Disturbing Effects of Chronic Low-dose 4-Nonylphenol exposing on Gonadal Weight and Reproductive Outcome over One-generation

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ABSTRACT: 4-Nonylphenol (NP) is a surfactant that is a well-known and widespread estrogenic endocrine disrupting chemical (EDC). Although it has been known that the affinity of NP to ERs is low, it has been suggested that low-dose NP has toxicity. In the present study, the endocrine disrupting effects on reproduction, and the weight of gonads, epididymis, and uterus were evaluated with the chronic lower-dose NP exposing. This study was designed by following the OECD test guideline 443 and subjected to a complete necropsy. In male, NP had an effect on the weight of the testis and epididymis in both F_0 and F_1 . In females, NP decreased the weight of ovary and uterus in F_0 but not in pre-pubertal F_1 pubs. Fertility of male and female in F_0 or F_1 was no related with NP administration. The number of caudal-epididymal sperm by body weight (BW) was not different between groups in both F_0 and F_1 . Besides, the difference of the sperm number between generations was not detected. The number of ovulated oocytes was similar between groups in F_0 , but significantly decreased in NP 50 group of F_1 . The litter size and sex ratios of offspring in F_1 and F_2 were not different. The accumulated mating rate and gestation period were not affected by the NP administration. Those results shows that chronic lower-dose NP administration has an effect of endocrine disruptor on the weight of gonads and epididymis of F_0 and F_1 but not in reproduction. Based on the results, it is suggested that chronic lower-dose NP exposing causes endocrine disruption in the weight of gonad and epididymis but not in the reproductive ability of next generations.

Key words: Chronic low-dose nonylphenol, Endocrine disrupting, Nonylphenol, Reproduction, Sperm count

INTRODUCTION

Endocrine disrupting chemical (EDC) is prevalent in the environment (Fox et al., 2004; Choi et al., 2016; Monneret, 2016), and animals can be exposed to EDC via their occuations, ingestion of food (Fox et al., 2004), dust and water (Rudel et al., 2009), and skin (Davey et al., 2007; Kabir et

al., 2015). Those EDCs can affect human and wildlife animal reproductive performance (Culty et al., 2008; Phillips et al., 2008; Robins et al., 2011). Based on the dogma of pharmacology, the dose makes the poison (Vandenberg et al., 2012), and most of the studies about EDC have been focused in the toxicological concepts for decades (Swedenborg, 2009; Vandenberg et al., 2012). However, some

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researches have suggested that EDCs can act in the nanomolar to micromolar range or at picomolar levels like in natural hormones and can damage human and/or animal health. Besides, research is needed to understand whether low-dose responses are lined to adverse effects and mechanisms. Therefore, it becomes more and more important to know the possible effects of endocrine disrupting chemicals (EDCs) in physiological process (Kawaguchi et al., 2015; Mantovani, 2017).

The environmental xenobiotics can get potential to hormonal signaling disruptors by itself or their metabolites (Bonde et al., 2017). So far, the information about metabolized EDCs and specific exposure is insufficient. The degradation products of nonlyphenol ethoxylate (NPE) included 4-nonylphenol (NP), nonylphenol ethoxylate (NPIEO), nonylphenol diethoxylate (NP2EO), nonylphenol triehoxylates (NP1-3EO), nonlyphenoxyacetic acid (NP1EC), and nonylphenolxyacetic acid (NP2EC) (Sheahan et al., 2002; Swartz et al., 2006). Those belong to aklyphenols and are considered to have weak estrogenic activities in vivo and in vitro testing system (Bøgh et al., 2001; Watanabe et al., 2004; Maruya et al., 2012). It is known that NP binds to estrogen receptors and mimics the work of 17β-estradiol, although the affinity of NP for ESR1 has less magnitude than the affinity of 17ß-estradiol (Preuss et al., 2006; Bonefeld-Jørgensen et al., 2007; Soares et al., 2008).

EDCs have been known to produce adversely embryonic development, reproductive, neurological and immune effects in animal (Diamanti-Kandarakis et al., 2009). The possible toxicity of EDC depends on the dosage, duration, and the expose of life stage. Maternal exposure of EDCs can be negatively influenced on parents as well as F₁ generation (Bøgh et al., 2001). The NP (100-300 mg/kg/day for 30 days; Aly et al., 2012) during perinatal period or adulthood (100 mg/kg; Duan et al., 2017) has a negative effect on spermatogenesis and sperm quality in vivo fertility. Neonatal expose delay testes decent and adult expose affect particularly on spermatogenesis (250 mg/kg, Dejager et al.,

1999a; 400 mg/kg, Dejager et al., 1999b; 8 mg/kg, Odum & Ashby, 2000). It also can give an effect on other organs and functions such as learning and memory (Jie et al., 2013; Kawaguchi et al., 2015). By such experimental results and warning of ECDs, REACH (registration, evaluation, authorization and restriction of chemicals), European chemical agency (ECHA), and Significant New Use Rules (SNURS) has restricted the percentage of NP as 0.1%.

Recently the possible effects of low dose EDCs were continuously suggested. Low-dose NP has effects in both cell and in vivo levels (Melnick et al., 2002; Xu et al., 2017). Yu et al (2011) explored that 100 µg/kg NP can modulate the production of progesterone in rat granulosa cells. 0.5 mg/kg NP slightly impairs special learning performance in male and female rat (Kawaguchi et al., 2015). In this study, we employed the OECD test guideline 443 "the expended one-generation reproductive toxicity study" to evaluate the possible adverse effects of chronic lowdose effects on the body weight, reproductive organ weight, and reproduction.

MATERIALS & METHODS

1. Experimental animals

All experimental animal(s) was studied according to the Guide for the Care and Use of Laboratory Animals published by the national Institutes of Health and to the Test Guideline 443 suggested by OECD and under the Experimental Animals Committee of Sungshin University. Animals (CD-1 mice) were maintained under standard condition; temperature (20–24 $^{\circ}$ C), humidity (45–55%), and light (14 hr light/ 10 hr dark) conditions. They were fed free phytoestrogen diet (2018 Teklad global 18% protein rodent diets; ENVIGO, Madison, WI, USA) and water ad libitum using glass bottles. To get F₀ generation, the mice fed phytoestrogen free diet were used. Their offspring with normal estrous cycle was chosen as F₀. The pubs were weaned at 21 post-natal days.

2. Chemical treatments

Nonylphenol (Sigma-Aldrich, Cat # 46018, CAS # 84852-15-3) was selected as an endocrine disrupting chemical and used at two concentration based on previous report (Kyselova et al., 2003), 50 μ g/L and 500 μ g/L in drinking water. The chemical stock solution was prepared in purified water and stored at $-20\,^{\circ}$ C. Treatment was proceded according on schedule suggested by OECD test guideline (TG) 443. Briefly, paternal generation (F0) were administered in drinking water during 10weeks (pre-mating; 2 weeks, mating; 2 weeks, post mating included pregnancy and lactation; 6 weeks) and then were anatomized for organ sampling (Fig. 1). F₁ were administered for experiment of organ sampling during postnatal day 51 (male) and about postnatal day 28 (female, after vagina opening 3 days).

3. Body and organ weight

The control and experimental groups of the same sex were sacrificed at the same time. The males and females F_0 were sacrificed at end of chemical treatment periods for 10 weeks, and the reproductive organ (testes and epididymis in males; ovary and uterus in females) were excised out and weighed individually. Relative organs weight was cal-

culated based on organ to body weight. In F₁ mice, they were sacrificed at postnatal day 51(male) and postnatal day 28 (female, after vagina opening 3 days) and examined as F₀.

4. Oocyte and sperm counting

To evaluate the number of ovulated oocytes, F₀ and F₁ generation female (6 weeks old) were superovulated by injection of 2.5 IU pregnant mare serum gonadotropin (PMSG, sigma) followed 48 hr later by injection 2.5 IU of human chronic gonadotropin (hCG, sigma). After 15 hr, the female mice were sacrificed and oocytes were collected from ampulla. The number of sperm was counted with Makler counter chamber (Sefi Medical Instruments LTD, Santa Ana, CA, USA). The caudal sperm collection was followed the method for mouse *in vitro* fertilization (Hogan et al., 1994).

5. Statistical analysis

The results represent means \pm SED. The data were analyzed using one-way analysis of variance (ANOVA) and *t*-test between control and experimental group. In all cases, values of P<0.05 were deemed to indicate statistical significance.

	Pre-mating	Mating	Post-mating		
P♂	2 weeks	2 weeks	6 weeks		
₽₽	2 weeks	2 weeks	Pregnancy	Lactation	
		F ₁ /F ₂	In-utero development	Pre-weaning	Post-weaning

Fig. 1. Scheme to assess the one-generation reproductive toxicity study. 8-week-old mice which were offspring of phytoestrogen free food-fed parents were examined their estrus cycle with vaginal smearing methods for 2 weeks. The normal female mice attend mating.

RESULT

1. Reproductive organ weight

In male mice, the weights of testis and epididymis were measured as mentioned in Materials and Methods. The weight of testis of F_0 male significantly decreased in NP 50 group but not in NP 500 group (Fig. 2). The weight of epididymis significantly increased in NP 50 group but not in NP 500 group (Fig. 2). In F_1 male, which were treated from gamete to until sampling, the weights of testis and epididymis significantly decreased and increased in NP 50 groups, respectively (Fig. 2).

In F₀ female mice, that were administered with NP as de-

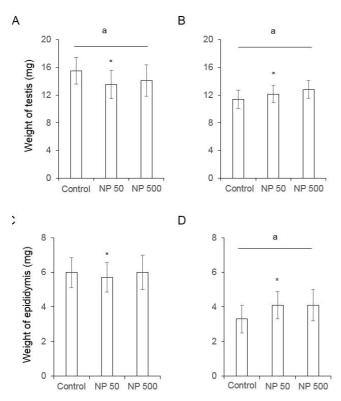


Fig. 2. The effects of nonylphenol (NP) on the weight of testis (A and B) and epididymis (C and D) in F₀ (A, C) and F₁ (B, D). NP was administered as shown Fig. 1. F₀ and F₁ males were sacrificed at 18-20 weeks and 7-8 weeks after birth, respectively. a: P<0.05 (One way ANOVA). *: P<0.05 (t-test, control vs. NPs).

picted in Fig. 1, the ovarian weight significantly increased in NP 50 group but not in NP 500 group. However, the weight of ovary was not changed in F_1 (Fig. 3). The weight of uterus in F_0 female was significantly increased NP 50 and NP 500 groups in concentration-dependent manner. However, in F_1 female, which were treated from gamete to sampling (4-5 weeks old), the weight of uterus was not changed (Fig. 3).

2. Number of caudal sperm and ovulated oocytes

The number of cauda epididymal sperm per body weight was same in controls of F_0 and F_1 males. Also, in NP 50 and NP 500 groups, it was not changed by generation and groups (Fig. 4). The numbers of ovulated oocytes in F_0 were not changed by administration of 50 and 500 μ g/L NP (Fig. 5A). In F_1 generation, the number of ovulated oocytes was significantly decrease in NP 50 group but not

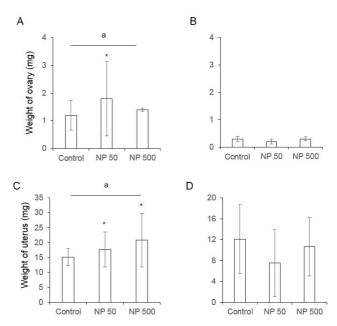


Fig. 3. The effects of nonylphenol (NP) on the weight of ovary (F₀, A; F₁, B) and uterus (F₀, C; F₁, D). NP was administered as shown Fig. 1. F₀ and F₁ females were sacrificed at 18-20 weeks and 4-5 weeks after birth, respectively. a: P<0.05 (One way ANOVA). *: P<0.05 (t-test, control vs. NPs).

in NP 500 group (Fig. 5B). However, the decrease or increase were not detected in F₂ female (data not shown).

3. Effect of nonylphenol on reproductive outcomeFor first 2 weeks we treated low-dose NP, and kept next2 weeks the males and females in the same cage to mate.There was no significant difference in litter size in both

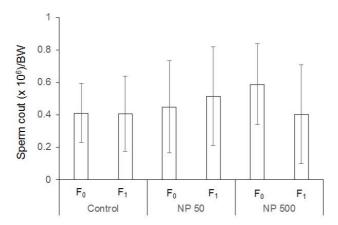


Fig. 4. The ratio of sperm count to body weight in caudal epididymis. The caudal epididymis was isolated and extracted the sperms by squeeze and mincing. The number of sperms were counted with Makler chamber.

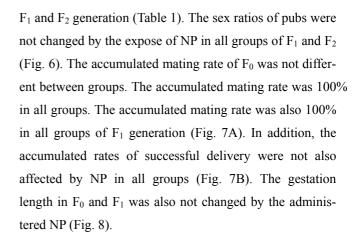


Table 1. Effect of nonylphenol on litter size over onegeneration

Generation	Group	No. female	No. delivery	Litter size
	Control	20	20	11.5±1.84
F_1	NP 50	20	20	12.0±3.29
	NP 500	20	20	10.8±5.34
	Control	35	35	12.0±6.75
F_2	NP 50	40	40	11.8±6.18
	NP 500	35	35	12.0±3.85

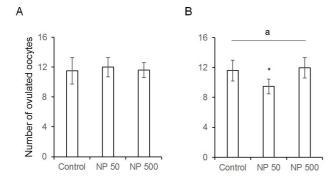


Fig. 5. The effects of nonylphenol (NP) on the ovulation of F₀ (A), and F₁ (B). Nonlyphnol (NP) was administered as shown Fig. 1 and ovulation was induced with gonadotrophins. The ovulated oocytes were collected from ampulla of oviduct at 15 hr after hCG injection. a: P<0.05 (One way ANOVA). *: P<0.05 (t-test, control vs. NPs).

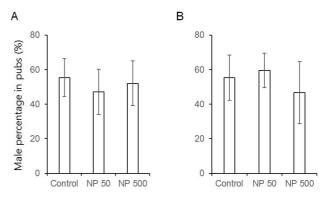


Fig. 6. The sex ratios in F₁ (A) and F₂ (B) pubs. The F₀ and F₁ parents were administered with nonylphenol (NP) containing water as depicted in Fig. 1. 10 weeks old mice were attended mating. NPs did not effect on sex ratios in both F₁ and F₂.

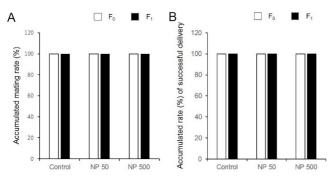


Fig. 7. The effects of nonylphenol (NP) on mating rate (A) and successful delivery rate (B). Nonlyphnol (NP) was administered according to the phytoestrogen free generation F₀ according to the OECD Test Guide line 443. Accumulated mating rates and the accumulated rate of successful delivery were not affected by NP administration in both F₀ and F₁. All pregnant mice delivered successfully.

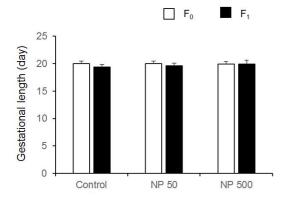


Fig. 8. The effects of nonylphenol (NP) on gestational length in F₀ and F₁. NP was administered as shown Fig. 1. Copulation plugs were checked every day morning. To keep the strict time for mating, the female which did not have copulation plug, were separated at morning and caging again with male at 5:00 pm. There was no difference in gestational periods.

DISCUSSION

Laboratory studies on animals have demonstrated adverse effects of NP, such as on reproduction, development, neurotoxicity and inflammation (Jie et al., 2013; Yücedağ et al., 2015). The responsibility of organisms are various

by the exposing dose and duration, and the exposing time of life cycle (Patisaul & Adewale, 2009; Solecki et al., 2017). One of the interesting things is that as anothema to the toxicological principle, the dose response of many hormones and EDCs appears to be nonmonotonic (Patisaul & Adewale, 2009). Lee (1998) suggested that NP (8 mg/ kg/ day, ip) administered from pnd 1 to 10 results in the decreased weight of reproductive organs including epididymis. However, Odum and Ashby (2000) does not conform to the results with Alpk (Wistar derived) rats. Such doseindependent results can be found in gene expression patterns. The genes are expressed in a dose-dependent manner at 0.5-50 µg/kg but not in 50 mg/kg NP (Watanabe et al., 2004). Consistences with those reports, our results show that the effects of NPs did not show dose dependent responsibility. The weight of testis and epididymis decreased or increased on only NP 50 group of F₀ and F₁, respectively. In ovary, the weight was increased only at 50 μ g/L NP in F₀.

Estradiol stimulates the proliferation of immature and adult uterine cells and increases the weight (Black and Goode, 1980; Maier et al., 1985). The relative binding affinity of NP to estrogen receptors is 0.035-0.037% and the mean IC₅₀ is $4.00 \times 10^{-7} \pm 0.10 \times 10^{-7} - 4.40 \times 10^{-7} \pm 0.40 \times 10^{-7}$ (Blair et al., 2000). In the uterus, high dose NP (50 mg/kg) has similar effect on most of the genes that are activated by estradiol, and low-dose NP (0.5 and 5 mg/kg) has little effect on the genes that are activated by estradiol (Watanabe et al., 2004). In adults, daily dosing 0.5 and 5 mg/kg NP are equal to 15 and 150 µg NP, respectively. On the other hand, 50 and 500 µg/L NP in daily drinking water (4-7 mL/day/adult) are equivalent to 0.275 and 2.75 µg NP, respectively. Interestingly the uterine weight of parent generation was increased in a concentration-dependent manner. However, in pre-pubertal F₁ female, the uterine weights were not different between groups. This means that lower dose than 15 µg/day can effect on the reproductive organs

in a nonmonotonic manner, although NP is 100,000 times less affinity to ERs than estradiol.

Studies show the dimorphic characteristics in males and females to EDCs and suggested female and male germ cells have different susceptibilities to ECD (Anway et al., 2005; Clifton, 2010). Such dimorphism is also supported by our results, as in male mice, the weights of testis and epididymis were affected only by 50 μ g/L in both F₀ and F₁. However in female, the ovary and uterus were affected only in F₀. Lower-dose NP did not effect on the weight of ovary or uterus in the pre-pubertal female F₁ mice, even though they were exposed from germ cell stage. Parental generation and F₁ males treated with 50 and 500 μ g/L had similar sperm counts compared with control but not in the number of ovulated oocytes. This means that as expected, the susceptibility to EDC effects of NP changed depending on sex-related endocrine regulation.

It has been suggested that exposing to prenatal and early postnatal EDC might be the causal path of male reproductive disorder such as cryptorchidism, hypospadias, low sperm count and testicular cancer (Bøgh et al., 2001; Rissman & Adli, 2014; Bonde et al., 2017). In addition, altering the epigenic programing of the germ line is observed in the male administered EDCs during fetal gonadal sex determination and such altered epigenic characters are inherited to the subsequent generations (Guerrero-Bosagna & Skinner, 2014). Spermatogenesis in adult male rat that are exposed with NP (100 mg/kg) is affected and the number of sperm decreased (De Jager et al., 1999). However, in meta-analysis, a total of 33 papers provide 89 risk estimates on which there is no strong support for a global effect as a whole or on any specific outcome in male reproduction disorders following prenatal and postnatal exposure (Bonde et al., 2017). Interestingly in this study, the responsibility of testis and epididymis were different between NP 50 and NP 500 groups but the number of offspring, sex ratio of pubs, and accumulated mating rate were not changed. This means that the responding patterns to lower-dose NP in reproduction are different by the character of organs.

Kyselova and colleagues (2003) examined with the same doses except the exposed duration (4 week in CD-1 mice (8 wks old)). Based on the results of Kyselova et al (2003), the detrimental effects on the fertility of sperm of F₁ which were exposed to NP according to OECD TG 443 could be suspected. Kyselova et al (2003) evaluated the detrimental effect on acrosomal stability. However, the F₁ generation mice had similar responsibility to the NPs in accumulated mating rates and gestational length compared with the control and NP groups. The litter size and sex ratio in F1 and F2 generation were also not different between groups and generations. However, interestingly we did not find any detrimental effects on male fertility. The number of pubs and their sex ratio in F₀ and F₁ generation were similar between groups and generation. Such difference may be the result of the difference in exposing duration.

The WHO defines the endocrine disruptors as an exogenous substance or mixture that alters the function(s) of the endocrine system and consequently causes adverse effects in an intact organism, or its progeny, or population. The adverse effect refers to "a change in morphology, physiology, growth, reproduction, development or lifespan of an organism which results in impairment of functional capacity or impairment of capacity to compensate for additional stress or increased susceptibility to the harmful effects of other environmental influences" (Solecki et al., 2017). Kyselova et al (2003) suggest that exposing NP (50 and 500 µg/L) to 2 months old male for 4 weeks has multigenerational effect on selective reproductive organ. From our results it is clear that lower-dose NP has characters of EDC in male and female of F₀ and F₁ generation. However, the chronic exposing of lower-dose NP has no detrimental effect on fertility of the subsequent generation.

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