ORIGINAL RESEARCH



Negative effects of nitrogen fertilization on herbivore fitness are exaggerated at warmer temperatures and in high-altitude populations

Ange Raharivololoniaina $^1 \cdot$ Roland Busch $^1 \cdot$ Franziska Deppe $^1 \cdot$ Anna Hitzler $^1 \cdot$ Eva Plath $^1 \cdot$ Tamara Rischen $^1 \cdot$ Mine Yilmazer $^1 \cdot$ Klaus Fischer $^1 \cdot$

Received: 26 October 2024 / Accepted: 28 February 2025 / Published online: 11 March 2025 © The Author(s) 2025

Abstract

Biodiversity is currently under strong pressure due to anthropogenic global change. Different drivers of global change may exert direct and indirect effects on biodiversity, and may furthermore interact with one another, but our respective knowledge is still very limited. We investigated indirect and interactive effects of two important drivers of global change, eutrophication and climate change, in replicated low- and high-altitude populations of an insect herbivore, the butterfly *Lycaena tityrus*, in a laboratory setting. We found local adaptation in developmental traits, with low-altitude populations being adapted to warmer temperatures and longer seasons. *Lycaena tityrus* responded negatively to agriculturally relevant levels of fertilization of its host plant, showing reduced body mass and prolonged development time. Negative effects were particularly pronounced at warmer temperatures and in high-altitude populations. Our study adds to the increasing knowledge that different drivers of global change may interact and thereby increase the overall level of threat to biodiversity. We suggest that populations inhabiting naturally nutrient-poor environments might be even more vulnerable to agricultural intensification than others, potentially applicable to many species. These findings may have important implications for protecting numerous vulnerable species in the face of rapid environmental change.

 $\textbf{Keywords} \ \ \text{Biodiversity conservation} \cdot \text{Butterfly} \cdot \text{Eutrophication} \cdot \text{Global change} \cdot \text{Plant-herbivore interaction}$

Introduction

In the current era, biodiversity is strongly threatened by global environmental change, which involves a variety of human-induced drivers (Pereira et al. 2012; Ceballos et al. 2015; Johnson et al. 2017; Wagner 2020). These drivers may directly or indirectly affect biodiversity, and may additionally interact with one another (Oliver and Morecroft 2014; Jackson et al. 2016; Gunderson et al. 2017; Hamann et al. 2021). While the direct effects of individual drivers have been quite intensively studied, our knowledge on indirect effects and interactions among drivers is still limited (Oliver

Communicated by Sebastian Seibold.

and Morecroft 2014; Sage 2020; Wagner 2020; Kuczyk et al. 2021a).

Important drivers of biodiversity loss are agricultural intensification and anthropogenic climate change (Pereira et al. 2012; Ceballos et al. 2015; WallisDeVries and Van Swaay 2017). With regard to agricultural intensification, adverse effects of eutrophication on species richness have been widely documented for plants (Bobbink et al. 2010; Stevens et al. 2010), but also for invertebrates (Van Dyck et al. 2009; Carvalheiro et al. 2019; Nessel et al. 2021). Negative effects of excess nitrogen on animals might be conveyed by microclimatic cooling, reduced host-plant quality, and the loss of host plants or nectar sources (Fischer and Fiedler 2000; WallisDeVries and Van Swaay 2006; Lebeau et al. 2016; Nijssen et al. 2017). However, our knowledge on the effects of nitrogen fertilization on animals is still limited (Nijssen et al. 2017). With regard to anthropogenic climate change, its effects on biodiversity have been widely demonstrated and discussed (Weiskopf et al. 2020; Pereira et al. 2024). Climate change may induce shifts in phenology or geographic ranges, and may also affect morphology,



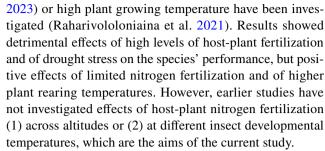
Department of Biology, Institute for Integrated Natural Sciences, University of Koblenz, Universitätsstraße 1, 56070 Koblenz, Germany

life history and behavior (Scheffers et al. 2016; Weiskopf et al. 2020). Both, eutrophication and climate change, likely play a significant role in biodiversity loss, specifically in the decline of insect diversity and abundance (Nessel et al. 2021; Raven and Wagner 2021).

Herbivores comprise particularly interesting organisms to study the consequences of global change, as they are sensitive to direct as well as indirect effects, the latter being mediated by alterations in host-plant quality or availability (Harvey et al. 2020; Kuczyk et al. 2021a). For instance, increased temperatures during plant growth may diminish host-plant quality for insect herbivores (Bauerfeind and Fischer 2013; Kuczyk et al. 2021a, b), which may result from increased C:N ratios, concentrations of secondary metabolites, and leaf toughness (Meihls et al. 2012; Block et al. 2019). Similarly, nitrogen fertilization, as abundantly applied in modern agriculture, affects host-plant chemistry (Kuczyk et al. 2021a, b). This may in turn diminish herbivore performance, e.g. by reducing survival rates, growth rates and body mass (Fischer and Fiedler 2000; Kurze et al. 2018; Raharivololoniaina et al. 2023).

In species with a wide geographical distribution, populations may respond differently to environmental stressors (Kuczyk et al. 2021b). Such within-species variation in the vulnerability to global change has been largely neglected thus far, although geographical variation in trait values is common (Kawecki and Ebert 2004; Blanquart et al. 2013). For instance, in the widespread butterfly *Pieris napi*, fastgrowing Italian populations are more sensitive to poor food quality than slower growing German ones (Kuczyk et al. 2021b). Such studies on local adaptation provide important insights into the power of natural selection relative to gene flow and other evolutionary forces (Kawecki and Ebert 2004), but add an additional level of complexity.

Here, we study indirect and interactive effects of two important global change drivers, namely eutrophication and a vital component of climate change, viz. increasing temperature. Specifically, we investigate effects of host-plant nitrogen fertilization and herbivore developmental temperature, on high- and low-altitude populations of the butterfly Lycaena tityrus (Poda 1761). This species has been repeatedly used as a model to study phenotypic plasticity, local and thermal adaptation as well as dispersal (e.g., Karl et al. 2008, 2009a, b; Fischer and Karl 2010; Fischer et al. 2011; Reim et al. 2018a, b). These studies revealed, amongst others, that high- compared to low-altitude populations showed longer development times, increased cold- but decreased heat-stress resistance, reduced plasticity in heat resistance, temperature-induced plasticity in developmental traits, and a pronounced sexual dimorphism. In terms of host-plant quality, effects of nitrogen fertilization of host plants (Fischer and Fiedler 2000; Kurze et al. 2018), and nitrogen fertilization in combination with drought stress (Raharivololoniaina et al.



We predict that (1) cold-adapted high-altitude populations of *L. tityrus*, which are not (yet) exposed to high levels of eutrophication, are more sensitive to host plant fertilization than low-altitude populations, and that (2) detrimental effects of fertilization are more pronounced at higher ambient temperatures, indicating an interactive effect between two important drivers of biodiversity loss, eutrophication and climate change.

Materials and methods

Study organism

Lycaena tityrus (Poda 1761) is a widespread, temperatezone butterfly ranging from Spain throughout large parts
of central and southern Europe to western Asia (Ebert and
Rennwald 1991; Bräu et al. 2013). In most parts of its range,
it has 2–3 generations per year (Ebert and Rennwald 1991;
Tolman and Lewington 2008). Two subspecies are recognized in central Europe, the widespread nominate form *L.*tityrus tityrus (Poda 1761) and the alpine subspecies *L.*tityrus subalpinus (Speyer 1851). The latter is confined to
the higher altitudes of the European Alps and some other
mountain ranges, where it has only one generation per year
(Tolman and Lewington 2008). The altitudinal distribution
of *L.* tityrus subalpinus ranges from 1200 to 2500 m a.s.l.
(Tolman and Lewington 2008).

Lycaena tityrus inhabits humid to moderately dry locations such as low-intensity grassland, alpine meadows, sandy heathland, bogs, and forest edges (Ebert and Rennwald 1991; Bräu et al. 2013). It uses different species of sorrel (Rumex spec., Polygonaceae) as larval host plant (Ebert and Rennwald 1991; Tolman and Lewington 2008). The primary hostplant is Rumex acetosa L., which is a perennial herb containing a variety of primarily non-nitrogenous secondary compounds, especially oxalic acids (Korpelainen and Pietiläinen 2020). Adults are rather opportunistic nectar feeders, using a wide array of plants (Ebert and Rennwald 1991). Lycaena tityrus has suffered substantial population losses due to grassland intensification, meadow conversion, and afforestation in recent decades (Bräu et al. 2013). It is projected to shift its distributional range northward and upward due to anthropogenic climate change (Settele et al. 2008).



Oecologia (2025) 207:51 Page 3 of 13 51

Population sampling

For this study, we collected in 2022 fecund females from four replicated low-altitude populations in northeastern and western Germany (L. tityrus tityrus; Greifswald, coordinates: 54.05°N, 13.44°E; Altwarp: 53.74°N, 14.26°E; Pottum: 50.59°N, 8.00°E; Westerburg: 50.54°N, 7.97°E; altitudinal range 0-420 m a.s.l.), and from four replicated high-altitude populations in Austria (*L. tityrus subalpinus*; Kühtai: 47.23°N, 10.97°E; Obernberg: 46.98°N, 11.42°E; Obergurgl: 46.86°N. 11.02°E; Venn: 47.01°N. 11.54°E; altitudinal range 1560–2030 m a.s.l.). All high-altitude populations were sampled on unfertilized alpine meadows in the central Alps, i.e., above crystalline primary rock (gneiss, slate) resulting in acidic, nutrient-poor soils. The low-altitude populations, in contrast, originate from slightly fertilized meadows from areas that are also naturally richer in nutrients (basalt mountains or marshland). Note that sampling locations not only differed in altitude but also in latitude. However, the impact of variation in altitude is expected to surpass the potential impact of latitude by far (Karl et al. 2008, 2009a). Per population, we sampled between 6 and 20 females, in total 34 from low- and 48 from high-altitude populations. Females were afterwards transferred to and randomly distributed among two climate cabinets (26 °C, 70% humidity, L19:D5 photoperiod; Panasonic MIR-554) at Koblenz University for egg laying. Therefore, females were kept individually in translucent plastic pots (1 L) covered with gauze. They were provided sucrose solution and fresh flowers for adult feeding, and leaves of Rumex acetosa as oviposition substrate. Deposited eggs were collected daily and transferred, separated by female, to small glass vials kept in one climate cabinet at 22 °C (70% humidity, L19:D5 throughout) until hatching (Panasonic MIR-554). Hatchlings were fed on field-collected R. acetosa leaves under the same conditions for 4 days, after which they were allocated to treatments.

Experimental design

On day 5 after hatching, larvae were individually transferred to transparent plastic boxes (250 ml) lined with moist tissue and containing a fresh cutting of *R. acetosa*. Larvae were randomly allocated to one out of four treatments, using a split-brood design. We employed a full-factorial design with two rearing temperatures and two host plant treatments. Rearing temperatures had a mean of 19.2 °C (light phase 20 °C, dark phase 16 °C; L19:D5 photoperiod) or 25.2 °C (light phase 26 °C, dark phase 22 °C), and we used one climate cabinet per temperature (Panasonic MIR-554). To rule out effects of individual cabinets, temperature regimes and according larvae were shifted among both cabinets every four days. Host plant treatments involved low versus

high levels of nitrogen fertilization, the latter mimicking levels typically used in modern agriculture (Raharivololoniaina et al. 2021, 2023). Host plants, 200 per treatment, were grown individually in standard pots $(9 \times 9 \times 10 \text{ cm})$ 0.5 L) in a greenhouse from commercially available seeds. Plants were provided with a common nitrogenous fertilizer (NH₄NO₃), and received a total amount of 1.225 ml (low N) or 3.675 ml (high N) of NH₄NO₃. The total amount of NH₄NO₃ was equally divided into five applications of 0.245 ml (low N) or 0.735 ml (high N) in 100 ml water each, corresponding to one application per week during the larval feeding phase. These nitrogen treatments have been shown to be highly effective in changing plant chemistry; specifically, higher nitrogen levels resulted in higher nitrogen and water contents but a reduced C:N ratio (Raharivololoniaina et al. 2021, 2023). All boxes were cleaned daily and supplied with a fresh leaf cutting and moist tissue. Leaf cuttings were taken from at least 20 plants per day and randomly allocated to the larvae within treatments. Boxes were shifted around daily to minimize potential temperature and light differences within the climate cabinets. All offspring (n = 579) were reared at the treatment they were allocated to until adult eclosion. Afterwards, the sex of each individual was determined and the animals frozen at -18 °C for later analyses.

Data collection and laboratory analyses

For each individual, we scored different developmental and morphological traits as well as mortality. Larval and pupal development time were measured in days. Pupal mass was weighed on the day following pupation to the nearest 0.1 mg (KERN ABJ-NM/ABS-N). Larval growth rate was calculated as mean weight gain per day (LN (pupal mass)/larval time). Frozen butterflies were first weighed to obtain total adult mass. Then, wings, head and legs were removed, and the abdomen was separated from the thorax. Abdomen and thorax were weighed separately, and thorax-abdomen ratio was calculated as a measure of the relative investment into flight versus reproduction. The length and area of the right forewings were measured using digital images and the software ImageJ. Wing loading was calculated as adult body mass divided by forewing area, and wing aspect ratio as 4×forewing length² divided by forewing area (Berwaerts et al. 2002). Finally, abdomen fat content was quantified, but due to the high work load only in a subset of individuals, following Fischer et al. (2003). In short, abdomens were dried at 60 °C for 24 h and then weighed. Thereafter, fat was extracted from the dried abdomens for 48 h on a laboratory shaker using 1.5 mL of dichloromethane (CH₂CL₂) for each abdomen. The extraction was repeated once, after which the abdomens were once again dried at 60 °C for 24 h and afterwards weighed. Abdomen fat content was then calculated by



51 Page 4 of 13 Oecologia (2025) 207:51

subtracting the fat-free abdomen dry mass from the initial abdomen dry mass and is given in percent.

Statistical analyses

We used general linear mixed models (GLMMs) to analyze all traits obtained (except survival, see below), with origin (low versus high altitude), sex, developmental temperature, and host-plant nitrogen treatment as fixed factors, and replicate population (nested within origin) and family (the offspring of a given female; nested within origin and population) as random factors. Minimum adequate models were constructed by sequentially removing non-significant interaction terms, in case no higher order interactions were significant. Larval time, pupal time, and larval growth rate were LN-transformed prior to analyses to meet GLMM requirements. Survival data were analyzed using a binary logistic regression with the same factors and model structure as above. All tests were computed using Statistica 12.0 (Statsoft, Tulsa, USA) or R 4.4.2 (R Core Team 2024) using the 'lme4' package (Bates et al. 2015). Throughout the text, we present means \pm standard error.

Results

Low- and high-altitude populations differed significantly in larval time, pupal time, larval growth rate, wing length, wing area, wing loading, wing aspect ratio, and abdomen fat content (Table 1). High- compared with low-altitude populations showed longer larval times $(33.4\pm0.4 \text{ vs. } 28.4\pm0.4 \text{ d})$, shorter pupal times $(10.5\pm0.1 \text{ vs. } 11.5\pm0.1 \text{ d})$, lower larval growth rates $(4.30\pm0.10 \text{ vs. } 5.04\pm0.08 \text{ mg/d})$, shorter forewings $(14.44\pm0.06 \text{ vs. } 15.26\pm0.05 \text{ mm})$, smaller forewing areas $(88.0\pm0.8 \text{ vs. } 96.5\pm0.6 \text{ mm}^2)$, higher wing loadings $(0.369\pm0.006 \text{ vs. } 0.350\pm0.005 \text{ mg/mm}^2)$, and lower wing aspect ratios $(9.59\pm0.06 \text{ vs. } 9.72\pm0.04)$ and abdomen fat contents $(9.6\pm0.9 \text{ vs. } 17.3\pm0.7\%)$. In addition, thorax masses tended to be lower in high- versus low-altitude populations $(14.7\pm0.3 \text{ vs. } 15.5\pm0.2 \text{ mg})$.

The above general patterns were modulated by interactions with other factors (Table 1). For six traits, the interaction between origin and rearing temperature was significant. These interactions show for larval time and larval growth rate that differences among altitudes were restricted to or more pronounced at the lower temperature (Fig. 1). Adult, thorax and abdomen mass were higher in high-altitude populations at the lower rearing temperature, but lower at the higher rearing temperature. Accordingly, wing loading was higher in high- than in low-altitude populations at the lower temperature, but the other way around at the higher temperature (Fig. 1f). The significant origin by nitrogen interactions for larval time and larval growth rate show that

the high-altitude populations responded much more strongly to the high nitrogen treatment, prolonging larval time and reducing larval growth rates (Fig. 2).

All measured traits differed significantly among sexes except from fat content, for which there was only a tendency towards a lower fat content in females (12.5 ± 0.8) vs. $14.4 \pm 0.7\%$). Females compared with males had longer larval $(32.2 \pm 0.4 \text{ vs. } 29.6 \pm 0.4 \text{ d})$ and pupal times $(11.2 \pm 0.1 \text{ vs. } 10.7 \pm 0.1 \text{ d})$, higher pupal masses $(126.5 \pm 1.3 \text{ vs. } 118.2 \pm 1.2 \text{ mg})$, lower larval growth rates $(4.57 \pm 0.10 \text{ vs. } 4.77 \pm 0.10 \text{ mg/d})$, higher adult $(38.2 \pm 0.6 \pm 0.00 \text{ mg/d})$ vs. 28.5 ± 0.5 mg), thorax $(15.5 \pm 0.3 \text{ vs. } 14.7 \pm 0.3 \text{ mg})$ and abdomen masses (15.6 ± 0.3 vs. 7.9 ± 0.2 mg), lower thoraxabdomen ratios $(1.04 \pm 0.03 \text{ vs. } 2.00 \pm 0.03)$, longer wings $(15.0 \pm 0.1 \text{ vs. } 14.8 \pm 0.1 \text{ mm})$, larger wing areas $(96.6 \pm 0.7 \text{ mm})$ vs. $87.9 \pm 0.7 \text{ mm}^2$), higher wing loadings (0.395 ± 0.005) vs. 0.323 ± 0.005 mg/mm²), and lower wing aspect ratios $(9.33 \pm 0.05 \text{ vs. } 9.99 \pm 0.05)$. The interaction between origin and sex was significant for thorax mass, wing length, and wing loading, showing that sex differences were restricted to or more pronounced in high-altitude (thorax mass, wing loading) or low-altitude populations (wing length; Table S1b, supplementary material).

All traits except from pupal mass and abdomen fat content differed among rearing temperatures. At the lower rearing temperature, larval $(41.9 \pm 0.5 \text{ vs. } 20.0 \pm 0.3 \text{ d})$ and pupal times $(13.4 \pm 0.1 \text{ vs. } 8.5 \pm 0.1 \text{ d})$ were longer, larval growth rate $(3.06 \pm 0.11 \text{ vs. } 6.28 \pm 0.06 \text{ mg/d})$ was lower, adult $(37.5 \pm 0.7 \text{ vs. } 29.2 \pm 0.4 \text{ mg})$, thorax $(16.7 \pm 0.3 \text{ mg})$ vs. 13.5 ± 0.2 mg) and abdomen masses (13.6 ± 0.3) vs. 9.9 ± 0.2 mg) were higher, thorax-abdomen ratio $(1.42 \pm 0.04 \text{ vs. } 1.59 \pm 0.02)$ was lower, wings (15.0 ± 0.1) vs. 14.7 ± 0.1 mm) were longer, wing area $(93.3 \pm 0.8 \text{ vs.})$ $91.1 \pm 0.5 \text{ mm}^2$) was larger, and wing loading (0.402 ± 0.006) vs. 0.316 ± 0.004 mg/mm²) and wing aspect ratio $(9.77 \pm 0.06 \text{ vs. } 9.54 \pm 0.04)$ were higher. The reductions in abdomen mass and concomitantly wing loading at the higher temperature were more pronounced in females than in males. However, the reductions in wing length and wing aspect ratio at the higher temperature were more pronounced in males than in females (significant sex by temperature interactions; Table S1a, supplementary material). The reduction in thorax mass, wing length and area at the higher temperature were more pronounced in the high nitrogen treatment (significant temperature by nitrogen treatment interactions; Fig. 3). For the latter two traits, this effect was especially pronounced in high altitude populations (significant threeway interactions; Table S2, supplementary material).

Host-plant nitrogen treatment significantly affected larval time, adult mass, abdomen mass, wing loading, and wing aspect ratio. At the high as compared with the low nitrogen treatment larval time $(32.0 \pm 0.4 \text{ vs. } 29.9 \pm 0.4 \text{ d})$ was longer, adult $(33.1 \pm 0.6 \text{ vs. } 33.6 \pm 0.5 \text{ mg})$ and abdomen masses



Table 1 Results of general linear mixed models for the effects of origin (low vs. high altitude), sex, rearing temperature, host-plant nitrogen fertilization (low vs. high; all fixed), replicate population (random, nested within origin), and family (random, nested within population and origin) on various traits of the butterfly *Lycaena tityrus*

Source	Effect	DFn	DFd	MS	F	p
Larval time						
Origin	Fixed	1	8	0.59	5.63	0.0470
Sex	Fixed	1	534	1.28	30.72	< 0.0001
Temperature	Fixed	1	534	50.03	1205.15	< 0.0001
Nitrogen	Fixed	1	534	0.30	7.15	0.0077
Origin*Temp	Fixed	1	534	1.04	25.00	< 0.0001
Origin*Nitro	Fixed	1	534	0.34	8.18	0.0044
Population (Origin)	Random	6	78	0.13	2.65	0.0217
Family (Origin*Pop)	Random	31	534	0.05	1.28	0.1472
Error		534		0.04		
Pupal time						
Origin	Fixed	1	7	0.39	24.68	0.0015
Sex	Fixed	1	525	0.39	45.01	< 0.0001
Temperature	Fixed	1	525	20.63	2402.60	< 0.0001
Nitrogen	Fixed	1	525	< 0.01	0.06	0.8133
Population (Origin)	Random	6	64	0.02	1.33	0.2563
Family (Origin*Pop)	Random	31	525	0.02	1.82	0.0049
Error		525		0.01		
Pupal mass						
Origin	Fixed	1	8	1175.78	1.88	0.2066
Sex	Fixed	1	529	7075.64	24.21	< 0.0001
Temperature	Fixed	1	529	694.77	2.38	0.1237
Nitrogen	Fixed	1	529	367.53	1.26	0.2626
Origin*Sex	Fixed	1	529	21.96	0.08	0.7841
Origin*Temp	Fixed	1	529	1051.88	3.60	0.0584
Sex*Temp	Fixed	1	529	1.29	< 0.01	0.9470
Origin*Nitro	Fixed	1	529	337.25	1.15	0.2832
Sex*Nitro	Fixed	1	529	510.61	1.75	0.1868
Temp*Nitro	Fixed	1	529	819.63	2.80	0.0946
Origin*Temp*Nitro	Fixed	1	529	1297.52	4.44	0.0356
Population (Origin)	Random	6	54	732.40	1.40	0.2303
Family (Origin*Pop)	Random	31	529	698.63	2.39	0.0001
Error		529		292.28	,	*****
Larval growth rate		02)		2>2.20		
Origin	Fixed	1	8	1.19	11.60	0.0094
Sex	Fixed	1	531	0.24	4.87	0.0277
Temperature	Fixed	1	531	49.76	1015.84	< 0.0001
Nitrogen	Fixed	1	531	0.18	3.67	0.0560
Origin*Temp	Fixed	1	531	0.59	12.14	0.0005
Origin*Nitro	Fixed	1	531	0.29	5.85	0.0159
Population (Origin)	Random	6	81	0.12	2.22	0.0494
Family (Origin*Pop)	Random	31	531	0.06	1.20	0.2111
Error		531		0.05	-1	******
Adult mass		001		0.00		
Origin	Fixed	1	15	19.93	0.43	0.5237
Sex	Fixed	1	532	10,338.96	187.37	< 0.0001
Temperature	Fixed	1	532	6496.44	117.73	< 0.0001
Nitrogen	Fixed	1	532	297.57	5.39	0.0206
Origin*Temp	Fixed	1	532	5645.05	102.30	< 0.0001
Population (Origin)	Random	6	50	39.72	0.34	0.9114
Family (Origin*Pop)	Random	31	532	165.61	3.00	< 0.0001
- anniy (Origin 1 op)	Randoni	J1		105.01	3.00	~ 0.0001



 Table 1 (continued)

Source	Effect	DFn	DFd	MS	F	p
Error		532		55.18		
Thorax mass						
Origin	Fixed	1	15	48.44	4.54	0.0504
Sex	Fixed	1	533	83.11	6.27	0.0125
Temperature	Fixed	1	533	1079.83	81.53	0.0000
Nitrogen	Fixed	1	533	18.84	1.42	0.2336
Origin*Sex	Fixed	1	533	87.40	6.60	0.0105
Origin*Temp	Fixed	1	533	960.27	72.50	< 0.0001
Temp*Nitro	Fixed	1	533	54.66	4.13	0.0427
Population (Origin)	Random	6	54	9.20	0.40	0.8776
Family (Origin*Pop)	Random	31	533	30.74	2.32	0.0001
Error		533		13.25		
Abdomen mass						
Origin	Fixed	1	9	2.44	0.13	0.7318
Sex	Fixed	1	535	6287.01	601.96	< 0.0001
Temperature	Fixed	1	535	1336.71	127.99	< 0.0001
Nitrogen	Fixed	1	535	47.68	4.56	0.0331
Origin*Temp	Fixed	1	535	560.00	53.62	< 0.0001
Sex*Temp	Fixed	1	535	110.82	10.61	0.0012
Population (Origin)	Random	6	49	22.12	1.00	0.4343
Family (Origin*Pop)	Random	31	535	30.97	2.97	< 0.0001
Error	Kandom	535	333	10.44	2.97	< 0.0001
		333		10.44		
Thorax-abdomen ratio	F' 1	1	7	0.44	0.71	0.4200
Origin	Fixed	1	7	0.44	0.71	0.4290
Sex	Fixed	1	536	136.19	727.77	< 0.0001
Temperature	Fixed	1	536	2.39	12.77	0.0004
Nitrogen	Fixed	1	536	0.18	0.95	0.3296
Population (Origin)	Random	6	63	0.68	2.52	0.0298
Family (Origin*Pop)	Random	31	536	0.33	1.77	0.0069
Error		536		0.19		
Wing length	-		4.0	20.55	26.76	0.0004
Origin	Fixed	1	10	28.75	36.56	0.0001
Sex	Fixed	1	526	2.69	4.70	0.0306
Temperature	Fixed	1	526	8.16	14.26	0.0002
Nitrogen	Fixed	1	526	< 0.01	< 0.01	0.9667
Origin*Sex	Fixed	1	526	4.63	8.09	0.0046
Origin*Temp	Fixed	1	526	0.48	0.85	0.3580
Sex*Temp	Fixed	1	526	2.29	4.00	0.0460
Origin*Nitro	Fixed	1	526	1.05	1.84	0.1755
Sex*Nitro	Fixed	1	526	0.22	0.39	0.5344
Temp*Nitro	Fixed	1	526	3.58	6.25	0.0127
Origin*Temp*Nitro	Fixed	1	526	3.16	5.52	0.0192
Population (Origin)	Random	6	52	0.84	0.76	0.6010
Family (Origin*Pop)	Random	31	526	1.50	2.62	< 0.0001
Error		526		0.57		
Wing area						
Origin	Fixed	1	9	2724.44	21.46	0.0011
Sex	Fixed	1	529	10,610.74	125.16	< 0.0001
Temperature	Fixed	1	529	393.02	4.64	0.0318
Nitrogen	Fixed	1	529	261.25	3.08	0.0798
Temp*Nitro	Fixed	1	529	872.54	10.29	0.0014



Oecologia (2025) 207:51 Page 7 of 13

Table 1 (continued)

Source	Effect	DFn	DFd	MS	F	p
Origin*Temp*Nitro	Fixed	3	529	246.84	2.91	0.0340
Population (Origin)	Random	6	55	138.86	0.95	0.4643
Family (Origin*Pop)	Random	31	529	191.55	2.26	0.0002
Error		529		84.78		
Wing loading						
Origin	Fixed	1	28	0.02	8.96	0.0057
Sex	Fixed	1	527	0.55	115.06	< 0.0001
Temperature	Fixed	1	527	0.74	155.66	< 0.0001
Nitrogen	Fixed	1	527	0.04	7.87	0.0052
Origin*Sex	Fixed	1	527	0.07	15.31	0.0001
Origin*Temp	Fixed	1	527	0.73	153.70	< 0.0001
Sex*Temp	Fixed	1	527	0.03	5.85	0.0159
Population (Origin)	Random	6	52	< 0.01	0.18	0.9805
Family (Origin*Pop)	Random	31	527	0.01	2.66	< 0.0001
Error		527		< 0.01		
Wing aspect ratio						
Origin	Fixed	1	13	2.33	10.70	0.0063
Sex	Fixed	1	532	48.57	99.97	< 0.0001
Temperature	Fixed	1	532	4.08	8.39	0.0039
Nitrogen	Fixed	1	532	2.81	5.79	0.0165
Sex*Temp	Fixed	1	532	2.19	4.50	0.0344
Population (Origin)	Random	6	71	0.17	0.28	0.9428
Family (Origin*Pop)	Random	31	532	0.69	1.43	0.0643
Error		532		0.49		
Fat content						
Origin	Fixed	1	7	1735.72	11.44	0.0110
Sex	Fixed	1	315	311.61	3.69	0.0556
Temperature	Fixed	1	315	119.53	1.42	0.2349
Nitrogen	Fixed	1	315	141.32	1.67	0.1966
Population (Origin)	Random	6	105	166.85	2.23	0.0456
Family (Origin*Pop)	Random	27	315	66.58	0.79	0.7662
Error		315		84.39		

Minimum adequate models were constructed by sequentially removing non-significant interaction terms, as far as no higher order interactions were significant. Significant *p*-values are given in bold

 $(11.6\pm0.3~{\rm vs.}~12.0\pm0.2~{\rm mg})$ were lower, and wing loading $(0.351\pm0.006~{\rm vs.}~0.368\pm0.005~{\rm mg/mm^2})$ and wing aspect ratio $(9.54\pm0.05~{\rm vs.}~9.76\pm0.05)$ were lower. Additionally, larval growth rate tended to be lower in the high nitrogen treatment $(4.62\pm0.10~{\rm vs.}~4.72\pm0.08~{\rm mg/day})$. The significant origin by temperature by nitrogen interaction indicates that nitrogen fertilization had generally little impact on pupal mass, except from the high-altitude populations at the lower rearing temperature, where high nitrogen resulted in heavier pupae (Table S2, supplementary material).

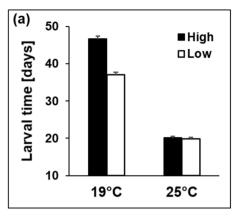
Significant variation between replicate populations were found in larval time, larval growth rate, thoraxabdomen ratio, and fat content, and between families in pupal time, pupal mass, adult mass, thorax mass, abdomen mass, thorax-abdomen ratio, wing length, wing area, and wing loading (Table 1). This suggests genetic variation in development time, body size, storage reserves, and the allocation to dispersal versus reproduction among populations and / or families.

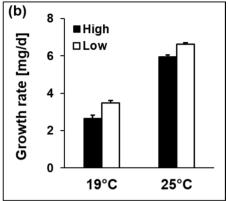
Survival rates, finally, only differed significantly among rearing temperatures (19 °C: 55.1%, 25 °C: 85.2%; z=-4.55, p<0.0001), but not in relation to origin (low altitude: 84.0%, high altitude: 72.2%; z=-1.37, p=0.1715) or nitrogen treatment (low N: 84.2%, high N: 71.1%; z=-1.06, p=0.2884). All interactions were nonsignificant (all p-values > 0.14).

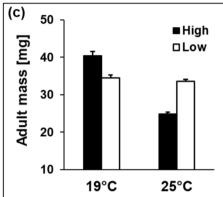


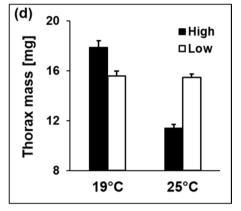
51 Page 8 of 13 Oecologia (2025) 207:51

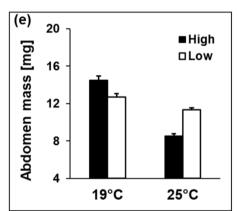
Fig. 1 Larval time (a), larval growth rate (b), adult mass (c), thorax mass (d), abdomen mass (e), and wing loading (f) in relation to origin (high vs. low altitude) and rearing temperature (19 vs. 25 °C) in the butterfly *Lycaena tityrus*. Given are means ± SE











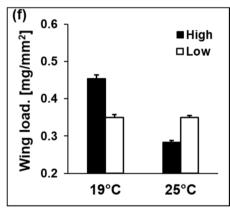
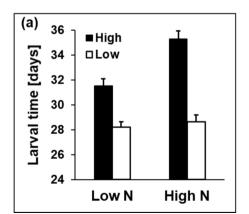
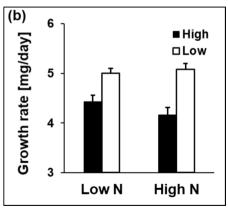


Fig. 2 Larval time (a) and larval growth rate (b) in relation to origin (high vs. low altitude) and host-plant nitrogen fertilization (low N vs. high N) in the butterfly $Lycaena\ tityrus$. Given are means \pm SE

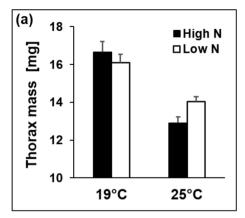


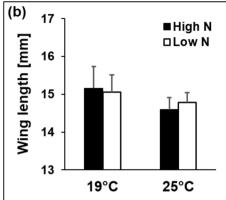


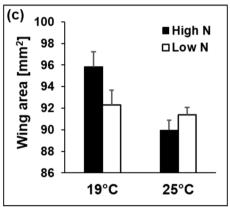


Oecologia (2025) 207:51 Page 9 of 13 51

Fig. 3 Thorax mass (a), wing length (b), and wing area (c) in relation to temperature (19 vs. 25 °C) and host-plant nitrogen treatment (low N vs. high N) in the butterfly *Lycaena tityrus*. Given are means ± SE







Discussion

Our study revealed evidence for local genetic adaptation in high- and low-altitude populations of the butterfly Lycaena tityrus. Specifically, low-altitude populations showed an overall shorter development time associated with higher larval growth rates (cf. Karl et al. 2008), presumably as an adaptation to having 2-3 generations per year in contrast to the high-altitude populations which are univoltine (Tolman and Lewington 2008). Thus, the time constraints imposed by having to fit in additional generations a year seem to be much more severe than those imposed by the shorter vegetation period in alpine environments (Roff 1980; Nylin and Gotthard 1998; Burke et al. 2005; Karl et al. 2008). Interestingly, Estonian populations of L. tityrus also show shorter development times (and smaller body size) compared to German populations, presumably as an adaptation to the short season length in Estonia (Reim et al. 2018b). However, Estonian populations have two generations a year, supporting the above conclusions (Reim et al. 2018b). Interestingly, the differences in larval development time were restricted to the lower rearing temperature, at which also the difference in larval growth rate was more pronounced. This may indicate that the lowaltitude populations speed up their development as much as possible if ambient conditions are challenging for quick development (Nylin and Gotthard 1998), a pressure univoltine high-altitude populations do not face.

Further evidence for local adaptation stems from the fact that high-altitude populations gained a higher body mass than the low-altitude ones at the lower rearing temperature, but vice versa at the higher rearing temperature. Thus, animals were most efficient under the thermal conditions they are presumably better adapted to, probably reflecting metabolic cold adaptation and associated costs at warmer temperatures in high-altitude populations (see Addo-Bediako et al. 2002; Terblanche et al. 2009).

High-altitude populations seem to invest less into dispersal, having smaller wings (despite a similar body mass) and concomitantly higher wing loadings, lower wing aspect ratios and relative abdomen fat contents, and tending to have lower thorax masses as compared with low-altitude populations. Typically, large thoraxes and wings as well as low wing loadings and aspect ratios indicate a high dispersal ability and an aerodynamically- and cost-efficient flight (Berwaerts et al. 2002; Almbro and Kullberg 2012; Lion et al. 2023), while abdominal fat is used to fuel dispersal (Arrese and Soulages 2010; Toprak et al. 2020). However, in general insects seem to have increased rather than decreased wing sizes at higher altitudes (Dillon et al. 2006). The opposite pattern revealed in our study, i.e., smaller wings and a lower investment into dispersal in high-altitude populations,



might result from inhabiting relatively continuous habitats along alpine streams (Trense et al. 2021), and/or from a lower dispersal propensity due to an exposure to stronger winds at higher altitudes (Hodkinson 2005).

In general, insect herbivores are often nitrogen limited, typically referred to as the nitrogen limitation hypothesis (White 1993). However, in our study, high levels of host plant fertilization as used in modern agriculture, thus exceeding natural nutrient levels, negatively affected fitness components in L. tityrus, reducing growth rates, body mass and concomitantly wing loading, and prolonging larval time (cf. Fischer and Fiedler 2000; Kurze et al. 2018; Raharivololoniaina et al. 2023). Note that low levels of host plant nitrogen fertilization often increases insect fitness in line with nitrogen limitation hypothesis (White 1993; Kurze et al. 2017; Raharivololoniaina et al. 2021), while high levels may exert negative effects. Thus, herbivore performance in response to host plant nitrogen levels likely follows a humpshaped curve (Han et al. 2014; Tao et al. 2014; Lebigre et al. 2018). Such negative effects may even include increased mortality (Fischer and Fiedler 2000), though here we could only find sublethal effects.

Consequently, L. tityrus seems highly vulnerable to agricultural intensification not only due to more frequent grassland cutting or more intensive grazing, but also due to increasing nutrient loads. Possible mechanisms behind negative effects of increased nitrogen fertilization on insect herbivores include host-plant changes in elemental stoichiometry, nutritional geometry, essential micronutrients, and allelochemicals (Vogels et al. 2023). As predicted, detrimental effects were more pronounced in high-altitude populations, such that these populations, likely adapted to low levels of host-plant nitrogen, are even more vulnerable to agricultural intensification. Note that all alpine populations were sampled in the central Alps on unfertilized meadows above crystalline primary rock, resulting in acidic, nutrient-poor soils. In contrast, sampling locations of low-altitude populations were on partly fertilized meadows above less acidic soils (basalt, marshland). This also suggests that our data probably do not reflect recent adaptations to different levels of nitrogen availability, but rather long-term local adaptation. In any case, our results are in agreement with the notion that herbivores from naturally nutrient-poor environments should be more strongly affected than those from richer environments (Vogels et al. 2023). Furthermore, detrimental effects of nitrogen fertilization were more pronounced at the higher rearing temperature, indicating increasing risks from climate change. Similarly, additive effects between host-plant nitrogen fertilization and drought were found in L. tityrus (Raharivololoniaina et al. 2023).

The effects of sex and temperature on life-history traits were as expected. Female insects are typically selected for large body size and a high investment into the abdomen, which is positively related to fecundity (Honek 1993; Nylin and Gotthard 1998; Teder et al. 2021). Males, in contrast, have a selective premium on rapid development to eclose before the females (protandry) and on flight capacity, both increasing their reproductive success (Wiklund et al. 1991; Berwaerts et al. 2006; Reim et al. 2018b). As generally found in ectotherms, development time was longer and growth rates lower at the lower developmental temperature (Von Bertalanffy 1960; Karl and Fischer 2008). According to the temperature-size rule on plasticity in body size (Atkinson and Sibly 1997; Karl and Fischer 2008), animals were larger at the lower temperature, also resulting in higher wing loadings. The higher mortality rates at the lower rearing temperature are likely associated with the concomitantly much longer development time.

In summary, we show local adaptation in developmental traits in the butterfly L. tityrus. Low-altitude populations were adapted to warmer temperatures and longer seasons, displaying a more rapid development and phenotypes indicative of a high dispersal capacity. Importantly, we also show that L. tityrus is vulnerable to agricultural intensification, responding negatively to agriculturally relevant levels of nitrogen fertilization of its host plant. According effects were particularly pronounced at warmer temperatures and in high-altitude populations. Thus, our study adds to the increasing knowledge that different drivers of global change, here nutrient loading and climate change, may interact and thereby increase the overall level of threat to biodiversity (Raven and Wagner 2021). Furthermore, our study suggests that populations inhabiting nitrogen-poor environments, as can be assumed for the alpine populations investigated here, might be even more vulnerable to agricultural intensification than others, which may be applicable to many species (see also Turlure et al. 2013). A much lower level of nitrogen input in the habitats of the alpine populations investigated here compared to those of the low-altitude populations is very likely, resulting from an absence of any fertilization of these remote alpine meadows in combination with acidic ground rocks and a low level of atmospheric nitrogen deposition in contrast to central and northern Germany (Walter 2007; European Environment Agency 2024).

In *L. tityrus*, high-altitude populations seem to be especially vulnerable, showing a high sensitivity to high nitrogen levels, reduced dispersal capacity, and reduced plasticity in heat resistance (Karl et al. 2008, 2009b). We think that our findings, specifically the detrimental effects of agriculturally relevant levels of nitrogen fertilization being exaggerated at warmer temperatures and in populations from nutrient-poorer environments, may have important implications for other vulnerable species in the face of rapid environmental change. In particular, nitrogen fertilization and deposition should be reduced especially in naturally nutrient-poore environments. Future research should investigate whether



Oecologia (2025) 207:51 Page 11 of 13 51

according species, populations, and communities are indeed more vulnerable to nutrient input, and which mechanisms may be responsible for detrimental effects.

Supplementary Information The online version contains supplementary material available at https://doi.org/10.1007/s00442-025-05690-8.

Acknowledgements We thank Jawaria Khalid and Supriya Pathak for help with rearing butterflies. This would have been the last paper of Ange's PhD thesis, but she passed away way too early. Rest in peace Ange!

Author contribution statement AR and KF conceived and designed the experiments. AR, RB, and KF conducted fieldwork. AR performed the experiments with the help of RB, FD, AH, EP, TR, and MY. AR and KF analyzed the data. AR and KF wrote the manuscript; other authors provided editorial advice.

Funding Open Access funding enabled and organized by Projekt DEAL. This research was funded by the research initiative 'Rheinland-Pfalz'.

Data availability The datasets used during the current study are available from the corresponding author on reasonable request.

Code availability Not applicable.

Declarations

Conflict of interest The authors declare that they have no conflict of interest.

Ethical approval All applicable institutional and/or national guidelines for the care and use of animals were followed.

Consent to participate Not applicable.

Consent for publication Not applicable.

Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons licence, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons licence, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons licence and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this licence, visit http://creativecommons.org/licenses/by/4.0/.

References

- Addo-Bediako A, Chown SL, Gaston KJ (2002) Metabolic cold adaptation in insects: a large-scale perspective. Funct Ecol 16:332–338
- Almbro M, Kullberg C (2012) Weight loading and reproductive status affect the flight performance of *Pieris napi* butterflies. J Insect Behav 25:441–452. https://doi.org/10.1007/s10905-011-9309-1
- Arrese EL, Soulages JL (2010) Insect fat body: energy, metabolism, and regulation. Annu Rev Entomol 55:207-225

Atkinson D, Sibly RM (1997) Why are organisms usually bigger in colder environments? Making sense of a life history puzzle. Trends Ecol Evol 12:235–239

- Bates D, Mächler M, Bolker B, Walker S (2015) Fitting linear mixedeffects models using lme4. J Stat Softw 67:1–48
- Bauerfeind SS, Fischer K (2013) Increased temperature reduces herbivore host-plant quality. Glob Change Biol 19:3272–3282
- Berwaerts K, Van Dyck H, Aerts P (2002) Does flight morphology relate to flight performance? An experimental test with the Butterfly *Pararge aegeria*. Funct Ecol 16:484–491
- Berwaerts K, Aerts P, Van Dyck H (2006) On the sex-specific mechanisms of butterfly flight: flight performance relative to flight morphology, wing kinematics, and sex in *Pararge aegeria*. Biol J Linn Soc 89:675–687
- Blanquart F, Kaltz O, Nuismer SL, Gandon S (2013) A practical guide to measuring local adaptation. Ecol Lett 16:1195–1205
- Block AK, Vaughan MM, Schmelz EA, Christensen SA (2019) Biosynthesis and function of terpenoid defense compounds in maize (*Zea mays*). Planta 249:21–30
- Bobbink R, Hicks K, Galloway J, Spranger T, Alkemade R, Ashmore M, Bustamante M, Cinderby S, Davidson E, Dentener F, Emmett B, Erisman J, Fenn M, Gilliam F, Nordin A, Pardo L, WallisDeVries MF (2010) Global assessment of nitrogen deposition effects on terrestrial plant diversity: a synthesis. Ecol Appl 20:30–59
- Bräu M, Bolz R, Kolbeck H, Nunner A, Voith J, Wolf W (2013) Tagfalter in Bayern. Ulmer, Stuttgart
- Burke S, Pullin AS, Wilson RJ, Thomas CD (2005) Selection for discontinuous life-history traits along a continuous thermal gradient in the butterfly *Aricia agestis*. Ecol Entomol 30:613–619
- Carvalheiro LG, Biesmeijer JC, Franzén M, Aguirre-Gutiérrez J, Garibaldi LA, Helm A, Michez D, Pöyry J, Reemer M, Schweiger O, van Leonden B, WallisDeVries MF, Kunin WE (2019) Soil eutrophication shaped the composition of pollinator assemblages during the past century. Ecography 43:209–221
- Ceballos G, Ehrlich PR, Barnosky AD, García A, Pringle RM, Palmer TM (2015) Accelerated modern human-induced species losses: entering the sixth mass extinction. Sci Adv 19:e1400253
- Dillon ME, Frazier MR, Dudley R (2006) Into thin air: physiology and evolution of alpine insects. Integr Comp Biol 46:49–61
- Ebert G, Rennwald E (1991) Die schmetterlinge baden-Württembergs. Ulmer, Stuttgart
- European Environment Agency (2024). Risk of eutrophication measured as exceedance of critical loads of nitrogen deposition in Europe, in 2022. www.eea.europa.eu.
- Fischer K, Fiedler K (2000) Response of the copper butterfly *Lycaena tityrus* to increased leaf nitrogen in natural food-plants: evidence against the nitrogen limitation hypothesis. Oecologia 124:235–241
- Fischer K, Karl I (2010) Exploring plastic and genetic responses to temperature variation using Copper butterflies. Climate Res 43:17–30
- Fischer K, Brakefield PM, Zwaan BJ (2003) Plasticity in butterfly egg size: why larger offspring at lower temperatures? Ecology 84:3138–3147
- Fischer K, Kölzow N, Höltje H, Karl I (2011) Assay conditions in laboratory experiments: is the use of constant rather than fluctuating temperatures justified when investigating temperature-induced plasticity? Oecologia 166:23–33
- Gunderson AR, Tsukimura B, Stillman JH (2017) Indirect effects of global change: from physiological and behavioral mechanisms to ecological consequences. Integr Comp Biol 57:48–54
- Hamann E, Blevins C, Franks SJ, Jameel MI, Anderson JT (2021) Climate change alters plant-herbivore interactions. New Phytol 229:1894–1910



- Han P, Lavoir AV, le Bot J, Amiens-Desneux E, Desneux N (2014) Nitrogen and water availability to tomato plants triggers bottomup effects on the leafminer *Tuta absoluta*. Sci Rep 4:1–8
- Harvey JA, Heinen R, Gols R, Thakur MP (2020) Climate changemediated temperature extremes and insects: from outbreaks to breakdowns. Glob Change Biol 26:6685–6701
- Hodkinson ID (2005) Terrestrial insects along elevation gradients: species and community responses to altitude. Biol Rev 80:489–513
- Honek A (1993) Intraspecific variation in body size and fecundity in insects: a general relationship. Oikos 66:483–492
- Jackson MC, Loewen CJG, Vinebrooke RD, Chimimba CT (2016) Net effects of multiple stressors in freshwater ecosystems: a metaanalysis. Glob Change Biol 22:180–189
- Johnson CN, Balmford A, Brook BW, Buettel JC, Galetti M, Guangchun L, Wilmshurst JM (2017) Biodiversity losses and conservation responses in the Anthropocene. Science 356:270–275
- Karl I, Fischer K (2008) Why get big in the cold? Towards a solution of a life-history puzzle. Oecologia 155:215–225
- Karl I, Janowitz SA, Fischer K (2008) Altitudinal life-history variation and thermal adaptation in the Copper butterfly Lycaena tityrus. Oikos 117:778–788
- Karl I, Schmitt T, Fischer K (2009a) Genetic differentiation between alpine and lowland populations of a butterfly is related to PGI enzyme genotype. Ecography 32:488–496
- Karl I, Sørensen JG, Loeschcke V, Fischer K (2009b) HSP70 expression in the Copper butterfly Lycaena tityrus across altitudes and temperatures. J Evol Biol 22:172–178
- Kawecki TJ, Ebert D (2004) Conceptual issues in local adaptation. Ecol Lett 7:1225–1241
- Korpelainen H, Pietiläinen M (2020) Sorrel (*Rumex acetosa* L.): not only a weed but a promising vegetable and medicinal plant. Bot Rev 86:234–246. https://doi.org/10.1007/s12229-020-09225-z
- Kuczyk J, Raharivololoniaina A, Fischer K (2021a) High temperature and soil moisture reduce host-plant quality for an insect herbivore. Ecol Entomol 46:889–897. https://doi.org/10.1111/een.13025
- Kuczyk J, Raharivololoniaina A, Fischer K (2021b) Population-specific responses of an insect herbivore to variation in host-plant quality. Ecol Evol 11:17963–17972. https://doi.org/10.1002/ece3.8392
- Kurze S, Heinken T, Fartmann T (2017) Nitrogen enrichment of host plants has mostly beneficial effects on the life-history traits of nettle-feeding butterflies. Acta Oecol 85:157–164
- Kurze S, Heinken T, Fartmann T (2018) Nitrogen enrichment in host plants increases the mortality of common Lepidoptera species. Oecologia 188:1227–1237
- Lebeau J, Wesselingh RA, Van Dyck H (2016) Floral resource limitation severely reduces butterfly survival, condition and flight activity in simplified agricultural landscapes. Oecologia 180:421–427
- Lebigre C, Vanderbeken C, Turlure C, Schtickzelle N (2018) Host plant nitrogen enrichment has both positive and negative effects on the larval growth of a specialist butterfly. Ecol Entomol 43:494–505
- Lion MB, Reis S, de Brito MRM, Cardoso MZ (2023) The effect of habitat amount on flight-related traits in the butterfly *Hamadryas* februa is sex-dependent. Ecol Entomol 48:135–144. https://doi. org/10.1111/een.13208
- Meihls LN, Kaur H, Jander G (2012) Natural variation in maize defense against insect herbivores. Cold Spring Harb Symp Quant Biol 77:269–283
- Nessel MP, Konnovitch T, Romero GQ, González AL (2021) Nitrogen and phosphorus enrichment cause declines in invertebrate populations: a global meta-analysis. Biol Rev Cambr Philos Soc 96:2617–2637
- Nijssen ME, WallisDeVries MF, Siepel H (2017) Pathways for the effects of increased nitrogen deposition on fauna. Biol Conserv 212:423–431

- Nylin S, Gotthard K (1998) Plasticity in life-history traits. Annu Rev Entomol 43:63–83
- Oliver TH, Morecroft MD (2014) Interactions between climate change and land use change on biodiversity. Attribution problems, risks, and opportunities. Wires Clim Change 5:317–335. https://doi.org/10.1002/wcc.271
- Pereira HM, Navarro LM, Martins IS (2012) Global biodiversity change: the bad, the good, and the unknown. Annu Rev Environ Res 37:25–50
- Pereira HM, Martins IS, Rosa IMD (2024) Global trends and scenarios for terrestrial biodiversity and ecosystem services from 1900 to 2050. Science 384:458–465. https://doi.org/10.1126/science.adn3441
- R Core Team (2024): A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria
- Raharivololoniaina A, Berweiler S, Fischer K (2021) Nitrogen fertilization and high plant growing temperature increase herbivore performance. Ecosphere 12:e03891. https://doi.org/10.1002/ecs2.3891
- Raharivololoniaina A, Moewert E, Fischer K (2023) Simulating effects of agricultural intensification and climate change: nitrogen fertilization and drought stress decrease insect herbivore performance. Ecol Entomol 48:325–335. https://doi.org/10.1111/een.13224
- Raven PH, Wagner DL (2021) Agricultural intensification and climate change are rapidly decreasing insect biodiversity. Proc Natl Acad Sci USA 118:1–6
- Reim E, Baguette M, Günter F, Fischer K (2018a) Emigration propensity and flight performance are decoupled in a butterfly. Ecosphere 9:e02502
- Reim E, Blesinger S, Förster L, Fischer K (2018b) Successful despite poor flight performance: range expansion is associated with enhanced exploratory behaviour and fast development. J Evol Biol 31:1165–1179
- Roff DA (1980) Optimizing development time in a seasonal environment—the ups and downs of clinal variation. Oecologia 45:202–208
- Sage RF (2020) Global change biology: a primer. Glob Change Biol 26:3–30
- Scheffers BR, de Meester L, Bridge TCL, Hoffmann AA, Pandolfi JM, Corlett RT, Butchart SHM, Pearce-Kelly P, Kovacs KM, Dudgeon D, Pacifici M, Rondinini C, Foden WB, Martin TG, Mora C, Bickford D, Watson JEM (2016) The broad footprint of climate change from genes to biomes to people. Science 354:1–10
- Settele J, Kudrna O, Harpke A, Kühn I, van Swaay C, Verovnik R, Warren M, Wiemers M, Hanspach J, Hickler T, Kühn E, van Halder I, Veling K, Vliegenthart A, Wynhoff I, Schweiger O (2008) Climatic risk atlas of European butterflies. BioRisk 1:1–712
- Stevens CJ, Duprè C, Dorland E, Gaudnik C, Gowing DJG, Bleeker A, Diekmann M, Alard D, Bobbink R, Fowler D, Corcket E, Mountford JO, Vandvik V, Aarrestad PA, Muller S, Dise NB (2010) Nitrogen deposition threatens species richness of grasslands across Europe. Environ Poll 158:2940–2945
- Tao L, Berns AR, Hunter MD (2014) Why does a good thing become too much? Interactions between foliar nutrients and toxins determine performance of an insect herbivore. Funct Ecol 28:190–196
- Teder T, Kaasik A, Taits K, Tammaru T (2021) Why do males emerge before females? Sexual size dimorphism drives sexual bimaturism in insects. Biol Rev 96:2461–2475
- Terblanche JS, Clusella-Trullas S, Deere JA, van Vuuren BJ, Chown SL (2009) Directional evolution of the slope of the metabolic rate-temperature relationship is correlated with climate. Physiol Biochem Zool 82:495–503
- Tolman T, Lewington R (2008) Die Tagfalter Europas und Nordwestafrikas. Franckh-Kosmos, Stuttgart



Oecologia (2025) 207:51 Page 13 of 13 51

Toprak U, Hegedus D, Doğan C, Güney G (2020) A journey into the world of insect lipid metabolism. Arch Insect Biochem Physiol 104:e21682

- Trense D, Schmidt TL, Yang Q, Chung J, Hoffmann AA, Fischer K (2021) Anthropogenic and natural barriers affect genetic connectivity in an Alpine butterfly. Mol Ecol 30:114–130. https://doi.org/10.1111/mec.15707
- Turlure C, Radchuk V, Baguette M, Meijrink M, van den Burg A, WallisDeVries MF, van Duinen G-J (2013) Plant quality and local adaptation undermine relocation in a bog specialist butterfly. Ecol Evol 3:244–254
- van Dyck H, van Strien AJ, Maes D, van Swaay CAM (2009) Declines in common, widespread butterflies in a landscape under intense human use. Conserv Biol 23:957–965
- Vogels JJ, van de Waal DB, WallisDeVries MF, van den Burg AB, Nijssen M, Bobbink R, Berg MP, Olde Venterink H, Siepel H (2023) Towards a mechanistic understanding of the impacts of nitrogen deposition on producer–consumer interactions. Biol Rev 98:1712–1731. https://doi.org/10.1111/brv.12972
- Von Bertalanffy L (1960) Principles and theory of growth. In: Nowinski WW (ed) Fundamental aspects of normal and malignant growth. Elsevier, New York, pp 137–259
- Wagner DL (2020) Insect declines in the anthropocene. Annu Rev Entomol 65:457–480

- WallisDeVries MF, van Swaay CAM (2006) Global warming and excess nitrogen may induce butterfly decline by microclimatic cooling. Glob Change Biol 12:1620–1626
- WallisDeVries MF, van Swaay CAM (2017) A nitrogen index to track changes in butterfly species assemblages under nitrogen deposition. Biol Conserv 212:448–453. https://doi.org/10.1016/j.biocon. 2016.11.029
- Walter R (2007) Geologie von Mitteleuropa. Schweizerbart'sche Verlagsbuchhandlung, Stuttgart
- Weiskopf SR, Rubenstein MA, Crozier LG, Gaichas S, Griffis R, Halofsky JE, Hyde KJW, Morelli TL, Morisette JT, Muñoz RC, Pershing AJ, Peterson DL, Poudel R, Staudinger MD, Sutton-Grier AE, Thompson L, Vose J, Weltzin JF, Whyte KP (2020) Climate change effects on biodiversity, ecosystems, ecosystem services, and natural resource management in the United States. Sci Total Environ 733:137782. https://doi.org/10.1016/j.scitotenv. 2020.137782
- White TCR (1993) The inadequate environment. Nitrogen and the abundance of animals. Springer, Berlin
- Wiklund C, Nylin S, Forsberg J (1991) Sex-related variation in growth rate as a result of selection for large size and protandry in a bivoltine butterfly, *Pieris napi*. Oikos 60:241–250

