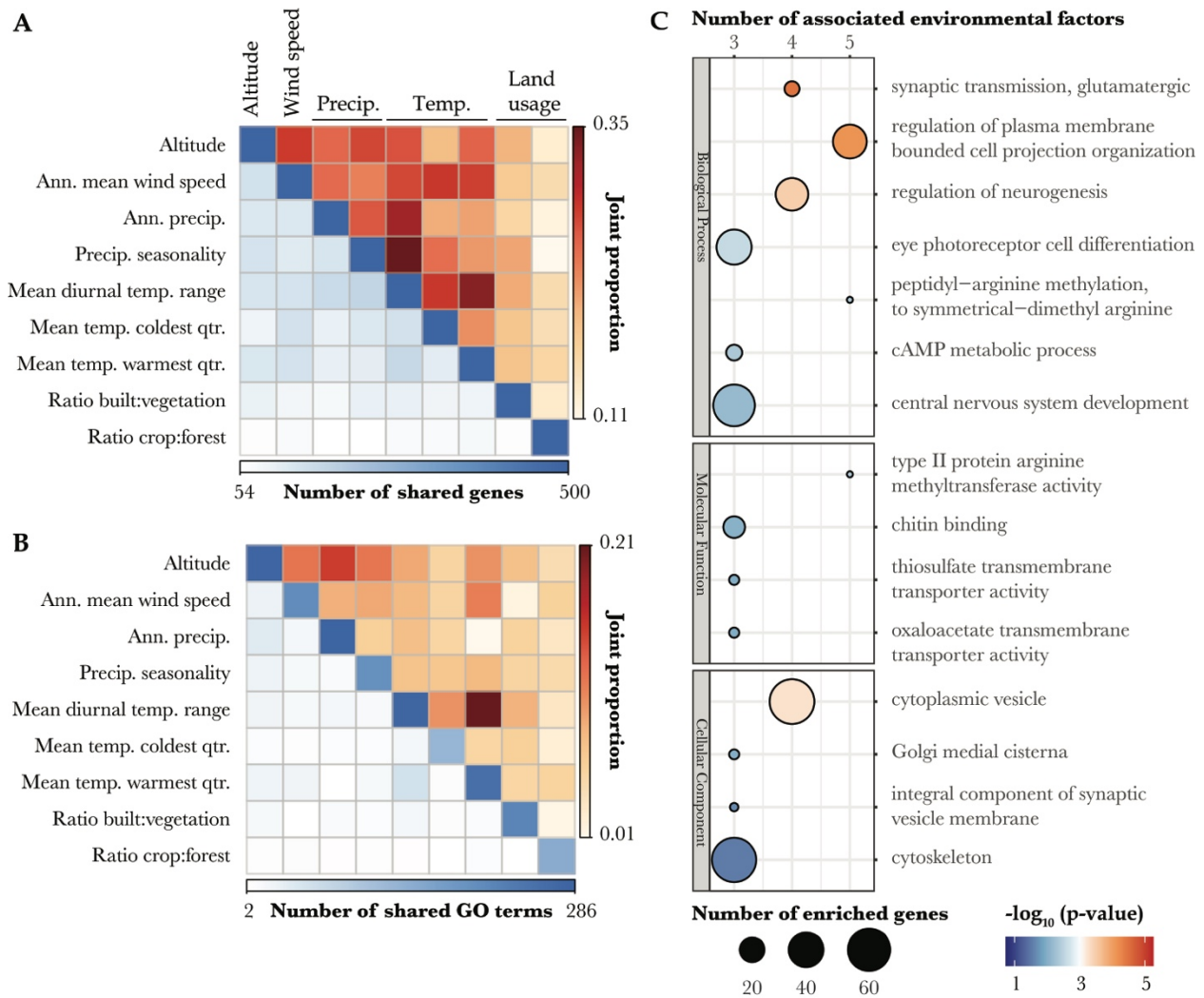
435 category. Descriptions of GO categories are colored by their GO class (see legend at top right). Only GO
436 categories including more than five associated genes are listed here. For a full list of enriched GO categories, see
437 Table S6.

438

439 As broader evidence for a shared (or biologically similar) underlying genetic basis of
440 adaptation to multiple environmental factors, we examined the overlap of the most significant
441 genes and most enriched GO categories between different environmental variables. We found
442 the top candidate genes to be mostly associated with both temperature and precipitation. The
443 gene sets showed relatively greater overlap among climatic variables (including altitude), whereas
444 the two land usage variables had less overlap with climatic variables or with each other (Figure
445 6A). Since the patterns of shared genes cannot be fully explained by correlations between
446 environmental values (Figure 4B), at least some of the genes may have been responding to
447 multiple selective pressures. The overall proportions of shared GO categories were lower than
448 those of shared genes, indicating that the shared genes do not necessarily lead shared functional
449 categories between environmental variables. Relatively higher GO term sharing was observed
450 between altitude and either wind speed or precipitation, and between diurnal temperature range
451 and temperature of the warmest quarter (Figure 6B). Based on the shared genes and GO terms
452 observed, it is possible that during the rapid range expansion of *D. sukukii*, pleiotropy may have
453 facilitated local adaptation to multiple selective pressures (Hämälä et al. 2020; Kinsler et al.
454 2020).

455 Consistent with our gene-based analyses of universal adaptive function, we found three of
456 the top shared biological processes clearly related to nervous system functions, including the
457 topmost “synaptic transmission, glutamatergic” (shared by altitude, wind speed, diurnal

458 temperature range, and temperature of the warmest quarter), “regulation of neurogenesis”
459 (altitude, diurnal temperature range, temperature of the warmest quarter, and ratio of built area
460 to vegetation), and “central nervous system development” (precipitation seasonality, diurnal
461 temperature range, and temperature of the warmest quarter) (Figure 6C). Further, as mentioned
462 above, “cAMP metabolic process” (shared by temperature of the coldest quarter, precipitation
463 seasonality and ratio of built area to vegetation) could entail neurologically modulated changes in
464 circadian behavior. Each of these traits was associated with at least three environmental
465 variables, which suggests a multifaceted role for nervous system evolution in facilitating the
466 invasion success of *D. sukuzii* under multiple environmental challenges.



467

468 Figure 6. Overlapping genes and GO categories among environmental factors reveal the shared genetic and
 469 functional basis of environmental adaptation in *D. sukukii*. The numbers and proportions of shared (A) environment-
 470 associated genes and (B) enriched GO categories among environmental factors are shown in heatmaps. Here, joint
 471 proportion represents the fraction of the genes or GO terms associated with either of two environmental variables
 472 that are associated with both variables. (C) Top GO categories of each type are depicted as bubbles. Bubbles are
 473 colored by the negative logarithm of the combined p-value of enrichment across all environmental variables, and are
 474 scaled by the number of enriched genes. The number of environmental variables that enrich a given GO category is
 475 indicated by the top horizontal axis.

476

477 **Discussion**

478 We performed population genomic analyses of 29 population samples of *D. suzukii* to
479 investigate the genomic diversity, invasion history, and environmental adaptation of this highly
480 invasive species across its worldwide distribution. Our data shows a genetic grouping of these
481 populations into four primary geographic regions: eastern Asia (containing the native range),
482 Hawaii, the Americas, and Europe. We estimated the highest genetic diversity in eastern China
483 (Ningbo), suggesting this region could be within or near the ancestral range of *D. suzukii*. We
484 confirmed moderate founder event bottlenecks in introduced populations, and our analysis
485 suggested that admixture during the expansion of *D. suzukii* populations might be more complex
486 than previously estimated, but that further sampling in Asia is needed to address the precise
487 origins of introduced populations. We then identified the environmental drivers for local
488 adaptation of *D. suzukii*, including altitude, wind speed, precipitation, temperature, and human
489 land usage, where temperature of the coldest quarter shows the clearest evidence of
490 environmental adaptation. Our results suggested extensive local adaptation in response to
491 specific environmental challenges. We also revealed appreciable sharing of genes and functional
492 pathways underlying invasion success across multiple environmental pressures, which were most
493 obvious with nervous system genes.

494

495 **Insights and mysteries regarding the history of *D. suzukii***

496 Surprisingly, the relative levels of diversity of *D. suzukii* populations have been the subject
497 of conflicting findings. Here, we detected the highest overall genetic diversity for *D. suzukii* in
498 eastern Asia, especially in the eastern China Ningbo population (Figure 1B; Figure S1). Our
499 results contrast with some prior findings, such as higher diversity in several North American
500 populations relative to a Japanese population (Adrion et al. 2014), the species' highest diversity

501 being in north-central China (Fraitout et al. 2017), or alternatively in Japan (Gautier et al.
502 2022). These conflicts may reflect challenges in accurate estimation of genetic diversity,
503 especially when the number of investigated loci is limited (Adrion et al. 2014; Fraitout et al.
504 2017), or when rare alleles are filtered by a relatively strict allele count threshold (Gautier et al.
505 2022), given that rare variants are especially likely to be lost during bottlenecks. In this study, we
506 accounted for such biases by applying stringent filtering by sequencing read base quality and
507 mapping quality, by including putative rare alleles in genome-wide analyses, and by focusing on
508 synonymous diversity which has a relatively higher ratio of signal to noise (more true variation
509 relative to sequencing errors).

510 Beyond the eastern China population's maximal diversity, we found that no pairwise D_{XY}
511 values meaningfully exceeded Ningbo π , implying that two Ningbo *D. suzukii* genomes are as
512 different from each other as any pair of genomes worldwide (paralleling a result for a Zambia
513 population of the human commensal *D. melanogaster*; Pool et al. 2012). Those findings are
514 consistent with all sampled *D. suzukii* populations originating from a Ningbo-like ancestor in
515 recent population genetic time. This inference could conceivably imply that *D. suzukii* only
516 expanded across portions of its current eastern Asian range due to its relationship with human
517 agriculture. Alternatively, in light of the apparently large effective population size of *D. suzukii*,
518 the plausible time frame of divergence among Asian populations could extend back toward the
519 last glacial maximum (LGM). We note that Ningbo is near the historical northern limit of
520 temperate woodlands during the LGM (Ray and Adams 2001). If this region is indeed part of a
521 limited overlap between the continental habitats of this temperate species during glacial and
522 interglacial periods, then this stability of occupation could help explain the Ningbo population's

523 elevated genetic diversity, whereas other eastern Asian populations' somewhat lesser diversity
524 might stem from mild founder event bottlenecks during post-glacial range expansions.

525 Further sampling and analysis of Asian *D. suzukii* populations is urgently needed to
526 advance our understanding of the history of this species. Even the current distribution of *D.*
527 *suzukii* in eastern Asia is not entirely clear. According to TaxoDros
528 (<https://www.taxodros.uzh.ch/>), *D. suzukii* has been sampled from China, Japan, North and
529 South Korea, Burma, Cambodia, Thailand, and Taiwan. It is unclear whether the lack of
530 records from neighboring countries such as Laos and Vietnam, where the species is predicted to
531 thrive (Santos et al. 2017), reflects a genuine absence or insufficient sampling effort.
532 Understanding the history of *D. suzukii* in Asia will require the analysis of genomic diversity from
533 dozens of population samples from across its Asian range, including from countries beyond
534 China, Japan, and Korea, where all population genetic study to date has focused.

535 Geographically denser population genomic data from Asian *D. suzukii* would also
536 facilitate clearer conclusions regarding the origins of lineages that invaded regions such as
537 Hawaii, the mainland US, and Europe. Above, we noted some contrasts between our TreeMix
538 results, which placed mainland US and Europe populations closer to sampled Japan populations
539 than to China, and the model of Fraimout et al. (2017), which primarily drew upon China source
540 populations for these invasions. However, our autosomal TreeMix results suggested that neither
541 of these invasive lineages falls within the clades of sampled populations from either China or
542 Japan (and our X chromosome analysis echoes this finding for Europe). Our results are
543 compatible with the possible origins of continental *D. suzukii* invasions from one or more Asian
544 populations that are somewhat genetically differentiated from any of those analyzed here. At

545 present, we suggest that the geographic origin of these economically significant introductions
546 remains an open question.

547 Our TreeMix analysis also suggested that worldwide patterns of admixture may be more
548 complex than suggested by previously inferred scenarios (Adrion et al. 2014; Fraimout et al.
549 2017; Gautier et al. 2022). However, caution is required to interpret admixture events that were
550 inferred solely from this analysis, and there may be several reasons why some inferences differ
551 between ours and prior analyses. First, as indicated above, we are still working with a very
552 incomplete panel of populations, especially from Asia, and methods may differ in terms of how
553 they respond to the absence of historically important populations such as the sources of the
554 invasive lineages. Second, whereas the study of Fraimout et al. (2017) restricted its consideration
555 to founder events compatible with the observed chronology of species expansion, our TreeMix
556 results represent an investigation of population history through genomic data alone, without
557 assuming that each worldwide invasion was detected rapidly upon its occurrence. In species with
558 complex population structure, multiple histories could lead to identical covariance matrices, thus
559 allowing for several different invasion histories compatible with the data (Pickrell and Pritchard
560 2012). Finally, as mentioned by Adrion et al. (2014) and Fraimout et al. (2017), the additional
561 migration events predicted in our analysis could indicate that genome-wide data is more sensitive
562 for discriminating among complex invasion scenarios and population structures than a limited
563 number of loci.

564 As indicated above, our TreeMix results also differed somewhat between autosomal and
565 X-linked loci. To some extent, demographic history is expected to differ between these loci,
566 including due to different levels of bottleneck severity at distinct effective population sizes, and
567 the unbalanced sex ratio of this species, especially when the multiple mating is present or sex

568 ratios depart from null expectations (Pool and Nielsen 2009; Tait et al. 2018; Weißinger et al.
569 2019; Manko et al. 2021). However, it is not clear that tree topology or gene flow presence
570 should differ substantially on such grounds. In addition to concerns listed above such as model
571 identifiability and missing populations, the consistency of autosomal versus X-linked inferences
572 could also be influenced by natural selection, particularly if it influences a longer stretch of the
573 genome with reduced recombination, such as a centromeric or telomeric region, or a
574 polymorphic inversion.

575 A key limitation of our pool-seq data is that it cannot provide haplotype information. In
576 light of the challenges in performing demographic inference from pooled sequencing data, there
577 would be considerable benefit in high quality individual sequencing of multiple wild-caught flies
578 from each of a broad array of populations (particularly in Asia), in order to facilitate quantitative
579 inferences regarding population history. Such data, especially if combined with efficient model
580 selection methods (*e.g.*, ABC-RF, (Pudlo et al. 2016; Fraimout et al. 2017), could provide a more
581 confident and precise demographic model which could both inform species history and serve as
582 null hypothesis for the detection of natural selection.

583

584 **Environmental Drivers of Adaptation in *D. suzukii***

585 Here, we presented a GEA analysis that investigated the most geographically and
586 genetically diverse set of *D. suzukii* populations and the most comprehensive set of environmental
587 factors to date, which enabled unprecedented power to capture even minor adaptive genetic
588 differentiation in response to distinct environmental challenges during the species' rapid
589 invasions. In addition to our identification of climatic factors including temperature,
590 precipitation-related variables and wind speed as the most frequently correlated with putatively

591 local adaptative variants (consistent with previous GEA analysis in *D. melanogaster*, e.g., Bogaerts-
592 Márquez et al. 2021), we also for the first time identified large numbers of genome-wide variants
593 associated with altitude and human land usage-related variables (Table S4), which were not
594 included in most GEA studies despite their potential significance to local adaptation. In
595 particular, the detected associations with ratios of developed land to vegetation and of cropland
596 to forests highlights the ecological impacts of urbanization and agriculture on natural populations
597 of insects. Since the prior selection of environmental variables is critical for successful GEA
598 analyses, we also provided an instructive example for correlation-based selection to identify the
599 most relevant and least redundant environmental factors (Rellstab et al. 2015).

600

601 **Nervous System Evolution is Ubiquitous in Environmental Adaptation of** 602 ***Drosophila***

603 In *D. suzukii*, we found nervous system and related sensory and behavior annotations
604 associated with top genes for all nine environmental variables studied. Concordantly, we found
605 that GO categories related to the nervous system were among the most shared across
606 environmental variables (Figure 6C). In *D. melanogaster*, related GO categories like ‘neuron
607 development’, ‘nervous system development’ and ‘eye development’ were also enriched among
608 genes associated with environmental variation among natural populations within North America
609 or Europe, and across seasons within Europe (Bogaerts-Márquez et al. 2021). GO categories
610 associated with the nervous system have also shown evidence of positive selection in various
611 genome scans of *D. melanogaster* (Langley et al. 2012; Pool et al. 2012; Pool 2015), including a
612 study of parallel evolution in cold-adapted populations (Pool et al. 2017). Given the
613 morphological evidence of neuron-muscular junction evolution across the entire *Drosophila*

614 phylogeny (Campbell and Ganetzky 2012), we therefore propose a broad adaptive importance of
615 the nervous system in *Drosophila* species and potentially other insects. Such evolutionary
616 processes may have either maintained ancestral neural functions in novel challenging
617 environments, or created novel phenotypes that better fit the new optima arising from complex
618 combinations of environmental factors.

619

620 **Cautions with Interpretations of Association Results and Future Directions**

621 While we have generated intriguing hypotheses about gene functions that may underlie
622 the environmental adaptation of *D. suzukii*, it is difficult to distinguish between correlated
623 environmental selective pressures that may have driven the detected associations, including not
624 only the 17 environmental factors that were excluded in the process of variable reduction, but
625 also correlated biotic or abiotic factors not represented in global databases. As an intrinsic
626 limitation of GEA analysis that cannot be accounted for by applying stricter thresholds,
627 associations observed with a particular environmental factor might stem from adaptation to other
628 covarying factors (Rellstab et al. 2015). For example, the two tracheal branching genes *crp* and
629 *Mrtf* (Han et al. 2004; Wong et al. 2015) associated with mean temperature of the warmest
630 quarter could represent adaptations to reduce water loss under conditions of elevated water
631 vapor pressure (Telonis-Scott et al. 2012), which is closely related to humidity and has a
632 significant positive correlation with mean temperature of the warmest quarter (Figure 4B).

633 Therefore, expanded characterization of the relationships between genotype, phenotype,
634 and fitness in this species is needed to further clarify the functional and phenotypic
635 interpretations associated with certain environmental factors and genes. Experimental
636 validations that leverage RNA interference (Boutros and Ahringer 2008) and/or transgenic

637 overexpression (Prelich 2012) to modify the expression of associated genes, and/or genome
638 editing techniques (Stern 2014; Turner 2014; Shalem et al. 2015) to target putatively adaptive
639 variants would also bring a more solid understanding about the invasive biology of this species in
640 distinct environments. Such functional studies could be complemented by population
641 experiments under controlled laboratory environments or field conditions (*e.g.*, Behrman et al.
642 2015; Rudman et al. 2022), in order to more clearly demonstrate the connections between
643 specific selective pressures and alleles or traits of interest.

644

645 **Broader Impacts and Significance**

646 Our work integrates genetic and environmental data to improve the reconstruction of the
647 invasion genomics of a crop pest carrying significant economic costs (Knapp et al. 2021), which
648 will hopefully inspire future studies on developing diverse pest control methods given the
649 adaptive and neutral genetic differentiation among *D. suzukii* populations. Understanding the
650 extent of local adaptation and its potential environmental drivers will also help predict the spread
651 and future distributions of invasive species (Colautti and Lau 2016). More broadly, the enhanced
652 understanding of how organisms may adapt to geographical, climatic and artificial selective
653 pressures from this study is also of value in assessing the susceptibility of natural populations to
654 climate change (Kellermann et al. 2012) and human activities (Barange et al. 2010).

655 Our analysis of genome-wide diversity has uncovered novel patterns (such as the maximal
656 diversity of an eastern China population), while reinforcing other prior findings and adding new
657 perspectives on the invasive history of *D. suzukii*. As suggested above, our findings indicate a
658 strong need for the sequencing and analysis of individual genomes from a larger number of
659 worldwide populations, especially from additional regions of eastern and southern Asia. Such

660 efforts could lead to a vastly improved understanding of both the evolution of *D. suzukii* within its
661 native continent and the geographic origins of worldwide invasions.

662

663 **Materials and Methods**

664

665 **Fly Collection, DNA Preparation, and Pooled Sequencing**

666 Fly samples from 29 populations were used, seven of which were sequenced for the
667 present study. The fly samples sequenced in this study were collected from wild *D. suzukii*
668 populations in two states of the USA, two provinces of Japan, and three European countries
669 (Figure 1; Table S1). Pooled whole adult flies ($n = 100 \sim 183$) from each population (Table S1)
670 were used for DNA extraction as previously described (Langley et al. 2011). Library
671 preparations were conducted at the Next Generation Sequencing Core of University of
672 Wisconsin Madison Biotechnology Center (<https://dnaseq.biotech.wisc.edu>), where pair-end
673 (PE) reads at the length of 150bp were then generated for each of seven pooled DNA samples on
674 an Illumina NovaSeq 6000.

675 Pool-sequenced reads of 22 additional *D. suzukii* population samples, including from
676 Europe, the Americas, and Asia, were obtained from public data provided by Olazcuaga et al.
677 (2020) at EBI's SRA (Figure 1; Table S1). Taken together, we formed a comprehensive dataset
678 of 29 populations sampled from native and invasive ranges of *D. suzukii*.

679

680 **Quality Control, Alignment, and Variant Calls from Pool-seq Data**

681 To maximize the quality of our analyzed data, we built a high-throughput assembly and
682 quality control pipeline *poolWGS2SNP* with optimized performance, stringent filtering,
683 compatibility with large numbers of genomic contigs, and customized functions to call high-
684 confidence single-nucleotide variants from pool-sequenced data in *D. suzukii* (Figure S9), in part
685 by utilizing resources from the DrosEU bioinformatics pipeline (Kapun et al. 2020).

686 As an initial quality control of raw PE reads, adapters were removed, and the 3' end of
687 reads with base quality < 20 were trimmed using *fastp* (Chen et al. 2018). Further trimming was
688 performed using a self-developed python program *filter_PE_length_mem.py* (see Data Availability),
689 where any pair of forward and reverse reads with less than a total of 150 bases with base quality
690 (BQ) ≥ 20 , as well as any individual reads with less than 25 bases with BQ ≥ 20 were discarded.

691 The trimmed and qualified reads were then mapped against the recently released near-
692 chromosome level *D. suzukii* genome assembly Dsuz-WT3_v2.0 that covers autosomes and the X
693 chromosome (Paris et al. 2020) using *bwa mem* (Li 2013). Reads with a mapping quality below 20
694 were then removed using *Samtools* (Li et al. 2009). We used Picard's *SortSam* to sort BAM files,
695 and used Picard's *MarkDuplicates* to mark PCR duplicates to avoid false variant calls
696 (<http://broadinstitute.github.io/picard>). Indel identification and realignment around indels were
697 performed using GATK's *RealignerTargetCreator* and *IndelRealigner* (Auwera and O'Connor 2020).
698 Finally, alignments in BAM format were checked for formatting errors using Picard's
699 *ValidateSamFile*. Summary statistics for quality checking of BAM files were generated using *bamdst*
700 (<https://github.com/shiquan/bamdst>).

701 To call SNPs, we merged the quality-checked BAM files of all population samples into
702 one file using *Samtools mpileup*, only retaining alignments with mapping quality no less than 20
703 and sites with base quality no less than 20. Variant calling was then performed on the *mpileup*

704 file using the heuristic SNP caller *PoolSNP* (Kapun et al. 2020). We used a nominally low value
705 for the parameter miss-frac (0.001) to require for each population sample individually, that depth
706 of coverage at a given site be 12 or greater (min-cov = 12), and that this site not be in the top 1%
707 of sites genome-wide for depth of coverage (max-cov = 0.99; calculated separately for each
708 population and for autosomal and X-linked contigs), in order to filter sites subject to copy
709 number variation. In the initial data set used for analysis of genome-wide diversity, we avoided
710 potential biases from allele frequency filters by using min-count = 1 and min-freq = 0. We
711 termed the resulting high-quality sites as ‘analyzed sites’ for brevity.

712

713 **Identifying Autosomal and X-linked Contigs**

714 We chose to perform all population genomic analyses and whole-genome scans separately
715 for SNPs from autosomes and the X chromosome for the following reasons: 1) autosomal and X-
716 linked variants have different allelic sample sizes as samples were obtained from both male and
717 female flies; 2) autosomes and the X chromosome could reflect different demographic histories
718 and outcomes of natural selection, *e.g.*, the lower effective population size of the X chromosome
719 than autosomes could lead to a higher impact of bottlenecks and selection on genomic diversity;
720 3) unbalanced sex ratios and male-biased dispersal could further differentiate autosomal versus X
721 chromosome variation (Clemente et al. 2018; Olazcuaga et al. 2020).

722 Since the assembly of *D. suzukii* reference genome is still at the contig level, chromosomal
723 identities of each contig are needed to perform separate analyses. However, 497 contigs that
724 represent ~43% of the assembly length have not been unambiguously mapped onto chromosome
725 arms of the *D. melanogaster* dm6 genome assembly. Although 264 of the unplaced contigs had
726 been assigned to autosomes and the X chromosome based on a female-to-male read depth ratio,

727 233 contigs that represent ~5% of the genome remained unassigned due to the lack of statistical
728 power (Paris et al. 2020).

729 Given our interest in accurately analyzing a larger proportion of the euchromatic
730 genome, we identified ~70% of these 233 unassigned contigs as autosomal and X-linked based
731 on the correlation between the mean read depth of each contig (among population samples) and
732 that across unambiguously aligned autosomal or X-linked contigs. We chose Spearman's rank
733 correlation instead of the Pearson correlation, as the distribution of depth data failed the
734 assumption of bivariate normality. A contig that has mean depth significantly correlated with
735 that of either known autosomal or X-linked contigs was assigned to the chromosome with a
736 higher correlation coefficient. Our method completely confirmed all prior mapping-based
737 assignments and had a ~96% consistency with the previous assignment based on female-to-male
738 read depth ratios. Inconsistent assignments for four contigs were corrected according to our
739 method (Table S2). The eight previously-assigned contigs that could not be assigned using our
740 method, as well as other unassigned contigs using all methods (totalling ~2.7 mb), were excluded
741 from downstream GEA analyses, because the assignment information is needed for estimating
742 effective sample size that were used to correct allele count data as an input to GEA.

743

744 **Annotating Genomic Features and Estimating Divergence**

745 To explore genomic diversity at synonymous sites and selective constraint for other site
746 types in *D. sukukii*, we classified the reference genome into nine exclusive categories of site
747 degeneracy and function (Lange and Pool 2018), including: non-degenerate (i.e.,
748 nonsynonymous) sites, two-, three-, and four-fold degenerate (i.e., synonymous) sites, 3' and 5'
749 UTRs, RNA-coding genes, introns, and intergenic regions. From input files including the

750 eukaryotic codon table, the published genome sequence and GFF3 annotation obtained at NCBI
751 RefSeq, we generated a letter-coded annotation (in FASTA format) mirroring both strands of the
752 whole genome sequence of *D. suzukii* and a coordinate-based annotation (in BED format) that
753 combines adjacent sites of the same category into a single row. Degeneracy was determined
754 based on the standard codon table. 5' UTRs were defined as regions between the start of the
755 first exon and the start of the first coding sequence (CDS), while 3' UTRs were defined as regions
756 between the end of the last exon and the end of the last CDS. In cases of overlapping genes and
757 alternative splicing that raise annotation conflicts, we followed an annotation priority in the
758 category order listed above.

759 We then estimated the divergence between *D. suzukii* and its close relative *D. biarmipes* in
760 each of these categories (Suvorov et al. 2022). We obtained results of multiple sequence
761 alignment between the current reference genomes of *D. suzukii* and *D. biarmipes* (Paris et al. 2020).
762 For each site category of *D. suzukii*, the unpolarized divergence was estimated as the number of
763 substitutions over the total number of sites within aligned blocks of reference genome sequences.

764

765 **Estimating Nucleotide Diversity, F_{ST} D_{XY}**

766 To compare genome-wide polymorphism among populations, we estimated nucleotide
767 diversity (π) across SNPs at four-fold degenerate sites (π_S) in addition to that at all categories of
768 sites (π_A), as π_S estimation is relatively less affected by sequencing errors than nucleotide diversity
769 estimated from other site categories (due to a higher ratio of real variation to errors). To
770 calculate π for each population sample, we adopted an unbiased estimator of nucleotide diversity
771 ($\hat{\theta}_\Pi$) based on heterozygosity (Π), which has been optimized for pool-seq data (Ferretti et al.
772 2013). Numerically,

$$\pi_S = \hat{\theta}_\Pi = \frac{n_c}{n_c - 1} \frac{\Pi}{L} = \frac{n_c}{n_c - 1} \frac{1}{L} \sum_l \frac{2}{n_r(l)(n_r(l) - 1)} m_l (n_r(l) - m_l) \quad (1)$$

774 Here, L represents the total number of genome-wide analyzed sites. Of a given
775 population sample, $n_r(l)$ represents the read depth of the top two alleles at the l th site (i.e., SNP)
776 and m_l represents the minor allele count. n_c as a normalization factor represents the haploid
777 sample size for either autosomes or X chromosome in a pool (Table S1). Strictly speaking, n_c as
778 a normalization factor should represent equally contributing chromosomes in a pool.
779 Nevertheless, for our data it is sufficient to use haploid sample size for either autosomes or X
780 chromosome to approximate n_c in the above equation, as the estimation of $\hat{\theta}_\Pi$ is not substantially
781 affected by the precise value of n_c when the number of individuals in the pool is large. The
782 above formula is a simplified version for SNP data, based on equation 3 in Ferretti et al. (2013).

783 To examine patterns of polymorphism across chromosome arms, we also estimated
784 window nucleotide diversity (π_W) for all polymorphic sites. Each window was defined as a
785 continuous genomic region that includes 125,000 analyzed sites (Figure 3). Since chromosomal
786 identity was required in this analysis, we only took windows from 32 major contigs that contain
787 at least one full-size window and were unambiguously mappable to a chromosome arm of the *D.*
788 *melanogaster* dm6 genome assembly. Although such contigs only make up 57% of the *D. suzukii*
789 genome assembly, they contain a relatively larger proportion of all identified SNPs (72%), and
790 thus are still representative of genome-wide polymorphism.

791 To estimate genome-wide pairwise F_{ST} between populations, we adopted an unbiased
792 multi-loci estimator known as Reynolds' estimator of the coancestry coefficient, which accounts
793 for unequal sample sizes among populations and is applicable for more than two alleles at a site
794 (Reynolds et al. 1983). We first heuristically partitioned the genome into windows that exceeded

795 a cross-sample average accumulated heterozygosity threshold of 100. The genome-wide F_{ST} was
 796 then calculated as an average of window- F_{ST} $F_{ST}(w)$ weighted by the number of analyzed sites
 797 within each window ($L(w)$), where $F_{ST}(w)$ was calculated as a weighted average of single-site
 798 ratio estimators. Numerically,

$$799 \quad F_{ST} = \frac{1}{L} \sum_w L(w) F_{ST}(w) \quad (2)$$

$$800 \quad F_{ST}(w) = (\sum_l a_l) / \sum_l (a_l + b_l) \quad (3)$$

801 where

$$802 \quad a_l = \frac{1}{2} \sum_u (\tilde{p}_{1lu} - \tilde{p}_{2lu})^2 - \frac{(n_{1l} + n_{2l})(n_{1l}\tilde{\alpha}_{1l} + n_{2l}\tilde{\alpha}_{2l})}{4n_{1l}n_{2l}(n_{1l} + n_{2l} - 1)} \quad (4)$$

$$803 \quad a_l + b_l = \frac{1}{2} \sum_u (\tilde{p}_{1lu} - \tilde{p}_{2lu})^2 + \frac{(4n_{1l}n_{2l} - n_{1l} - n_{2l})(n_{1l}\tilde{\alpha}_{1l} + n_{2l}\tilde{\alpha}_{2l})}{4n_{1l}n_{2l}(n_{1l} + n_{2l} - 1)} \quad (5)$$

804 At the l th site in each population, \tilde{p}_{1lu} and \tilde{p}_{2lu} represent the frequency of the u th allele at the
 805 l th site; $\tilde{\alpha}_{1l}$ and $\tilde{\alpha}_{2l}$ represent the heterozygosity; n_{1l} and n_{2l} represent the sample size. Unlike
 806 the sequencing of individual genomes, pool-seq induces an uncertainty in the number of
 807 individual alleles actually sequenced at a locus (i.e., effective sample size), and this uncertainty
 808 decreases slowly even at high read depth (Ferretti et al. 2013). Since the sample size is an
 809 important parameter for F_{ST} estimation, we took measures to obtain an estimate of the effective
 810 sample size, n_{il} , at each given site (Ferretti et al. 2013). Numerically,

$$811 \quad n_{il} = \sum_j^{n_c} j P_c(j | n_r, n_c) \quad (6)$$

812 where

813

$$814 \quad P_c(j | n_r, n_c) = \frac{n_c! S(n_r, j)}{(n_c - j)! n_c^{n_r}} \quad (7)$$

815 Here, we explicitly estimated the probability of the number of j unique lineages sampled at a site
816 given n_r sampled reads and n_c equally contributing chromosomes in a pool, where $S(n_r, j)$ are
817 the Stirling numbers of the second kind, defined as the number of ways to partition n_r reads into
818 j non-empty sets (Ferretti et al. 2013). We then estimated n_{il} as the expected number of lineages
819 for each n_r and n_c . Ideally, n_c should be estimated as $2n_e$, where n_e is the effective pool size
820 representing the number of diploid individuals contributing the same amount of reads to a pool
821 (Gautier et al. 2013; Lange et al. 2022). Although we lack sample replicates to estimate n_e and
822 therefore used haploid sample size for n_c as an approximation, the probability estimation is still
823 reasonable given that our number of lineages for each pool is large (Table S1) (Ferretti et al.
824 2013).

825 Lastly, we estimated genome-wide pairwise D_{XY} as an absolute measure of population
826 differentiation that is independent of levels of within-population diversity. It was calculated as
827 pairwise differences per site between two populations, divided by L total analyzed sites (Nei 1987;
828 Hahn 2018). Numerically,

$$829 \quad D_{XY} = \frac{1}{L} \sum_l \sum_{ij} x_i y_j k_{ij} \quad (8)$$

830 where x_i and y_j represent frequencies of the i th allele from population X and the j th allele from
831 population Y , and k_{ij} is either 1 or 0, depending on whether or not the alleles differ at the l th
832 site. Calculations in this section were all implemented with Python and Shell scripts (see Data
833 Availability).

834

835 **Building Population Trees and Admixture Graphs**

836 To infer the population structure assuming no migrations, we used MEGA X (Stecher et
837 al. 2020) to build neighbor-joining population trees (Figure S5) for autosomes and X
838 chromosome from Reynolds' distance, which is a transformation of pairwise $F_{ST} : D = -\ln(1 -$
839 $F_{ST})$ (Reynolds et al. 1983). Since the position of the tree root is not identifiable through this
840 analysis, and the China-Ningbo population was identified through diversity and distance analyses
841 as a potentially basal lineage among the sampled populations, we rooted the tree along the
842 branch to Ningbo.

843 To further investigate population splits and mixtures, we used TreeMix (Pickrell and
844 Pritchard 2012) to infer population trees using allele counts that were corrected based on
845 effective sample size. The input SNPs were a subset of our total identified SNPs whose total
846 minor allele count across all samples was no less than ten, as it would allow us to complete this
847 computationally intensive analysis in a reasonable amount of time and avoid biases on tree
848 construction from false positive SNPs. We combined SNPs into blocks of 54 (-k 54) to account
849 for linkage disequilibrium across windows that have a median length of ~500 bp.

850 To determine the number of migration events (-m) that improved the model fit while
851 avoiding overfitting and losing interpretability, we ran TreeMix on our actual SNP data with no
852 migration and with -m from one to 20. Since the position of the root is only partially identifiable
853 through TreeMix inference, we rooted the tree at China-Ningbo, as suggested by Pickrell &
854 Pritchard (2012). We then estimated the fraction of the variance in relatedness between
855 populations that is accounted for by each model (f). We found that f began to plateau for both
856 the autosomal and X-linked models at $m = 11$. For the X chromosome model, we also
857 confirmed that all migration edges were statistically significant ($p < 0.05$) based on calculating
858 their p-values from the jackknife estimates of the migration weights and their standard error

859 (Pickrell and Pritchard 2012), whereas this analysis remained computationally intractable for the
860 autosomal case. We interpreted the migration weights w as an admixture proportion, as it was
861 correlated with ancestry fraction in previous simulations, although it tends to be a more
862 conservative approximation for high admixture proportions (Pickrell and Pritchard 2012).

863

864 **Preparing Environmental Data**

865 To generate environmental data for GEA, we selected a preliminary set of 26 candidate
866 environmental variables representing geographic, climatic and land cover-related factors (Figure
867 4A) that may be relevant in the adaptation process of *D. sukukii* based on prior knowledge
868 (Kellermann et al. 2012; Bogaerts-Márquez et al. 2021). With R packages ‘raster’ (v. 3.5.2) and
869 ‘SpatialPoints’ (v. 1.4.6), we retrieved environmental data of high spatial resolution ($\sim 100 \text{ km}^2$)
870 in batch for the sampling locations of our 29 populations from online databases WorldClim (Fick
871 and Hijmans 2017) and Esri 2020 Land Cover (Karra et al. 2021). Annual mean values of
872 monthly climatic variables, including mean wind speed, solar radiation, and water vapor
873 pressure, were derived by averaging across 12 months of data.

874 Due to the large number of statistical tests that would result from running GEA on all the
875 environmental variables one by one, there is an increased difficulty in controlling rates of false
876 discovery. Additionally, including multiple highly correlated variables in a model would lead to
877 multicollinearity issues (Rellstab et al. 2015). To avoid these problems, we calculated a pairwise
878 Pearson correlation matrix from values of environmental factors across sampled locations (Figure
879 4A; Table S3), and then selected a subset of nine least correlated environmental variables for
880 one-by-one GEA analyses (Figure 4B). To avoid scale inconsistencies between estimated GEA
881 statistics, the environmental differentiation of each population was calculated as the absolute

882 difference between the environmental value of that population and the average across all
883 populations, standardized by the standard deviation (de Villemereuil and Gaggiotti 2015). This
884 standardized differentiation was then input to GEA (Table S3).

885

886 **Environmental Association Analyses**

887 To characterize the environmental adaptation of *D. sukukii*, we scanned the whole
888 genome for adaptive loci using the F_{ST} -based GEA method BayesScEnv (de Villemereuil and
889 Gaggiotti 2015). We chose this specific approach over other GEA methods mainly because it has
890 a lower false positive rate than other GEA approaches in the presence of hierarchical population
891 structure. It also allows for detecting patterns of allele frequency that are not linearly dependent
892 on environmental factors (Rellstab et al. 2015; de Villemereuil and Gaggiotti 2015).

893 For each environmental variable, the association analyses tested the relationship between
894 environmental and genetic differentiation among populations, for 5,752,156 genome-wide SNPs
895 with a MAF higher than 5%. To control for false positives, we chose stringent model parameters
896 expected to yield extremely conservative results, setting the prior probability of non-neutral
897 models as 0.02 (-pr_jump 0.02) and the prior probability of the competing environment-
898 unrelated locus-specific model as 0.9 (-pr_pref 0.9). These parameters correspond to assumptions
899 that genetic differentiation reflects the action of natural selection in just 2% of the genome, and
900 the focal environmental variable is only expected to be involved at 10% of the non-neutral loci.

901 To make this GEA analysis computationally feasible with our large SNP set, while still
902 analyzing all qualifying SNPs, we applied a split-run strategy: we subsampled SNPs across
903 concatenated sequences of contigs within the autosomes and the X chromosome separately, and

904 then ran subsamples with BayesScEnv in parallel. Since the null model of population structure is
905 estimated separately in each run, we subsampled non-adjacent SNPs at a fixed interval to limit
906 locus-specific biases in that estimation, where the length of the interval between jointly analyzed
907 SNPs was equal to the total number of subsamples. With a targeted subsample/interval size of
908 up to 10,000 SNPs, we divided the concatenated autosomal contigs into 490 subsamples (with
909 actual subsample sizes of 9982-9983 SNPs), and the concatenated X-linked contigs into 87
910 subsamples (with actual subsample sizes of 9893-9894 SNPs). Hence, the first autosomal
911 subsample contained SNP #1, SNP #491, and so on.

912 Convergence of each run was confirmed with the R package ‘CODA’. Individual runs
913 were then merged across autosomes and X chromosome to calculate the genome-wide q -value
914 (q) of locally-estimated posterior error probability (*PEP*) across all sites, where we targeted a false
915 discovery rate (*FDR*) of 5% by setting the q threshold at 0.05 (Storey 2003; Muller et al. 2006).
916 For downstream analyses, to remove redundancy due to linkage disequilibrium, we pared down
917 closely linked candidate sites by maintaining the site with the lowest q within each 20 kb genomic
918 window. To assess the relative levels of support for associations between SNPs and a given
919 environmental variable, we ranked all candidate loci first by q and then by the estimated g
920 parameter as a tie-breaker, which measures the sensitivity of a locus to environmental
921 differentiation.

922

923 **Identifying Candidate Genes**

924 For each candidate SNP, the closest gene in each direction within a 200-exon flanking
925 region that overlapped with it was considered to be associated with that variant, in order to
926 encompass both potential coding and regulatory adaptation. To facilitate clear comparisons

927 among environmental variables with different numbers of significant variants, we focused on the
928 top 500 candidate genes that were linked to variants with the lowest significant q and highest g
929 within each environmental variable (Table S5).

930

931 **GO Enrichment and Semantic Clustering**

932 Gene ontology (GO) enrichment of the top 500 candidate genes associated with
933 candidate SNPs was performed via genomic permutation of outlier SNP positions (100,000,000
934 replicates), which accounts for the variability of gene length and the clustering of functionally
935 related genes, as described in previous work (Pool et al. 2017). For each GO category, a p-value
936 indicated the proportion of permutation replicates in which an equal or greater number of genes
937 was implicated.

938 We then prioritized the most informative and significant GO terms and removed
939 redundant terms that potentially share similar groups of genes by clustering GO terms based on
940 their semantic similarity and ranking representative terms of each cluster by their p-value
941 (Reijnders and Waterhouse 2021). For GO terms that were shared among associations with
942 multiple environmental variables, a combined p-value was calculated from the p-values of
943 independent enrichment tests using Fisher's method (Fisher 1938).

944

945 **Supplementary Material**

946 Supplementary figures S1–S9 and tables S1–S6 are available at Molecular Biology and
947 Evolution online.

948

949 **Data Availability**

950 All sequence data generated for this project are available from the NIH Short Read
951 Archive under project PRJNA973110, with specific sample information given in Table S1,
952 Supplementary Material online. All computational scripts created for this study have been
953 uploaded to <https://github.com/Sfeng666/poolWGS2SNP> and
954 https://github.com/Sfeng666/Dsuz_popgen_GEA (last accessed May 27, 2023).

955

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965

966

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