

Review on invasion of microplastic in our ecosystem and implications

Science Progress

2022, Vol. 105(4) 1–23

© The Author(s) 2022

Article reuse guidelines:

sagepub.com/journals-permissions

DOI: 10.1177/00368504221140766

journals.sagepub.com/home/sci



Minha Lee , Heejung Kim , Han-Sun Ryu,
Jinah Moon , Naing Aung Khant ,
Chaerim Yu and Ji-Hee Yu

Department of Geology, Kangwon National University, Chuncheon,
Republic of Korea

Abstract

Today the world is going through the “Plastic Age.” Nowadays, it is difficult to find a commonly used convenient item that is nonplastic. Plastic production and consumption, thus, increased exponentially and plastic emerged as one of the major concerns for waste management. Recent studies confirmed a faster rate of plastic degradation than previously believed under various conditions (e.g. saltwater, UV, soil interaction) that microplastic has become a new type of health-hazardous pollution source. Much research has been conducted since the discovery of the “Pacific Garbage Patch,” and the scope has expanded from marine to soil, groundwater, air, and food chain. This article underwent a substantial amount of literature review to verify the degree of microplastic pollution progression in major pillars of the environment (aqueous, terrestrial, airborne, bio-organism, and human). Multiple kinds of literature indicated a high possibility of vigorous interaction among the pillars that microplastic is not stationary at the point of contamination but travels across the nation (transboundary) and medium (transmedium). Thus, only the waste reduction policy (i.e. production and consumption reduction) would be effective through a single national or local effort, while pollution and contamination management require more of a collective, if not global, approach. For these characteristics, this article proposes two most urgently required actions to combat microplastic pollution: (a) global acknowledgement of microplastic as transboundary and transmedium pollution source that require international collective action and (b) standardization of microplastic related research including basic definition and experimental specification to secure global comparativeness among data analysis. Without resolving these two issues, it could be very difficult to obtain an accurate global status mapping of microplastic pollution to design effective and efficient global microplastic pollution management policies.

Keywords

Microplastic pollution, soil, marine, air, food chain, groundwater

Corresponding author:

Heejung Kim, Department of Geology, Kangwon National University, Chuncheon 200-701, Republic of Korea.
Email: hydroqueen@kangwon.ac.kr



Creative Commons Non Commercial CC BY-NC: This article is distributed under the terms of the Creative Commons Attribution-NonCommercial 4.0 License (<https://creativecommons.org/licenses/by-nc/4.0/>)

which permits non-commercial use, reproduction and distribution of the work without further permission provided the original work is attributed as specified on the SAGE and Open Access page (<https://us.sagepub.com/en-us/nam/open-access-at-sage>).

Introduction

Since the commercialization of plastic materials to replace animal-derived materials such as elephants' ivory or tortoiseshell in 1862,¹ plastic penetrated every corner of modern life. Owing to its cheap and easy mass production capability than conventional materials (e.g. paper, glass, and metals), plastic products have created a "single-use culture." Fueled by the post-World War modern fast-moving lifestyle, it now is impossible to spend a day without using plastic of some kind. The outbreak of COVID-19 further expanded the use of plastic to "sanitation and safety."² The world now faces a new type of daily "safety measure" waste (i.e. personal protective equipment, test kits, vaccine syringes, and single-use cutlery) in addition to the conventional high volume of plastic waste traded for its convenience, low-cost, durability, unbreakable, heat resistance, and lightweight.

Due to the traditional belief around the "durability" of plastics, the major issues focused on how to secure enough landfill space to address rapidly increasing plastic wastes and how to capture illegally dumped plastic products in the early years. The agenda, however, got stirred with the discovery of the "Pacific Garbage Patch," also known as the Pacific Trash Vortex, in 1997 by Charles Moore on his sailing way back to California from Hawaii.³ Unlike traditionally imagined plastic trash floating on the ocean surface or sunk in deep-sea sediments, the Garbage Patch revealed a "cloudy soupy zone in the middle of the Pacific Ocean" made of tiny bits of plastics, now termed microplastic.⁴ As microplastic pollution gained research interest among academia, the scope of studies assessing the spread and impacts of microplastic also expanded from marine environment to fresh-to-groundwater, terrestrial habitats, and air. Reinforced by enhanced public awareness and international pressure, it also took less than a decade for many governments to adopt a microplastic reduction policy stance.²

The research, however, lacked time to fully ripe in depth that the state of global knowledge on microplastic is rather fragmentary and sparse than interdisciplinary and comprehensive.⁵ Although any research is meaningful for its own purpose, it would be even better if each puzzle piece can collectively contribute toward a global knowledge map on microplastic. This paper, thus, attempts to review the nature of microplastic pollution progress observed and analyzed in various environmental fields (marine, soil, groundwater, air, and food chain) to deduce possible policy implications to best support research advancement in this newly emerged environmental challenge of this "Plastic Age."

Invasion of microplastic pollution in the ecosystem

Marine environment

The most common types of plastics found in marine environments are polyethylene (PE), polypropylene (PP), polystyrene (PS), and polyvinyl chloride (PVC).^{6,7} There are two main processes, through which microplastics reach the marine environment. First, large plastics naturally degrade into microplastics in the marine environment, and second, the microplastics flow into seawater through rivers and other water cycles from the land.^{6,8} The major source of large plastics in the marine environment is the

fishery industry: fishing nets contributes approximately 640 thousand tons (10% of total marine wastes).⁹ These fishing nets undergo biodegradation, photodegradation, or hydrolysis into microplastic less than 5 mm.^{6,10,11}

Yet, beach waste is the largest contributor to marine microplastic pollution through plastic degradation.¹² This is because the low water temperature of saltwater slows the degradation process that the microplastic formation rate is lower in the far deep ocean. Contrarily, beach sand temperature rise up to 40 °C creating a faster photodegradation for plastic bottles dumped on the shore rather than floating in the Pacific Ocean.^{6,12,13} Microplastics that reach the ocean floor from land usually are comprised of the wash-off of personal products, fabric from laundry, and particles from vehicle tires.¹⁴ This finding matches the analysis result on marine microplastics collected from Playa Grande Beach, Canary Islands of Spain in 2020 to have comprised of PP and PE in the form of microplastic fragments, pellets, fiber and films.^{13,15} As the gravity of marine plastic pollution was only revealed in 1997 with Pacific Garbage Patch, the history of marine microplastic pollution research is not very long. Old observation records on plastics of a few millimeters in size both on the ocean surface and inside of many saltwater fishes on the East Coast of the United States and the Southern part of the United Kingdom went unnoticed for almost 30 years.^{16–18} It was Thompson's article published in *Science* in 2004 that first introduced the term, "microplastic,"¹⁹ and triggered much research efforts in related fields.

Promisingly, the scope of marine microplastic research has been expanded to marine ecological microorganisms and food chains. Animals exposed to microplastic-concentrated water has been found to produce unhealthy egg and sperms, which resulted in 41% fewer larvae than the animals in clean water.⁸ The Pacific Krill (*Euphasea pacifica*) was observed to ingest its staple algae, as well as PE beads, ground to about the same size range with no evident foraging bias.²⁰ This infers a possibility of microplastic invasion into the food chain as krill, whether it is ingested via algae or plastic beads, is favored diet by predators and seagulls nearby. Some researchers further introduced the concept of "secondary nanoplastics" resulting from the biodegradation of microplastic through digestion and food chain accumulation.¹⁴ Yet, many laboratory experiments and related research efforts still lean heavily toward "how much" and "what kind" of microplastic has polluted the marine environment. Thus, there currently is insufficient research on the effects of microplastic formation in marine environments that understanding on the possible consequences on marine ecology and human health is yet too low to be translated into effective policy.

Soil

Since the introduction of microplastic soil contamination²¹ related studies expanded both in scope and volume.^{22–27} The United Nations Environment Program (UNEP) further recognized the need to verify the consequences of microplastic pollution in the soil environment on World Environment Day.²⁸ Insofar, researchers focused on the hazardous effects of microplastics in soil and terrestrial ecosystems,^{29–31} with emphasis on its rapid spreading rate in soil environments.^{21,29,31,32} Recent studies worryingly indicated the possibility of 4 to 23 times higher microplastic concentration on land than in the

ocean²⁴ due to many entry points such as microfibers from clothing, plastic beads from personal care products, biosolids,^{33–37} reclamation sites in urban and industrial centers,³² lake water flooding, road waste and illegal waste dumping,³⁸ tire wear,^{39–41} atmospheric particles,⁴² and fertilizers.^{24,31}

Yet, there is a growing consensus that the agricultural ecosystem is the major pathway for microplastic in terrestrial environment^{43,44}; the largest concerns involve wastewater irrigation and mulching used in agricultural activities.^{45–48} The application of PE plastics in agricultural processes in 1938 revolutionized commercial crop productivity^{49,50} but resulted in heavy microplastic soil pollution (approximately 10% of all agricultural soil).⁵¹ The situation was exacerbated when the usage of sewage sludge as agricultural fertilizer became a standard management process. Studies estimate approximately 63,000 to 430,000 tons and 44,000 to 300,000 tons of microplastics have been discharged annually in Europe and North America, respectively.^{31,32} Accounting for the vast physical coverage, the agricultural soil is also described as microplastic-contaminated large reservoirs in which airborne particles drop,⁴² mostly likely in form of precipitation.^{25,38,52}

Microplastic soil pollution has been found to alter the physical properties of soil, reduce soil fertility, and destroy resident microbial communities. Altogether, the contamination impairs soil quality and nutritional circulation^{53–56} to infertile the crops,⁵⁷ damages the reproductive growth of plants,⁵⁸ and exhibits plant ecotoxicity and genetic toxicity.⁵⁹ It has been shown that macro and microplastic residues in soil have a negative effect on wheat (*Triticum aestivum*) growth.⁵⁸ In addition, the germination rate decreased when exposed to microplastics using Cress seeds, indicating that microplastics can physically block the pores of the seeds and interfere with moisture absorption, delaying germination.⁵⁷ Intracellular absorption of nanoplastics (20 and 40 nm) in tobacco BY-2 plant cell culture was confirmed by fluorescent microplastic images.⁶⁰ In addition, evidence of attachment, absorption, accumulation, and migration of microplastics within edible plant lettuce was observed using fluorescent markers.⁶¹ As such, there is still not much information on microplastic accumulation by plants in the soil, and research on quantifying microplastics in more plant species is needed.⁶²

In addition, microplastics ingested by earthworms were found to travel through the food chain, which may pose a potential threat to terrestrial predators and even humans.⁶³ Despite the seriousness of microplastic pollution hazard posed to soil biodiversity, ecosystem function, and food security, there currently is limited knowledge on the environmental impact of microplastics on actual soil, or how microplastic-polluted soil impedes nutritional circulation.

Groundwater

Groundwater is another important water resource as it accounts for 97% of accessible freshwater^{64,65} that deserves an increase in quality management efforts. Yet, microplastic pollution research on groundwater only began in 2017 as groundwater traditionally was considered as impenetrable.⁶⁶ The first study subjected karst areas, the most vulnerable to aquifer pollution, to verify the existence of microplastic contamination in groundwater.⁶⁶ Since then, the experiments focused on the microplastic pollution progress investigation (e.g. concentration, polymer type, shape, and color) by the surrounding environment and

geology of the aquifer.⁶⁷ Among various inflow pathways of microplastics into the aquifer, ranging from sinkholes, groundwater wells (pumping), seawater infiltration, and vertical transport by soil organisms or rainfall,^{44,56,68} surface water is known to be the critical passage.^{69,70} Although research on this important carrier, surface water, took its course a decade prior to groundwater in 2007, it still is in the early stage.^{71–73} The findings deduced four major cases through which microplastic enters surface water: discharged residuals from unsuccessful decomposition or filtration during the wastewater treatment processes, run-offs, flooding and tidal inflows.^{24,74,75}

To understand the transport and retention of microplastics in the hyporheic zone (i.e. groundwater-surface water interaction zone), studies adopted colloidal transport model analysis under various environmental conditions (e.g. chlorine concentration, ion strength, and microplastic particle size change).^{76–81} Research has also been expanded to develop a new upgraded microplastic transport model tailored for groundwater passage investigation from the conventional colloidal model.^{76,82} The analysis results reported increasing average microplastic concentration in deeper layers of sediments with decreasing particle size.⁸³ As PP and PE are the dominant microplastic polymers collected from surface water, followed by PS, PET, PA, PCV, and PU, the two are the most common types detected also in groundwater.^{80,84} The difference is observed from shape: while the majority takes fiber form in surface water, most are decomposed into fragments under 50 μm in groundwater.^{65,67} Such difference satisfies research conclusions that surface water contributes the most to the microplastic pollution in groundwater and the geological environment surrounding groundwater acts as a natural filter against large pollutants.

Accounting that the major uses of groundwater are agricultural and drinking purposes, microplastic contamination poses a serious threat to human health, possibly worse than food. Furthermore, there is a growing concern about the effects of microplastic pollution on smaller organisms (microorganisms, troglofauna, and stygofauna) inhabiting in groundwater environment.^{85–87} Yet, research on microplastic pollution in groundwater is limited to the verification of causes and subjects rather than the aftereffects or remedies to contamination.^{88,89} Furthermore, much of the currently conducted microplastic transport modeling experiments are bounded within the laboratory that their actual application is still questionable. In this light, microplastic pollution-related research on groundwater environments can be categorized as being in its primitive stage and possess much room for expansion and development.

Air

Interest in research on microplastic atmospheric pollution increased when the first signs were detected in Grand Paris in 2015.⁹⁰ Since this identification, primary research efforts acknowledged the possible causal relation between airborne microplastic fallout and urban wastewater/freshwater microplastic contamination.^{42,90,91} Another strand of comparative studies in remote areas with almost no anthropogenic activities (such as the Pyrenees, a nature-preserved mountain in France) further suggested the likelihood of microplastic atmospheric transport in a similar pattern with particulate matters.^{42,52,92,93} Specifically, the meteorological factors and cloud formation process

(atmospheric currents) are found to be the most vital conditions that determine the length and time of microplastic atmospheric transport.^{94–96} The atmospheric microplastic fallout concentration is observed to be higher on rainy days and Winter seasons compared to Summer and Spring.^{42,90,91,94,97}

Their low density and very small size allow long buoyant time,^{42,98} and there seems to be free interchange occurring between airborne, terrestrial, and aqueous microplastic. Strong wind and spray emission releases significant amounts of microplastics formed either through degradation in seawater or various economic activities on land into the atmosphere,^{42,90,99–101} while the airborne falls into vast land or oceans under appropriate climate conditions (e.g. precipitation).¹⁰² Although the exact interaction between extreme weather conditions (e.g. hurricanes, typhoons, cyclones, and monsoons) and airborne microplastics has yet to be identified, dust storms have been identified to play a key role in atmospheric microplastic transportation.¹⁰³ Erosion of synthetic materials (e.g. tires, fibers, building materials) is found to be the major source of airborne microplastic,^{42,90,100} followed by building materials, waste incineration, fertilizers made of sludge, abrasive powders, and 3D printing.^{104–106} Accordingly, fragments, films, and fibers^{97,106–110} made of PE, PP, PS, and PET are the most common shapes and types of microplastic found in the atmosphere.^{106,108}

Recently, the research scope of microplastic pollution in the air has expanded to include the impacts of airborne microplastics on human health and the environment.^{108,110–113} A number of studies suggest a higher risk from indoor microplastic contamination than outdoor as is the case with many other typical air pollutants' health risks.^{42,108,112,113} This is because pollutants typically require a certain inhalation concentration to exert health-hazardous reactions such as interstitial lung diseases.¹¹² Although such findings are still too broad to understand the movement and consequences of airborne microplastics, one thing clear is that microplastic air pollution is transboundary, transmittable to other mediums, and health hazardous. The world, thus, requires more collaborative research in the field to deduce policy recommendations for global cooperation. In this light, standardizing key experiment definitions, protocols, and specifications is an urgent task, especially in aspects regarding air pollution.

Food

Food research began when nonpollen particulates spotted in honey and sugar were identified as microfibers and fragments,¹¹⁴ which raised the possibility of microplastic existence in nature as much as being inserted during the food manufacturing process. In 2014, microplastic was detected in local beer sold in Germany,¹¹⁵ which triggered related research together with the findings that ocean creatures have been exposed to marine plastic waste for years. It was found that floating marine plastic waste is ingested by planktonic organisms and eventually travel across the food chain from trophic level to higher level.^{6,116,117} Other research revealed that microplastic can reach marine mammals both directly from surrounding particles and indirectly through contaminated prey,¹¹⁸ which may result in decreased body defenses through cell damage.¹¹⁹ As microplastic ingestion by copepod and Euphausiacea (krill) were verified,^{19,120–125} experiments on bivalves raised for human consumption were conducted to understand how

close microplastic pollution has approached human's everyday lives.¹²⁵ Food items from which microplastic wastes are detected include: bivalves and various fishes,¹²⁶ salt,¹²⁷ dry fishes,¹²⁸ mussels,¹²⁹ tap water and beer,¹³⁰ and milk.¹³¹

Since the identification of microplastics from salt through a case study in China,¹²⁷ micro-Fourier Transform Infrared Spectroscopy (μ -FT-IR) analysis has been adopted to distinguish the type of plastic present in food and beverage. Supported by technical advancement, the related research scope has been expanded to verify that plastic residuals penetration into food occurs throughout the entire food preparation period: mulching film used in agriculture is absorbed by wheat,⁵⁸ microplastic present in groundwater is transferred to drinking water,¹³² microplastic exist in various drink packed in plastics bottles,¹³³ plastic tea bags released microplastic in hot water,¹³⁴ raw meat absorb microplastic from pressed PS used in packaging,¹³⁵ and microplastic was released from various plastic containers, plates, cutlery, and cups under various conditions.^{136,137} Other researchers raised further concerns about the possibility of microplastic turning poisonous once ingested internally (adhesive to other chemicals and cohesive to metals or bacteria).^{121,138–141}

So far, microplastic-related research on food has expanded much in both quantity and subjects. Yet, the research scope and purpose are still limited to existence verification in items that are commonly consumed by people, either directly ingested food or plastics used in food packaging and cutlery.^{135,142,143} The aftereffects of the microplastic present in these eatable conditions are still unknown as much as the process or degree of absorption once these microplastics successfully reach the human body (Table 1).

Microplastic pollution as global agenda

Transboundary and transmedium characteristics of microplastic pollution

As elaborated in the previous section, microplastic pollution is not a point-specific phenomena that remain at the point of origin until managed, but more of a transboundary, transmedium like conventional particulate matter air pollution or dichloro-diphenyl-trichloroethane (DDT) pollution.^{106,157} Recent analytical results that reported 28% of marine microplastic pollution originated from road vehicle movements (tire wearing) further emphasized the need to understand the relationship between air circulation and the transportation of airborne microplastics.^{158,159} Thus, there now is a growing concern calling for a comprehensive study on the dynamic interaction and transport among different environment mediums and locations to identify microplastic pollution cruising through the ecosystems¹⁵⁶ possibly taking different forms.

While the sea spray process and air bubble bursting on the ocean surface contribute to atmospheric microplastic pollution,^{106,160,161} airborne microplastics also are deposited in terrestrial or aquatic environments that may end up in the human body through food chain accumulation^{104,106} (Figure 1). Microplastic pollution from precipitation can spread over long distances through atmospheric transport, which can also spread microplastic pollution to remote uninhabited areas.^{92,162} Soil and water basins act as a sink for microplastic¹⁴ receiving airborne microplastic fallouts and more generated from various anthropogenic activities. Microplastics, thereon, invade hidden pillars of the ecosystem like the food chain and groundwater.^{22,38,162} Raindrops falling from the atmosphere

Table 1. Literature review on microplastic pollution in research fields (marine, soil, groundwater, air and food).

| Location | Sampling | Polymer types | Size | Shape | Color | Concentration (abundance) | Reference |
|---|--------------|--------------------------------|--------------------|--------------------------------|---|--|-----------|
| Marine Maine Bay, USA | Sea surface | Rayon and PE | 5–282 mm | Fiber | Black (62%), blue (15%), red (13%), and colorless | 0.60 ± 0.25 n/m ³ | 14 |
| Great Barrier Reef Marine Park, Australia | Sea surface | PE, PP, PS, and EVA | 0.4–82.6 mm | Pellet | White, colorless, and blue | 4.96 ± 0.27 n/m ³ | 144 |
| The North Atlantic | Sea surface | PE, PS, and PP | 1–100 mm | N/A | N/A | 1.6 ± 0.5 n/m ³ | 145 |
| Daebu Island, South Korea | Beach | PP, PE, PS, EVA, and PUF | 0.3–5 mm | N/A | N/A | Summer: 272,349 ($\pm 357,535$) n/m ³ and winter: 10,318 (± 6247) n/m ³ | 146 |
| Soil Franconia, Germany | Farmland | PE, PP, and PS | 1–5 mm | Film, fragment and fiber | N/A | 0.34 ± 0.36 n/kg ⁻¹ | 147 |
| Shanghai, China | Farmland | PP, PE, and PES | 1.91 ± 0.13 mm | Fiber, fragment and film | N/A | 78.0 ± 12.9 n/kg ⁻¹ | 148 |
| | Agricultural | N/A | 20 mm (width) | Fiber | N/A | 600–10,400 n/kg ⁻¹ | 149 |

(Continued)

Table 1. (continued)

| Location | Sampling | Polymer types | Size | Shape | Color | Concentration (abundance) | Reference |
|---------------------------|-------------------------------------|--|---|---|--|-------------------------------------|-----------|
| Mellipilla, Chile | Paddy | PE, PP, and PVC | 0.97 mm (length) 0.02–1 mm (domain) | Fiber, fragment and film | N/A | 10.3 ± 2.2 n/kg ⁻¹ | 150 |
| Groundwater Illinois, USA | Karst aquifer | PE | < 1.5 mm | Fibers | Blue and/or colorless (65%), red (15%), and gray (13%) | 15.2 n/L (max) | 66 |
| Victoria, Australia | Alluvial aquifer | PE, PS, PP, PVC, PET, PC, PMMA, and PA | 18–491 μm | Fragments and fibers | N/A | 38 ± 8 n/L | 67 |
| Tamil Nadu, India | Wells & borewells | Nylon (PA, 35%), PE (55%), and PET (10%) | 0.11–12.5 mm (mean: 0.6 ± 1.4 mm) and < 1 mm (34% domain) | Fibers, foam, pellets, films, and fragments | Colorless (53%), white (30%), blue (10%), gray (5%), and yellow (2%) | 4.2 n/L (median) and 10.1 n/L (max) | 151 |
| Krakow, Poland | Deep well (untreated potable water) | N/A | 0–0.045 mm | Fragments | Blue and green | N/A | 152 |

(Continued)

Table 1. (continued)

| Location | Sampling | Polymer types | Size | Shape | Color | Concentration (abundance) | Reference |
|-------------------------------------|---|--|--|-------------------------------------|------------------------------------|--|-----------|
| Air | | | | | | | |
| Greater Paris, France | Urban outdoor atmospheric fallout | N/A | 100–5000 μm | Fibers and fragments | N/A | 30–100 n/100 m^3 | 90 |
| Asaluyeh County, Iran | Outdoor | N/A | 100–1000 μm | Fiber, spherules, fragment and film | White-transparent; >70% blue-green | 0.3–1.1 n/ m^{-3} | 153 |
| Central London | Outdoor | PAN, PES, PA, PP, PVC, PE, PET, PE, PUR, and PPR | 75–100 μm | Fragments, films, granules and foam | N/A | 59 \pm 32 n/ m^2 /d (mean) | 154 |
| South Korea | Indoor/outdoor | PE, PP, PA, PES, AR, and PS | Indoor: 20.1–680 μm and outdoor: 20.3–4497.4 μm | Fibers, fragments, and film | N/A | Indoor: 0.49–6.64 (3.02 \pm 1.77) n/ m^3 and outdoor: 0.45–5.16 (1.96 \pm 1.65) n/ m^3 | 155 |
| Food | | | | | | | |
| North Sea, Germany/Brittany, France | Bivalves (<i>Mytilus edulis</i> and <i>Crassostrea gigas</i>) | N/A | 5–10 mm/11–15 mm | N/A | Red and blue | 0.36 \pm 0.07 n/g (<i>Mytilus edulis</i>) and 0.47 \pm 0.16 n/g (<i>Crassostrea gigas</i>) | 125 |
| China | Table salt (sea salt, lake salt, and rock/well salt) | PET, PE, and cellophane | 45 μm –4.3 mm | Fiber and fragment | Black, red, blue, and white | 550–681 n/kg (sea salts), 43–364 n/kg (lake salts), 7–204 n/kg (rock/well salts) | 127 |

(Continued)

Table 1. (continued)

| Location | Sampling | Polymer types | Size | Shape | Color | Concentration (abundance) | Reference |
|---------------------|--|------------------|--------------------------|--------------------|------------------------------------|--|-----------|
| Spain | Table salt (sea salt and rock/well salt) | PET, PP, and PE | 30 μm –3.5 mm | Fiber | Black, red, blue, and white | 50–280 n/kg (sea salt), 115–185 n/kg (well/rock salt) | 156 |
| Mexico City, Mexico | Beer, soft drinks, energy drinks, and cold tea | PA, PEA, and ABS | 0.1–3 mm | Fiber and fragment | Blue, red, brown, black, and green | ND–28 \pm 5.29 n/L (beer), ND–7 \pm 3.21 n/L (soft drinks), ND–6 \pm 1.53 n/L (energy drinks), and 1 \pm 0.57–6 \pm 2 n/L (cold tea) | 133 |

N/A: not available/reported; ND: not detected; PA: polyamide; PE: polyethylene; PES: polyester; PP: polypropylene; PS: polystyrene; PVC: polyvinyl chloride. Source: Author tabulated.

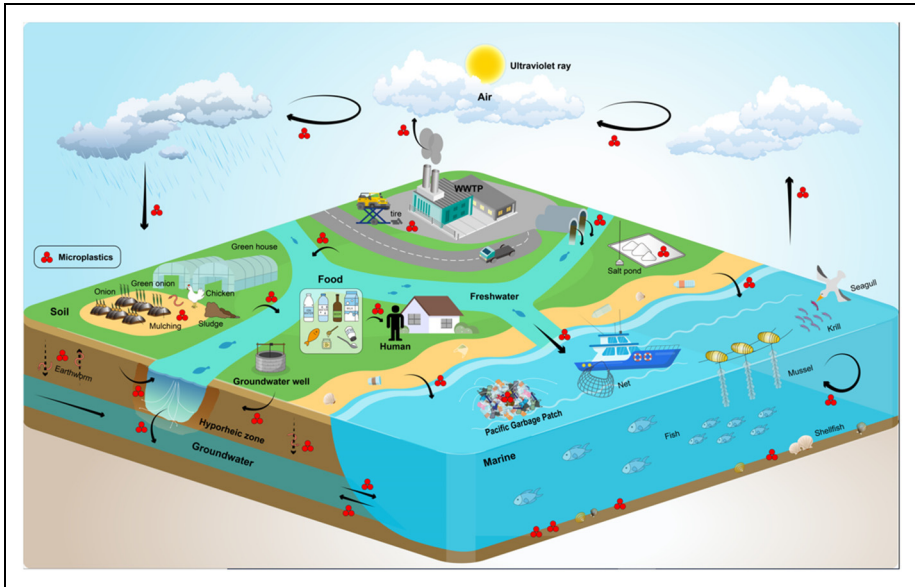


Figure 1. Microplastic transport and circulation throughout environmental pillars.

are recharged into aquifers through soil or stream bed sediments (hyporheic zone).^{56,163} In addition, in areas where tidal motion is active, microplastic enter the aquifer through seawater infiltration.¹⁶⁴

The natural adaptation patterns have also been observed, which, in turn, confirm the prolonged effects of microplastic pollution: marine microorganisms are now utilizing plastic parts floating on marine surfaces as a new type of habitat¹¹ as the portion of plastic waste surface coverage increased exponentially for them to persist on traditional ways. Several marine ecologists have further experimented on the consequences of microplastic ingestion by marine animals and reported that the survival instincts of perch larvae exposed to microplastic pollution were weakened.¹⁶⁵ The observed microplastic concentration from Southern Bluefin tuna in Tasmania revealed that microplastic accumulates along the marine food chain in a similar manner to the famous mercury poisoning.¹⁴⁴ A study on the effects of microplastic-contaminated sewage sludges and mulching also revealed different onion growth patterns depending on the type of microplastic exposed.¹⁶⁶ It has also been found that plastics on surface soil can be integrated into the deep soil by the latent activity of earthworms,²¹ which indicated the possibility of debris plastics and microplastics combined in the surface soil can be transported into deeper layers by the activity of soil organisms such as collembolans, insects, and plants.^{44,54,167–169} Lwanga et al.⁶³ reported that the highest microplastic concentration was observed in chicken feces, especially in chicken sandbags (a popular food ingredient in many countries), in the nutritional migration study on microplastic in house garden soil, earthworms, and chicken excrement.

Prerequisite to address microplastic pollution together

Despite the growing number of literature addressing various aspects of microplastic pollution, it is still difficult to “stand on the shoulders of giants (Isaac Newton)” due to the absence of standard operating protocols for microplastic analysis.¹⁵⁷ Researchers across the world are adopting different terminology, size, meaning and apparatus that a direct comparison and contrast among the results is limited. One prime example of a currently lacking global consensus or standard is the absence of a global definition/classification of microplastic. While the most frequently cited definition goes “microplastics are plastics that are <5 mm in length,” as proposed by Zhu et al.,¹⁶⁹ there is no clear lower boundary that many researchers indicate only the lower limit of the sampling method or processing apparatus. This variation results in incomparable or incompatible data collection, which consequently poses risks to global knowledge sharing for complete microplastic pollution mapping. Another example is the inconsistency in an itemized subject definition of microplastic: while rubber, some fibers, and paint particles are not considered as plastic by many polymer chemists,¹⁷⁰ others name abrasion of vehicle tires (19% natural rubber and 24% synthetic plastic rubber polymer) as major contributors to microplastic pollution. This inconsistency may cause confusion in data collection with different total numbers undermining analysis results reliability.

Additionally, there still are many missing puzzle pieces in each environment medium for a comprehensive microplastic pollution mapping. In marine microplastic pollution, for example, it is very likely for microplastics to sink below and endanger deep sea or marine ecosystems beyond the surface water as density increases by adhering to marine microorganisms or others.^{6,8,10} Despite a growing consensus, there are not many studies addressing the deep ocean floor or its potential consequences on marine ecosystems. Airborne microplastics are transported a long distance through air circulation system and are deposited on the terrestrial or aquatic environment under appropriate meteorological conditions such as rainfall and snowfall.^{92,94,161,171,172} Yet, the effects of extreme weather events (e.g. cyclone, hurricane, and monsoon) and climate change on microplastic pollution have yet been scrutinized. Similarly, despite the occurrence of microplastic in soil being ingested¹⁷³ or transferred to other soil organisms,³² the possible toxic side effects of microplastics on living organisms, including human health, have yet to be fully explored. As demonstrated, the research in groundwater contamination is in such an early stage to discuss the loopholes.

The most urgently required step, however, is to achieve global consensus to treat microplastic pollution as another kind of transboundary and transmedium pollutant that calls for a global approach. The absence of the exact term, “microplastic pollution,” on the global agenda (such as the Sustainable Development Goals of the United Nations or the Intergovernmental Oceanographic Commission of UNESCO) impedes cooperative international efforts including the collection and exchange of related data and knowledge.¹⁷⁴ It is also important to realize there is no reason for microplastic pollution to be less severe in developing or less-developed countries than in developed regions. Accounting for the effects of environmental policy in general terms, it is a logical deduction to expect a worse scenario under relatively fewer sensitive governments (i.e. heavier use of plastics for cost reduction). Yet not much research has been conducted in these

regions for low interest. This is another reason to push microplastic pollution on the global agenda for collective efforts: such global attention will pressurize international organizations and local governments to monitor microplastic pollution circulation in these regions, thus, contributing toward global microplastic generation and circulation reduction.

Conclusion

The discovery of the “Pacific Garbage Patch” in 1997 changed the entire view on plastic pollution and the world realized the presence of microplastics in our daily lives. Nothing comes for free, and convenience demands payment of some sort. The world enjoyed the Plastic Age of convenience and now faces the Climate Change Age and the Pollution Management Age. Microplastic pollution is one of the items that require global efforts to Reduce (initial generation), Stop (transboundary and transmedium transport), Recover (damages done), and Shift (toward a sustainable living style). The world needs more information to design these international collaborative actions into policies and regulations. The first step for this is to acknowledge transboundary and transmedium characteristics of microplastic pollution and to stipulate the term in global actions. Global acknowledgment will trigger the standardization of research protocols, scopes, and experimental specifications to enable more concrete and comparative research in related fields. Once research data and volume reach a threshold, the research interest will naturally expand to developing and less-developed countries, beyond currently interested governments, for global status mapping of microplastic pollution. Microplastic pollution, in short, calls for more comprehensive and comparative scientific data and analysis interpretation to be translated into effective and efficient policy drivers, another cornerstone toward a Sustainable Development Age to avoid future environmental crises.

Author contributions

HK designed most of the research analysis input, and HSR, JM, NAK, CY, JHY, and ML wrote the manuscript. HK review and editing. HK resources, project administration, and funding acquisition. ML, HSR, JM, NAK, CY, and JHY edited the manuscript. All authors have read and agreed to the published version of the manuscript.

Declaration of conflicting interests


The author(s) declared no potential conflicts of interest with respect to the research, authorship, and/or publication of this article.


Funding


This research was funded by the Korea Environment Industry & Technology Institute (KEITI) through the Measurement and Risk Assessment Program for the Management of Microplastics Program, funded by the Korea Ministry of Environment (MOE) (Grant No. 2020003110010); Korea Ministry of Environment, with the strategic EcoSSSoil Project (Grant No. 2019002820004); and the Basic Science Research Program through the National Research Foundation of Korea (NRF) funded by the Ministry of Education (Grant No. 2019R111A2A01057002 and 2019R1A61A03033167).

ORCID iDs

Minha Lee  <https://orcid.org/0000-0001-8991-7189>

Heejung Kim  <https://orcid.org/0000-0003-2652-3160>

Jinah Moon  <https://orcid.org/0000-0002-1684-4254>

Naing Aung Khant  <https://orcid.org/0000-0002-6953-3601>

References

1. Science Museum. The age of plastic: from parkesine to pollution. <https://www.sciencemuseum.org.uk/objects-and-stories/chemistry/age-plastic-parkesine-pollution> (2019, accessed 14 May, 2022).
2. Lee M and Kim H. COVID-19 pandemic and microplastic pollution. *Nanomaterials* 2022; 12: 51.
3. National Geographic Society. School educational encyclopedic entry on “Great Pacific Garbage Patch”. 2022. <https://www.nationalgeographic.org/encyclopedia/great-pacific-garbage-patch> (accessed 14 May 2022).
4. International Union for Conservation of Nature (IUCN) Issues Briefs. Marine plastic pollution review. <https://www.iucn.org/resources/issues-briefs/marine-plastic-pollution> (2021, accessed 14 May 2022).
5. Welden NA and Lusher A. Microplastics: From origin to impacts. In *Plastic waste and recycling*. Cambridge, MA: Academic Press, 2020, pp.223–249.
6. Andrady AL. Microplastics in the marine environment. *Mar Pollut Bull* 2011; 62: 1596–1605.
7. Barboza LG and Gimenez BC. Microplastics in the marine environment: current trends and future perspectives. *Mar Pollut Bull* 2015; 97: 5–12.
8. Cressy D. The plastic ocean: scientists know that there is a colossal amount of plastic in the oceans. But they don’t know where it all is, what it looks like or what damage it does. *Nature* 2016; 536: 263–265.
9. Macfadyen G, Huntington T and Cappell R. Abandoned, lost or otherwise discarded fishing gear. Report, Courtesy of the National Oceanic and Atmospheric Administration, USA, January 2009.
10. Eriksen M, Lebreton LC, Carson HS, et al. Plastic pollution in the world’s oceans: more than 5 trillion plastic pieces weighing over 250,000 tons afloat at sea. *PLoS ONE*. December 10, 2014. DOI: 10.1371/journal.pone.0111913.
11. Rogers KL, Carreres-Calabuig JA, Gorokhova E, et al. Micro-by-micro interactions: how microorganisms influence the fate of marine microplastics. *Limnol Oceanogr Lett* 2020; 5: 18–36.
12. Pedrotti ML, Petit S, Elineau A, et al. Changes in the floating plastic pollution of the Mediterranean Sea in relation to the distance to land. *PLOS ONE*. August 24, 2016. DOI: 10.1371/journal.pone.0161581.
13. Pelamatti T, Fonseca-Ponce IA, Rios-Mendoza LM, et al. Seasonal variation in the abundance of marine plastic debris in Banderas Bay, Mexico. *Mar Pollut Bull* 2019; 145: 604–610.
14. Lindeque PK, Cole M, Coppock RL, et al. Are we underestimating microplastic abundance in the marine environment? A comparison of microplastic capture with nets of different mesh-size. *Environ Pollut* 2020; 265: 114721.
15. Napper IE and Thompson RC. Plastic debris in the marine environment: history and future challenges. *Global Chall* 2020; 4: 1900081.
16. Carpenter EJ, Anderson SJ, Harvey GR, et al. Polystyrene spherules in coastal waters. *Science* 1972; 178: 749–750.

17. Carpenter EJ and Smith KL. Plastics on the sargasso sea surface. *Science* 1972; 175: 1240–1241.
18. Morris AW and Hamilton EI. Polystyrene spherules in the Bristol channel. *Mar Pollut Bull* 1974; 5: 26–27.
19. Thompson RC, Olsen Y, Mitchell RP, et al. Lost at sea: where is all the plastic? *Science* 2004; 304: 838–838.
20. Andrady A. Fate of plastics debris in the marine environment. In: Arthur C, Baker J and Bamford H (eds.), *NOAA marine debris program*, WA, USA, 9–11 Sept 2008, pp.78. US: NOAA.
21. Rillig MC. Microplastic in terrestrial ecosystems and the soil? *Environ Sci Technol* 2012; 46: 6453–6454.
22. Rillig MC, Ingraffia R and de Souza Machado AA. Microplastic incorporation into soil in agroecosystems. *Front Plant Sci* 2017; 8: 8–11.
23. Chae Y and An YJ. Current research trends on plastic pollution and ecological impacts on the soil ecosystem: a review. *Environ Pollut* 2018; 240: 387–395.
24. Horton AA, Walton A, Spurgeon DJ, et al. Microplastics in freshwater and terrestrial environments: evaluating the current understanding to identify the knowledge gaps and future research priorities. *Sci Total Environ* 2017; 586: 127–141.
25. Hurley RR and Nizzetto L. Fate and occurrence of micro(nano)plastics in soils: knowledge gaps and possible risks. *Curr Opin Environ Sci Health* 2018; 1: 6–11.
26. Mai L, Bao LJ, Wong CS, et al. Microplastic contamination in aquatic environments: an emerging matter of environmental urgency. In: Zeng E. Y. (ed) *Microplastics in the terrestrial environment*. Amsterdam: Elsevier, 2018, pp.365–378.
27. Ng EL, Huerta Lwanga E, Eldridge SM, et al. An overview of microplastic and nanoplastic pollution in agroecosystems. *Sci Total Environ* 2018; 627: 1377–1388.
28. UNEP. Plastic planet: how tiny plastic particles are polluting our soil. <https://www.unenvironment.org/news-and-stories/story/plastic-planet-how-tinyplasticparticles-are-polluting-our-soil> (2018, accessed 28 May 2022).
29. Liu EK, He WQ and Yan CR. ‘White revolution’ to ‘white pollution’ agricultural plastic film mulch in China. *Environ Res Lett*. 2014; 9: 091001.
30. Rochman CM, Kross SM, Armstrong JB, et al. Scientific evidence supports a ban on microbeads. *Environ Sci Technol* 2015; 49: 10759–10761.
31. Nizzetto L, Futter M and Langaas S. Are agricultural soils dumps for microplastics of urban origin? *Environ Sci Technol*. 2016; 50: 10777–10779.
32. Nizzetto L, Bussi G, Futter MN, et al. A theoretical assessment of microplastic transport in river catchments and their retention by soils and river sediments. *Environ Sci Process Impacts* 2016; 18: 1050–1059.
33. Carr SA, Liu J and Tesoro AG. Transport and fate of microplastic particles in wastewater treatment plants. *Water Res*. 2016; 91: 174–182.
34. Mason SA, Garneau D, Sutton R, et al. Microplastic pollution is widely detected in US municipal wastewater treatment plant effluent. *Environ Pollut* 2016; 218: 1045–1054.
35. McCormick AR, Hoellein TJ, London MG, et al. Microplastic in surface waters of urban rivers: concentration, sources, and associated bacterial assemblages. *Ecosphere* 2016; 7. DOI: 10.1002/ecs2.1556.
36. Talvitie J, Mikola A, Setälä O, et al. How well is microliter purified from wastewater? – a detailed study on the stepwise removal of microliter in a tertiary level wastewater treatment plant. *Water Res* 2017; 2017: 164–172.
37. Ziajahromi S, Neale PA, Rintoul L, et al. Wastewater treatment plants as a pathway for microplastics: development of a new approach to sample wastewater-based microplastics. *Water Res* 2017; 112: 93–99.

38. Blasing M and Amelung W. Plastics in soil: analytical methods and possible sources. *Sci Total Environ* 2018; 612: 422–435.
39. Dubaish F and Liebezeit G. Suspended microplastics and black carbon particles in the jade system, southern north sea. *Water Air Soil Pollut* 2013; 224: 1352.
40. Foitzik MJ, Unrau HJ, Gauterin F, et al. Investigation of ultra fine particulate matter emission of rubber tires. *Wear* 2018; 394: 87–95.
41. Wagner S, Hüffer T, Klöckner P, et al. Tire wear particles in the aquatic environment – a review on generation, analysis, occurrence, fate and effects. *Water Res* 2018; 139: 83–100.
42. Dris R, Gasperi J, Saad M, et al. Synthetic fibers in atmospheric fallout: a source of microplastics in the environment? *Mar Pollut Bull* 2016; 104: 290–293.
43. Nizzetto L. Do microplastics spill on to farm soils? *Nature* 2016; 537: 88.
44. Rillig MC, Ziersch L and Hempel S. Microplastic transport in soil by earthworms. *Sci Rep* 2017; 7: 1362.
45. Kasirajan S and Ngouajio M. Polyethylene and biodegradable mulches for agricultural applications: a review. *Agron Sustain Dev* 2012; 32: 501–529.
46. Li C, Moore-Kucera J, Lee J, et al. Effects of biodegradable mulch on soil quality. *Appl Soil Ecol* 2014; 79: 59–69.
47. Farmer J, Zhang B, Jin X, et al. Long-term effect of plastic film mulching and fertilization on bacterial communities in a brown soil revealed by high through-put sequencing. *Arch Agron Soil Sci.* 2017; 63: 230–241.
48. Sintim HY and Flury M. Is biodegradable plastic mulch the solution to agriculture’s plastic problem? *Environ Sci Technol* 2017; 51: 1068–1069.
49. William JL. Plastic mulches for the production of vegetable crops. *Hort Technol* 1993; 3: 35–39.
50. Steinmetz Z, Wollmann C, Schaefer M, et al. Plastic mulching in agriculture. Trading short-term agronomic benefits for long-term soil degradation? *Sci Total Environ* 2016; 550: 690–705.
51. Ramos L, Berenstein G, Hughes EA, et al. Polyethylene film incorporation into the horticultural soil of small periurban production units in Argentina. *Sci Total Environ* 2015; 523: 74–81.
52. Bergmann M, Mutzel S, Primpke S, et al. White and wonderful? Microplastics prevail in snow from the Alps to the Arctic. *Sci Adv* 2019; 5: 57.
53. Awet TT, Kohl Y, Meier F, et al. Effects of polystyrene nanoparticles on the microbiota and functional diversity of enzymes in soil. *Environ Sci Eur* 2018; 30: 11.
54. Machado AA dS, Lau CW, Till J, et al. Impacts of microplastics on the soil biophysical environment. *Environ Sci Technol* 2018; 52: 9656–9665.
55. Zhu BK, Fang YM, Zhu D, et al. Exposure to nanoplastics disturbs the gut microbiome in the soil oligochaete *Enchytraeus crypticus*. *Environ Pollut* 2018; 239: 408–415.
56. Wan Y, Wu C, Xue Q, et al. Effects of plastic contamination on water evaporation and desiccation cracking in soil. *Sci Total Environ* 2019; 654: 576–582.
57. Bosker T, Bouwman LJ, Brun NR, et al. Microplastics accumulate on pores in seed capsule and delay germination and root growth of the terrestrial vascular plant *Lepidium sativum*. *Chemosphere* 2019; 226: 774–781.
58. Qi Y, Yang X, Pelaez AM, et al. Macro- and micro-plastics in soil-plant system: effects of plastic mulch film residues on wheat (*Triticum aestivum*) growth. *Sci Total Environ* 2018; 645: 1048–1056.
59. Jiang X, Chen H, Liao Y, et al. Ecotoxicity and genotoxicity of polystyrene microplastics on higher plant *Vicia faba*. *Environ Pollut* 2019; 250: 831–838.
60. Bandmann V, Müller JD, Köhler T, et al. Uptake of fluorescent nano beads into BY2-cells involves clathrin-dependent and clathrin-independent endocytosis. *FEBS Lett* 2012; 586: 3626–3632.
61. Li L, Zhou Q, Yin N, et al. Uptake and accumulation of microplastics in an edible plant. *Chin Sci Bull* 2019; 64: 928–934.

62. Ebere EC, Wirnkor VA and Ngozi VE. Uptake of microplastics by plant: a reason to worry or to be happy? *World Sci News* 2019; 131: 256–267.
63. Huerta Lwanga E, Mendoza Vega J, Ku Quej V, et al. Field evidence for transfer of plastic debris along a terrestrial food chain. *Sci Rep* 2017; 7: 14071.
64. European Commission. Directive 2006/118/EC of the European Parliament and of the ... - eur-lex. 2006. https://eur-lex.europa.eu/legal_content/EN/LSU/?uri=celex%3A32006L0118 (accessed 16 July 2022).
65. Shi J, Dong Y, Shi Y, et al. Groundwater antibiotics and microplastics in a drinking-water source area, northern China: occurrence, spatial distribution, risk assessment, and correlation. *Environ Res* 2022; 210: 112855.
66. Panno SV, Kelly WR, Scott J, et al. Microplastic contamination in karst groundwater systems. *Groundwater* 2019; 57: 189–196.
67. Samandra S, Johnston JM, Jaeger JE, et al. Microplastic contamination of an unconfined groundwater aquifer in Victoria, Australia. *Sci Total Environ* 2022; 802: 149727.
68. Viaroli S, Lancia M and Re V. Microplastics contamination of groundwater: current evidence and future perspectives: a review. *Sci Total Environ* 2022; 824: 153851.
69. Padervand M, Lichtfouse E, Robert D, et al. Removal of microplastics from the environment: a review. *Environ Chem Lett* 2020; 18: 807–828.
70. Walker TR. (Micro)plastics and the UN sustainable development goals. *Curr Opin Green Sustain Chem* 2021; 30: 100497.
71. Shruti VC, Jonathan MP, Rodriguez-Espinosa PF, et al. Microplastics in freshwater sediments of atoyac river basin, Puebla city, Mexico. *Sci Total Environ* 2019; 654: 154–163.
72. Liu Y, Zhang JD, Cai CY, et al. Occurrence and characteristics of microplastics in the haihe river: an investigation of a seagoing river flowing through a megacity in northern China. *Environ Pollut* 2020; 262: 114261.
73. Yin L, Wen X, Du C, et al. Comparison of the abundance of microplastics between rural and urban areas: a case study from East Dongting lake. *Chemosphere* 2020; 244: 125486.
74. Simon M, van Alst N and Vollertsen J. Quantification of microplastic mass and removal rates at wastewater treatment plants applying focal plane array (fpa)-based fourier transform infrared (FT-IR) imaging. *Water Res* 2018; 142: 9.
75. Barchiesi M, Chiavola A, Di Marcantonio C, et al. Presence and fate of microplastics in the water sources: focus on the role of wastewater and drinking water treatment plants. *J Water Proc Eng* 2021; 40: 101787.
76. Bradford SA and Leij FJ. Modeling the transport and retention of polydispersed colloidal suspensions in porous media. *Chem Eng Sci* 2018; 192: 972–980.
77. Dong Z, Qiu Y, Zhang W, et al. Size-dependent transport and retention of micron-sized plastic spheres in natural sand saturated with seawater. *Water Res* 2018; 143: 518–526.
78. Jin G, Jiang Q, Tang H, et al. Density effects on nanoparticle transport in the hyporheic zone. *Adv Water Res* 2018; 121: 406–418.
79. Chu X, Li T, Li Z, et al. Transport of microplastic particles in saturated porous media. *Water* 2019; 11: 2474.
80. Li J, Song Y and Cai Y. Focus topics on microplastics in soil: analytical methods, occurrence, transport, and ecological risks. *Environ Pollut* 2020; 257: 113570.
81. He L, Rong H, Li M, et al. Bacteria have different effects on the transport behaviors of positively and negatively charged microplastics in porous media. *J Hazard Mater* 2021; 415: 125550.
82. Ryu H-S, Moon J, Kim H, et al. Modeling and parametric simulation of microplastic transport in groundwater environments. *Appl Sci* 2021; 11: 7189.
83. Niu L, Li Y, Li Y, et al. New insights into the vertical distribution and microbial degradation of microplastics in urban river sediments. *Water Res* 2021; 188: 116449.

84. Plastics Europe. *Plastics - the facts 2020*. Brussels: Plastics Europe, December 2020. <https://plasticseurope.org/knowledge-hub/plastics-the-facts-2020/> (2022, accessed 10 June 2022).
85. Momot O, Synzynys B, Kozmin G, et al. Pollution of ground sources of drinking water with technogenic tritium. In: Hlavinek P, Bonacci O, Marsalek J and Mahrikova I (eds) *Dangerous pollutants (xenobiotics) in urban water cycle*. NATO Science for Peace and Security Series. Dordrecht: Springer, 2008, pp.331–340. DOI: 10.1007/978-1-4020-6795-2_31.
86. Huang J, Chen H, Zheng Y, et al. Microplastic pollution in soils and groundwater: characteristics, analytical methods and impacts. *Chem Eng J* 2021; 425: 131870.
87. Oladoja NA and Unuabonah IE. The pathways of microplastics contamination in raw and drinking water. *J Water Proc Eng* 2021; 41: 102073.
88. Qi R, Jones DL, Li Z, et al. Behavior of microplastics and plastic film residues in the soil environment: a critical review. *Sci Total Environ* 2020; 703: 134722.
89. Chia RW, Lee J-Y, Kim H, et al. Microplastic pollution in soil and groundwater: a review. *Environ Chem Lett* 2021; 19: 4211–4224.
90. Dris R, Gasperi J, Rocher V, et al. Microplastic contamination in an urban area: a case study in greater Paris. *Environ Chem* 2015; 12: 592–599.
91. Dris R, Gasperi J, Mirande C, et al. A first overview of textile fibers, including microplastics, in indoor and outdoor environments. *Environ Pollut* 2017; 221: 453–458.
92. Allen S, Allen D, Phoenix VR, et al. Atmospheric transport and deposition of microplastics in a remote mountain catchment. *Nat Geosci* 2019; 12: 339–344.
93. Allen S, Allen D, Baladima F, et al. Evidence of free tropospheric and long-range transport of microplastic at pic du midi observatory. *Nat Commun* 2021; 12: 1–10.
94. Ferrero L, Scibetta L, Markuszewski P, et al. Airborne and marine microplastics from an oceanographic survey at the baltic sea: an emerging role of air-sea interaction? *Sci Total Environ* 2022; 824: 153–709.
95. Takahashi N and Hayasaka T. Air–sea interactions among oceanic low-level cloud, sea surface temperature, and atmospheric circulation on an intraseasonal time scale in the summertime north pacific based on satellite data analysis. *J Clim* 2020; 33: 9195–9212.
96. Gouin T, Roche N, Lohmann R, et al. Thermodynamic approach for assessing the environmental exposure of chemicals absorbed to microplastic. *Environ Sci Technol* 2011; 45: 1466–1472.
97. Xumiao L, Prata JC, Alves JR, et al. Airborne microplastics and fibers in indoor residential environments in Aveiro, Portugal. *Environ Adv* 2021; 6: 100–134.
98. Cai L, Wang J, Peng J, et al. Characteristic of microplastics in the atmospheric fallout from Dongguan City, China: preliminary research and first evidence. *Environ Sci Pollut Res* 2017; 24: 24928–24935.
99. Wright SL and Kelly FJ. Plastic and human health: a micro issue? *Environ Sci Technol* 2017; 51: 6634–6647.
100. Klein M and Fischer EK. Microplastic abundance in atmospheric deposition within the metropolitan area of Hamburg, Germany. *Sci Total Environ* 2019; 685: 96–103.
101. Allen S, Allen D, Moss K, et al. Examination of the ocean as a source for atmospheric microplastics. *PLoS One* 2020; 15: e0232746.
102. Liu K, Wang X, Fang T, et al. Source and potential risk assessment of suspended atmospheric microplastics in Shanghai. *Sci Total Environ* 2019; 675: 462–471.
103. Abbasi S, Rezaei M, Ahmadi F, et al. Atmospheric transport of microplastics during a dust storm. *Chemosphere* 2022; 292: 133–456.
104. Catarino AI, Macchia V, Sanderson WG, et al. Low levels of microplastics (MP) in wild mussels indicate that MP ingestion by humans is minimal compared to exposure via household fibers fallout during a meal. *Environ Pollut* 2018; 237: 675–684.

105. Hatinoğlu MD and Sanin FD. Sewage sludge as a source of microplastics in the environment: a review of occurrence and fate during sludge treatment. *J Environ Manag* 2021; 295: 113–028.
106. Munyaneza J, Jia Q, Qaraah FA, et al. A review of atmospheric microplastics pollution: in-depth sighting of sources, analytical methods, physiognomies, transport and risks. *Sci Total Environ* 2022; 153–339.
107. Chen G, Fu Z, Yang H, et al. An overview of analytical methods for detecting microplastics in the atmosphere. *TrAC, Trends Anal Chem* 2020; 130: 115–981.
108. Chen G, Feng Q, Wang J, et al. Mini-review of microplastics in the atmosphere and their risks to humans. *Sci Total Environ* 2020; 703: 135–504.
109. Zhang Y, Kang S, Allen S, et al. Atmospheric microplastics: a review on current status and perspectives. *Earth Sci Rev* 2020; 203: 103–118.
110. Sridharan S, Kumar M, Singh L, et al. Microplastics as an emerging source of particulate air pollution: a critical review. *J Hazard Mater* 2021; 418: 126–245.
111. Prata JC. Airborne microplastics: consequences to human health? *Environ Pollut* 2018; 234: 115–126.
112. Amato-Lourenço LF, dos Santos Galvão L, de Weger LA, et al. An emerging class of air pollutants: potential effects of microplastics to respiratory human health? *Sci Total Environ* 2020; 749: 141–676.
113. Torres-Agullo A, Karanasiou A, Moreno T, et al. Overview on the occurrence of microplastics in air and implications from the use of face masks during the COVID-19 pandemic. *Sci Total Environ* 2022; 800: 149–555.
114. Liebezeit G and Liebezeit E. Non-pollen particulates in honey and sugar. *Food Addit Contam A* 2013; 30: 2136–2140.
115. Liebezeit G and Liebezeit E. Synthetic particles as contaminants in German beers. *Food Addit Contam A* 2014; 31: 1574–1578.
116. Laist DW. Overview of the biological effects of lost and discarded plastic debris in the marine environment. *Mar Pollut Bull* 1987; 18: 319–326.
117. Setälä O, Fleming-Lehtinen V and Lehtiniemi M. Ingestion and transfer of microplastics in the planktonic food web. *Environ Pollut* 2014; 185: 77–83.
118. Nelms SE, Galloway TS, Godley BJ, et al. Investigating microplastic trophic transfer in marine top predators. *Environ Pollut* 2018; 238: 999–1007.
119. Wang T, Hu M, Xu G, et al. Microplastic accumulation via trophic transfer: can a predatory crab counter the adverse effects of microplastics by body defence? *Sci Total Environ* 2021; 754: 142099.
120. Hart MW. Particle captures and the method of suspension feeding by echinoderm larvae. *Biol Bull* 1991; 180: 12–27.
121. Graham ER and Thompson JT. Deposit- and suspension-feeding sea cucumbers (echinodermata) ingest plastic fragments. *J Exp Mar Biol Ecol* 2009; 368: 22–29.
122. Murray F and Cowie PR. Plastic contamination in the decapod crustacean nephrops norvegicus (linnaeus, 1758). *Mar Pollut Bull* 2011; 62: 1207–1217.
123. Cole M, Lindeque P, Fileman E, et al. Microplastic ingestion by Zooplankton. *Environ Sci Technol* 2013; 47: 6646–6655.
124. Lee K-W, Shim WJ, Kwon OY, et al. Size-dependent effects of micro polystyrene particles in the marine copepod tigrionus japonicus. *Environ Sci Technol* 2013; 47: 11278–11283.
125. Van Cauwenbergh L and Janssen CR. Microplastics in bivalves cultured for human consumption. *Environ Pollut* 2014; 193: 65–70.
126. Rochman CM, Tahir A, Williams SL, et al. Anthropogenic debris in seafood: plastic debris and fibers from textiles in fish and bivalves sold for human consumption. *Sci Rep*. September 24, 2015. DOI: 10.1038/srep14340.

127. Yang D, Shi H, Li L, et al. Microplastic pollution in table salts from China. *Environ Sci Technol* 2015; 49: 13622–13627.
128. Karami A, Golieskardi A, Ho YB, et al. Microplastics in eviscerated flesh and excised organs of dried fish. *Sci Rep*. Epub ahead of print 2017. DOI: 10.1038/s41598-017-05828-6.
129. Li J, Green C, Reynolds A, et al. Microplastics in mussels sampled from coastal waters and supermarkets in the United Kingdom. *Environ Pollut* 2018; 241: 35–44.
130. Kosuth M, Mason SA and Wattenberg EV. Anthropogenic contamination of tap water, beer, and Sea Salt. *PLoS One*. April 11, 2018. DOI: 10.1371/journal.pone.0194970.
131. Da Costa Filho PA, Andrey D, Eriksen B, et al. Detection and characterization of small-sized microplastics ($\geq 5 \mu\text{m}$) in milk products. *Sci Rep*. December 15, 2021. DOI: 10.1038/s41598-021-03458-7.
132. Mintenig SM, Löder MGJ, Primpke S, et al. Low numbers of microplastics detected in drinking water from ground water sources. *Sci Total Environ* 2019; 648: 631–635.
133. Shruti VC, Pérez-Guevara F, Elizalde-Martínez I, et al. First study of its kind on the microplastic contamination of soft drinks, cold tea and energy drinks - future research and environmental considerations. *Sci Total Environ* 2020; 726: 138580.
134. Hernandez LM, Xu EG, Larsson HC, et al. Plastic teabags release billions of microparticles and nanoparticles into tea. *Environ Sci Technol* 2019; 53: 12300–12310.
135. Kedzierski M, Lechat B, Sire O, et al. Microplastic contamination of packaged meat: occurrence and associated risks. *Food Packag Shelf Life* 2020; 24: 100489.
136. Fadare OO, Wan B, Guo L-H, et al. Microplastics from consumer plastic food containers: are we consuming it? *Chemosphere* 2020; 253: 126787.
137. Hee YY, Weston K and Suratman S. The effect of storage conditions and washing on microplastic release from food and drink containers. *Food Packag Shelf Life* 2022; 32: 100826.
138. Mato Y, Isobe T, Takada H, et al. Plastic resin pellets as a transport medium for toxic chemicals in the marine environment. *Environ Sci Technol* 2001; 35: 318–324.
139. Teuten EL, Saquing JM, Knappe DR, et al. Transport and release of chemicals from plastics to the environment and to wildlife. *Philos Trans R Soc B* 2009; 364: 2027–2045.
140. Rochman CM, Hentschel BT and Teh SJ. Long-term sorption of metals is similar among plastic types: Implications for plastic debris in aquatic environments. *PLoS One*. January 15, 2014. DOI: 10.1371/journal.pone.0085433.
141. McCormick A, Hoellein TJ, Mason SA, et al. Microplastic is an abundant and distinct microbial habitat in an urban river. *Environ Sci Technol* 2014; 48: 11863–11871.
142. Schymanski D, Goldbeck C, Humpf H-U, et al. Analysis of microplastics in water by micro-Raman spectroscopy: release of plastic particles from different packaging into mineral water. *Water Res* 2018; 129: 154–162.
143. Öbmann BE, Sarau G, Holtmannspötter H, et al. Small-sized microplastics and pigmented particles in bottled mineral water. *Water Res* 2018; 141: 307–316.
144. Reisser J, Shaw J, Wilcox C, et al. Marine plastic pollution in waters around Australia: characteristics, concentrations, and pathways. *PLoS One*. November 27, 2013. DOI: 10.1371/journal.pone.0080466.
145. Everaert G, De Rijcke M, Lonneville B, et al. Risks of floating microplastic in the global ocean. *Environ Pollut* 2020; 267: 115499.
146. Seung-Kyu K and In-Sung K. Seasonal distribution characteristics of microplastics in the sand beach of the Daebu Island, Gyeonggi-do. *J Environ Anal, Health Toxicol* 2015; 18: 221–231.
147. Piehl S, Leibner A, Löder MGJ, et al. Identification and quantification of macro- and microplastics on an agricultural farmland. *Sci Rep* 2018; 8: 17950.
148. Liu M, Lu S, Song Y, et al. Microplastic and meso-plastic pollution in farmland soils in suburbs of Shanghai, China. *Environ Pollut* 2018; 242: 855–862.

149. Corradini F, Meza P, Eguiluz R, et al. Evidence of microplastic accumulation in agricultural soils from sewage sludge disposal. *Sci Total Environ* 2019; 671: 411–420.
150. Lv W, Zhou W, Lu S, et al. Microplastic pollution in rice-fish co-culture system: a report of three farmland stations in Shanghai, China. *Sci Total Environ* 2019; 652: 1209–1218.
151. Selvam S, Jesuraja K, Venkatramanan S, et al. Hazardous microplastic characteristics and its role as a vector of heavy metal in groundwater and surface water of coastal south India. *J Hazard Mater* 2021; 402: 123786.
152. Poleć M, Aleksander-Kwaterczak U, Wątor K, et al. The occurrence of microplastics in freshwater systems – preliminary results from Krakow (Poland). *Geol Geophys Environ* 2018; 44: 91.
153. Abbasi S, Keshavarzi B, Moore F, et al. Distribution and potential health impacts of microplastics and micro-rubbers in air and street dusts from Asaluyeh county, Iran. *Environ Pollut* 2019; 244: 153–164.
154. Wright SL, Ulke J, Font A, et al. Atmospheric microplastic deposition in an urban environment and an evaluation of transport. *Environ Int* 2020; 136: 105–411.
155. Choi H, Lee I, Kim H, et al. Comparison of microplastic characteristics in the indoor and outdoor air of urban areas of South Korea. *Water Air Soil Pollut* 2022; 233: 1–10.
156. Iñiguez ME, Conesa JA and Fullana A. Microplastics in Spanish table salt. *Sci Rep*. August 17, 2017. DOI: 10.1038/s41598-017-09128-x.
157. Enyoh CE, Verla AW, Verla EN, et al. Airborne microplastics: a review study on method for analysis, occurrence, movement and risks. *Environ Monit Assess* 2019; 191: 1–17.
158. Panko JM, Chu J, Kreider ML, et al. Measurement of airborne concentrations of tire and road wear particles in urban and rural areas of France, Japan, and the United States. *Atmos Environ* 2013; 72: 192–199.
159. Kooi M, Besseling E, Kroeze C, et al. Modeling the fate and transport of plastic debris in freshwaters. *Rev Guidance* 2018; 125–152.
160. Lewis ER, Lewis R, Lewis ER, et al. Sea salt aerosol production: mechanisms, methods, measurements, and models. In *Geophysical Monograph Series*. Washington, DC: American Geophysical Union, Volume 152, 2004.
161. Trainic M, Flores JM, Pinkas I, et al. Airborne microplastic particles detected in the remote marine atmosphere. *Commun Earth Environ* 2020; 1: –9.
162. Van der Does M, Knippertz P, Zschenderlein P, et al. The mysterious long-range transport of giant mineral dust particles. *Sci Adv* December 12, 2018; 4(12). DOI: 10.1126/sciadv.aau2768.
163. Frei S, Piehl S, Gilfedder BS, et al. Occurrence of microplastics in the hyporheic zone of rivers. *Sci Rep* October 24, 2019. DOI: 10.1038/s41598-019-51741-5.
164. Li M, Zhang M, Rong H, et al. Transport and deposition of plastic particles in porous media during seawater intrusion and groundwater-seawater displacement processes. *Sci Total Environ* 2021; 781: 146752.
165. Lönnstedt OM and Eklöv P. Environmentally relevant concentrations of microplastic particles influence larval fish ecology. *Science* 2016; 352: 1213–1216.
166. de Souza Machado AA, Lau CW, Kloas W, et al. Microplastics can change soil properties and affect plant performance. *Environ Sci Technol* 2019; 53: 6044–6052.
167. Maaß S, Daphi D, Lehmann A, et al. Transport of microplastics by two collembolan species. *Environ Pollut* 2017; 225: 456–459.
168. Rillig MC, Ingraffia R and Machado AA. Microplastic incorporation into soil in agroecosystems. *Front Plant Sci* 2017b; 8: 1805.
169. Zhu D, Bi QF, Xiang Q, et al. Trophic predator-prey relationships promote transport of microplastics compared with the single hypoaspis aculeifer and folsomia candida. *Environ Pollut* 2018; 235: 150–154.

170. Lusher AL, McHugh M and Thompson RC. Occurrence of microplastics in the gastrointestinal tract of pelagic and demersal fish from the English channel. *Mar Pollut Bull* 2013; 67: 94–99.
171. Perry JS. Atmospheric transport and dispersion of air pollutants associated with vehicular emissions. In: Watson AY, Bates RR and Kennedy D (eds) *Air pollution, the automobile, and public health*. Washington DC: National Academies Press (US), 1988, pp. 77–98.
172. Evangelidou N, Grythe H, Klimont Z, et al. Atmospheric transport is a major pathway of microplastics to remote regions. *Nat Commun* 2020; 11: 1–11.
173. Peng S, Wu D, Ge Z, et al. Influence of graphene oxide on the transport and deposition behaviors of colloids in saturated porous media. *Environ Pollut* 2017; 225: 141–149.
174. Intergovernmental Oceanographic Commission of UNESCO (IOC-UNESCO). UNESCO and the Ocean Decade Community rally at 2022 UN Ocean Conference to advocate for science-driven ocean action. <https://ioc.unesco.org/> (2022, accessed 25 July 2022).

Author biographies

Minha Lee received her PhD from Korea University in 2021. She is a post-Doc researcher at the Kangwon National University conducting researches on emerging environmental issues like water security, microplastic pollution, and climate change.

Heejung Kim is an associate professor in the Department of Geology at Kangwon National University, focusing on groundwater resources and environment.

Han-Sun Ryu is a Ph.D candidate in the Department of Geology at Kangwon National University, Korea. She mainly researched on hydrochemistry in karst area, the relationship between earthquake and groundwater, and the change in groundwater level in deep geothermal wells.

Jinah Moon is a PhD candidate in the Department of Geology at the Kangwon National University. Her current research interest is in geomicrobiology and environment.

Naing Aung Khant is a PhD candidate in Department of Geology at Kangwon National University. His research mainly focuses on microplastic pollution in the environments and environmental geology.

Chaerim Yu is a Master candidate in the Department of Geology at Kangwon National University, Korea. Her current research interest is in geomicrobiology.

Ji-Hee Yu is a Master candidate in Department of Geology at Kangwon National University, Korea. Her current research interest is in geological heritage.