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Heavy metal tolerance traits of filamentous fungi isolated from gold and gemstone mining sites



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ABSTRACT

Increased environmental pollution has necessitated the need for eco-friendly clean-up strategies. Filamentous fungal species from gold and gemstone mine site soils were isolated, identified and assessed for their tolerance to varied heavy metal concentrations of cadmium (Cd), copper (Cu), lead (Pb), arsenic (As) and iron (Fe). The identities of the fungal strains were determined based on the internal transcribed spacer 1 and 2 (ITS 1 and ITS 2) regions. Mycelia growth of the fungal strains were subjected to a range of (0–100 Cd), (0–1000 Cu), (0–400 Pb), (0–500 As) and (0–800 Fe) concentrations (mgkg^{-1}) incorporated into malt extract agar (MEA) in triplicates. Fungal radial growths were recorded every three days over a 13-days' incubation period. Fungal strains were identified as *Fomitopsis meliae*, *Trichoderma ghanense* and *Rhizopus microsporus*. All test fungal exhibited tolerance to Cu, Pb, and Fe at all test concentrations ($400\text{--}1000\text{ mgkg}^{-1}$), not differing significantly ($p > 0.05$) from the controls and with tolerance index >1 . *T. ghanense* and *R. microsporus* demonstrated exceptional capacity for Cd and As concentrations, while showing no significant ($p > 0.05$) difference compared to the controls and with a tolerance index >1 at 25 mgkg^{-1} Cd and 125 mgkg^{-1} As. Remarkably, these fungal strains showed tolerance to metal concentrations exceeding globally permissible limits for contaminated soils. It is envisaged that this metal tolerance trait exhibited by these fungal strains may indicate their potentials as effective agents for bioremediative clean-up of heavy metal polluted environments.

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Introduction

Increased heavy metal contamination of soil and water environments¹ has necessitated the need for clean-up strategies. Recently, diverse eco-friendly remediation options have been explored for the restoration of contaminated environments. These remediation options, among others, include the use of plants (phytoremediation),² bacteria (bacterial bioremediation)³ and fungi (mycoremediation).⁴ The employability of these bio-resources (plants, bacteria and fungi) for effective bioremediation has been well reported.^{2–4}

At present of these options, mycoremediation strategy has received increased attention in the bioremediation of contaminated/polluted environments due to its reasonably low cost implications and significant success outcomes.^{5–8} Filamentous fungal species have been identified for their distinct attributes (ability to thrive under extreme pH, temperature and nutrient variability conditions, as well as tolerance to high metal concentrations)^{9–11} and hence their effective remediation traits of contaminated sites.

Metal tolerance/resistance has been defined as the ability of an organism to survive metal toxicity by means of one or more mechanisms devised in direct response to the metal(s) concerned.^{7,12} Metal tolerance by filamentous fungi has been associated with their sites of isolation, toxicity of the metal tested, its concentration in medium, and on the isolate's competence.¹⁰ Contaminated sites are known as principal sources of metal-resistant species^{18–22} with indigenous fungal strains isolated from heavy metal contaminated sites exhibiting notable tolerance for high heavy metal concentrations.^{9,21,23–25}

However, of more importance is the specific and non-specific heavy metal tolerance mechanisms adopted by fungal species. According to Vadkertiova and Slavikova¹³ the introduction of heavy metals into the environment has induced physiological and morphological adaptation strategies in the microbial community. Specifically, fungal species adopt one or more metal tolerance strategies which include extracellular metal sequestration and precipitation, suppressed influx, enhanced metal efflux, production of intracellular/extracellular enzymes, metal binding to cell walls, intracellular sequestration and complexation.^{14–17}

Several metal-tolerant filamentous fungi (*Rhizopus*, *Trichoderma*, *Aspergillus*, *Penicillium*, and *Fusarium*) have been isolated from multiple heavy metal contaminated soils.⁷ Zafar et al.⁷ reported that *Rhizopus* sp., isolated from metal-contaminated agricultural soils tolerated Cd and Cr concentrations. In addition, Volesky²⁶ observed that the mycelium of a *Rhizopus* specie was biosorbent towards Pb, Cd, Cu and Zn. *Trichoderma* species have also been known to exhibit tolerance to a range of toxicants^{27–29} and Cu, Cd, As and Zn heavy metals in vitro conditions.^{8,23,27,30–34}

However, there is a dearth of knowledge of the growth response and heavy metal tolerance of filamentous fungal species isolated from gold and gemstone mining sites. This study was therefore designed to isolate, identify and assess the growth response and tolerance/resistance of filamentous fungi isolated from gold and gemstone mining sites to varied

concentrations of selected heavy metals associated with mining sites.

Materials and methods

Study sites and soil sampling

Mine site soils used in this study were obtained from gemstone and gold mining sites in Southwestern, Nigeria namely: Awo (7°46' N, 4°24' E) and Itagunmodi (7°30' N, 4°49' E) as described.^{1,4,35} From previous studies,^{1,4,35} soil preliminary heavy metal analysis of the sites recorded elevated concentrations of 0.20–0.35 mgkg⁻¹ Cd; 3.68–48.60 mgkg⁻¹ Cu; 19.05–35.00 mgkg⁻¹ Pb; 20.45–34.80 mgkg⁻¹ As and 240.24–296.18 mgkg⁻¹ Fe.

Isolation of soil fungi

Isolation of soil fungi was performed by serial dilution and the spread plate method using malt extract agar (MEA) medium and incubated at 30°C for five days as previously described.^{4,35} Streptomycin (35 mgmL⁻¹) was added as a supplement into the medium to inhibit bacterial growth. After incubation, isolates of single spores were successively sub-cultured on MEA to obtain pure isolates. Fungal species were characterized on the basis of phenotypical/macrosopic observation (pigmentation, shape, diameter, colony appearance and texture) and microscopic examination (septation of mycelium, shape, form, diameter and texture of spore/conidia). The cultural and morphological characteristic features of the fungal species were compared with those described.³⁶ Fungal species were then selected for genotypic-based identification.

DNA extraction and PCR amplification

The ZR fungal/bacterial DNA kit (Zymo Research, Irvine, CA, USA) was used to extract genomic DNA from pure 5-day old fungal cultures according to the manufacturer's manual. About 40 mg (wet wt.) mycelium was harvested aseptically into the ZR BashingBead™ lysis tube and lysed in 750 µL of lysis buffer by bead beating. The lysate was then centrifuged at 13,400 rpm for 300 s to obtain clear supernatant. Further protocols, which are binding, wash steps and elution of DNA were performed as instructed by the manufacturer. The quality and integrity of the extracted DNA were verified on 1% agarose gel, while DNA concentration and purity were verified using NanoDrop spectrophotometer (ND-1000, NanoDrop Technologies Inc., Wilmington, Delaware USA).

Taxonomic identification of isolates was between the internally transcribed spacer regions – 1 (ITS1) and 2 (ITS2). DNA amplification was done using primer sets ITS1 (5'-TCCGTAGGTGAACCTGCGG-3') and ITS4 (5'-TCCTCCGCTTATTGATATGC-3').³⁷ Each PCR reaction contained 12.5 µL of 2× Dream Master mix (Thermo Scientific Technologies, Waltham, MA, USA), 50 ng DNA template, 0.2 M of each forward and reverse primers and nuclease-free water to a final volume of 25 µL. PCR was performed in a C1000™ thermal cycler (Bio-Rad, Hercules, CA, USA) involving an initial denaturation at 95°C for 5 min, 29 cycles of denaturation at

95°C for 30 s, annealing at 55°C for 30 s and extension at 72°C for 60 s. The amplification process was terminated by a final extension of 72°C for 5 min. The PCR amplicons were then verified on 1.5% agarose gel after electrophoresis.

Sequencing and phylogenetic reconstruction

Purified PCR amplicons were sequenced by using forward primer ITS1 and the Big Dye terminator v3.1 cycle sequencing kit (Applied Biosystems, Warrington, UK) on a 3130 Genetic analyzer (Applied Biosystems/Hitachi, Tokyo, Japan). Sequence electropherograms were inspected manually and edited with FinchTV (v. 1.4.0; <http://www.geospiza.com/Products/finchtv.shtml>). For taxonomic assignment, edited sequences were aligned with sequences on the UNITE ITS database (<https://unite.ut.ee/index.php>) while for, phylogenetic reconstruction, the sequences, together with closely related sequences in the GenBank were selected. Multiple sequence alignment of the obtained sequences was done using MUSCLE³⁸ integrated in MEGA V. 6.0.³⁹ The resulting multiple sequence alignments were then edited manually and rectified for gaps using DAMBE software.⁴⁰ Phylogeny dendrograms were constructed using the neighbour-joining method of the Tamura–Nei substitution model and a thousand bootstrap replications in MEGA.

Heavy metal tolerance assay

Isolated filamentous fungi were assessed for heavy metal tolerance at varying Cd, Cu, Pb, As and Fe concentrations. Filter (0.25 µm pore size) sterilized heavy metal salts of CdCl₂, CuSO₄, PbSO₄, AsSO₄ and Fe₂SO₄ were incorporated into sterile MEA. Media were supplemented with 35 mgmL⁻¹ streptomycin and pH was maintained at 5.6 by the addition of 3 M NaOH. The experiment was conducted in triplicates with control and four other varied test concentrations. Heavy metal concentrations (mgkg⁻¹) were: (25, 50, 75 and 100) cadmium, (125, 250, 500 and 1000) copper, (100, 200, 300 and 400) lead, (125, 250, 375 and 500) arsenic and (200, 400, 600 and 800) iron. The non-amended medium served as a control.

Test fungal strain of 8 mm diameter disks from 7-day old pure culture each were individually inoculated into an 8 mm well aseptically bored at the centre of control and test MEA plates. All plates were incubated at 29 ± 1°C for 13 days, during which mycelial radial growth was monitored and recorded every three days. Heavy metal tolerance potential of the fungal species in the test medium was calculated in relation to the control radial growths (Eq. (1)). Fungi heavy metal tolerance was rated thus: 0.00–0.39 (very low tolerance), 0.40–0.59 (low tolerance), 0.60–0.79 (moderate tolerance), 0.80–0.99 (high tolerance) and 1.00–>1.00 (very high tolerance), the higher the values, the higher the fungal tolerance to the heavy metal.

$$\text{Tolerance index} = \frac{\text{Radial growth (mm) of test fungus in heavy metal incorporated medium}}{\text{Radial growth (mm) of fungus in non-incorporated medium}} \quad (1)$$

Statistical analysis

Statistical analysis of data obtained was done using one-way analysis of variance (ANOVA) at 5% level of significance using

the Statistical Package for Social Sciences (SPSS) version 23 (IBM, Armonk, NY, USA). A post hoc test was performed using the Duncan's New Multiple Range Test.

Results

Fungi identification

Three indigenous fungal species isolated from gold and gemstone mining sites were identified. The ITS-based taxonomic assignment of the fungal strains confirmed the identities of *Fomitopsis meliae*, *Trichoderma ghanense* (two isolates) and *Rhizopus microsporus* (Table 1). The isolation sources of the species revealed the presence of two genera – *Fomitopsis* and *Trichoderma* from Itagunmodi, the gold mining site and *Rhizopus* from the gemstone mine site. The evolutionary relatedness of the isolates with similar GenBank sequences further confirmed the identities of the strains (Fig. 1).

Growth response of tested fungal strains in heavy metal-rich media

Mycelia growth response of *F. meliae*, *T. ghanense* and *R. microsporus* to varied concentrations of cadmium, copper, lead, arsenic and iron differed among the species (Fig. 2).

On exposure to all cadmium and arsenic concentrations, *F. meliae* exhibited inhibited growth, with mycelia growths differing significantly ($p < 0.05$) compared to the control. Although, when exposed to varied Cu, Pb and Fe-enriched media, a divergent trait was displayed as *F. meliae* revealed no statistical ($p > 0.05$) difference in radial growth compared to the control. With respect to *T. ghanense* and *R. microsporus* strains, no statistical ($p > 0.05$) differences were obtained in the radial growth of the strains compared to their controls in Cd (25–100 mgkg⁻¹), Cu (125–1000 mgkg⁻¹), Pb (100–400 mgkg⁻¹), As (125–500 mgkg⁻¹) and Fe (200–800 mgkg⁻¹) enriched media.

Overall, a growth response trend of Fe = Cu = Pb > As = Cd was observed among the fungal strains to heavy metal concentrations. In terms of their response in heavy metal rich-media, a trend showing *T. ghanense* > *R. microsporus* > *F. meliae* was observed. Generally, *F. meliae* tolerated 400 mgkg⁻¹ Pb, 800 mgkg⁻¹ Fe and 1000 mgkg⁻¹ Cu concentrations, but revealed high inhibition to all Cd and As concentrations. On the other hand, *T. ghanense* and *R. microsporus* showed tolerance to elevated Cd (100 mgkg⁻¹), Pb (400 mgkg⁻¹), Fe (500 mgkg⁻¹), As (800 mgkg⁻¹) and Cu (1000 mgkg⁻¹) concentrations. Comparing the response of these fungal species to heavy metal limits for contaminated soils, it was observed that the fungal species far exceeded the set permissible limits (Table 2) except *F. meliae* which was intolerant to all Cd and As concentrations.

Table 1 – Taxonomic identification of fungal species with similarity on the UNITE ITS database.

LAB-ID	Sample origin (mining site)	Closest relative	Sequence similarity (%)	Accession number
FUG-07	Itagunmodi	<i>Trichoderma ghanense</i>	99.5	KT819140
FUG-08	Itagunmodi	<i>Trichoderma ghanense</i>	99.8	KT819141
FUG-09	Itagunmodi	<i>Fomitopsis meliae</i>	97.5	KT819142
FUG-14	Awo	<i>Rhizopus microsporus</i>	100	KT819147

Tolerance index rating of *F. meliae*, *T. ghanense* and *R. microsporus* to Cd, Cu, Pb, As and Fe concentrations

In ascertaining the tolerance of the test fungal strains to heavy metal concentrations (Fig. 2), we further evaluated the heavy metal tolerance levels of the fungal species. This was obtained by calculating the tolerance index of the test fungal species relative to their controls using the mycelia radial growth data in heavy metal enriched media (Eq. (1)).

The tolerance rating of *F. meliae* to 25–100 mgkg⁻¹ cadmium and 125–500 mgkg⁻¹ arsenic concentrations were observed to be very low, with tolerance indices ranging between 0.17 and 0.32 (Table 3). However, in Cu and Pb concentrations, *F. meliae* indicated high tolerance revealed by the very high

tolerance index values of 1.02–1.32 in 125–500 mgkg⁻¹ Cu and 0.96–1.28 in 100–300 mgkg⁻¹ Pb. At higher Pb (400 mgkg⁻¹) and Cu (1000 mgkg⁻¹) concentrations, a decreased tolerance was indicated by *F. meliae* with tolerance index values of 0.67 and 0.78 respectively.

For *T. ghanense* and *R. microsporus* in all Cd concentrations, an indication of high to very high tolerances of 0.80–1.13 and a 0.84–1.01 tolerance index were obtained respectively. In addition, in Cu and Pb concentrations, the species indicated very high tolerance with index ranges of 1.02–1.27. In arsenic enriched-media, *T. ghanense* indicated moderate (0.72) to high (0.98) tolerance while *R. microsporus* indicated high to very high tolerance at 375 and 500 mgkg⁻¹ and 125 and 250 mgkg⁻¹ concentrations. Exceptionally, the three fungal species

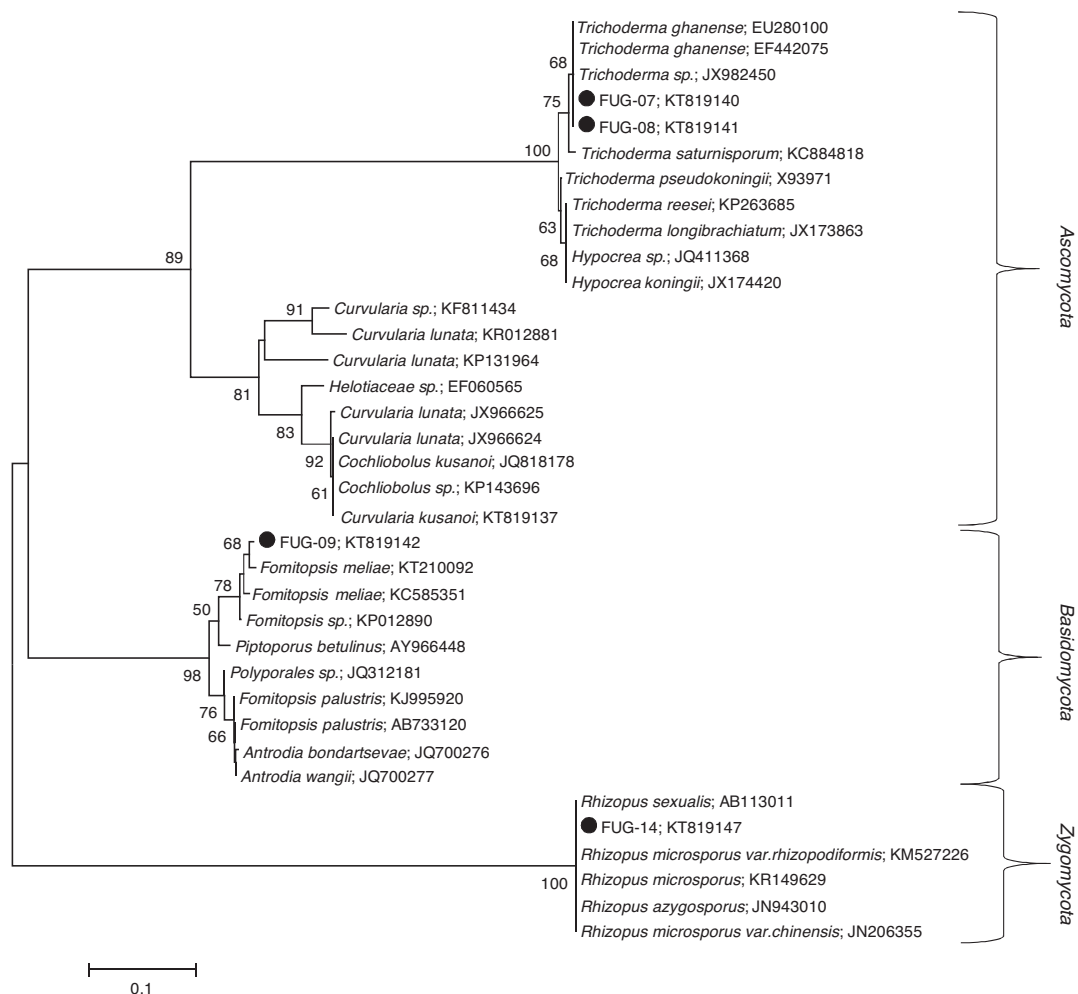


Fig. 1 – Unrooted neighbour-joining tree of fungal species. Sequences obtained in this study are indicated shaded circles (●). Neighbour-joining tree was constructed in Mega (Version 6) by using the Tamura–Nei substitution model and a thousand bootstrap replications. Bootstrap values below 50 are not shown.

Table 2 – Tolerance capability of fungal species to heavy metal concentrations (mgkg⁻¹).

Heavy metals	Fungi	Highest metal concentration (mgkg ⁻¹) tolerated in media	^a World permissible limit in soils (mgkg ⁻¹)
Cadmium	<i>F. meliae</i>	^b NT	0.41
	^d <i>T. ghanense</i>	100	
	<i>R. microsporus</i>	100	
Copper	<i>F. meliae</i>	1000	38.90
	^d <i>T. ghanense</i>	1000	
	<i>R. microsporus</i>	1000	
Lead	<i>F. meliae</i>	400	27.0
	^d <i>T. ghanense</i>	400	
	<i>R. microsporus</i>	400	
Arsenic	<i>F. meliae</i>	^b NT	20.0
	^d <i>T. ghanense</i>	500	
	<i>R. microsporus</i>	500	
Iron	<i>F. meliae</i>	800	^c
	^d <i>T. ghanense</i>	800	
	<i>R. microsporus</i>	800	

^a FAO⁴¹ and Kabata-Pendias.⁴²

^b 'NT' – not tolerated at any concentration (mgkg⁻¹).

^c Not available. Dependent on different soil parental constituents.

^d Mean concentrations of taxonomically similar fungal identities in the study was used.

Table 3 – Tolerance index levels of fungal strains in metal-rich media concentrations.

Heavy metals	Fungi	Concentrations (mgkg ⁻¹)			
Cadmium		25	50	75	100
	<i>F. meliae</i>	0.17	0.17	0.17	0.17
	^a <i>T. ghanense</i>	1.13	0.85	0.96	0.80
	<i>R. microsporus</i>	1.01	0.99	0.99	0.84
Copper		125	250	500	1000
	<i>F. meliae</i>	1.32	1.12	1.02	0.78
	^a <i>T. ghanense</i>	1.25	1.27	1.27	1.27
	<i>R. microsporus</i>	1.02	1.02	1.02	1.02
Lead		100	200	300	400
	<i>F. meliae</i>	1.27	1.28	0.96	0.67
	^a <i>T. ghanense</i>	1.20	1.25	1.25	1.27
	<i>R. microsporus</i>	1.02	1.02	1.02	1.02
Arsenic		125	250	375	500
	<i>F. meliae</i>	0.32	0.21	0.17	0.17
	^a <i>T. ghanense</i>	0.98	0.91	0.72	0.87
	<i>R. microsporus</i>	1.02	1.02	0.99	0.86
Iron		200	400	600	800
	<i>F. meliae</i>	1.16	1.32	1.38	1.45
	^a <i>T. ghanense</i>	1.25	1.27	1.27	1.25
	<i>R. microsporus</i>	1.02	1.02	1.02	1.02

Tolerance index rating values indicate:

0.00–0.39 – very low metal tolerance.

0.40–0.59 – low metal tolerance.

0.60–0.79 – moderate metal tolerance.

0.80–0.99 – high metal tolerance.

1.00–>1.00 – very high metal tolerance.

^a Mean concentrations of taxonomically similar fungal identities in the study was used.

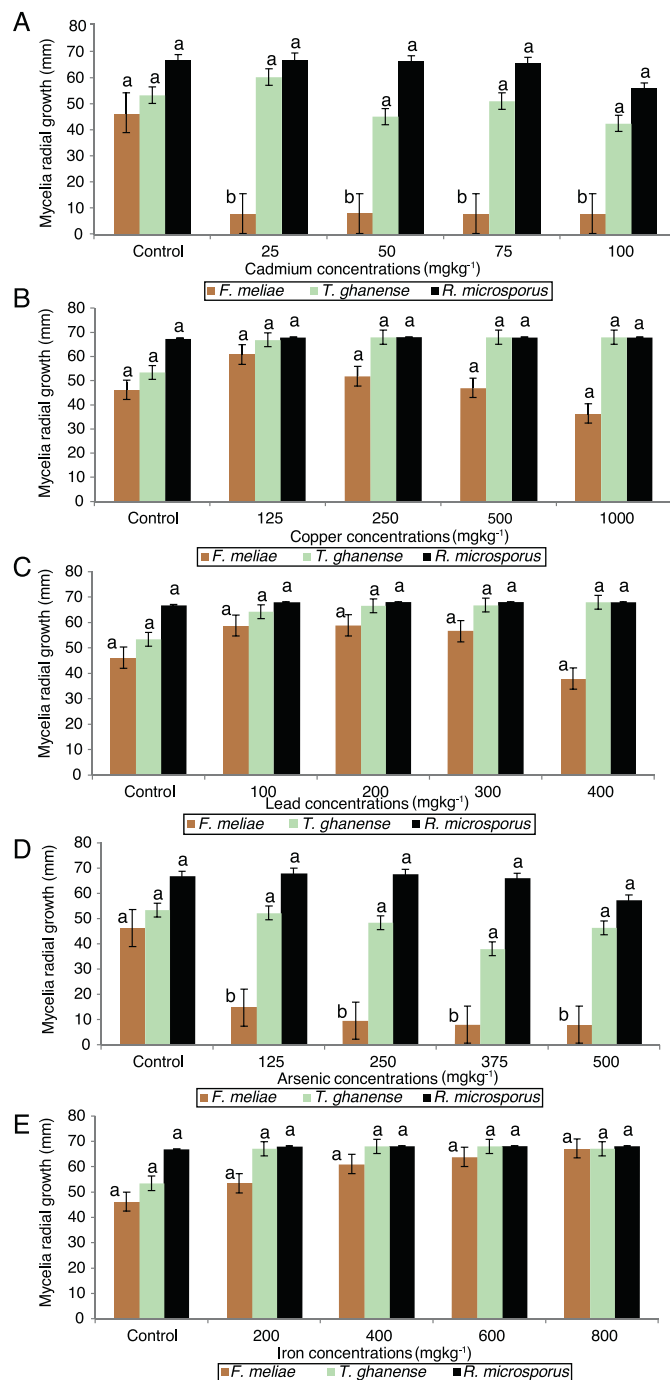


Fig. 2 – Effect of varied concentrations of (A) Cd, (B) Cu, (C) Pb, (D) As and (E) Fe on fungi radial growth (mm) over 13 days incubation period. Key – *F. meliae* (*Fomitopsis meliae*); *T. ghanense* (*Trichoderma ghanense*) and *R. microsporus* (*Rhizopus microsporus*). Means of 3 replicates (\pm SE). Bars of the same fungal species with different letters are significantly different ($p < 0.05$) according to Duncan's New Multiple Range Test.

indicated remarkably high tolerances (tolerance index ranges of 1.02–1.45) in all varied Fe enriched-media concentrations, with *F. meliae* demonstrating the highest tolerance index of 1.45 at 800 mgkg⁻¹.

When assessing the tolerance index of the fungal species, *R. microsporus* exhibited high to very high tolerance in all five heavy metal concentrations tested closely followed by *T.*

ghanense which revealed high to very high tolerance in Cd, Cu, Pb and Fe concentrations except arsenic. *F. meliae*, on the other hand, indicated high to very high tolerance in Cu, Pb and Fe concentration exposures but very low tolerance in Cd and As concentrations. On the whole, the tolerance levels of the species to the heavy metals showed a decreasing trend of Fe > Cu > Pb > As > Cd.

Discussion

The occurrence of *F. meliae*, *T. ghanense* and *R. microsporus* on heavy metal contaminated gold and gemstone mining sites was confirmed in this study. The presence of fungal species in various contaminated/polluted sites with elevated heavy metal concentrations has been well documented. Specifically, Zafar et al.⁷ and Fazli et al.⁴³ reported the occurrence of fungal strains in soils with elevated Cd, Cu, As and Zn concentrations. In addition, Anand et al.⁹ and Karcprzak and Malina²⁹ confirmed the presence of fungi in heavy metal polluted soils. Iram et al.,¹² Iskandar et al.,²¹ and López and Vázquez²³ also affirmed the occurrence of fungal species in sewage sludge water plants, heavy metal contaminated freshwater ecosystem and sewage and industrial waste waters respectively. Furthermore, Mo et al.,⁴⁴ Srivastava et al.⁴⁵ and Babu et al.⁴⁶ confirmed the existence of fungal strains in Pb and As polluted sites and mine tailings soils.

Of more importance is the marked tolerance displayed by these fungal species to heavy metals. Fungal species tolerate metals^{6,15,47} and thrive at elevated metal concentrations.^{9,24,48} In particular, indigenous filamentous fungi isolated from contaminated sites have shown tolerance to heavy metals.^{12,18,19,49} This exceptional trait may be attributed to the isolates' tolerance strategies to elevated heavy metal contaminations. Fomina et al.,¹⁴ Turnau et al.,¹⁵ Gadd¹⁶ and Vala and Sutariya¹⁷ reported that these tolerance mechanisms include metal binding to cell walls, production of intracellular/extracellular enzymes, intracellular sequestration, extracellular metal sequestration and precipitation, suppressed influx, enhanced metal efflux, and complexation.

Remarkable heavy metal tolerance was demonstrated by *R. microsporus* and *T. ghanense* species. *Trichoderma* and *Rhizopus* species have been widely reported for their notable tolerance to various heavy metals at varied concentrations.^{7,23,31,50} Some strains of *Rhizopus* and *Trichoderma* revealed high resistance to a range of heavy metals, such as Cd,^{11,23,26,44,50,51} Cu,^{21,26,46} Pb^{26,52} and As.^{17,45} Vala and Sutariya¹⁷ reported that *Rhizopus* species were highly tolerant to 25 and 50 mgkg⁻¹ As concentrations, which confirms the findings of this study. In addition, strains of *Trichoderma* tolerated Cd at 100 and 125 mgkg⁻¹,^{53,54} and Cu at 300 mgkg⁻¹,²³ 500 mgkg⁻¹,⁹ 800 mgkg⁻¹¹²¹ and 1000 mgkg⁻¹¹⁴⁵ concentrations. Furthermore, a strain of *Trichoderma* was found to tolerate Pb concentrations of 1000 mgkg⁻¹ in medium.²¹

All three fungal species demonstrated extraordinary preference for Fe at all concentrations as observed in their tolerance index values. This may be ascribed to the fact that iron serves as a micronutrient and is crucial in many metabolic processes.⁵⁵ In addition, Kosman,⁵⁶ Philpott⁵⁷ and Johnson⁵⁸ found that fungal species have a high affinity and capacity to take up Fe in various forms and variety. Aznar and Dellagi⁵⁵ and Neilands⁵⁹ stated that most fungal strains synthesize and secrete siderophores (small organic compounds that bind ferric Fe with high affinity and specificity) which they utilize to extract Fe from their environment. Furthermore, according to Kosman⁵⁶ microorganisms including fungi basically deploy three main strategies to increase iron solubility by acidifying

the environment, reducing ferric iron to a more soluble ferrous form and secreting soluble iron-chelating molecules.

Overall, *T. ghanense* and *R. microsporus* exhibited higher tolerance in Cd, Cu, Pb and As-enriched media compared to *F. meliae* which specifically displayed sensitivity to all Cd and As concentrations. Studies confirm that differing levels of metal resistance have been demonstrated by different fungal species isolated from the same source of metal-contaminated sites.^{12,60–64} This may be ascribed to variations in the tolerance mechanism utilized by the fungal species⁷ which is individually dependent.⁶⁵ In addition, the evident sensitivity to all Cd and As concentrations displayed by *F. meliae* may be attributed to the known toxicity of these heavy metals as reported.^{66–68}

Conclusion

Indigenous filamentous fungal species from gold and gemstone mine sites exhibited remarkable tolerance in heavy metal-rich media. Exposures of *F. meliae*, *T. ghanense* and *R. microsporus* to elevated Cu, Pb and Fe levels revealed high tolerance with index values >1. Furthermore, *T. ghanense* and *R. microsporus* demonstrated extraordinary tolerance for As and Cd concentrations with a tolerance index >1 at 25 mgkg⁻¹ Cd and 50 mgkg⁻¹ As. These exceptional traits displayed by these fungal species to elevated heavy metal levels may indicate the bioremediative potentials inherent in the indigenous filamentous fungal species.

Conflicts of interest

The authors declare no conflicts of interest.

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REFERENCES

1. Oladipo OG, Olayinka A, Awotoye OO. Maize (*Zea mays* L.) performance in organically remediated mine site soils. *J Environ Manage.* 2016;182:435–442.
2. Chirakkara R, Comeselle C, Reddy K. Assessing the applicability of phytoremediation of soils with mixed organic and heavy metal contaminants. *Rev Environ Sci Biotechnol.* 2016;15(2):299–326.
3. de la Cueva SC, Rodríguez CH, Cruz NOS, Contreras JAR, Miranda JL. Changes in bacterial populations during bioremediation of soil contaminated with petroleum hydrocarbons. *Water Air Soil Pollut.* 2016;227(3):1–12.
4. Oladipo OG, Awotoye OO, Olayinka A, Ezeokoli OT, Maboeta MS, Bezuidenhout CC. Heavy metal tolerance potential of *Aspergillus* strains isolated from mining sites. *Bioresour J.* 2016;(20):287–297.
5. Fu Y, Viraraghavan T. Fungal decolorization of dye wastewaters: a review. *Bioresour Technol.* 2001;79:251–262.

6. Baldrian P. Interactions of heavy metals with white-rot fungi. *Enzyme Microb Technol.* 2003;32:78–91.
7. Zafar S, Aqil F, Ahmad I. Metal tolerance and biosorption potential of filamentous fungi isolated from metal contaminated agricultural soil. *Bioresour Technol.* 2007;98:2557–2561.
8. Shukla D, Vankar PS. Role of *Trichoderma* species in bioremediation process: biosorption studies on hexavalent chromium. In: Gupta VK, Schmoll M, Herrera-Estrella A, Upadhyay RS, Druzhinina I, Tuohy MG, eds. *Biotechnology and Biology of Trichoderma*. Vol 30. USA: Elsevier Publishers; 2014:405–414.
9. Anand P, Isar J, Saran S, Saxena RK. Bioaccumulation of copper by *Trichoderma viride*. *Bioresour Technol.* 2006;97:1018–1025.
10. Ruta L, Paraschivescu C, Matache M, Avramescu S, Farcasanu IC. Removing heavy metals from synthetic effluents using “kamikaze” *Saccharomyces cerevisiae* cells. *Appl Microbiol Biotechnol.* 2010;85:763–771.
11. Puglisi I, Faedda R, Sanzaro V, Lo Piero AR, Petrone G, Cacciola SO. Identification of differentially expressed genes in response to mercury I and II stress in *Trichoderma harzianum*. *Gene.* 2012;506:325–330.
12. Iram S, Zaman A, Iqbal Z, Shabbir R. Heavy metal tolerance of fungus isolated from soil contaminated with sewage and industrial wastewater. *Pol J Environ Stud.* 2013;22(3): 691–697.
13. Vadkertiova R, Slavikova E. Metal tolerance of yeasts isolated from water, soil and plant environments. *J Basic Microbiol.* 2006;46:145–152.
14. Fomina MA, Alexander IJ, Colpaert JV, Gadd GM. Solubilization of toxic metal minerals and metal tolerance of mycorrhizal fungi. *Soil Biol Biochem.* 2005;37:851–866.
15. Turnau K, Orłowska E, Ryszka P, et al. Role of mycorrhizal fungi in phytoremediation and toxicity monitoring of heavy metal rich industrial wastes in southern Poland. *Soil Water Pollut.* 2006;23(3):533–551.
16. Gadd GM. Geomycology: biogeochemical transformations of rocks, minerals, metals and radionuclides by fungi, bioweathering and bioremediation. *Mycol Res.* 2007;111:43–49.
17. Vala AK, Sutariya V. Trivalent arsenic tolerance and accumulation in two facultative marine fungi. *Jundishapur J Microbiol.* 2012;5(4):542–545.
18. Gadd GM, Sayer GM. Fungal transformation of metals and metalloids. In: Lovely DR, ed. *Environmental Microbe–Metal Interactions*. American Soc Microbiol; 2000:237–256.
19. Malik A. Metal bioremediation through growing cells. *Environ Int.* 2004;30:261–278.
20. Machado MD, Santos MSF, Gouveia C, Soares HMVM, Soares EV. Removal of heavy metals using a brewer’s yeast strain of *Saccharomyces cerevisiae*: the flocculation as a separation process. *Bioresour Technol.* 2008;99:2107–2115.
21. Iskandar NL, Zainudin NA, Tan SG. Tolerance and biosorption of copper (Cu) and lead (Pb) by filamentous fungi isolated from a freshwater ecosystem. *J Environ Sci (China).* 2011;23:824–830.
22. Muñoz AJ, Ruiz E, Abriouel H, et al. Heavy metal tolerance of microorganisms isolated from wastewaters: identification and evaluation of its potential for biosorption. *Chem Eng J.* 2012;210:325–332.
23. López EE, Vázquez C. Tolerance and uptake of heavy metals by *Trichoderma atroviride* isolated from sludge. *Chemosphere.* 2003;50:137–143.
24. Deng Z, Cao L, Gaur A, Adholeya A. Prospects of arbuscular mycorrhizal fungi in phytoremediation of heavy metal contaminated soils. *Curr Sci.* 2004;86(4):528–534.
25. Cecia A, Maggia O, Pinzarib F, Persiani AM. Growth responses to and accumulation of vanadium in agricultural soil fungi. *Appl Soil Ecol.* 2012;58:1–11.
26. Volesky B. Advances in biosorption of metals: selection of biomass types. *FEMS Microbiol Rev.* 1994;14:291–302.
27. Harman GE, Lorito M, Lynch JM. Uses of *Trichoderma* spp. to remediate soil and water pollution. *Adv Appl Microbiol.* 2004;56:313–330.
28. Ezzi MI, Lynch JM. Biodegradation of cyanide by *Trichoderma* spp. and *Fusarium* spp. *Enzyme Microb Technol.* 2005;36:849–854.
29. Karcprzak M, Malina G. The tolerance and Zn²⁺, Ba²⁺ and Fe²⁺ accumulation by *Trichoderma atroviride* and *Mortierella exigua* isolated from contaminated soil. *Can J Soil Sci.* 2005;85:283–290.
30. Guillermina M, Romero M, Cazau M, Bucsinszky A. Cadmium removal capacities of filamentous soil fungi isolated from industrially polluted sediments, in La Plata (Argentina). *World J Microbiol Biotechnol.* 2002;18(9):817–820.
31. Harman GE, Howell CR, Viterbo A, Chet I, Lorito M. *Trichoderma* species opportunistic, a virulent plant symbionts. *Nat Rev Microbiol.* 2004;2:43–56.
32. Akhtar K, Khalid A, Akhtar M, Ghauri M. Removal and recovery of uranium from aqueous solutions by Ca-alginate immobilized *Trichoderma harzianum*. *Bioresour Technol.* 2009;100(20):4551–4558.
33. Zeng X, Su S, Jiang X, Li L, Bai L, Zhang Y. Capability of pentavalent arsenic bioaccumulation and biovolatilization of three fungal strains under laboratory conditions. *Clean: Soil Air Water.* 2010;38:238–241.
34. Tripathi P, Singh AM, Chauhan PS, et al. *Trichoderma*: a potential bioremediator for environmental clean up. *Clean Technol Environ Policy.* 2013;15(4):541–550.
35. Oladipo OG, Olayinka A, Awotoye OO. Ecological impact of mining on soils of Southwestern Nigeria. *Environ Exp Biol.* 2014;12:179–186.
36. Samson RA, Hoekstra ES, van Oorschot CAN. *Introduction of Food-Borne Fungi*. Amsterdam, The Netherlands: Institute of the Royal Netherlands Academy of Arts Science; 1984:248 pp.
37. White TJ, Bruns T, Lee S, Taylor J. Amplification and direct sequencing of fungal ribosomal RNA genes for phylogenetics. In: Innis MA, Gelfand DH, Sninsky JJ, White TJ, eds. *PCR Protocols: A Guide to Methods and Applications*. New York: Academic Press, Inc.; 1990:315–322.
38. Edgar RC. MUSCLE: multiple sequence alignment with high accuracy and high throughput. *Nucleic Acids Res.* 2004;32:1792–1797.
39. Tamura K, Stecher G, Peterson D, Filipski A, Kumar S. MEGA6: Molecular Evolutionary Genetics Analysis version 6.0. *Mol Biol Evol.* 2013;30:2725–2729.
40. Xia X. DAMBE5: a comprehensive software package for Data Analysis in Molecular Biology and Evolution. *Mol Biol Evol.* 2013;30:1720–1728.
41. Food and Agricultural Organization of the United Nations (FAO). *Plant Production and Protection Series – Agroclimatology Data for Africa. Vol 1: Countries North of the Equator*. Rome, Italy: FAO; 1984.
42. Kabata-Pendias A. *Trace Elements in Soils and Plants*. 4th ed. Florida: CRC Press, Taylor & Francis Gp; 2011, 534 pp.
43. Fazli MM, Soleimani N, Mehraabi M, Darabian S, Mohammadi J, Ramazani A. Highly cadmium tolerant fungi: their tolerance and removal potential. *J Environ Health Sci Eng.* 2015:13–19.
44. Mo MH, Chen WM, Zhang KQ. Heavy metal tolerance of nematode-trapping fungi in lead-polluted soils. *Appl Soil Ecol.* 2006;31:1–11.

45. Srivastava PK, Vaish A, Dwivedi S, Chakrabarty D, Singh N. Biological removal of arsenic pollution by soil fungi. *Sci Total Environ.* 2011;409:2430–2442.
46. Babu AG, Shea PJ, Oh BT. *Trichoderma* sp. PDR1-7 promotes *Pinus sylvestris* reforestation of lead-contaminated mine tailing sites. *Sci Total Environ.* 2014;476–477:561–567.
47. Qazilbash AA. Isolation and characterization of heavy metal tolerant biota from industrially polluted soils and their role in bioremediation. *Biol Sci.* 2004;41:210–256.
48. Adriaensen K, Vrålstad T, Noben JP, Vangronsveld J, Colpaert JV. Copper-adapted *Suillus luteus*, a symbiotic solution for pines colonizing Cu mine spoils. *Appl Environ Microbiol.* 2005;71:7279–7284.
49. Ashida J. Adaptation of fungi to metal toxicants. *Annu Rev Phytopathol.* 1965;3:153–174.
50. Babich H, Stotzky G. Effect of cadmium on fungi and on interactions between fungi and bacteria in soil: influence of clay minerals and pH. *Appl Environ Microbiol.* 1977;33:1059–1066.
51. Townsley CC, Ross IS, Atkins AS. Biorecovery of metallic residues from various industrial effluents using filamentous fungi. In: Lawrence RW, Branion RMR, Ebner HG, eds. *Fundamental Appl Biohydrometall.* Amsterdam: Elsevier; 1986.
52. Babu AG, Shim J, Bang KS, Shea PJ, Oh BT. *Trichoderma virens* PDR-28: a heavy metal-tolerant and plant growth-promoting fungus for remediation and bioenergy crop production on mine tailing soil. *J Environ Manag.* 2014;132:129–134.
53. Baldrian P, Gabriel J, Nerud F. Effect of cadmium on the ligninolytic activity of *Stereum hirsutum* and *Phanerochaete chrysosporium*. *Folia Microbiol.* 1996;41:363–367.
54. Datta B. Heavy metal tolerance of filamentous fungi isolated from metal-contaminated soil. *Asian J Microbiol Biotechnol Environ Sci.* 2015;17(4):965–968.
55. Aznar A, Dellagi A. New insights into the role of siderophores as triggers of plant immunity: what can we learn from animals? *J Exp Bot.* 2015;66(11):3001–3010.
56. Kosman DJ. Molecular mechanisms of iron uptake in fungi. *Mol Microbiol.* 2003;47(5):1185–1197.
57. Philpott CC. Iron uptake in fungi: a system for every source. *Biochim Biophys Acta.* 2006;1763(7):636–645.
58. Johnson L. Iron and siderophores in fungal–host interactions. *Mycol Res.* 2008;112:170–183.
59. Neilands JB. Siderophores: structure and function of microbial iron transport compounds. *J Biol Chem.* 1995;270(45):26723–26726.
60. Vala AK, Anand N, Bhatt PN, Joshi HV. Tolerance and accumulation of hexavalent chromium by two seaweed associated *Aspergilli*. *Mar Pollut Bull.* 2004;48(9–10):983–985.
61. Taboski MA, Rand TG, Piorko A. Lead and cadmium uptake in the marine fungi *Corollospora lacera* and *Monodictys pelagica*. *FEMS Microbiol Ecol.* 2005;53(3):445–453.
62. Sanyal A, Rautaray D, Bansal V, Ahmad A, Sastry M. Heavy metal remediation by a fungus as a mean of lead and cadmium carbonate crystals. *Langmuir.* 2005;2(21):7220–7224.
63. Iram S, Ahmad I, Javed B, et al. Fungal tolerance to heavy metals. *Pak J Bot.* 2009;41(5):2583–2594.
64. Li T, Liu MJ, Zhang XT, Zhang HB, Sha T, Zhao ZW. Improved tolerance of maize (*Zea mays* L.) to heavy metals by colonization of a dark septate endophyte (DSE) *Exophiala pisciphila*. *Sci Total Environ.* 2011;409:1069–1074.
65. Srinath T, Verma T, Ramteke P, Garg K. Chromium biosorption and bioaccumulation by chromate resistant bacteria. *Chemosphere.* 2002;48:427–435.
66. DalCorso G, Farinati S, Maistri S, Furini A. How plants cope with cadmium: staking all on metabolism and gene expression. *J Integr Plant Biol.* 2008;50:1268–1280.
67. Bhattacharya SK, Gupta S, Debnath UC, Ghosh D, Chattopadhyay, Mukhopadhyay A. Arsenic bioaccumulation in rice and edible plants and subsequent transmission through food chain in Bengal basin: a review of the perspectives for environmental health. *Toxicol Environ Chem.* 2012;94:429–441.
68. Tamás MJ, Sharma SK, Ibstedt S, Jacobson T, Christen P. Heavy metals and metalloids as a cause for protein misfolding and aggregation. *Biomolecules.* 2014;4:252–267.