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Enhanced Bio-hydrogen Production from Protein Wastewater by Altering Protein Structure and Amino Acids Acidification Type

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Enhanced bio-hydrogen production from protein wastewater by altering protein structure and amino acids acidification type via pH control was investigated. The hydrogen production reached 205.2 mL/g-protein when protein wastewater was pretreated at pH 12 and then fermented at pH 10. The mechanism studies showed that pH 12 pretreatment significantly enhanced protein bio-hydrolysis during the subsequent fermentation stage as it caused the unfolding of protein, damaged the protein hydrogen bonding networks, and destroyed the disulfide bridges, which increased the susceptibility of protein to protease. Moreover, pH 10 fermentation produced more acetic but less propionic acid during the anaerobic fermentation of amino acids, which was consistent with the theory of fermentation type affecting hydrogen production. Further analyses of the critical enzymes, genes, and microorganisms indicated that the activity and abundance of hydrogen producing bacteria in the pH 10 fermentation reactor were greater than those in the control.

ydrogen has been deemed the future energy carrier due to its high energy content and the lack of a
greenhouse effect. The use of biotechnology to produce hydrogen is becoming an important strategy
and attracting the attent greenhouse effect. The use of biotechnology to produce hydrogen is becoming an important strategy hydrogen production methods^{1,2}. Bio-hydrogen production by dark fermentation is a process in which anaerobic bacteria use organic wastes to produce hydrogen in the absence of a light source³. This process has several benefits with respect to the environment such as energy recover and resource reuse⁴. Additionally, it has a high hydrogen production rate with a simple reactor configuration⁵.

Dark fermentative hydrogen can be produced from many organic wastes, such as wastewater. Carbohydrate in wastewater often accounts for the majority of the organic load to produce bio-hydrogen in the literature⁶. Nevertheless, protein is also the main organic substrate in some wastewaters. For example, the wastewaters generated during the productions of abattoir, whey, cheese, casein, fish, and silk usually contain significant amount of protein^{7,8}. As seen in Figure 1 during anaerobic fermentation protein is firstly hydrolyzed to peptides and amino acids by extracellular enzyme (step-I), and then fermented to volatile fatty acids and hydrogen (step-II). Unlike carbohydrates, protein has a unique three-dimensional structure. Usually, the step-I is very slow as protein is unsusceptible to protease cleavage in its native folded conformation^{9,10}. Thus, hydrolysis is the ratelimiting step during hydrogen production from protein wastewater.

Although it has been reported that bio-hydrogen can be produced from protein wastewater via anaerobic dark fermentation at neutral pH, the reported maximal hydrogen yield is less than 20 mL/g-protein at a protein concentration of 2 $g/L^{2.11,12}$. Moreover, in all these studies the hydrogen-producing substrate was peptone, which was formed during partial hydrolysis of protein and was distinguished from protein on the basis of size and threedimensional structure. The hydrolysis rate of protein could be accelerated by the change of its folded structure, which however, has never been documented in previous studies. In addition, previous studies found that the fermentation pH is an important factor affecting bio-hydrogen production, and the optimal pH for bio-hydrogen production from protein-rich waste (such as sewage sludge) is alkaline pH (about 8.5–10) owing to the improved solubilization of sludge organic matters and the inhibitory effect on hydrogen consuming microorganisms^{2,13}, but the impacts of pH on protein structure and protein hydrolyzate (amino acids) conversion involved in step-II during hydrogen production from protein wastewater have seldom been investigated.

Figure 1 [|] Proposed metabolic pathway for anaerobic bio-hydrogen production from protein.

In this study a new method (i.e., pretreating protein wastewater at pH 12 and then fermenting it at pH 10) for enhancing bio-hydrogen production from protein wastewater by altering protein structure and amino acids acidification type was reported. Firstly, the effects of different pretreatment and fermentation pH values on the production of bio-hydrogen from protein wastewater were studied. Then, the reasons for pH 12 pretreatment and pH 10 fermentation showing significantly higher hydrogen production were explored by advanced analytical techniques from the aspects of protein unfolding, protein hydrogen bonding networks damage, disulfide bridge breakage, the reactions of amino acids fermentation type, critical enzyme activity and genes, and key microbes.

Results

Effects of pretreatment and fermentation pH on bio-hydrogen production. In this study the influence of pretreatment pH on fermentative hydrogen production was investigated firstly. It was observed that the hydrogen production from synthetic protein wastewater was increased with fermentation time between 12 and 72 h (Figure 2A), which, however, was not significantly enhanced when the time was further prolonged to 84 h ($p > 0.05$). The comparison of pretreatment pH values affecting hydrogen production was therefore made at fermentation time of 72 h. It can be seen from Figure 2A that pH 12 or pH 12.5 pretreatments resulted in a much higher hydrogen production than other pHs and the control (pH 7) at a fermentation time of 72 h. Nevertheless, there was no significant difference in the hydrogen production between pH 12 and pH 12.5 pretreatments $(p > 0.05)$. Thus, the optimal pretreatment pH was 12 under the condition of fermentation pH 7.

The hydrogen production was also observed to be significantly affected by fermentation pH. With the use of pH 12 pretreated wastewater as the substrate the time courses of hydrogen production being affected by fermentation pH are shown in Figure 2B. There was no detectable hydrogen produced at fermentation pH 4 and pH 11. Although bio-hydrogen was produced between pH 5 and pH 10, the maximal hydrogen production and suitable fermentation time were different. In the pH range of 6–10 the optimal fermentation time was 72 h, whereas it was 96 h at pH 5 (statistical analysis is shown in Table S1). The hydrogen production at optimal fermentation time was respectively 155.6 (pH 5), 155.7 (pH 6), 117.1 (pH 7), 96.2 (pH 8), 162.8 (pH 9), and 205.2 (pH 10) mL/g-protein. Obviously, the maximal hydrogen production from pH 12 pretreated synthetic protein wastewater occurred at fermentation pH 10.

Effects of pH 12 pretreatment on the characters of protein wastewater. During pH 12 pretreatment there was no significant change in the concentration of protein ($p > 0.05$), and no hydrogen was generated. However, pH 12 pretreatment significantly changed the characters of protein wastewater. In this study the influence of pH 12 pretreatment on the α -helix content of protein molecule was firstly investigated by circular dichroism. Compared with the control, as seen in Figure 3A, pH 12 pretreatment induced a significant decrease in the α -helix content of protein from 69.1% to 42.4%. The effect of pH 12 pretreatment on the hydrodynamic diameter of protein was further investigated with BSA as a model protein. The data in Figure 3B showed that the average hydrodynamic diameter of BSA was 8.0 nm in the control (pH 7 pretreatment), while it was increased to 12.3 nm after pH 12 pretreatment.

Protein has endogenous fluorescence since it contains aromatic amino residues, and the numbers of tryptophan, tyrosine and phenylalanine residues in the BSA molecule are 2, 21 and 26, respectively¹⁴. The effect of pH 12 pretreatment on the changes of protein structure was further assayed by EEM fluorescence spectroscopy, which had been applied as a useful and powerful methodology to extract information about the structural features of protein¹⁵. As seen

Figure 2 | The time courses of effects of pretreatment pH (A) and fermentation pH (B) on the accumulated hydrogen production. The fermentation pH was 7 (A) and pretreatment pH was 12 (B). Error bars represent standard deviations of triplicate tests.

in Figure 4 the fluorescence maximum peak intensity of BSA in the control test was 5.18 \times 10⁶ a.u., which was sharply dropped to 1.79 \times 106 a.u. after pH 12 pretreatment. Obviously, pH 12 pretreatment caused considerable decrease in the fluorescence intensity of protein. At the same time the maximum excitation and emission wavelengths of protein were shifted after pH 12 pretreatment, i.e., the excitation was 272 nm and emission was 350 nm in the control, while the corresponding data were 278 nm and 340 nm for the pH 12 pretreated protein. It implied that the main reason for pH 12 pretreatment resulting in the decrease of fluorescence intensity was due to the quenching of aromatic residues fluorescence since there was a significant shift in the fluorescence maximum¹⁶.

Protein structure is influenced and stabilized to a great extent by disulfide bridges. Raman spectroscopy has been employed extensively to probe the conformation of disulfide bridges in protein because the frequency of S-S stretching mode of the disulfide bridge is sensitive to its conformation¹⁷. Raman spectra of the control and pH 12 pretreated BSA in the 200-1800 cm^{-1} region after being normalized against the phenylalanine band at 1004 cm^{-1} are shown in Figure S1 (Supplementary information). The region of 500– 550 cm^{-1} in the Raman spectrum belongs to the stretching vibration mode of S-S bridges, and the peaks appearing around 500–510, 515– 525, and 535–545 $\rm cm^{-1}$ are the features of the gauche-gauche-gauche (ggg), gauche-gauche-trans (ggt), and trans-gauche-trans (tgt) conformation, respectively^{17,18}. According to the curve fitting of the band, the conformations of 17 disulfide bridges in BSA can be determined. The analyses of S-S Raman bands of BSA before and after pH 12 pretreatment are illustrated in Figure 5. The main conformation of BSA in the control was ggg (506 cm⁻¹, Figure 5A). pH 12 pretreatment decreased the relative intensity of S-S Raman bands and meanwhile induced a prominent increase of ggt conformation of disulfide bridges (Figures 5B). Figure 5C shows the composition of the S-S bridge conformations of BSA after pH 12 pretreatment. There were respectively 11, 3 and 3 out of 17 disulfide bridges of BSA with the ggg, ggt and tgt conformations in the control, which were 3, 6, and 2 after pH 12 pretreatment.

Figure 3 | Effects of pH 12 pretreatment on the CD spectra and hydrodynamic diameter of protein.

Figure 4 | Effects of pH 12 pretreatment on the fluorescence spectra of protein wastewater.

Effects of pH 12 pretreatment on the susceptibility of protein to protease and protein hydrolysis efficiency. To investigate the effects of pH 12 pretreatment on the susceptibility of protein to protease, the protein enzyme hydrolysis experiments were conducted, and the protein enzymolysis efficiency was used to express the susceptibility of protein to protease. As seen in Figure 6A the protein enzymolysis efficiency of synthetic protein wastewater was significantly affected by pH 12 pretreatment. After enzymolysis for 1 h, the protein enzymolysis efficiency with the pH 7 pretreated wastewater was only 39.9%, whereas it was significantly increased to 71.9% with the pH 12 pretreated one.

As seen in Figure 6B the hydrolysis of protein during anaerobic hydrogen production was significantly affected by pH 12 pretreatment. The hydrolysis efficiency in the control (pH 7 pretreatment) was only 28.8%, whereas it bacame 86.7% after the wastewater was pretreated at pH 12. Further investigation showed that during fermentation for hydrogen production the activity of protease was also remarkably improved after the protein wastewater was pretreated by pH 12 (Figure 6B).

Effects of pH 10 fermentation on amino acids conversion. From the proposed metabolic pathway illustrated in Figure 1 it can be seen that the hydrogen production from protein wastewater comes from the conversion of amino acids to volatile fatty acids. Thus, the synthetic amino acids wastewater was used to further investigate the influence of pH 10 fermentation on bio-hydrogen production. As shown in

Figure S2A (Supplementary information) the hydrogen generation in the pH 7 (control) fermentation test of synthetic amino acids wastewater was much lower than that in the pH 10 one (63.1 versus 139.9 mL/g-amino acids). The data in Figure S2A indicated that the concentration of VFA at fermentation pH 7 test was slightly higher than that at fermentation pH 10 (2166.8 versus 1836.4 mg COD/L). However, the concentration and percentage of acetic acid in pH 10 reactor were significantly greater than those in pH 7 reactor (940.0 against 572.0 mg COD/L, and 51.2% versus 26.4%), and the opposite observations were made with propionic acid (207.0 against 632.8 mg COD/L, and 11.3% versus 29.2%, see Figure S2B).

Effects of pH 12 pretreatment plus pH 10 fermentation on the activities of key enzymes and microorganisms during hydrogen production. According to the supposed metabolic pathway for biohydrogen production from protein (Figure 1), the NADH-Fd reductase and hydrogenase are the key metalloenzyme responsible for hydrogen formation. To recycle NAD^+ , $NADH$ is oxidized by NADH-Fd reductase, while hydrogenase oxidizes reduced ferredoxin to produce molecular hydrogen^{19,20}. Oxaloacetate transcarboxylase (OAATC) is a key enzyme involved in propionic acid synthesis and NADH consumption²¹. As illustrated in Figure 7A the relative activity of NADH-Fd reductase and hydrogenase was improved by 34% and 71% when the wastewater fermentation pH changed from pH 7 to pH 10. The OAATC, however, was significantly restrained to 58.5% of the control at fermentation pH 10. Then

Figure 5 | Curve fitting of the S-S Raman band of pH 7 pretreated (A) and pH 12 pretreated (B) BSA, and the composition of S-S bridge conformations (C) of BSA by the analysis of the 480-560 cm⁻¹ Raman regions. Black line: experimental spectrum; red line: the fitting curve; blue line: peaks.

Figure 6 | Effects of pH 12 pretreatment on the susceptibility of protein to protease (A), and the protein hydrolysis efficiency during protein wastewater anaerobic fermentation for hydrogen production at pH 7 for 72 h (B). The concentration of pronase E in the synthetic protein wastewater was 50 mg/L and the enzymolysis time was 1 h. Error bars represent standard deviations of triplicate tests.

the real-time polymerase chain reaction (PCR) analysis (see Supplementary information of "Real-time PCR" and Table S2) was conducted to explore the intrinsic reason. As seen in Figure 7B the expression of Fd (ferredoxin) gene in the pH 10 fermentation reactor was about 1.44- fold of that in the control reactor. The same observation was made with the expression of hydA (Fe-hydrogenase) gene, which was 1.85-fold of the control when the anaerobic fermentation pH was 10. The fluorescence in situ hybridization (FISH) assay was conducted to investigate the influences of pH 10 on the abundance of hydrogen producing bacteria (see Supplementary information of ''Fluorescence in situ hybridization''). The images are shown in Figure 8. Further images analysis indicated that the abundance of hydrogen producing bacteria in pH 7 and pH 10 fermentation reactors was 24.5% and 32.1%, respectively.

Hydrogen production from real protein wastewater enhanced by pH 12 pretreatment and pH 10 fermentation. The examination of applying pH 12 pretreatment plus pH 10 fermentation to enhance hydrogen production from whey protein wastewater, a real high protein content organic wastewater, was conducted. As seen in Figure S3A (Supplementary information) no matter whether the hydrogen production from whey protein wastewater pretreated or not was increased with fermentation time, the accumulative hydrogen production of control (pH 7 pretreatment and pH 7 fermentation) was only 44.8 mL/g-COD at fermentation time of 72 h, while it was significantly improved to 106.9 mL/g-COD after the whey protein wastewater was pretreated at pH 12 and fermented at pH 10. The same observations were made with the protein utilization efficiency and VFA production, which was increased

Figure 7 | Comparison between pH 7 and pH 10 on the relative activities of OAATC, NADH-Fd reductase, and hydrogenase (A), and the expression of ferredoxin (Fd) and Fe-hydrogenase (hydA) gene during wastewater fermentation for hydrogen production (B). Error bars represent standard deviations of triplicate tests.

Figure 8 [|] Microscopy of sludge from the hydrogen production reactor of fermentation pH 7 and fermentation pH 10 at the time of 72 h by FISH (310). Hydrogen producing bacteria were labeled with HEX (yellow, A1 for pH 7 and A2 for pH 10), and domain bacteria were labeled with FITC (green, B1 for pH 7 and B2 for pH 10).

respectively by 154% and 67% compared with the control (Figure S3B, Supplementary information).

Discussion

In the literature protein wastewater is directly fermented at approximately pH 7 to produce hydrogen without any pretreatment, and the hydrogen yield is less than 20 mL/g-protein at a protein concentration of around 2.0 $g/L^{2,11}$. The data in Figure 2A indicated that if the protein wastewater was unpretreated and directly fermented at pH 7 (control) the bio-hydrogen production was 39.0 mL/g-protein, which was higher than the results of some previous studies because the the activity of homoacetogenesis was inhibited by $CHCl₃$ in the current study. However, the hydrogen production was significantly increaed to 205.2 mL/g-protein when the protein wastewater was pretreated at pH 12 and then fermented at pH 10, which was more than 5-fold of the control. The reasons for pH 12 pretreatment plus pH 10 fermentation showing significantly higher bio-hydrogen production were explored and discussed next.

It is well known that under anaerobic fermentation conditions protein is firstly hydrolyzed to peptides and amino acids, and then fermented to volatile fatty acids and hydrogen (Figure 1). As the generation of hydrogen is relevant to the metabolism of protein, the influence of pH 12 pretreatment on protein hydrolysis is investigated firstly. In this study it was observed that during pH 12 pretreatment there was no significant change in the concentration of protein, which suggested that pH 12 pretreatment did not directly cause the hydrolysis of protein.

The bio-hydrolysis behavior of protein has been reported to be affected by the changes of its secondary structure^{10,22}. The α -helix and b-sheet are the two main shapes in the secondary structure of protein, and the native BSA is dominated by α -helix without β -sheet²³. The decrease in α -helix content of protein indicates the breakage of hydrogen bonding networks, which has been reported to be positively consistent with the degradation of protein^{24,25}. In this study pH 12 pretreatment induced a significant decrease in the α -helix content of protein from 69.1% to 42.4% (Figure 3A), which was consistent with the results of protein hydrolysis efficiency during anaerobic fermentation (Figure 6B).

The unfolding of protein is often a prerequisite for its bio-hydrolysis⁹. Dynamic light scattering can probe the unfolding of protein by giving the size properties of protein in a three-dimensional structure26. The unfolding of protein has been reported to play a critically important role in protein bio-hydrolysis by ATP-dependent protease as it influences the susceptibility of protein to proteases²⁷. The data in Figure 3B showed that pH 12 pretreatment unfolded BSA from 8.0 nm to 12.3 nm. The unfolding of protein would change the microenvironment of tryptophan, which could decrease protein fluorescence intensity and shift the fluorescence maximum^{15,28}. This observation consistented with the changes of protein fluorescence after pH 12 pretreatment (Figure 4). Then the unfolding of protein enhanced susceptibility of protein to proteases (Figure 6A) and the protease activity was therefore increased by almost 28% compared with the control (Figure 6B). Further investigation indicated that pretreating wastewater at pH 12 not only destroyed the disulfide bridges of protein but changed their conformations (Figure 5). Thus the hydrolysis of protein during anaerobic fermentation was significantly increaced from 28.8% to 86.7% after the wastewater was pretreated at pH 12. Since more protein was bio-hydrolyzed, more amino acids were provided for hydrogen producers, and the hydrogen production was increased.

As stated above, there was no hydrogen produced during pH 12 pretreatment, and pH 12 pretreatment did not cause protein hydrolysis directly. However, it significantly altered protein structure and the alteration remained stable when the pH was adjusted to pH 10 for

Under anaerobic conditions the evolution of hydrogen through a NADH pathway is driven by the necessity of reoxidizing the residual NADH of metabolic reactions as NADH + H^+ \rightarrow NAD⁺ + H_2^{29} . As shown in Figure 1, there is $1, -2$, and 0 molecules residual NADH produced when one molecule pyruvate is respectively converted to acetic, propionic, and butyric acid, which indicates that higher hydrogen production corresponds to greater acetic acid but less propionic acid generation. It should be noted that the generation of butyric acid has no connection with the production of hydrogen but only contributes to reoxidation of NADH. There was no detectable ethanol and lactic acid produced in the current study. As seen in Figure S2B, the percentage of acetic acid in the pH 10 fermentation reactor was significantly greater than that in the pH 7 reactor (51.2% versus 26.4%), but the opposite observations was made with propionic acid (11.3% versus 29.2%). Obviously, the result was in line with the theory of fermentation type affecting hydrogen production.

The NADH-Fd reductase and hydrogenase are the key enzymes to oxidize NADH and reduced ferredoxin to produce molecular hydrogen, while OAATC is a key enzyme for propionic acid synthesis^{19,20,21}. As illustrated in Figure 7A pH 10 fermentaion improved the relative activitives of NADH-Fd reductase and hydrogenase, but it restrained the relative activitiy of OAATC compared with pH 7 fermentaion. The real-time PCR further indicated that the expressions of Fd (ferredoxin) and hydA (Fe-hydrogenase) genes both were improved in the pH 10 fermentation reactor (Figure 7B). The same observation was made with the abundance of hydrogen produing bacteria, which was increased from 24.5% (in the control reactor) to 32.1% (in the pH 10 fermentation reactor) according to the FISH results (Figure 8). It can be easily understood therefore that amino acids fermentation at pH 10 had higher hydrogen production than that at pH 7.

In summary, we presented a new strategy to significantly increase bio-hydrogen production from protein wastewater by altering protein structure and amino acids acidification type via pH control. The hydrogen production reached 205.2 mL/g-protein, which was much higher than that reported in the literature². pH 12 pretreatment induced the unfolding of protein, damaged the protein hydrogen bonding networks, destroyed the disulfide bridges and changed their conformations, which increased the susceptibility of protein to protease and enhanced the hydrolysis of protein⁹. Moreover, when the fermentation pH varied from pH 7 to pH 10 the anaerobic metabolic reaction of amino acids shifted from propionic acid to acetic acid, and the critical enzyme activities and abundance of hydrogen producing bacteria were improved, which resulted in higher hydrogen yield in the fermentative reactor of pH 10. Finally, the application of 12 pretreatment and pH 10 fermentation to enhance bio-hydrogen evolution from real protein wastewater was testified to be feasible.

Methods

Experiments of pretreatment pH affecting hydrogen production from protein wastewater. The synthetic protein wastewater used in this study for bio-hydrogen production was prepared with bovine serum albumin (BSA, diagnostic grade power of BAH 68, obtained from Equitech-Bio) as carbon source, and its concentration was maintained at 2.0 g/L. The pretreatment was conducted in 12 identical reactors, each with working volume of 600 mL, outside diameter of 85 mm, height of 180 mm, and cap of 25 mm. For each reactor 300 mL protein wastewater was added and the pH value was adjusted to 2, 3, 4, 5, 6, 7 (control), 8, 9, 10, 11, 12, 12.5, respectively, by adding 6 M hydrochloric acid (HCl) or 6 M sodium hydroxide (NaOH) and maintained in an air-bath shaker (100 rpm) with temperature of 21 \pm 1°C for 24 h. There was no microorganism during the pretreatment of protein wastewater. After pretreatment, 1.0 g NaHCO₃ (buffer salt), 35 mg K₂HPO₄·3H₂O (phosphorus source), 1 mL trace element solution and 30 mL seed microorganism (see Supplementary Information for details) were added to each reactor. The mixture was then anaerobically fermented to produce hydrogen after its pH value was adjusted to 7 by adding 6 M HCl or 6 M NaOH. According to the literature 60 ul CHCl₃ (0.02%, $\mathrm{V}/\mathrm{V})$ was added to the mixture to inhibit homoacetogenesis 30 . After being flushed with nitrogen gas for 2 min to remove oxygen, all reactors were capped with rubber stoppers, sealed and placed in an air-bath shaker (140 rpm) with the temperature of 36 ± 1 °C. The total gas volume was measured by releasing the pressure in the reactor using a glass syringe (100 mL) to equilibrate with the room pressure according to the previous publication¹³. Gas was sampled with a 5 mL syringe for composition assay. As the pH value of protein wastewater was around 7, the pH 7 pretreated test could be regarded as the un-pretreated one.

Experiments of fermentation pH affecting hydrogen production from pH 12 pretreated protein wastewater. The bio-hydrogen production experiments were conducted in 8 identical reactors with the pH 12 pretreated protein wastewater as substrate. The fermentation pH in these 8 reactors was respectively 4, 5, 6, 7, 8, 9, 10 and 11 and was kept constant by adding 6 M HCl or 6 M NaOH during the entire fermentation process. All other operations were the same as described above.

Experiments of pH 10 affecting hydrogen generation from amino acids fermentation. In this study the comparison of fermentation pH 7 and pH 10 affecting amino acids fermentation was conducted with synthetic amino acids wastewater. The synthetic amino acids wastewater consisted of 0.25 g/L L-leucine, L-lysine, Lglutamic acid, L-alanine, L-aspartic acid, L-valine, L-cysteine and L-threonine, respectively. Batch experiments of bio-hydrogen production from synthetic wastewater were conducted in 2 identical reactors, where fermentation pH was adjusted to 7 and 10, respectively, by adding 6 M HCl or 6 M NaOH. All other operations were the same as described above.

Experiments of hydrogen production from real protein wastewater. The real wastewater containing high concentration of whey protein was obtained from a milk powder processing plant in Hangzhou, Zhejiang Province, China. The main characteristics of whey protein wastewater in experiments are as follows: pH 6.56 \pm 0.04, total chemical oxygen demand (TCOD) 3436 \pm 80 mg/L, protein 2709 \pm 72 mg COD/L, and carbohydrate 554 \pm 22 mg COD/L. Comparison of bio-hydrogen production from pH 7 and pH 12 pretreated whey protein wastewater was conducted in 2 identical reactors, where fermentation pH was controlled at 7 and 10, respectively, by adding 6 M HCl or 6 M NaOH. All other operations were the same as described above.

Experiments of protein enzyme hydrolysis. To investigate the effects of pH 12 pretreatment on susceptibility of protein to protease, the protein enzyme hydrolysis experiments were conducted and the protein enzymolysis efficiency was used to express the susceptibility of protein to protease. The experiments were conducted in 2 identical reactors with the pH 7 pretreated protein wastewater (control) and pH 12 pretreated protein wastewater as substrate, respectively. The concentration of pronase E (model protease, purchased from Merck) in each reactor was 50 mg/L and the hydrolysis pH was adjusted to 7 by adding 6 M HCl or 6 M NaOH during the entire hydrolysis process. After being flushed with nitrogen gas for 2 min to remove oxygen, both reactors were capped with rubber stoppers, sealed and placed in an airbath shaker (140 rpm) with the temperature of 36 \pm 1°C. At set intervals 5 mL sample was took out and immediately being heat-treated at 100°C for 10 min to inhibit enzyme activity. After cooling down to room temperature the residual protein in sample was precipitated by adding 5 mL 15% trichloroacetic acid (TCA) and the mixture was centrifuged at 4500 rpm for 15 min. The total organic nitrogen concentration in supernatant was determined using a TOC analyzer (TOC-V CPH, Shimadzu, Tokyo, Japan). In this study the enzymolysis efficiency of protein was expressed by the following equation.

Protein enzymes ificiency (
$$
\degree
$$
) = $\frac{C_{initial} - C_{remaining}}{C_{initial}} \times 100$
= $\frac{TN}{C_{initial} \cdot 100} \times 100$ (1)

where C_{initial} and C_{remaining} is respectively the concentration of protein in wastewater before and after experiment, TN is the concentration of nitrogen in supernatant, and 16% is the average value of nitrogen content in protein³¹.

Circular dichroism spectrum measurement of protein. The circular dichroism (CD) spectrum of protein was measured on a JASCO J-715 automatic recording spectropolarimeter (Tokyo, Japan) controlled by the JASCO software with a 0.1 cm quartz cell at room temperature. The sample was diluted to a concentration of 45 mg BSA/L and transferred to a quartz cell with a 1 cm light-path length. Molecular ellipticity was measured in the range of 190–250 nm with a bandwidth of 1 nm and the scan rate was 50 nm/min. The distilled water was selected as the blank and was automatically subtracted from each spectrum during scanning. The a-helix content of BSA was calculated according to equations (2) and $(3)^{32}$.

$$
MRE = \frac{ObservedCD \ (m \text{ deg})}{C_p \bullet n \bullet L} \tag{2}
$$

where MRE is the mean residue ellipticity (deg.cm².dmol⁻¹), C_P is the molar concentration of protein, n is the number of amino acid residues (583 for BSA), and L is the path length of the cell (mm).

$$
a - Helix \; (%) = \frac{-MR_{208} - 4000}{33000 - 4000} \times 100 \tag{3}
$$

where $MRE₂₀₈$ is the observed MRE at 208 nm, 4000 is the MRE of the β -form and random coil conformation cross at 208 nm, and 33000 is the MRE value of a pure ahelix at 208 nm.

Fluorescence spectroscopy analysis of protein. Excitation emission matrix (EEM) fluorescence spectroscopy of protein was obtained with a luminescence spectrometry (Fluoromax-4 Spectrofluorometer, HORIBA Scientific, France). Prior to the EEM test, the sample was filtered through a $0.45 \mu m$ hydrophilic filtration membrane and diluted to an approximate concentration. To get the fluorescence of EEM, the excitation and emission wavelengths were gradually increased from 200 to 400 nm at 2 nm increments and 250 to 450 nm at a 5 nm step, respectively.

Raman spectrum analysis of protein. Before Raman measurements, the pH 7 (control) or pH 12 pretreated synthetic protein wasterwaters was frozen at $-20^{\circ}\mathrm{C}$ and then lyophilized. Raman spectrum was recorded on a Dilor-Jobin Yvon-Spex Raman spectrometer equipped with an Olympus BX 40 system microscope with $100\times$ objectives. A He–Ne laser system with excitation line at 632.8 nm and an excitation power of 6 mW was used in the measurements. The recording range was 200–1800 nm with a resolution of 1.2 cm^{-1} and the spectrum was the accumulated averages of 3 exposures at room temperature (25 $^{\circ}$ C). The average spectrum was smoothed by a $\frac{1}{5}$ cm⁻¹ filter and was corrected by a polynomial baseline using LabSpec (Dilor-Jobin Yvon-Spex, France) software. To calculate the corresponding content of the protein conformation, spectra shown in the main text were normalized against the phenylalanine band at 1004 cm^{-1} as its intensity (set as 1000) was unaffected by conformational changes in protein according to the leaterature¹⁸. The curve fitting of Raman spectrum was made by Origin 8.5 (Origin, USA) using multiple peak fit (Gaussian model).

Other analytical methods. The gas component was measured using a gastight syringe (0.2 mL injection volume) and a gas chromatograph (GC112A, China) equipped with a thermal conductivity detector (TCD) and a 4 mm \times 2 m GDX-102 stainless column¹³. The temperatures of the injection port, column, and detector were set at 40, 40 and 80 $^{\circ}$ C, respectively. Nitrogen was used as the carrier gas at a flow rate of 30 mL/ min. The pH value was measured by a pH meter. The determinations of protein, SS, VSS, and volatile fatty acids (VFA) were the same as those described previously³³. The total VFA was calculated as the sum of measured acetic, propionic, n-butyric, isobutyric, n-valeric, and iso-valeric acids. The detailed analytical procedures of fluorescence in situ hybridization (FISH), real-time polymerase chain reaction (PCR), protease, oxaloacetate transcarboxylase (OAATC), NADH-Fd reductase, and hydrogenase activities are presented in Supplementary information.

Statistical analysis. All tests were conducted in triplicate and the results were expressed as mean \pm standard deviation. An analysis of variance (ANOVA) was used to test the significance of results and $p < 0.05$ was considered to be statistically significant.

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Author contributions

N.D.X. carried out the experiments and wrote the paper, Y.G.C. conceived the study design and co-wrote the paper, A.H.C. analyzed the data, L.Y.F. characterized the protein structure. All authors contributed to the scientific planning and discussions.

Additional information

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