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RESEARCH ARTICLE

Conservation tillage increases carbon sequestration of winter wheat-summer maize farmland on Loess Plateau in China

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Abstract

The idea of mitigating anthropogenic CO_2 emissions by increasing soil organic carbon (SOC) is notable. However, the estimation of the net ecosystem carbon balance after conversion from conventional tillage to conservational tillage has been poorly quantified for the Loess Plateau in China. A 2-year field experiment was conducted to estimate the agroecosystem carbon balance of a winter wheat–summer maize rotation system using a full carbon cycle analysis. The results showed that a positive net ecosystem carbon balance value in the cases of rotary tillage with straw incorporation, chisel plow tillage with straw incorporation, and no tillage with straw mulching treatments. Note that a negative value was detected for the conventional tillage to conservational tillage substantially enhanced the carbon sink potential from 0.84 t C ha⁻¹ yr⁻¹ to 2.69 t C ha⁻¹ yr⁻¹ in both years. Our findings suggest that the expansion of conservational tillage could enhance the potential carbon sink of the rain-fed land in China.

Introduction

Agriculture accounts for approximately 10.0%-12.0% of the total global anthropogenic emissions of greenhouse gases (GHGs) [1]. The direct emission of CO₂ included soil respiration or indirect emission of CO₂ induced by the production of agriculture inputs (fertilizers and pesticides), fuel combustion, and application of machinery on the farm that is increasing year on year [2]. The winter wheat–summer maize rotation system under a rain-fed condition is one of the major grain productions in North China [3]. Therefore, it is important to study carbon balance in rain-fed fields to select appropriate tillage methods to develop low-carbon agriculture and promote the development of sustainable agriculture.

Previous studies have been conducted to evaluate the carbon source or sink by using several methods such as net carbon flux [4–5], net ecosystem productivity [6], and carbon sustainability [7]. Moreover, previous studies on the carbon balance were primarily focused on forest, grassland, and wetland ecosystems [8–11]. The carbon balance of an agricultural ecosystem is manuscript. The specific roles of these authors are articulated in the 'author contributions' section.

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Competing interests: Xingneng Lu received support in the form of a salary from Yinchuan Provincial Sub-branch, The People's Bank of China. This does not alter our adherence to all the PLOS ONE policies on sharing data and materials. primarily observed in a rice paddy field [12-14]. Conservation tillage treatments (e.g., reduced tillage, no-tillage, and straw returning) are often suggested to improve the potential negative effects of crop residue removal, which may refer to the reduction of soil organic carbon (SOC), the increase in soil compaction, disruption of soil aggregates, and deterioration of soil health [15–17]. Conservation tillage operations have often been reported to enhance soil organic carbon sequestration whereas simultaneously mitigate the carbon (C) emissions associated with agricultural inputs such as fertilizers and on-farm fuels [5]. However, there is considerably uncertainty in the estimation of the carbon sink/source of an agricultural system. For example, Snyder *et al.* [18] has showed that the agricultural fields not only are a carbon sink but also a carbon source because of the application of tillage and fertilizer treatments. Tillage and fertilizer methods always support food, energy, and air for the development of soil organisms, thus increasing the decomposition rate of residues and soil respiration and ultimately resulting in that the stable soil organic carbon is awkward [19]. Li et al. [20] measured the carbon balance of winter wheat $(-1.98 \text{ t hm}^{-2})$ and summer maize $(-1.38 \text{ t hm}^{-2})$ in the North China Plain, suggesting that without considering the harvest grain carbon part, this ecosystem is a carbon source; however, this ecosystem is a carbon sink when considering the harvest grain carbon part. However, Zhao et al. [21] found that the carbon uptake of a farmland in China's coastal areas is significantly higher than that of C emissions. In addition, Chen et al. [22] showed that rotary tillage with straw incorporation and no tillage with straw mulching display a C sink, while moldboard plow tillage with or without straw shows a carbon source in paddy soil. These differences may be attributed to the difference in the level of regional economic development, production layout, and agricultural management practices such as tillage, fertilizer use, and herbicide use.

Soil respiration plays a key role in determining the carbon balance [23]. The lack of available and comprehensive carbon balance data revealed an urgent need to increase the research on the effect of conservation tillage on the net ecosystem carbon balance for the Loess Plateau in China. Thus, the goals of our study are as follows: (1) to estimate the effects of different tillage treatments on soil respiration and its components and (2) to evaluate the effects of different tillage on the net ecosystem carbon balance in the winter wheat–summer maize rotation system.

Materials and methods

Study site

The study was conducted at the Dry-land Experimental Station of Northwest A&F University, Yangling Town, Shaanxi province, in the northwestern part of China ($34^{\circ}21$ 'N and $108^{\circ}10$ 'E). The soil is classified as silt loam (19% sand, 77% silt, and 4% clay) based on the USDA Texture Classification System. The surface soil (0-20 cm) bulk density before the start of the experiment (in 2009 year) was 1.30 g cm⁻³. The study area belong to a semiarid climate, and the annual average temperature, the mean annual precipitation, and the annual potential evaporation are 13° C, 622 mm, and 993 mm, respectively. The weather conditions including mean daily air temperature and daily precipitation are presented in Fig.1.

Experimental design

The experiment had a randomized block design with three replications. Four tillage systems including conventional moldboard plowing tillage without crop straw (CT), rotary tillage with straw incorporation (RTS), chisel plow tillage with straw incorporation (STS), and no tillage with straw mulching (NTS) were included. Thus, 12 plots were designed; the plot size was 48 m² (3.2 m × 15 m). Moreover, three root-free plots with the same size (these were placed at



Fig 1. The average of daily air temperature (a) and daily precipitation (b) during the experimental period (from Oct 2013 to Oct 2015).

approximately 10-m intervals adjacent to the whole-soil plots) were designed. The root-free plots were established to evaluate microbial respiration (R_h). The root-free plots were kept free of vegetation by cutting the plants manually throughout the sampling period.

Four tillage treatments, namely, STS, NTS, RTS, and CT, were arranged in three repetitions. The tillage treatments were the same every year since 2009. After harvest, crop straw was removed by hand from the field for the CT treatment before the tillage application. While for the STS, NTS, and RTS treatments, the crop straw was left in the field. The soil tillage was operated twice each year; one was operated on June 17 before planting the summer maize, and the other was operated on October 17 before planting the winter wheat. In the CT plot, the soil was plowed up to a depth of 20–25 cm, and a rotavator was then used to plow the soil up to a depth of 15 cm (Fig 2). In the RTS plot, a rotavator (15 cm) driven by a 95-horsepower tractor (Dong fanghong-LX954, Luoyang, China) was used. In the STS plot, chisel plow machinery was used (30–35 cm). The NTS plots were not disturbed by using the tillage machine either before or after the establishment of the experiment except during sowing with a planter.

Crop management

Crop cultivation and management were applied in the experiment from October 2013 to October 2015 (Table 1). Winter wheat cultivar Shanmai-139 was sown by using a wheat drill at the rate of 208–210 kg ha⁻¹ on October 18, 2013/2014. Summer maize (CV Shandan-609) was planted using a maize drill at the rate of 30 kg ha⁻¹ on June 17, 2014/2015. Every year after tillage in June before planting summer maize, all the plots received applications of P_2O_5 (172 kg P_2O_5 ha⁻¹) and N (68 kg N ha⁻¹) as the diammonium phosphate fertilizer broadcast. At the seven-leaf stage of summer maize, 172 kg N ha⁻¹ was applied as the urea fertilizer by broadcasting according to local recommendations. Similarly, every year after tillage in October, all the plots received applications of P_2O_5 (172 kg P_2O_5 ha⁻¹) and N (68 kg N ha⁻¹) as the diammonium phosphate fertilizer by broadcasting according to local recommendations. Similarly, every year after tillage in October, all the plots received applications of P_2O_5 (172 kg P_2O_5 ha⁻¹) and N (68 kg N ha⁻¹) as the diammonium phosphate fertilizer broadcast. At the same time, 160 kg N ha⁻¹ in the form of urea was applied. Weeds were killed using herbicides. The distance between the rows of wheat was 16 cm. The row spacing was 70 cm, and the plant spacing was 20 cm in the summer maize field. No irrigation was applied at any other time during the entire crop growing season.

Crop yields and root biomass measurements

At the winter wheat maturity stage, the grain yield of the winter wheat was determined by harvesting three $1-m^2$ sampling areas per treatment by hand. At the summer maize maturity stage, 20 plants from the middle rows per subplot were randomly selected and harvested by hand to determine the yield, and the straw was cut for the crop straw biomass determination.



Fig 2. Season variations of soil respiration (a, 2013–14; b, 2014–15), microbial respiration (c, 2013–14; d, 2014– 15) during the cycle of wheat-maize rotation in 2013–2015 in Yangling, China. (NTS: no tillage with straw mulching; RTS: rotary tillage with straw incorporation; STS: chisel plow tillage with straw incorporation; CT: conventional moldboard plowing tillage without crop straw).

Tabl	le 1.	Summar	y of e	xperimenta	l d	lesign witł	1 f	our ti	llage	e treatments	on	Loess	Plateau	in	China.
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Туре		Oper	ations		
Soil tillage	Conventional moldboard plowing tillage	Rotary tillage	Chisel plow tillage	No tillage	
Straw methods	No straw	Straw incorporation	Straw incorporation	Straw mulching	
Winter crops	Wheat	Wheat	Wheat	Wheat	
Sowing time	October	October	October	October	
Harvest time	June	June	June	June	
Fertilization	$\begin{array}{l} 324 \ \text{kg} \ (\text{NH}_4)_2 \text{HPO}_4 \ (172 \ \text{kg} \ \text{P}_2 \text{O}_5 \\ \text{ha}^{\text{-1}}; \ 68 \ \text{kg} \ \text{N} \ \text{ha}^{\text{-1}}) \ 348 \ \text{kg} \ \text{CO} \\ (\text{NH}_2)_2 \ (160 \ \text{kg} \ \text{N} \ \text{ha}^{\text{-1}}) \end{array}$	$\begin{array}{l} 324 \ \text{kg} \ (\text{NH}_4)_2 \text{HPO}_4 \ (172 \ \text{kg} \ \text{P}_2 \text{O}_5 \\ \text{ha}^{-1}; \ 68 \ \text{kg} \ \text{N} \ \text{ha}^{-1}) \ 348 \ \text{kg} \ \text{CO} \\ (\text{NH}_2)_2 \ (160 \ \text{kg} \ \text{N} \ \text{ha}^{-1}) \end{array}$	$\begin{array}{c} 324 \ \text{kg} \ (\text{NH}_4)_2 \text{HPO}_4 \ (172 \ \text{kg} \ \text{P}_2 \text{O}_5 \\ \text{ha}^{-1}; \ 68 \ \text{kg} \ \text{N} \ \text{ha}^{-1}) \ 348 \ \text{kg} \ \text{CO} \\ (\text{NH}_2)_2 \ (160 \ \text{kg} \ \text{N} \ \text{ha}^{-1}) \end{array}$	$\begin{array}{l} 324 \ \text{kg} \ (\text{NH}_4)_2 \text{HPO}_4 \ (172 \ \text{kg} \ \text{P}_2 \text{O}_5 \\ \text{ha}^{-1}; \ 68 \ \text{kg} \ \text{N} \ \text{ha}^{-1}) \ 348 \ \text{kg} \ \text{CO} \\ (\text{NH}_2)_2 \ (160 \ \text{kg} \ \text{N} \ \text{ha}^{-1}) \end{array}$	
Summer crops	Maize	Maize	Maize	Maize	
Sowing time	June	June	June	June	
Harvest time	October	October	October	October	
Fertilization	$\begin{array}{c} 324 \ \text{kg} \ (\text{NH}_4)_2 \text{HPO}_4 \ (172 \ \text{kg} \ \text{P}_2 \text{O}_5 \\ \text{ha}^{\text{-1}}; \ 68 \ \text{kg} \ \text{N} \ \text{ha}^{\text{-1}}) \ 374 \ \text{kg} \ \text{CO} \\ (\text{NH}_2)_2 \ (172 \ \text{kg} \ \text{N} \ \text{ha}^{\text{-1}}) \end{array}$	$\begin{array}{c} 324 \ \text{kg} \ (\text{NH}_4)_2 \text{HPO}_4 \ (172 \ \text{kg} \ \text{P}_2 \text{O}_5 \\ \text{ha}^{-1}; \ 68 \ \text{kg} \ \text{N} \ \text{ha}^{-1}) \ 374 \ \text{kg} \ \text{CO} \\ (\text{NH}_2)_2 \ (172 \ \text{kg} \ \text{N} \ \text{ha}^{-1}) \end{array}$	$\begin{array}{c} 324 \ \text{kg} \ (\text{NH}_4)_2 \text{HPO}_4 \ (172 \ \text{kg} \ \text{P}_2 \text{O}_5 \\ \text{ha}^{-1}; \ 68 \ \text{kg} \ \text{N} \ \text{ha}^{-1}) \ 374 \ \text{kg} \ \text{CO} \\ (\text{NH}_2)_2 \ (172 \ \text{kg} \ \text{N} \ \text{ha}^{-1}) \end{array}$	$\begin{array}{c} 324 \ \text{kg} \ (\text{NH}_4)_2 \text{HPO}_4 \ (172 \ \text{kg} \ \text{P}_2\text{O}_5 \\ \text{ha}^{-1}; \ 68 \ \text{kg} \ \text{N} \ \text{ha}^{-1}) \ 374 \ \text{kg} \ \text{CO} \\ (\text{NH}_2)_2 \ (172 \ \text{kg} \ \text{N} \ \text{ha}^{-1}) \end{array}$	
Experimental year	Oct.2013-Oct.2015	Oct.2013-Oct.2015	Oct.2013-Oct.2015	Oct.2013-Oct.2015	

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The remaining crop was harvested mechanically. In addition, wheat and corn roots were collected at the ripening stage. To get soil cores, a soil auger (diameter: 8 cm) was applied at three different locations, i.e., at the plant spots, intra-plant spots, and intra-row spots. Each core was taken from a depth of 0 to 100 cm in the soil profile and was incremented by 10 cm, i.e., 0–10, 10–20, 20–30, 30–40, 40–50, 50–60, 60–70, 70–80, 80–90, and 90–100 cm. The soil cores were soaked in a plastic container overnight, and then the root was carefully washed by swirling water through it. The soil material and old dead root debris were separated from the live roots manually. The aboveground and root dry weights were determined after drying the root samples in an oven at 105°C for 30 min and then at 75°C until constant dry weight. After weighing the dry weight, we crushed the aboveground and underground of crops and then, sieved the dry samples by using a 0.25-mm sieve. Then, we placed the samples in clean plastic bags to measure the carbon content. The carbon content of the plant samples was measured using potassium dichromate (K₂Cr₂O₄) and sulfuric acid (H₂SO₄) oxidation and ferrous sulfate (FeSO₄) [24].

Soil respiration measurements and estimation of root respiration

After planting the crop including winter wheat and summer maize, the PVC chambers (height: 20 cm; inner diameter: 11 cm) were placed and pressed by hand into the soil to a depth of 5 cm for the measurements of R_s and R_h by the closed chamber method using an infrared gas analyzer (GXH-3010EI, Beijing Huayuan Gas Chemical Industry Co., Ltd., Beijing, China). The samples were placed in field twice each year; one was placed on June 18 after plating the summer maize, and the other was placed on October 20 after planting the winter wheat. The date of measurement of the R_s and R_h values were shown in S1 and S2 Tables. In each plot, one chamber was located to measure R_s in the entire soil. As a result, three chambers were placed in each treatment to measure the R_s value in the entire soil for the three replications. Another three chambers were placed in the no-root zone for each root-free plot. The samples were kept free of vegetation by cutting the plants manually throughout the sampling period.

All R_s and R_h measurements were performed between 9:00 AM and 11:00 AM. local time to avoid the highest CO₂ emission at noon. The increase in the concentration of CO₂ within the chamber was measured after three minutes [25]. Root respiration (R_a) was estimated by subtracting the microbial respiration in the non-root zone from the soil respiration (R_s) in the whole soil.

The R_s was calculated using Eq (1):

$$F = \frac{K(X_2 - X_1)H}{\Delta t} \tag{1}$$

where *F* is the R_s value (mg CO₂ m⁻² h⁻¹); *K* is the reduction coefficient, which is equal to 1.80 at 25°C and 1 Pa; *H* is the height inserted in soil; and $\frac{X_2-X_1}{\Delta t}$ is the time rate of the change in CO₂ concentration in the air within the chamber (mg CO₂ m⁻³ h⁻¹). The total R_s and its components were calculated as follows:

$$FCO_2 = \sum_{i}^{n} \frac{F_{i+1} + F_i}{2} \times (t_{i+1} - t_i) \times 24 \times 10^{-4}$$
⁽²⁾

Where FCO₂ is the total emission of CO₂-C (t ha⁻¹), F_i is the first CO₂ emission value (mg CO₂-C m⁻² h⁻¹) at time t_i (h), and F_{i+1} is the following value at time t_{i+1} (h); n is the total number of CO₂ emission values.

Measurements of carbon balance

System dynamics. The net ecosystem carbon balance from the winter wheat-summer maize production system to the atmosphere was calculated using a full carbon cycle analysis [8–9]. Both carbon fixation within wheat-maize production system and emissions from agricultural practices were considered.

Net primary production (NPP). The NPP (t $C ha^{-1}$) of crops was calculated as follows [26]:

$$NPP = NPP_{grain} + NPP_{straw} + NPP_{root} + NPP_{litter} + NPP_{rhizodebosit}$$
(3)

Grain, straw, and root biomass NPP were converted by applying the dry biomass weight at harvest. Litter was calculated to account for 5% of the aboveground and root dry biomass [27], while rhizodeposits accounted for 18% [28] and 12% [29] of the aboveground and root dry biomass of wheat and maize.

Net ecosystem productivity (NEP). The net ecosystem productivity (NEP) was assessed according to [30] as follows:

$$NEP = NPP - R_{\rm h} \tag{4}$$

where R_h is the microbial respiration measured using the root exclusion technique.

Calculation of net ecosystem carbon balance (NCF). The net ecosystem carbon balance without considering the carbon emission from the farm inputs (NECB) was calculated according to Smith *et al.* [26].

$$NECB = NEP - Harvest \tag{5}$$

where NEP is the net ecosystem productivity and harvest means the grain harvest for the STS, NTS, and RTS treatments, while for the CT treatment, harvest means grain + straw.

The net ecosystem carbon balance considering the carbon emission from farm inputs (NCF) was then calculated as follows:

$$NCF = NEP - Harvest - C_{AP} = NECB - C_{AP}$$
(6)

where C_{AP} is the carbon emission from the agricultural input and the data are taken from Lu and Liao [31].

Carbon productivity. Carbon productivity (CP) can be calculated by using the following equation [32]:

С

$$P = \frac{Y_c}{C_{AP}} \tag{7}$$

where Y_c is the grain carbon content (kg C ha⁻¹) and C_{AP} is the carbon emission from the agricultural input (kg C ha⁻¹).

Data analysis

Data are shown as the mean values ± standard error. The two-way analysis of variance (ANOVA) with the SAS version 8 software package (SAS Institute, Cary, NC, USA) was used for analyzing the effects of the cropping years and the tillage treatments on the total soil respiration and its components, crop yield, carbon productivity, NPP, NECB, and NCF. When significant, the difference between treatments was determined at the 5% level by applying the least significant difference (LSD) test.

Results

Crop production

The grain yields of wheat and maize strongly showed the treatment differences (Table 2). The grain yields ranged from 6.14 to 6.87 t ha⁻¹ for wheat and 7.96 to 9.51 t ha⁻¹ for maize in 2013–14. No difference in the wheat yield was recorded among the different tillage treatments in 2013–14, while the STS significantly (p < 0.05) increased the maize yield by 19.5% as compared with CT. In 2014–15, the STS treatment significantly (p < 0.05) increased the wheat yield among the NTS, RTS, and CT treatments was recorded. Similar to the 2013–14 cropping year, the STS significantly (p < 0.05) increased the maize yield by 20.6% as compared to CT.

Seasonal variations in R_s, R_h, and R_a

Fig 2 shows the seasonal variation in R_s and R_h in the 2013–2015 cropping seasons. The seasonal patterns of R_s for both seasons showed a similar trend, with the peak appearing in August (at the heading stage of the maize crop) in both the cropping years. After winter wheat was sown, R_s was high because of the disturbance of the tillage application. Then, it decreased rapidly and the lowest values during the entire cropping year was recorded at the wintering stage (January 12–19, 2014/2015) because of the low temperature in the winter season. R_s then exhibited a dramatic increase when the air temperature recovered and fluctuated till the wheat harvest.

During the maize crop season, R_s increased rapidly from July to August and the highest value was recorded at the heading stage (August 10, 2014/2015). The peak value of R_s was 395.0 mg CO₂ m⁻² h⁻¹ for NTS, 557.8 mg CO₂ m⁻² h⁻¹ for RTS, 698.7 mg CO₂ m⁻² h⁻¹ for STS, and 696.2 mg CO₂ m⁻² h⁻¹ for CT in the 2013–14 cropping year. While in the 2014–15 cropping year, the peak value of R_s was recorded to be 427.8 mg CO₂ m⁻² h⁻¹ for NTS, 591.0 mg CO₂ m⁻² h⁻¹ for RTS, 602.0 mg CO₂ m⁻² h⁻¹ for STS, and 652.0 mg CO₂ m⁻² h⁻¹ for CT. R_s then reduced and varied as waves till the maize harvest (Fig 2).

 R_h showed a similar trend to that of R_s when the crop was small, but a contrasting trend to R_s was observed when the crop became big. The R_a and R_a/R_s values showed a similar trend,

Table 2. Total soil respiration (R_s), microbial respiration (R_h), and wheat and maize grain yields, straw yields and aboveground biomass during the wheat- and maize- growing seasons during the 2013–2015 rotation.

Treatments		Wheat season					Maize season					Annual				
	Rs (t CO ₂ -C ha ⁻¹ yr ⁻¹)	Rh (t CO ₂ -C ha ⁻¹ yr ⁻¹)	Grain yield (t ha ⁻¹)	Straw yield (t ha ⁻¹)	Aboveground biomass (t ha ⁻¹)	Rs (t CO ₂ -C ha ⁻¹ yr ⁻¹)	Rh (t CO ₂ -C ha ⁻¹ yr ⁻¹)	Grain yield (t ha ⁻¹)	Straw yield (t ha ⁻¹)	Aboveground biomass (t ha ⁻¹)	Rs (t CO ₂ -C ha ⁻¹ yr ⁻¹)	Rh (t CO ₂ -C ha ⁻¹ yr ⁻¹)	Grain yield (t ha ⁻¹)	Straw yield (t ha ⁻¹)	Aboveground biomass (t ha ⁻¹)	
2013-14																
NTS	2.02b	1.60c	6.44a	6.50a	12.94ab	2.05b	1.12c	8.20b	8.12ab	16.32b	4.07c	2.72d	14.64b	14.62ab	29.26bc	
RTS	3.28a	2.58ab	6.36a	6.76a	13.12ab	3.32a	1.70a	8.17b	8.24ab	16.41b	6.60b	4.28a	14.54b	14.99a	29.53b	
STS	3.51a	2.67a	6.87a	6.56a	13.43a	3.55a	1.46b	9.51a	8.67a	18.18a	7.05a	4.12b	16.38a	15.23a	31.61a	
СТ	3.12a	2.41b	6.14a	6.22a	12.36b	3.17a	1.39b	7.96b	7.73b	15.69b	6.29b	3.80c	14.09b	13.96b	28.05c	
2014-15																
NTS	2.03b	1.60b	6.54ab	6.70a	13.25ab	2.07b	1.11c	8.26b	8.13ab	16.20b	4.10c	2.71c	14.81b	14.83ab	29.44b	
RTS	3.32a	2.59a	6.36ab	6.27a	12.63b	3.43a	1.79a	8.31b	8.31ab	16.41b	6.75a	4.38a	14.66b	14.58bc	29.04bc	
STS	3.52a	2.61a	7.20a	6.78a	13.98a	3.56a	1.34b	9.61a	8.87a	18.48a	7.08a	3.95b	16.81a	15.65a	32.46a	
СТ	3.13a	2.42a	6.24b	6.26a	12.49b	3.19a	1.36b	7.97b	7.72b	15.69b	6.32b	3.77b	14.21b	13.97c	28.19c	
LSD0.05	0.166	0.131	0.267	0.311	0.521	0.283	0.139	0.193	0.423	0.554	0.389	0.213	0.667	0.822	1.244	

The different lowercase letters following the same column represent significant difference at 5% levels. CT: conventional moldboard plowing tillage without crop straw; RTS: rotary tillage with straw incorporation; STS: chisel plow tillage with straw incorporation; NTS: no tillage with straw mulching; Rs: soil respiration; Rh: microbial respiration.

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Fig 3. Season variations of root respiration (a, 2013–14; b, 2014–15) and its ratio (c, 2013–14; d, 2014–15) during the cycle of wheat-maize rotation in 2013–2015 in Yangling, China. (NTS: no tillage with straw mulching; RTS: rotary tillage with straw incorporation; STS: chisel plow tillage with straw incorporation; CT: conventional moldboard plowing tillage without crop straw).

and the season pattern of R_a and R_a/R_s was a unimodal curve, corresponding to the growing seasons of wheat and maize (Fig 3). The seasonal pattern of R_a/R_s resembled that of Ra and corresponded clearly to the development of crop (Fig 3). R_a/R_s for wheat exhibited a peak at approximately the flowering stage (April 13–27, 2014/2015). For the maize crop, R_a/R_s exhibited a peak value at approximately the heading stage (August 10–14, 2014/2015). At this time, R_s increased and R_h decreased, which resulted in an increase in R_a/R_s .

Cumulative R_s, and R_h

The cumulative R_s emissions from the wheat–maize rotation ranged from 4.07 t C ha⁻¹ for NTS to 7.05 t C ha⁻¹ for STS (Table 2) in the 2013–14 cropping season. While in the 2014–15 cropping season, the STS treatment significantly (p < 0.05) increased the cumulative R_s emissions by 12.0% and 72.7% as compared to CT and NTS, respectively. When compared with CT, the NTS significantly (p < 0.05) reduced the total R_s emissions by 54.1%, no difference in the cumulative R_s emissions was found between the RTS and STS treatments (Tables 2 and 3).

Similarly, the lowest cumulative R_h emissions were recorded in the NTS treatment. The NTS significantly (p < 0.05) reduced the cumulative R_h emissions from 28.4% to 36.4% as compared to the other three treatments in the 2013–14 cropping season. While in the 2014–15 cropping season, this reduction percentage varied from 28.1% to 38.1% (Table 2).

Net ecosystem carbon balance under conservation tillage treatments

The calculations of the NCF from the estimates of the potential carbon inputs from the aboveground biomass, root biomass, negative cumulative carbon loss via R_h , and agricultural input emissions resulted in differences among the different tillage treatments. The NPP values



Effect d.f. ^a				Maize season				Annual								
		Rs (t CO ₂ - C ha ⁻¹ yr ⁻¹)	Rh (t CO ₂ - C ha ⁻¹ yr ⁻¹)	Grain yield (t ha ⁻¹)	Straw yield (t ha ⁻¹)	Aboveground biomass (t ha ⁻¹)	Rs (t CO ₂ - C ha ⁻¹ yr ⁻¹)	Rh (t CO ₂ - C ha ⁻¹ yr ⁻¹)	Grain yield (t ha ⁻¹)	Straw yield (t ha ⁻¹)	Aboveground biomass (t ha ⁻¹)	Rs (t CO ₂ - C ha ⁻¹ yr ⁻¹)	Rh (t CO ₂ -C ha ⁻¹ yr ⁻¹)	Grain yield (t ha ⁻¹)	Straw yield (t ha ⁻¹)	Aboveground biomass (t ha ⁻¹)
Block	2	65.498**	38.379**	19.366**	8.401**	15.748**	19.355**	2.231	92.241**	9.29**	31.017**	41.895**	9.814**	69.687**	23.199**	57.795**
Year (Y)	1	0.246	0.108	1.102	0.001	0.268	0.464	0.29	0.758	0.12	0.03	0.5	0.307	1.811	0.098	0.347
2013-14		2.98a	2.32a	6.45a	6.51a	12.96a	3.02a	1.42a	8.46a	8.19a	16.65a	6.00a	3.73a	14.91a	14.70a	29.61a
2014-15		3.00a	2.31a	6.59a	6.50a	13.09a	3.07a	1.40a	8.54a	8.258a	16.70a	6.07a	3.70a	15.12a	14.76a	29.78a
LSD0.05		0.083	0.065	0.267	0.311	0.521	0.142	0.07	0.193	0.423	0.554	0.195	0.106	0.333	0.411	0.622
Tillage (T)	3	294.097**	250.871**	8.705**	1.706	4.825*	103.681**	63.449**	63.513**	4.78*	19.666**	216.317**	201.742**	48.218**	9.909**	32.733**
STS		3.51a	2.64a	7.04a	6.67a	13.71a	3.55a	1.40b	9.56a	8.77a	18.33a	7.07a	4.04b	16.60a	15.44a	32.04a
NTS		2.03d	1.60c	6.49b	6.60a	13.10ab	2.06c	1.11c	8.23bc	8.13b	16.26b	4.09d	2.72d	14.72b	14.73b	29.35b
RTS		3.30b	2.59a	6.36b	6.51a	12.87b	3.38ab	1.74a	8.24b	8.27ab	16.41b	6.68b	4.33a	14.59bc	14.79b	29.29b
СТ		3.13c	2.41b	6.19b	6.24a	12.43b	3.18b	1.37b	7.97c	7.73b	15.69b	6.31c	3.79c	14.16c	13.97c	28.12c
LSD0.05		0.117	0.092	0.378	0.439	0.737	0.20	0.098	0.273	0.598	0.783	0.275	0.15	0.471	0.582	0.879
Y×T	3	0.033	0.233	0.328	1.319	0.843	0.121	1.71	0.079	0.058	0.12	0.108	1.254	0.229	0.852	0.891
LSD0.05		0.166	0.131	0.535	0.311	1.042	0.283	0.139	0.386	0.846	1.108	0.195	0.213	0.667	0.822	1.244

Table 3. ANOVA of total soil respiration (R_s), microbial respiration (R_h), wheat and maize grain yields, straw yields and aboveground biomass during the wheatand maize- growing seasons during the 2013–2015 rotation.

The different lowercase letters following the same column represent significant difference at 5% levels

** is significant at the P \leq 0.01 level

 * is significant at the P \leq 0.05 level.

CT: conventional moldboard plowing tillage without crop straw; RTS: rotary tillage with straw incorporation; STS: chisel plow tillage with straw incorporation; NTS: no tillage with straw mulching; Rs: soil respiration; Rh: microbial respiration.

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ranged from 10.4 to 11.7 t C ha⁻¹ yr⁻¹ in 2013–14 and from 10.5 to 12.0 t C ha⁻¹ yr⁻¹ in 2014–15 (Table 4). In the 2013–14 and 2014–15 cropping years, the carbon loss through R_h in the winter wheat–summer maize ecosystem accounted for NPP 25.1%–39.1% and 24.7%–40.4%, and the harvest part accounted for NPP 28.8%–90.5% and 45.6%–90.4%.

The overall NCFs were significantly (p < 0.05) affected by the tillage practices (Tables <u>4</u> and <u>5</u>). The lowest NCF value was found under the conventional moldboard plow tillage treatment

rable 4. The agro-ecosystem C balance (NCF) and its main components for the annual cycle of wheat-maize ro	ta-
ion in 2013–2015 (t C ha ⁻¹ yr ⁻¹).	

Year	Treatment	NPP	R _h	Harvest	NECB	C _{AP}	NCF
	NTS	10.85bc	2.72d	4.95c	3.12a	0.59	2.53a
2013-14	RTS	10.95b	4.28a	4.91c	1.72b	0.67	1.05b
	STS	11.71a	4.12b	5.53b	1.91b	0.67	1.24b
	СТ	10.40c	3.80c	9.41a	-2.85c	0.76	-3.61c
	NTS	10.99b	2.71c	5.01c	3.28a	0.59	2.69a
2014-15	RTS	10.84bc	4.38a	4.95c	1.51c	0.67	0.84c
	STS	12.02a	3.95b	5.68b	2.39b	0.67	1.72b
	СТ	10.45c	3.77b	9.45a	-2.77d	0.76	-3.53d
LSD0.05		0.461	0.213	0.292	0.390		0.390

The different lowercase letters following the same column represent significant difference at 5% levels. CT: conventional moldboard plowing tillage without crop straw; RTS: rotary tillage with straw incorporation; STS: chisel plow tillage with straw incorporation; NTS: no tillage with straw mulching; NPP: net primary productivity; Rh: microbial respiration; NECB: net ecosystem carbon balance without considering the carbon emission from farm inputs; C_{AP}: the carbon emission from agricultural input; NCF: the net ecosystem carbon balance with considering the carbon emission from farm inputs.

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Effect	D.f. ^a	NPP	Harvest	NECB	NCF
Block	2	57.912**	59.387**	28.738**	28.738**
Year (Y)	1	0.866	1.135	1.966	1.966
2013-14		10.98a	6.201a	0.974a	0.302a
2014-15		11.08a	6.273a	1.102a	0.429a
LSD0.05		0.231	0.146	0.195	0.195
Tillage (T)	3	31.523**	998.385**	848.846**	891.298**
STS		11.87a	5.607b	2.15b	1.48b
NTS		10.92b	4.978c	3.20a	2.61a
RTS		10.90b	4.933c	1.61c	0.94c
СТ		10.43c	9.430a	-2.81d	-3.57d
LSD0.05		0.326	0.206	0.276	0.276
Y×T	3	0.688	0.134	2.458	2.458
LSD0.05		0.461	0.292	0.390	0.390

Table 5. ANOVA of net primary productivity (NPP), harvest, net ecosystem carbon balance and net carbon flux during the wheat- and maize- growing seasons during the 2013–2015 rotation.

The different lowercase letters following the same column represent significant difference at 5% levels ** is significant at the P $\!\leq\!0.01$ level.

CT: conventional moldboard plowing tillage without crop straw; RTS: rotary tillage with straw incorporation; STS: chisel plow tillage with straw incorporation; NTS: no tillage with straw mulching; NPP: net primary productivity; Rh: microbial respiration; NECB: net ecosystem carbon balance without considering the carbon emission from farm inputs; CAP: the carbon emission from agricultural input; NCF: the net ecosystem carbon balance with considering the carbon emission from farm inputs.

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in the rain-fed winter wheat-summer maize field on the Loess Plateau in China at -3.61 t C ha⁻¹ yr⁻¹ in 2013–14, and -3.53 t C ha⁻¹ yr⁻¹ in 2014–15, whereas the highest NCF value was found under NTS at 2.53 t C ha⁻¹ yr⁻¹ in 2013–14, and 2.69 t C ha⁻¹ yr⁻¹ in 2014–15. This may be attributed to the lower carbon loss via R_h and higher carbon returning in the NTS treatment. There was no significant difference in the NCF values between the RTS and the STS treatments. Over the cycle of wheat-maize rotation, all of the treatments, except for the CT, led to carbon gains of 1.05 to 2.53 t C ha⁻¹ yr⁻¹ in 2013–14, and 0.84 to 2.69 t C ha⁻¹ yr⁻¹ in 2014–15. The CT treatment had a negative NCF value, mainly attributed to the higher harvest part as compared to the other straw returning treatments. The carbon in an agroecosystem mainly depended on the addition of organic matter; in the present study, positive NCF under three conservation tillage treatments was mainly attributed to the amount of crop residues in the soil. These results pointed out the importance of using crop straw in no tillage and reduced tillage for increasing the carbon input in the wheat-maize field on the Loess Plateau in China.

Effect of conservation tillage on carbon productivity

Tillage significantly (p < 0.05) affected the carbon productivity of the winter wheat (Fig 4). When compared with CT, STS and NTS significantly increased the carbon productivity of winter wheat by 28.3% and 36.0% in 2013–14, respectively. This increase percentage was 32.4% and 36.0% in the 2014–15 cropping year. Similarly, STS and NTS significantly improved the carbon productivity of summer corn from 31.1% to 36.8% in both the years. When compared with the CT treatment, STS, NTS, and RTS significantly increased the annual carbon productivity by 32.4%, 33.3%, and 17.6% in 2013–14, while in 2014–15, STS, NTS, and RTS significantly increased the annual carbon productivity by 34.6%, 33.7%, and 17.5% as compared to the CT treatment.





Fig 4. Effect of different tillage treatments on carbon productivity of winter wheat (a), summer corn (b), and annual (c) in both cropping years. (The different lowercase letters above the error bars represent significant difference between different tillage treatments within a two-year period at 5% levels according to the LSD test. CT: conventional moldboard plowing tillage without crop straw; RTS: rotary tillage with straw incorporation; STS: chisel plow tillage with straw incorporation; NTS: no tillage with straw mulching).

Discussion

Effect of conservation tillage on crop yields

In the present study, STS significantly increased the wheat yield by 15.4% as compared to the CT treatment in the 2014–15 cropping year. For the maize crop, when compared with CT, the STS treatment significantly improved the maize yield by 19.5% in 2013–14 and 20.6% in 2014–15, respectively. Similarly, Wang *et al.* [33] reported that chisel plow tillage is beneficial for reducing the soil bulk density, improving soil structure, increasing soil water availability, and aeration. Xu *et al.* [34] also reported that when compared to conventional tillage, chisel plow could improve the yield of crops by loosening the soil and promoting the root growth of the crop. In the present study, no difference in the grain yield was recorded between the NTS and the CT treatments. However, Liu *et al.* [35] reported that no tillage is beneficial for improving the crop yield and water use efficiency in Weibei Highland, China. These inconsistent results may be attributed to the difference in the cropping years, crop, climate conditions, and the

duration of conservation tillage. Our results suggest that long-term studies are required to identify the initial and long-term yield constraints of conservation tillage.

Effect of conservation tillage on R_s and its components

The two-year data showed that NTS reduced R_s by 13.0% as compared to the other tillage treatments, which was mainly attributed to the lack of disturbance in NTS after the harvest of summer maize. In addition, when compared with CT, the STS and RTS treatments increased the total annual R_s by 12.0% and 6.8%, respectively. This was mainly attributed to the organic carbon input and the tillage disturbance. Li *et al.* [36] also reported that R_s increases when the organic carbon is input into the soil. Our results suggested that the organic carbon input in the STS, RTSm and NTS treatments promoted the activity of microbial and ultimately affected the decomposition of the organic matter and the release of soil CO₂, and influenced the carbon balance. Although STS and RTS increased the R_h value, the less carbon input (i.e., diesel fuel and the carbon loss by residue removal) compensated for the higher carbon emissions; finally, STS and RTS showed the carbon sink. In addition, the highest R_h value was recorded in RTS, which was mainly attributed to the fact that the tillage depth was only 15 cm, which made it easier for the microbes to touch the residue.

NCF in winter wheat-summer corn ecosystem under different tillage treatments

This is a very important method to mitigate the atmospheric CO_2 emission by increasing the soil carbon pool [18]. This study focused on the effect of different tillage treatments on the carbon balance of the winter wheat–summer maize ecosystem. The carbon storage or loss mainly relied on the balance of carbon fixed into the soil through the addition of residues and carbon loss through the respiration in a dryland agricultural ecosystem. Our results showed that the carbon loss through R_h and agricultural inputs accounted for NPP 24.7%–40.4% in the different tillage treatments in the 2013–14 and 2014–15 cropping years, respectively. In addition, conservation tillage practices such as the NTS, RTS, and STS treatments were beneficial for carbon sequestration irrespective of the carbon input from the agricultural inputs, which agreed with the other reported results [37–39]. Moreover, the contribution of the agricultural input to the total carbon emission varied from 5.1% to 17.8% in the different tillage treatments, which indicated that the carbon emission from the agricultural input should also be included when evaluating the carbon sink.

Moreover, the positive value of NCF was mainly attributed to the large amount input of the crop straws in the three conservation tillage treatments. In the experiment, the carbon input under the NTS, RTS, and STS treatments varied from $5.90-6.34 \text{ t C ha}^{-1} \text{ yr}^{-1}$ in both years. Han *et al.* [40] also showed that SOC sequestration increases with the annual input rate of straw C input rate. The winter wheat–summer maize ecosystem showed a carbon sink for all the treatments without considering the carbon harvest part. However, when the carbon harvest part was derived from NPP – R_h, this ecosystem was a carbon source for the CT treatment. Our results were similar to those of Li *et al.* [41] who showed that the winter wheat–summer maize rotation ecosystem was a carbon sink without considering the carbon part; however, the carbon sink changed into a source when the harvest carbon part was included. Moreover, when considering the carbon emission from the agricultural input, the carbon balance decreased for all the treatments. Thus, our results suggested that the purpose of the carbon sink could be realized by increasing the carbon inputs of the crop residues and roots.

Carbon productivity

When combining the winter wheat and the summer maize, the highest annual carbon productivity was observed in the STS and NTS treatments, and the lowest value in the CT treatment. In the NTS treatment, the higher carbon productivity was mainly attributed to the lower carbon emission from the farm input because of the similar grain yield to that in the CT treatment. While for the STS treatment, the higher carbon productivity was mainly attributed to the higher crop productivity. Similarly, van den Putte et al. [42] reported that deep conservation tillage performs better than superficial conservation tillage. This indicated that plants could benefit from the increased pore space and aeration at deeper depths. In addition, studies have already shown that the root growth conditions for cereals are less favorable in the case of the NT treatment than in the case of the CT treatment [43–44]. However, Boomsma et al. [45] showed that as compared to the moldboard plow, conservation tillage reduced the crop yields. This was mainly attributed to the relatively poor seedbed conditions, delayed seedling emergence, and crop development in the case of conservation tillage with respect to those in the case of the moldboard plow tillage treatment. Chisel plow tillage may therefore be expected to be more favorable than superficial tillage for crops. Our results also suggested that the conversion of treatment from conventional tillage to conservation tillage could increase the annual carbon productivity.

Conclusions

Our results showed that heterotrophic (microbial) respiration was lower in the NTS treatment than in other three tillage treatments. In the case of the CT treatment in the winter wheat–summer maize field on China's Loess Plateau, carbon added as the aboveground biomass and root biomass was not sufficient to compensate for the loss of carbon from organic matter decomposition, rendering the rain-fed winter wheat–summer maize field as the net sources of atmospheric CO_2 . The conversion from conventional tillage to conservational tillage substantially enhanced the carbon sink potential from 0.84 t C ha⁻¹ yr⁻¹ in RTS to 2.69 t C ha⁻¹ yr⁻¹ in the NTS treatment. Thus, the expansion of conservational tillage could enhance the potential carbon sink of rain-fed land in China's Loess Plateau. Our results also showed the importance of the returning of crop straw to the field in order to change the winter wheat–summer maize ecosystem from carbon source to sink.

Supporting information

S1 Table. Soil respiration under different tillage treatments. (DOCX)

S2 Table. Soil microbial respiration under different tillage treatments. (DOCX)

S3 Table. Root Respiration under different tillage treatments. (DOCX)

S4 Table. The ratio of root respiration to total soil respiration under different tillage treatments.

(DOCX)

S5 Table. The total respiration (R_s), microbial respiration (R_h), wheat and maize grain yields, straw yields and aboveground biomass during the wheat- and maize- growing seasons during the 2013–2015 rotation. (DOCX) S6 Table. The agro-ecosystem C balance (NCF) and its main components for the annual cycle of wheat-maize rotation in 2013–2015. (DOCX)

S7 Table. The carbon productivity under different treatments. (DOCX)

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References

- Tubiello FN, Salvatore M, Ferrara AF, House J, Sandro FS, Rossi S, et al. The contribution of agriculture, forestry and other land use activities to global warming, 1990–2012. Global Change Biol. 2015; 21: 2655–2660.
- Metz B, Davidson OR, Bosch PR, Dave R, Meyer LA. Climate change 2007-mitigation of climate change: contribution of working group III to the fourth assessment report of the intergovernmental panel on climate change. United Kingdom and New York: Cambridge University Press; 2007.
- Peng X, Wu X, Wu F, Wang X, Tong X. Life cycle assessment of winter wheat-summer maize rotation system in Guanzhong region of Shaanxi province. Journal of Agro-Environment Science. 2015; 34: 809–816 (In Chinese with English abstract).
- Han B, Kong F, Zhang H, Chen F. Effects of tillage conversion on carbon sequestration capability of farmland soil doubled cropped with wheat and corn. Chinese Journal of Applied Ecology. 2010; 21:91– 98 (In Chinese with English abstract). PMID: 20387429
- 5. West T, Marland G. Net carbon flux from agricultural ecosystems: methodology for full carbon cycle analyses. Environ Pollut. 2002; 116: 439–444. PMID: 11822723
- Wang Y, Xu H, Wu X, Zhu Y, Gu B, Niu X, et al. Quantification of net carbon flux from plastic greenhouse vegetable cultivation: a full carbon cycle analysis. Environ Pollut. 2011; 159: 1427–1434. https://doi.org/10.1016/j.envpol.2010.12.031 PMID: 21277056
- Zhang MY, Wang FJ, Chen F, Malemela MP, Zhang HL. Comparison of three tillage systems in the wheat-maize system on carbon sequestration in the North China Plain. J Clean Prod. 2013; 54: 101– 107.
- 8. Lin H, Wang J, Xu Z, Chen Z. Research progress and trend of the carbon cycle in grassland agroecosystem. Pratacultural Science. 2005; 22: 59–62 (In Chinese with English abstract).
- Tian Y. Advance in research on carbon cycling in wetland soils. Journal of Yangtze University (Nat Sci Edit). 2005; 2: 1–4 (In Chinese with English abstract).
- Wei H, Ma X, Liu A, Feng L, Huang Y. Review on carbon cycle of forest ecosystem. Chinese Journal of Eco-Agriculture, 2007, 15: 43–46 (In Chinese with English abstract).
- Zha T, Zhang Z, Zhu J, Cui L, Zhang J, Chen J, et al. Carbon storage and carbon cycle in forest ecosystem. Science of Soil and Water Conservation. 2008; 6: 112–119 (In Chinese with English abstract).
- Li L, Wu F, Zhang H, Chen F. Organic carbon and carbon pool management index in soil under conversation tillage in two- crop paddy field area. Journal of Agro-Environment Science. 2008; 27: 0248–0253 (In Chinese with English abstract).

- Luo Y, Wang Z, Gao M, Wei C. Effects of different tillage systems on soil labile organic matter and carbon management index of purple paddy soil. Journal of Soil and Water Conservation. 2007; 21: 55–58 (In Chinese with English abstract).
- Rui W, Zhou B, Zhang W. A brief assessment of carbon sequestration effects of conservational farming systems in paddy soils of Yangtze delta plain. Resources and Environment in the Yangtze Basin. 2006; 15: 207–212 (In Chinese with English abstract).
- Blanco-Canqui H, Lal R. Soil and crop response to harvesting corn residues for biofuel production. Geoderma. 2007; 141:355–362.
- Graham RL, Nelson R, Sheehan J, Perlack RD, Wright LL. Current and potential US corn stover supplies. Agron J. 2007; 99: 1–11.
- Wienhold BJ, Varvel GE, Johnson JMF, Wilhelm WW. Carbon source quality and placement effects on soil organic carbon status. Bioenerg Res. 2013; 6: 786–796.
- Snyder CS, Bruulsema TW, Jensen TL, Fixen PE. Review of greenhouse gas emissions from crop production systems and fertilizer management effects. Agr Ecosyst Environ. 2009; 133: 247–266.
- Wang C, Pan G, Tian Y. Characteristics of cropland topsoil organic carbon dynamics under different conservation tillage treatments based on long-term agro-ecosystem experiments across mainland China. Journal of Agro-Environment Science. 2009; 28: 2464–2475 (In Chinese with English abstract).
- Li X, Toma Y, Yeluripati J, Iwasaki S, Bellingrath-Kimura SD, Jones EO, et al. Estimating agro-ecosystem carbon balance of northern Japan, and comparing the change in carbon stock by soil inventory and net biome productivity. Sci Total Environ. 2016; 554: 293–302. https://doi.org/10.1016/j.scitotenv.2016.02.151 PMID: 26956176
- Zhao R, Liu Y, Ding M, Jiao S. Research on carbon source and sink of farmland ecosystem in Henan province. Journal of Henan Agricultural Sciences. 2010; 40–44 (In Chinese with English abstract).
- 22. Chen ZD, Dikgwatlhe SB, Xue JF, Zhang HL, Chen F, Xiao XP. Tillage impacts on net carbon flux in paddy soil of the Southern China. J Clean Prod. 2015; 103: 70–76.
- Li XD, Fu H, Guo D, Li XD, Wan CG. Partitioning soil respiration and assessing the carbon balance in a setaria italica (L.) beauv. cropland on the Loess Plateau, Northern China. Soil Biol Biochem. 2010; 42 (2): 337–346.
- 24. Bao SD. Soil Agro-chemistrical Analysis. 3rd ed. Beijing: China Agriculture Press, 2008. (In Chinese).
- Gao CD, Sun XY, Gao J, Luan YN, Hao HD, Li ZJ, et al. A method and apparatus of measurement of carbon dioxide flux from soil surface in situ. Journal of Beijing for University. 2008; 30: 102–105 (In Chinese with English abstract).
- Smith P, Lanigan G, Kutsch WL, Buchmann N, Eugster W, Aubinet M, et al. Measurements necessary for assessing the net ecosystem carbon budget of croplands. Agr Ecosyst Environ. 2010; 139: 302– 315.
- Kimura M, Murase J, Lu Y. Carbon cycling in rice field ecosystems in the context of input, decomposition and translocation of organic materials and the fates of their end products (CO₂ and CH₄). Soil Biol Biochem. 2004; 36: 1399–1416.
- Gregory P. Roots, rhizosphere and soil: the route to a better understanding of soil science? Eur J Soil Sci. 2006; 57: 2–12.
- Nguyen C. Rhizodeposition of organic C by plants: mechanisms and controls. Agronomie. 2003; 23: 375–396.
- 30. Grace J. Understanding and managing the global carbon cycle. J Ecol. 2004; 92: 189–202.
- **31.** Lu X, Liao Y. Effect of tillage practices on net carbon flux and economic parameters from farmland on the Loess Plateau in China. J Clean Prod. 2016; 162: 1617–1624.
- Zhang MY, Wang FJ, Chen F, Malemela MP, Zhang HL. Comparison of three tillage systems in the wheat-maize system on carbon sequestration in the North China Plain. J Clean Prod. 2013; 54: 101– 107.
- Wang Q, Lu C, Li H, He J, Sarker KK, Rasaily RG, et al. The effects of no-tillage with subsoiling on soil properties and maize yield: 12-Year experiment on alkaline soils of Northeast China. Soil Till Res. 2014; 137: 43–49.
- Xu D, Schmid R, Mermoud A. Effects of tillage practices on the variation of soil moisture and the yield of summer maize. Transactions of the Chinese Society of Agricultural Engineering. 1999; 15(3): 101–106 (in Chinese with English abstract).
- Liu D, Zhang, LI J, Wang X.D. Effects of different tillage patterns on soil properties, maize yield and water use efficiency in weibei highland, China. Chinese Journal of Applied Ecology. 2018; 29(2): 573– 582 (in Chinese with English abstract). https://doi.org/10.13287/j.1001-9332.201802.023 PMID: 29692073

- Li Y, Wang G, Li W. Soil respiration and carbon cycle. Earth Science Frontiers. 2002; 9: 351–357 (In Chinese with English abstract).
- 37. Norton JB, Mukhwana EJ, Norton U. Loss and recovery of soil organic carbon and nitrogen in a semiarid agroecosystem. Soil Sci Soc Am J. 2012; 76: 505–514.
- Roldan A, Salinas-Garcia J, Alguacil M, Caravaca F. Changes in soil enzyme activity, fertility, aggregation and C sequestration mediated by conservation tillage practices and water regime in a maize field. Appl Soil Ecol. 2005; 30: 11–20.
- West TO, Post WM. Soil organic carbon sequestration rates by tillage and crop rotation. Soil Sci Soc Am J. 2002; 66: 1930–1946.
- **40.** Han X, Xu C, Dungait JAJ, Bol R, Wang XJ, Wu WL, et al. Straw incorporation increases crop yield and soil organic carbon sequestration but varies under different natural conditions and farming practices in China: a system analysis. Biogeosciences. 2018; 15: 1933–1946.
- Li J, Yu Q, Sun X, Tong X, Ren C, Wang J, et al. Carbon exchange and its regulating mechanism in farmland ecosystem of North China plain. Science in China Series D: Earth Sciences. 2006; 36: 210– 223 (In Chinese with English abstract).
- 42. Van den Putte A, Govers G, Diels J, Gillijins K, Demuzere M. Assessing the effect of soil tillage on crop growth: a meta-regression analysis on European crop yields under conservation agriculture. Eur J Agron. 2010; 33: 231–241.
- **43.** Pietola LM. Root growth dynamics of spring cereals with discontinuation of mouldboard ploughing. Soil Till Res. 2005; 80: 103–114.
- 44. Qin R, Stamp P, Richner W. Impact of tillage on root systems of winter wheat. Agron J. 2004; 96: 1523–1530.
- 45. Boomsma CR, Santini JB, West TD, Brewer JC, Mcintyre LM, Vyn TJ. Maize grain yield responses to plant height variability resulting from crop rotation and tillage system in a long-term experiment. Soil Till Res. 2010; 106: 227–240.