

Pesticide Exposures during Pregnancy and Health Outcomes in Latin America and the Caribbean: A Scoping Review of Human Observational Studies

Ricardo Rohweder,^{1,2} Shirley Salcedo Arteaga,^{1,3} Vithória Luz da Silva Gomes,^{2,4} Paulo Alfredo Casanova Schulze,⁵ and Lavinia Schuler-Faccini^{1,2} 

¹Programa de Pós-Graduação em Genética e Biologia Molecular, Universidade Federal do Rio Grande do Sul, Porto Alegre, Rio Grande do Sul, Brazil

²Sistema Nacional de Informação sobre Agentes Teratogênicos, Hospital de Clínicas de Porto Alegre, Porto Alegre, Rio Grande do Sul, Brazil

³Grupo de Investigaciones Biomedicas y Biología Molecular, Universidad del Sinú Elias Bechara Zainum, Montería, Colombia

⁴Faculdade de Medicina, Universidade Federal do Rio Grande do Sul, Porto Alegre, Rio Grande do Sul, Brazil

⁵Escola de Medicina, Pontifícia Universidade Católica do Rio Grande do Sul, Porto Alegre, Rio Grande do Sul, Brazil

BACKGROUND: Latin America and the Caribbean (LAC) are regions with intense pesticide use. Numerous studies have demonstrated the adverse health effects associated with pesticide exposure. The embryonic and fetal periods are particularly susceptible to xenobiotics, with pesticides exhibiting potentially teratogenic effects.

OBJECTIVES: The objective was to review the scientific literature on outcomes associated with prenatal pesticide exposure, identifying challenges and gaps in this field.

METHODS: We conducted a scoping review using terms related to pesticides, LAC, and pregnancy across six databases. The final search was conducted on 5 March 2024. The inclusion criteria for the studies were as follows: *a*) being human observational studies involving pregnant women of any age or gestational age duration, newborns from these pregnancies, or both; *b*) reporting any exposure to pesticide and any adverse outcome; *c*) having been conducted in any country of LAC; *d*) having been published between 1 January 2000 and 5 March 2024; *e*) having the full text available in English, Spanish, or Portuguese; and *f*) presenting original results. Nonoriginal research papers, such as qualitative studies, reviews, critical analyses, and opinion papers, were excluded. The included studies were categorized and presented based on the outcomes they evaluated.

RESULTS: We included 80 studies conducted in 13 countries; the included studies encompassed obstetric outcomes, anthropometric parameters, congenital anomalies, neurodevelopment, respiratory infections, and childhood leukemia, as well as molecular effects. Organochlorines were the primary type of exposure investigated among the included studies. Many studies relied on indirect measures of pesticide exposure.

DISCUSSION: Adverse outcomes associated with prenatal pesticide exposure have been observed in Latin American and Caribbean populations, consistent with the global literature. Significant knowledge gaps remain, especially regarding groups of pesticides other than organochlorines. Less than half of the countries in LAC have conducted any study on the potential effects of prenatal exposure. Ongoing research into the risks of prenatal exposure is imperative. It is essential to consider the region's unique characteristics, particularly when investigating the risks associated with pesticides authorized exclusively in this region. <https://doi.org/10.1289/JHP1043>

Introduction

Latin America and the Caribbean (LAC) are home to 660 million inhabitants and encompass a land surface of 2 billion hectares, of which 38% is dedicated to agriculture.^{1,2} The countries within LAC are major exporters of soybeans, pork, corn, poultry, animal feed, sugar, coffee, fruits, and vegetables.² Brazil, Argentina, and Mexico are this region's largest agricultural and food exporters.² Substantial agricultural growth has been observed over the past five decades.³ Most of this growth has resulted from productivity improvements rather than expanding the agricultural land under cultivation.⁴ At the same time, there has been an increase in the production and use of pesticides over the last two decades.^{4–6} A pesticide is defined as any substance, mixture of substances, or microorganisms intended for repelling, destroying, or controlling any pest.⁷ This includes vectors of human or animal disease and

unwanted species of plants or animals that cause harm or otherwise interfere with the production, processing, storage, transport, or marketing of food, agricultural commodities, wood, and wood products or animal feed.

Since the 1960s, the scientific community has closely examined the environmental impact of pesticides. Evidence has shown that pesticides can persist in soil and water, accumulate in invertebrate and vertebrate tissues, traverse food chains, and impact apex predators.⁸ An extensive body of literature examines the link between pesticide exposure and detrimental effects on human health, including type 2 diabetes,⁹ neurological disorders,¹⁰ cancer,^{11,12} respiratory diseases,¹³ and allergies.¹⁴ Given the teratogenic potential of xenobiotics for pregnant women, the impact of pesticides on this population has received considerable attention. Pesticide exposure during pregnancy has been associated with an increased risk of childhood leukemia.¹⁵ In addition, potential links to clinical manifestations of autism spectrum disorder have been explored,¹⁶ whereas findings regarding association with obesity remain inconclusive.¹⁷ Furthermore, there is a lack of consistency among the studies evaluating the association between maternal pesticide exposure and obstetric and neonatal outcomes.^{18,19} Moreover, a substantial number of studies assess the effects of prenatal pesticide exposure on congenital anomalies.^{20–32}

A recent scoping review investigated the current state of research on the health effects of pesticide exposure in populations from LAC.³³ That review underscored the potential health impacts of pesticide exposure, gathering findings pertaining to genotoxicity, neurobehavioral effects, cancer risks, reproduction, placental outcomes and teratogenicity, birth and child outcomes, and other effects. Moreover, it delineated significant methodological limitations within the scientific output on this subject. To address these limitations and research gaps within LAC, the

Address correspondence to Lavinia Schuler-Faccini. Email: lavinia.faccini@ufrgs.br

Supplemental Material is available online (<https://doi.org/10.1289/JHP1043>).

The authors declare they have no conflicts of interest related to this work to disclose.

Conclusions and opinions are those of the individual authors and do not necessarily reflect the policies or views of EHP Publishing or the National Institute of Environmental Health Sciences.

Received 4 December 2023; Revised 15 June 2024; Accepted 26 June 2024; Published 15 August 2024.

Note to readers with disabilities: *JHP* strives to ensure that all journal content is accessible to all readers. However, some figures and Supplemental Material published in *JHP* articles may not conform to 508 standards due to the complexity of the information being presented. If you need assistance accessing journal content, please contact jhp submissions@niehs.nih.gov. Our staff will work with you to assess and meet your accessibility needs within 3 working days.

authors advocated for increased investment in research and development, collaborative initiatives within the region, and enhanced methodological rigor in epidemiological studies. The aforementioned review has provided a comprehensive depiction of the research landscape, showing substantial heterogeneity among the included studies concerning exposure and outcomes. Considering the susceptibility of embryo and fetus development to xenobiotic agents throughout gestation and the adverse effects that exposure can induce, we conducted a scoping review further focused on prenatal pesticide exposure and adverse outcomes. In addition, our review expands the search to six databases, including one dedicated to studies specific to the LAC region, and extends the time frame to encompass publications on the theme over more than the last two decades. Our aim is to identify the main challenges and research gaps in pregnancy-related pesticide exposure and outcomes within LAC and offer insights for future studies.

Methods

The present study was conducted in accordance with the Preferred Reporting Items for Systematic Reviews and Meta-Analyses extension for Scoping Reviews (PRISMA-ScR).³⁴ A protocol was prepared prior to the identification step and this protocol has been registered within the Open Science Framework.³⁵

The selection of terms for the search strategy was based on the Population, Concept, and Context (PCC) framework,³⁶ and this search strategy was expanded to include related terms according to the Medical Subject Headings (MeSH) terms. The search strategy protocol, including all selected keywords and index terms, was adapted for each database or information source (see “Search Terms,” in the Supplemental Material). The development of the full search strategy protocol was facilitated by the assistance of a research librarian.

The following databases were searched to identify potentially relevant studies: Embase, PubMed, Scielo, Scopus, Virtual Health Library, and Web of Science. The first search was conducted on 14 November 2022. The final search was conducted on 5 March 2024. The inclusion criteria for the studies were as follows: *a*) being human observational studies involving pregnant women of any age or gestational age duration, newborns from these pregnancies, or both (population); *b*) reporting any exposure to pesticide and any adverse outcome (concept); *c*) having been conducted in any country of LAC (context), as the grouping of countries presented by the OECD-FAO document²; *d*) having been published between 1 January 2000, and 5 March 2024; *e*) having the full text available in English, Spanish, or Portuguese; and *f*) presenting original results. Nonoriginal research papers, such as qualitative studies, reviews, critical analyses, and opinion papers, were excluded. To capture the scientific literature concerning prenatal exposure to pesticides and its associated adverse outcomes, articles that did not gather data on exposure measured within the context of pregnancy were excluded. This decision was made because the target population of the study should consist of pregnant women, their newborns, or both. In addition, articles that did not directly assess adverse outcomes, such as those focusing on enzyme activity levels, were excluded.

All identified articles were collected and uploaded to Rayyan,³⁷ a web application that provides tools for conducting the initial screening of abstracts and titles during the article selection process. We used Rayyan’s tool to identify duplicate articles and then manually reviewed each possible duplicate identified, excluding those articles that were duplicate records. Subsequently, we conducted a test of titles and abstracts screening 20 articles, followed by a discussion to ensure consistent understanding of relevance among the reviewers. Finally, we screened the titles and abstracts of all articles. Each article was screened by two reviewers. Relevant

studies were imported into Mendeley.³⁸ The full text of each selected paper was then thoroughly assessed by the independent reviewers to further ensure compliance with the study’s inclusion criteria. Finally, the reviewers recorded the reasons why the excluded papers did not meet inclusion criteria.

Data extraction from the selected studies was performed by two independent reviewers using a standardized spreadsheet that included specific details about each study’s participants, concept, context, methods, and key findings relevant to the present study’s review questions. Disagreements between the reviewers at any stage of the selection process or data extraction were resolved by an additional reviewer or discussions between reviewers.

The selected studies were organized into tables based on their outcomes and included descriptions of the author’s name, year, exposure, outcomes, and key findings. In each table, studies were grouped by country of origin, with studies within each group presented chronologically by year of publication. Finally, a narrative summary was performed.

Results

The search across the six databases yielded 1,765 records in the first search and an additional 165 records were returned in the final search, resulting in a total of 1,930 records. After removing duplicates, 1,021 articles were then screened by examining titles and abstracts (see the screening step in Figure 1). From this screening, 853 articles were excluded, and we kept 168 articles for further analysis. In the next step, the full-text assessment, 88 more articles were excluded, resulting in 80 human studies being included in this scoping review.

We identified studies from 13 distinct countries: Mexico ($n = 26$), Brazil ($n = 14$), Guadeloupe ($n = 13$), Argentina ($n = 6$), Costa Rica ($n = 7$), Ecuador ($n = 4$), Puerto Rico ($n = 2$), Colombia ($n = 2$), Chile ($n = 2$), Bolivia ($n = 1$), Paraguay ($n = 1$), Peru ($n = 1$), and Guatemala ($n = 1$) (Figure 2A). It is worth noting that we identified several articles originating from the same study population. Notably, the Costa Rican Childhood Leukemia Study (CRCLS) contributed 2 articles.^{39,40} Similarly, The Infants’ Environmental Health (ISA) study from Costa Rica gave rise to 5 articles.^{41–45} The EcoSalud Project in Ecuador produced 2 articles,^{46,47} and there was a pilot study and a subsequently another study conducted in the town of Tabacundo, Ecuador, resulting in 2 articles.^{48,49} Among the articles from Brazil, 2 evaluated data from the Multi-institutional Study of Infant Leukemia.^{50,51} In addition, we included 13 articles that originated from data analysis of the prospective mother–child cohort TIMOUN study conducted in Guadeloupe.^{52–64} In the case of Mexico, several groups of articles originated from the same study: Seven articles came from the prospective perinatal cohort study in the State of Morelos,^{65–71} 3 articles came from the Early Life Exposures in Mexico to Environmental Toxicants (ELEMENT) study,^{72–74} and another 3 articles originated from the study conducted in Tapachula, State of Chiapas.^{75–77} The predominant study designs among the articles included were cohort ($n = 37$, 46.2%), cross-sectional ($n = 18$, 22.5%), and case–control ($n = 17$, 21.2%) (Figure 2B). Most studies evaluated pesticide exposure directly using levels measured in biological samples, with maternal blood ($n = 19$, 23.7%), cord blood ($n = 13$, 16.2%) and maternal urine ($n = 10$, 12.5%) being the most common biological matrixes (Figure 2C). A notable portion of studies indirectly assessed pesticide exposure through questionnaires ($n = 27$, 33.8%), surveying the participants about proximity of residences to pesticide applications sites, occupational exposure, or domestic use. The articles were categorized into five groups based on their respective outcomes: *a*) adverse obstetric outcomes ($n = 20$, 25.0%), *b*) congenital anomalies ($n = 21$, 26.3%), *c*) anthropometric measures and sex ratio ($n = 18$,

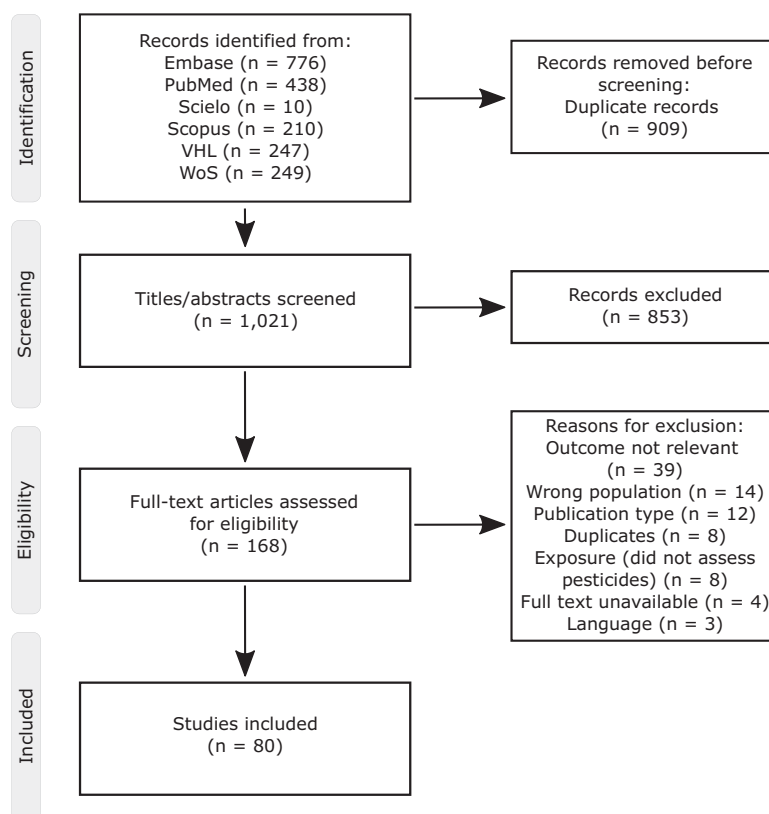


Figure 1. Flow diagram of the study selection process. Note: Embase, Excerpta Medica Database; Scielo, Scientific Electronic Library Online; VHL, Virtual Health Library; WoS, Web of Science.

22.5%), *d*) neurodevelopment and other outcomes throughout childhood ($n=29$, 36.3%), and *e*) molecular effects ($n=6$, 7.5%) (Figure 2D). We provide a summary of the findings from these articles in the subsequent sections, organized based on these outcome categories.

Obstetric Outcomes

Twenty studies explored the association between prenatal pesticide exposure and obstetric outcomes (Table 1). These studies include three sets of outcomes: premature birth, intrauterine growth restriction, and spontaneous abortion.

Premature birth. Thirteen studies examined the impact of pesticide exposure on preterm birth and gestational length. Among these, 6 did not yield statistically significant findings.^{44,79,80,82,88,93} The study by Wendel de Joode et al.⁴⁴ found no association between 3,5,6-trichloro-2-pyridinol (TCPY), 3-(2,2-dichlorovinyl)-2,2-dimethylcyclopropanecarboxylic acid (DCCA), 3-phenoxybenzoic acid (3-PBA), 2,4-dichlorophenoxyacetic acid (2,4-D), ethylenethiourea (ETU), hydroxy pyrimethanil (OHP), and length of gestation, even when stratifying levels in the first and second half of gestation. In the study by Rodriguez et al.,⁸⁰ a similar average gestational age at birth was observed when comparing pregnant women from urban and rural regions. However, differences in the concentration levels of chlorpyrifos were observed between the two groups. In the study by Souza et al.,⁷⁸ the average gestational age at birth in an Argentinian region of intense use of pesticides decreased by 2 wk during the spraying period. An ecological study with data from the Brazilian Live Birth Information System (SINASC) showed that regions with higher levels of per capita pesticide consumption presented a higher prevalence of births before 22 wk of gestation ($p < 0.001$).⁸³ A similar approach

showed that the prevalence ratio (PR) of prematurity was associated with per capita fungicide [PR = 1.23; 95% confidence interval (CI): 1.19, 1.28], acaricide (PR = 1.08; 95% CI: 1.05, 1.12), herbicide (PR = 1.24; 95% CI: 1.20, 1.28), and insecticide (PR = 1.25; 95% CI: 1.21, 1.30) use.⁸⁴ Other studies have examined the effects of specific pesticides on premature birth. In the TIMOUN Mother–Child Cohort Study conducted in Guadeloupe, a 1-log₁₀ increase in chlordecone concentration was significantly associated with a reduction in the length of gestation (−0.27 wk; 95% CI: −0.50, −0.03) and risk of preterm birth [hazard ratio (HR) = 1.6%; 95% CI: 1.1, 2.3].⁵² Arrebola et al.⁸¹ found a negative association between 1,1-dichloro-2,2-bis(*p*-chlorophenyl)ethylene (*p,p'*-DDE) but not its metabolite 2,2-(2-chlorophenyl-4'-chlorophenyl)-1,1-dichloroethene (*o,p'*-DDE) and gestation length ($\beta = -0.004$; 95% CI: −0.008, −0.001). Based on urine samples collected at ~26 wk of gestation, Silver et al.⁹⁴ reported that prenatal exposure to glyphosate (OR = 1.45; 95% CI: 1.00, 2.10) or aminomethylphosphonic acid (AMPA) (OR = 1.88; 95% CI: 1.35, 2.62) were associated with spontaneous preterm birth.

Intrauterine growth restriction. Evaluating data from 26 Brazilian states, Siqueira et al.⁸² showed an association between pesticide use and risk of low birth weight ($\beta = 0.26$, $p = 0.045$). Using national Brazilian data, Mendonça Guimarães et al.⁸⁴ observed higher PRs of low birth weight in the higher tertiles of fungicide (PR = 1.14; 95% CI: 1.10, 1.17), acaricide (PR = 1.07; 95% CI: 1.03, 1.10), herbicide (PR = 1.15; 95% CI: 1.12, 1.19), and insecticide (PR = 1.14; 95% CI: 1.11, 1.18) consumption. However, Cremonese et al.⁸³ reported null findings in a similar study that was limited to only three southern Brazilian states. Levorio-Carrillo et al.,⁸⁹ using data from Mexico, reported a significant association between prenatal exposure to pesticides and the presence of intrauterine growth restriction (IUGR) (OR = 2.3;

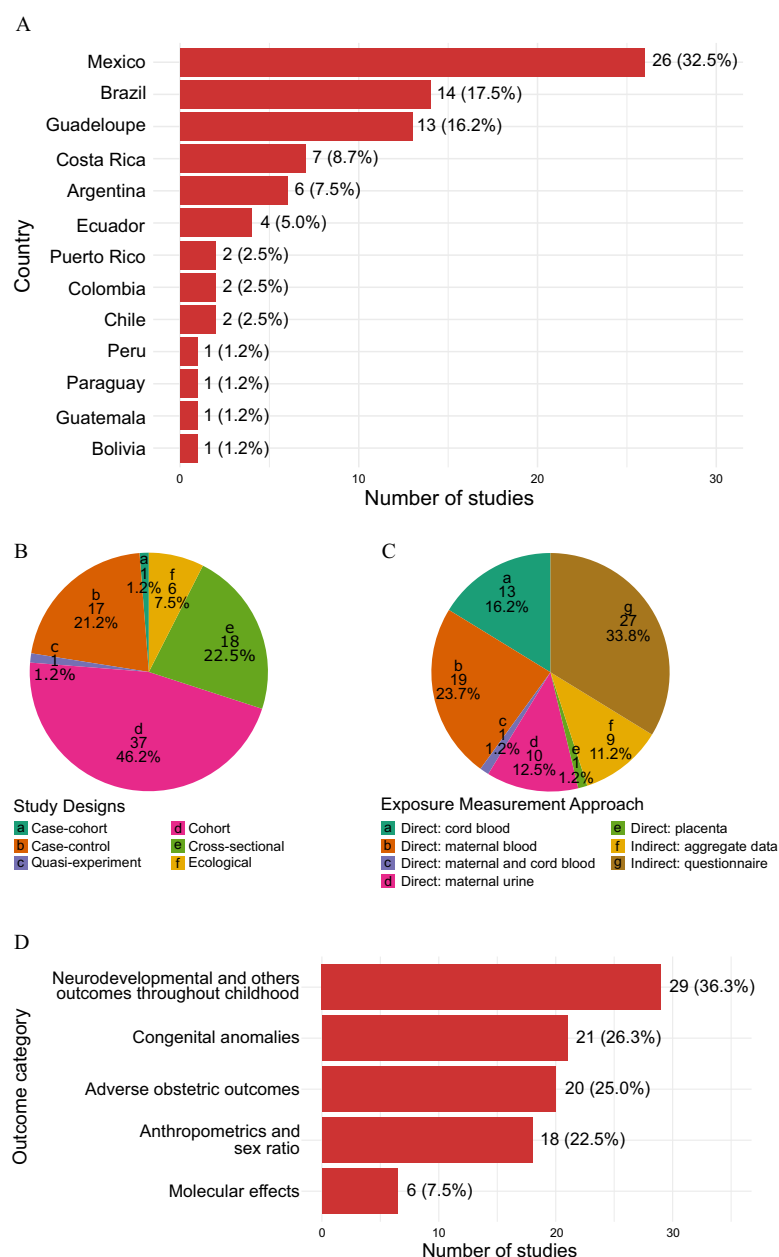


Figure 2. Characteristics of human observational studies assessing prenatal exposure to pesticides and adverse outcomes. Number of studies by (A) LAC country, (B) study design, (C) approaches employed to measure exposure, and (D) number of studies in each outcome category. Note that a study may be classified into multiple outcome categories.

95% CI: 1.0, 5.3). In Mexico, the study by Acosta-Maldonado et al.,⁹¹ which analyzed placentas, identified an association between pesticide exposure and morphometric parameters of placental maturity, which, in turn, is linked to birth weight. Cecchi et al.⁷⁹ studied the outcomes of pregnant women in Argentina based on their residential proximity to fruit croplands with intense pesticide applications. The study did not identify any urban/rural differences in the rates of IUGR or newborns classified as small for gestational age (SGA). In another study, prenatal exposure to *p,p'*-DDE was not associated with higher risk for SGA.⁹³

Spontaneous abortion. We identified seven studies evaluating exposure to pesticides and spontaneous abortion. DDT exposure was not related to spontaneous abortion among workers of the malaria control program in Mexico.⁹⁰ A rate of spontaneous abortion 8.5 times higher (95% CI: 6.72, 10.65) than average was

observed in an agricultural area in Chile than at the national level; besides that, there were statistically higher rates of agricultural workers among women with spontaneous abortion than among all women from the same agricultural area.⁸⁶ Braga et al.⁸⁵ observed higher rates of fetal death in regions with high agricultural production compared with regions with low agricultural production in the state of São Paulo, Brazil. In Ecuador, Handal and Harlow reported a significant association between maternal employment in the flower industry within the past 6 y and an increased likelihood of reporting a history of spontaneous abortion (OR = 2.6; 95% CI: 1.03, 6.73).⁴⁷ Conversely, in the study from Blanco-Muñoz et al.⁹² in Mexico, pregnant women who have worked as a floriculturist during the period of 3 months before pregnancy to 5 months into gestation did not have an increased risk of miscarriage compared with pregnant women who did not work as a floriculturist during the same period.

Table 1. Latin American and Caribbean studies on prenatal pesticide exposure and adverse obstetric outcomes published between 2000 and 2024 (*n* = 20).

Author, year	Country	Exposure	Outcome	Study design (<i>n</i>)	Key findings
Souza et al., 2006 ⁷⁸	Argentina	OP exposure estimated by maternal blood cholinesterase levels	Gestational length	Cohort (342)	Gestational age decreased by 2 wk during the fumigation period
Cecchi et al., 2021 ⁷⁹	Argentina	Maternal residential proximity to fruit croplands with intense pesticide applications	IUGR, preterm premature rupture of membranes, preterm birth (delivery at <37 wk gestation), fetal death, threat of premature labor, miscarriage, threatened miscarriage (a pregnancy-related bloody vaginal discharge or frank bleeding during the first half of pregnancy without cervical dilation), and SGA (weight-for-gestational age percentile <10 based on the INTERGROWTH-21st standards)	Cohort (776)	The rural population had a significantly higher rate of threatened miscarriage compared with the urban population (0.3% vs. 3%, <i>p</i> < 0.05). The percentage of intrauterine growth retardation, preterm premature rupture of membranes, intrauterine fetal death, fetal death, threat of premature labor, and miscarriage were similar in both groups
Rodriguez et al., 2023 ⁸⁰	Argentina	OCPs, chlorpyrifos, trifluralin, and chlorothalonil levels measured from placental samples collected immediately after delivery	Gestational age at birth	Cross-sectional (85)	The gestational age was similar between urban and rural groups, despite the rural group showing higher levels of chlorpyrifos and total OCPs + chlorpyrifos
Arreola et al., 2016 ⁸¹	Bolivia	<i>o,p'</i> -DDT and <i>p,p'</i> -DDE levels quantified in serum from cord blood samples collected after delivery	Length of gestation	Cohort (200)	<i>p,p'</i> -DDE was negatively associated with length of gestation ($\beta = -0.004$; 95% CI: $-0.008, -0.001$) and <i>o,p'</i> -DDT did not reach a significant association ^a
Siqueira et al., 2010 ⁸²	Brazil	Pesticide use (kg/hectare/y)	Prematurity (live births with pregnancy duration <37 wk) and low birth weight (live births with weight <2,500 g)	Ecological (3,154,233)	An association between pesticide use and low birth weight ($\beta = 0.26$; 95% CI: 0.006, 0.52). No association was found between pesticide use and prematurity ^b
Cremonese et al., 2012 ⁸³	Brazil	Per capita pesticide consumption	Premature birth (gestational age <22 wk), 1- and 5-min Apgar score, and low birth weight	Ecological (2,268,704)	In the upper quartile of pesticide consumption, higher PRs were observed for premature birth (<i>p</i> < 0.001) and low first (<i>p</i> < 0.001) and 5-min (<i>p</i> < 0.001) Apgar score. No significant findings were observed regarding low birth weight
Mendonça Guimarães et al., 2014 ⁸⁴	Brazil	Per capita pesticides consumption	Premature births (live births <37 wk) and low birth weight (live births <2,500 g)	Ecological (NA)	Associations between premature births and fungicide (PR = 1.23; 95% CI: 1.19, 1.28), acaricide (PR = 1.08; 95% CI: 1.05, 1.12), herbicide (PR = 1.24; 95% CI: 1.20, 1.28), and insecticide (PR = 1.25; 95% CI: 1.21, 1.30); associations between low birth weight and fungicide (PR = 1.14; 95% CI: 1.10, 1.17), acaricide (PR = 1.07; 95% CI: 1.03, 1.10), herbicide (PR = 1.15; 95% CI: 1.12, 1.19), and insecticide (PR = 1.14; 95% CI: 1.11, 1.18)
Braga et al., 2023 ⁸⁵	Brazil	Estimates of the amount of pesticides used within selected regions	Rates of fetal death before 22 wk and rates of fetal deaths weighing <500 g	Ecological (NA)	Higher ratios of fetal death rates at <500 g weight and higher ratios of fetal death rates before 22 wk in regions with high agricultural production compared with the region with the lowest agricultural production
Contreras-Levicoy et al., 2005 ⁸⁶	Chile	Living in agricultural area	Spontaneous abortion	Cross-sectional (975)	Spontaneous abortion rates of 8.5 (95% CI: 6.72, 10.65) times higher in the agricultural area than the national level. More agricultural workers among women with spontaneous abortion than among those from the same agricultural area
Camacho and Mejía, 2017 ⁸⁷	Colombia	Aerial spraying of glyphosate	Miscarriages	Quasi-experiment (687,735)	Aerial spraying significantly increases miscarriage rates ($\beta = 0.118$, <i>p</i> < 0.01)

Table 1. (Continued.)

Author, year	Country	Exposure	Outcome	Study design (n)	Key findings
Wendel de Joode et al., 2024 ⁴⁴	Costa Rica	Pesticide metabolites (ETU, TCPY, 2,4-D, 3-PBA, DCCA, and OHP) measured from maternal urine samples collected during the first half (<20 wk) and second half (≥20 wk) of gestation	Length of gestation	Cohort (386)	No statistically significant findings ^c
Handal and Harlow, 2009 ⁴⁷	Ecuador	Work in the flower industry	Spontaneous abortions	Cross-sectional (153)	Women in the flower industry had 2.6 times higher odds of reporting a history of spontaneous abortion compared with nonindustry workers (OR = 2.6; 95% CI: 1.03, 6.73) ^d
Kadhel et al., 2014 ⁵²	Guadeloupe	Chlordecone levels quantified from maternal blood samples collected in the third trimester	Length of gestation and preterm birth (birth before 37 completed wk of gestation)	Cohort (818)	Chlordecone levels were associated with a shorter length of gestation ($\beta = -0.27$ wk; 95% CI: $-0.50, -0.03$), as well as with preterm birth (HR = 1.6; 95% CI: 1.1, 2.3) ^e
Torres-Arreola et al., 2003 ⁸⁸	Mexico	Serum levels of <i>p,p'</i> -DDE, β -HCH and HCB from blood samples collected within 24 h after delivery	Preterm birth (<37 wk of gestation)	Case-cohort (233)	No statistically significant findings
Levario-Carrillo et al., 2004 ⁸⁹	Mexico	Environmental exposure to cholinesterase inhibitory pesticides during the gestational period	IUGR (newborns with weight-for-gestational age <10th percentile)	Case-control (371)	Association between pesticide exposure and the presence of IUGR (aOR = 2.3; 95% CI: 1.0, 5.3) ^f
Salazar-García et al., 2004 ⁹⁰	Mexico	DDT occupational exposure during pregnancy	Spontaneous abortion	Cross-sectional (9,187)	No statistically significant findings for spontaneous abortion
Acosta-Maldonado et al., 2009 ⁹¹	Mexico	Residing in agricultural communities, proximity to pesticide application zones, or cohabitation with agricultural worker during pregnancy	PMI and gestational age	Cross-sectional (54)	Exposure was associated with higher PMI. Significant correlation between PMI and birth weight ($\rho = 0.54$, $p < 0.01$) and between PMI and gestational age ($\rho = 0.44$, $p < 0.01$) from week 21 to 38
Blanco-Muñoz et al., 2013 ⁹²	Mexico	Floricultural work between 3 months before the beginning of pregnancy and the first 5 months of pregnancy	Miscarriage (pregnancies ending at a gestational age of <20 wk)	Cross-sectional (612)	Floricultural work during the periconceptional period and the first 5 months of pregnancy did not increase the risk of miscarriage (OR = 0.59; 95% CI: 0.26, 1.31)
Torres-Sánchez et al., 2023 ⁹³	Mexico	<i>p,p'</i> -DDE levels measured from maternal blood samples collected before 18 wk of pregnancy	Gestational age at birth, preterm birth (birth <37 complete wk), and SGA (weight-for-gestational age percentile <10 based on the INTERGROWTH-21st standards)	Cohort (170)	No statistically significant findings ^g
Silver, et al., 2021 ⁹⁴	Puerto Rico	Glyphosate and AMPA concentrations in maternal urine samples collected during the second and third trimesters	Preterm birth (delivery before 37 completed wk of gestation)	Case-control (247)	Preterm birth was significantly associated with glyphosate (aOR = 1.45; 95% CI: 1.00, 2.10) and AMPA (aOR = 1.88; 95% CI: 1.35, 2.62) levels from urine samples collected at ~26 wk of gestation, whereas associations with levels at ~18 wk of gestation were null or inconsistent ^h

Note: 2,4-D, 2,4-dichlorophenoxyacetic acid; 3-PBA, 3-phenoxybenzoic acid; AMPA, aminomethylphosphonic acid; aOR, adjusted odds ratio; β -HCH, beta-hexachlorocyclohexane; BMI, body mass index; CI, confidence interval; DCCA, 3-(2,2-dichlorovinyl)-2,2-dimethylcyclopropanecarboxylic acid; DDE, dichlorodiphenyldichloroethylene; DDT, dichlorodiphenyltrichloroethane; ETU, ethylenethiourea; HCB, hexachlorobenzene; HR, hazard ratio; IUGR, intrauterine growth restriction; NA, data not available; OCPs, organochlorine pesticides; OHP, hydroxy pyrimethanil; OP, organophosphates; OR, odds ratio; PMI, placental maturity index; PR, prevalence ratio; SGA, small for gestational age; TCPY, 3,5,6-trichloro-2-pyridinol.

^aAdjusted for gestational weight gain, maternal age, parity, smoking habit, log (newborn weight), and maternal BMI.

^bAdjusted for proportion of pregnant women with a low number of prenatal care visits.

^cAdjusted for newborn's sex, maternal age, parity, smoking during pregnancy, prepregnancy BMI, and country of birth.

^dAdjusted for maternal age.

^eAdjusted for maternal age, place of birth, enrollment site, marital status, educational level, BMI, high blood pressure during pregnancy, and total plasma lipid level (g/L).

^fAdjusted for fat-free mass, placental weight, and anti-cytomegalovirus antibodies.

^gAdjusted for newborn sex, gestational age at birth, number of previous pregnancies, tobacco consumption history, and BMI.

^hAdjusted for maternal age, education, prepregnancy BMI, and smoking.

Another study found a significantly higher risk of threatened miscarriage, defined as a pregnancy-related bloody vaginal discharge or frank bleeding during the first half of pregnancy without cervical dilation, among Argentinian women living in close proximity to areas with intensive pesticide application compared with those in urban areas (3% vs. 0.3%, $p < 0.05$).⁷⁹ However, that study found no significant differences in terms of miscarriages or fetal death. In a quasi-experimental study performed with data from Colombia, it was shown that exposure to aerial sprayed glyphosate in the months preceding miscarriage had a strong and statistically significant association with the reported miscarriage ($\beta = 0.118$, $p < 0.01$).⁸⁷

Congenital Anomalies

A total of 21 studies examining the potential association between prenatal pesticide exposure and congenital anomalies were included in the present review (Table 2). Among these studies, 11 had the primary objective of evaluating congenital anomalies in general, whereas the remaining 10 focused on specific groups of anomalies.

In the ecological study from Mendonça Guimarães et al.⁸⁴ using Brazilian data, the total pesticides and herbicides per capita consumption were correlated with the prevalence of congenital anomalies at birth. Moreover, when comparing the highest tertile of per capita pesticide consumption with the lowest tertile, higher PRs of congenital anomalies were observed in the highest tertile for fungicides (PR = 1.33; 95% CI: 1.28, 1.38), acaricides (PR = 1.25; 95% CI: 1.20, 1.30), insecticides (PR = 1.38; 95% CI: 1.33, 1.43), and herbicides (PR = 1.38; 95% CI: 1.33, 1.43). In another Brazilian study context, Siqueira et al.⁸² found that the ratio of pesticide use by cultivated area was not associated with the frequency of total congenital abnormalities at birth. In the ecological study by Braga et al.,⁸⁵ an increase in the incidence of congenital anomalies among live births was observed over time. However, a pattern of higher incidence of congenital anomalies in regions with higher pesticide use, compared with regions with lower use, was not identified. On the other hand, a study by Tipiana et al.¹¹⁰ in Peru found that women who reported going to sprayed fields (OR = 3.82; 95% CI: 1.92, 7.60), living near sprayed fields (OR = 3.07; 95% CI: 1.59, 5.92), and having a partner who works in sprayed fields (OR = 2.40; 95% CI: 1.21, 4.78) showed an increased risk of having a newborn with congenital anomalies compared with women who did not report these exposures. Similarly, Peruvian women who had direct exposure to pesticides during pregnancy presented a significantly increased risk of having a newborn with congenital anomalies compared with those who did not have such exposure (OR = 4.51; 95% CI: 1.77, 11.46).¹⁰⁹ Another study reported a higher proportion of parental exposure to pesticides and a higher proportion of maternal residence around fumigated crops among newborns with congenital anomalies than in newborns from the control group.¹⁰¹ Salazar-García et al.⁹⁰ showed an association between DDE levels from fathers and the risk of congenital anomalies (OR = 3.76; 95% CI: 1.24, 11.44 for the highest quartile). One study conducted in Brazil investigated parental residence and occupation, whereas another study in Guadeloupe assessed maternal and cord blood chlordecone levels in relation to congenital anomalies, but neither study found an association.^{56,96} Similarly, a case-control study from Argentina identified no difference in quantified pesticide levels from blood samples between mothers of newborns with congenital anomalies and mothers of control newborns.⁹⁵

Dutra and Ferreira used data from a Brazilian state to compare the prevalence of different groups of congenital anomalies between a region with low pesticide use and another region with high pesticide use.⁹⁹ They found that the prevalence of cleft lip and palate (OR = 1.64; 95% CI: 1.27, 2.12), circulatory system

anomalies (OR = 2.79; 95% CI: 1.87, 4.16), genitourinary system anomalies (OR = 1.62; 95% CI: 1.28, 2.06), and digestive system anomalies (OR = 1.56; 95% CI: 1.15, 2.13) were higher in the region with high pesticide use.

Considering more specific congenital anomalies, maternal occupation in agriculture from 3 months before to 1 month after the last menstrual period was identified as a risk factor for anencephaly (OR = 4.57; 95% CI: 1.05, 19.96).¹⁰⁵ Another study demonstrated that the use of pesticides in home gardens (OR = 5.66; 95% CI: 1.44, 22.20) and the use of pesticides in farms close to home (OR = 2.56; 95% CI: 1.21, 5.43) were associated with cleft lip and palate.¹⁰⁰ In addition, pesticide exposure during the first 3 months of pregnancy was associated with an increased risk of nonsyndromic cleft lip, cleft palate, or both (OR = 2.2; 95% CI: 1.1, 4.7).¹⁰⁷ In a cohort of orofacial cleft cases from a pediatric plastic surgery program in Guatemala, 49.1% of the mothers reported maternal pesticide or herbicide exposure during pregnancy.¹⁰² A study administered questionnaires to mothers of newborns with central nervous system (CNS) anomalies.¹⁰⁸ It was found that 40.2% reported exposure to pesticides during pregnancy, whereas 39.8% reported exposure to herbicides. Some studies investigated associations between pesticide exposure and male external genital malformations.^{84,97,103,104,106} Maternal levels of DDT and DDE were not associated with the risk of hypospadias.¹⁰³ Organochlorine pesticide levels from maternal blood samples were associated with cryptorchidism.^{104,106}

The mortality caused by congenital anomalies was also assessed. Siqueira et al.⁸² found an association between pesticide use and proportional mortality of congenital anomalies ($\beta = 1.52$, $p = 0.004$) and infant death rate ($\beta = 0.17$, $p = 0.039$) in infants <1 year of age. Cremonese et al.⁹⁸ investigated the mortality rate caused by anomalies in the CNS and cardiovascular system (CVS) in infants <1 y of age. They observed a trend of increased relative risk with an increase in pesticide use ($p < 0.05$) and higher relative risk for CNS (RR = 1.18; 95% CI: 1.01, 1.37) and CVS (RR = 1.15; 95% CI: 1.03, 1.29) in the highest quintile compared with the lowest quintile of exposure. The results were statistically significant in rural areas but not in urban areas.

Anthropometric Measures and Sex Ratio

We identified 16 studies assessing the association between prenatal pesticide exposure and anthropometric parameters in newborns and children (Table 3). In addition, 2 studies investigated the proportion of male births in populations with high pesticide exposure.

An ecologic study performed with data from the Brazilian state of Paraná did not find any correlation between the amount of pesticide sales and the proportion of male newborns.¹¹² There was no difference in the male-to-female sex ratio of children born from workers on the malaria control program in Mexico.⁹⁰

Longnecker et al.⁷⁵ investigated the effect of *in utero* exposure to DDE on penile and anogenital dimensions in newborn males. The penis length, width, and anogenital distance did not differ significantly across categories of serum DDE levels. Interestingly, Torres-Sánchez et al.⁶⁷ observed a significant reduction in the anogenital distance index among boys ($\beta = -0.02$; 95% CI: -0.03 , -0.003), but not among girls, associated with a 2-fold increase in maternal DDE serum levels during the first trimester of pregnancy. In another study, DDT was not associated with anogenital distance, but polychlorobiphenyls (PCBs) 28 ($\beta = -0.005$; 95% CI: -0.008 , -0.001), PCB 74 ($\beta = -0.003$; 95% CI: -0.006 , -0.0007), and PCB 170 ($\beta = -0.005$; 95% CI: -0.008 , -0.001) were associated with anogenital distance reduction in boys.¹¹⁴

Studies where the exposure variable was defined as the third trimester of pregnancy coinciding with the period of pesticide spraying,¹¹¹ maternal residential proximity to fruit croplands,⁷⁹

Table 2. Latin American and Caribbean studies on prenatal pesticide exposure and congenital anomalies published between 2000 and 2024 (*n* = 21).

Author, year	Country	Exposure	Outcome	Study design (<i>n</i>)	Key findings
Colussi et al., 2024 ⁹⁵	Argentina	Organochlorine, organophosphate, and pyrethroid pesticides quantified from maternal blood samples collected after delivery	Congenital anomalies	Case-control (290)	There were no significant differences in the concentrations of the pesticides assessed between cases and controls
Siqueira et al., 2010 ⁸²	Brazil	Pesticide use (kg/hectare/y)	Congenital abnormality at birth and infant death caused by congenital abnormality	Ecological (3,154,233)	Pesticide use was positively associated with proportional mortality by congenital abnormality in <1-y-old infants ($\beta = 1.52$, $p = 0.004$), as well as infant death rate by congenital abnormality ($\beta = 0.17$, $p = 0.039$). Associations with congenital abnormalities rates were null ^a
Gonçalves e Silva et al., 2011 ⁹⁶	Brazil	Occupational and environmental exposure to pesticides	Congenital defects	Case-control (126)	No statistically significant findings
Gaspari et al., 2012 ⁹⁷	Brazil	Occupational activities and lifestyles with potential exposure to EDCs before and during pregnancy	Male genital malformations	Cross-sectional (2,710)	High rate of occupational EDC exposure from mothers (>66%) and fathers (>55%) of newborns with male genital malformations. High fetal exposure to EDCs was reported (91% cases)
Cremoneze et al., 2014 ⁹⁸	Brazil	Per capita pesticide consumption	Infant mortality caused from CNS and CVS congenital malformations	Ecological (5,241)	Correlations between per capita pesticide consumption and mortality rates due to CNS and CVS defects in rural microregions ($p < 0.001$)
Mendonça Guimarães et al., 2014 ⁸⁴	Brazil	Per capita consumption of pesticides	Birth defects and cryptorchidism	Ecological (NA)	Higher PRs of congenital anomalies in the highest tertile of exposure to fungicides (PR = 1.33; 95% CI: 1.25, 1.38), acaricides (PR = 1.25; 95% CI: 1.20, 1.30), insecticides (PR = 1.38; 95% CI: 1.33, 1.43), and herbicides (PR = 1.38; 95% CI: 1.33, 1.43) compared with the lowest tertile
Dutra and Ferreira, 2017 ⁹⁹	Brazil	Regions with high and low pesticide use	Groups of congenital malformations	Cross-sectional (NA)	Region with higher exposure showed higher rates of cleft lip and palate (OR = 1.64; 95% CI: 1.27, 2.12), cleft-lary (OR = 2.79; 95% CI: 1.87, 4.16), genitourinary (OR = 1.62; 95% CI: 1.28, 2.06), and digestive (OR = 1.56; 95% CI: 1.15, 2.13) system anomalies
Silvestre et al., 2022 ¹⁰⁰	Brazil	Parents' environmental and occupational exposures	Cleft lip and palate	Case-control (375)	The use of pesticides in home gardens (OR = 5.66; 95% CI: 1.44, 22.20) and the use of pesticides in farms close to home (OR = 2.56; 95% CI: 1.21, 5.43) were associated with cleft lip and palate
Braga et al., 2023 ⁸⁵	Brazil	Estimates of the amount of pesticides used within selected regions	Prevalence of newborns with congenital anomalies per live birth	Ecological (NA)	An upward trend in the prevalence of congenital anomalies. Higher prevalence rates of congenital anomalies were not observed in regions with higher exposure
Rojas et al., 2000 ¹⁰¹	Chile	Occupational or environmental exposure to pesticides by parents during pregnancy	Congenital malformations	Case-control (858)	Higher frequency of agricultural labor or rural residency in the case group ($X^2 = 19.52$, $p < 0.00001$). Association between rural residency and congenital malformations (OR = 2.16; 95% CI: 1.55, 3.00)
Rouget et al., 2020 ⁸⁶	Guadeloupe	Chlordane concentrations in maternal blood samples collected at third-trimester and cord blood samples collected after delivery	Congenital anomalies	Cohort (1,068)	No statistically significant findings ^b
Card et al., 2023 ¹⁰²	Guatemala ^c	Pesticide and herbicide exposure during pregnancy through home use	Orofacial clefts	Cohort (55)	Pesticide or herbicide use during pregnancy was reported by 49.1% of the mothers
Flores-Luévano et al., 2003 ¹⁰³	Mexico	Maternal serum DDT/DDE levels from maternal blood samples collected <18 y after delivery	Hypopspadias	Case-control (69)	Maternal levels of DDT and DDE showed no association with hypospadias

Table 2. (Continued.)

Author, year	Country	Exposure	Outcome	Study design (n)	Key findings
Salazar-García et al., 2004 ⁴⁰	Mexico	DDT occupational exposure during pregnancy	Congenital malformations	Cross-sectional (9,187)	Parents with estimated high exposure to DDT exhibited an elevated risk of having a child with congenital malformations (quartile 3: OR = 4.15; 95% CI: 1.38, 12.46; quartile 4 OR = 3.76; 95% CI: 1.24, 11.44) ^d
Waliszewski et al., 2005 ¹⁰⁴	Mexico	Organochlorine pesticides in maternal blood serum lipids from samples collected postpartum	Undescended testis	Case-control (60)	The relative risk was higher for the case group compared with the control group for HCB (RR = 1.16; 95% CI: 1.07, 1.26), β -HCH (RR = 0.73; 95% CI: 0.66, 0.79), and total DDT (RR = 1.12; 95% CI: 1.09, 1.15)
Lacasaña et al., 2006 ¹⁰⁵	Mexico	Parental occupational exposure to agricultural work	Anencephaly	Case-control (302)	Mothers involved in agricultural work during the periconceptional period had a higher risk of having children with anencephaly (OR = 4.57; 95% CI: 1.05, 19.96) ^e
Montes et al., 2010 ¹⁰⁶	Mexico	Persistent organochlorine pesticides (HCB, β -HCH, p,p' -DDE, o,p' -DDT, p,p' -DDT) quantified from newborns' blood samples collected after delivery	Undescended testicles (cryptorchidism)	Case-control (82)	Association between serum levels of HCB (OR = 1.17; 95% CI: 1.08, 1.27) and total DDT (OR = 1.14; 95% CI: 1.09, 1.19) and cryptorchidism
Ibarra-Lopez et al., 2013 ¹⁰⁷	Mexico	Environmental and occupational exposure to pesticides during the first trimester	Nonsyndromic cleft lip and palate	Case-control (204)	Association between pesticide exposure and nonsyndromic cleft lip and palate (aOR = 2.2; 95% CI: 1.1, 4.7) ^f
Enríquez-Vera et al., 2023 ¹⁰⁸	Mexico	Types of pesticides reported by mothers	Congenital malformations of the CNS	Cross-Sectional (1113)	During pregnancy, 40.2% of mothers reported exposure to pesticides, whereas 39.8% reported exposure to herbicides
Ojeda and Leite, 2018 ¹⁰⁹	Paraguay	Occupational and environmental exposure to agrochemicals during pregnancy	Congenital malformations	Case-control (132)	Association between direct maternal exposure to agrochemicals and congenital malformations (OR = 4.51; 95% CI: 1.77, 11.46)
Tipiana et al., 2015 ¹¹⁰	Peru	Any contact with pesticides during pregnancy	Congenital malformations	Case-control (171)	Having exposure to pesticides (OR = 2.85; 95% CI: 1.46, 5.54), going to sprayed fields (OR = 3.82; 95% CI: 1.92, 7.60), living near sprayed fields (OR = 3.07; 95% CI: 1.59, 5.92), and having a partner working in sprayed fields (OR = 2.40; 95% CI: 1.21, 4.78) during pregnancy were associated with a higher risk of having a child with a congenital anomaly

Note: aOR, adjusted odds ratio; β -HCH, beta-hexachlorocyclohexane; CI, confidence interval; CNS, central nervous system; CVS, cardiovascular system; DDE, dichlorodiphenyl/dichloroethylene; DDT, dichlorodiphenyl/trichloroethane; EDCs, endocrine-disrupting chemicals; HCB, hexachlorobenzene; NA, data not available; OR, odds ratio; PR, prevalence ratio; RR, relative risk; X2, result for the chi-squared test.

^aAdjusted for proportion of pregnant women with a low number of prenatal care visits.

^bAdjusted for maternal place of birth, maternal education, parity, gestational age, maternal age, and total plasma lipids.

^cThis study compared patients born in Guatemala with those from the United States.

^dAdjusted for maternal age and application of malathion.

^eAdjusted for age of the mother, socioeconomic level (family income and education of the mother), adverse reproductive history with previous children, folate intake of the mother, and energy of the mother.

^fAdjusted for age at conception, years of schooling, region of origin of the mother, folic acid intake, and tobacco exposure.

Table 3. Latin American and Caribbean studies on prenatal pesticide exposure and anthropometrics or sex ratio published between 2000 and 2024 (*n* = 18).

Author, year	Country	Exposure	Outcome	Study design (<i>n</i>)	Key findings
Souza et al., 2006 ⁷⁸	Argentina	OP exposure estimated by maternal blood cholinesterase levels	HC	Cohort (342)	Increase in HC in newborns of the women who lived in rural areas ($p \leq 0.05$) and who sprayed OP at home ($p \leq 0.05$) No statistically significant findings ^a
Silvia et al., 2020 ¹¹¹	Argentina	Third-trimester pregnancy coinciding with the period of pesticides spraying	Birth weight, length at birth, HC and the HC/weight ratio	Cross-sectional (53)	Higher ponderal index ($p < 0.05$), lower length at birth ($p < 0.01$), and HC <5th percentile ($p < 0.05$) in the rural group compared with the urban group ^b No statistically significant findings
Cecchi et al., 2021 ⁷⁹	Argentina	Maternal residential proximity to fruit croplands with intense pesticide applications	Newborn parameters	Cohort (776)	
Rodriguez et al., 2023 ⁸⁰	Argentina	OCPs, chlorpyrifos, trifluralin, and chlorothalonil levels measured from placental samples collected immediately after delivery	Weight, length, and HC	Cross-sectional (85)	
Arrebola et al., 2016 ⁸¹	Bolivia	<i>o,p'</i> -DDT and <i>p,p'</i> -DDE levels quantified in serum from cord blood samples collected after delivery	Birth weight, HC, height, ponderal index	Cohort (200)	<i>p,p'</i> -DDE was associated with birth weight ($\beta = -0.014$; 95% CI: $-0.028, -0.0010$) ^c
Gibson and Koifman, 2008 ¹¹² Villar et al., 2022 ¹¹³	Brazil Brazil ^d	Volume of pesticide sales Metabolite signatures untargeted from maternal blood samples collected around the second trimester and from umbilical cord venous blood samples collected at delivery	Proportion of male births Fetal abdominal growth, postnatal growth, adiposity, and neurodevelopment	Ecological (NA) Cohort (3,206)	No statistically significant findings Maternal chlorothalonil was positively associated with an early accelerating growth phenotype and negatively associated with a faltering growth phenotype ^e
Wendel de Joode et al., 2024 ⁴⁴	Costa Rica	Pesticide metabolites (ETU, TCPY, 2,4-D, 3-PBA, DCCA, and OHP) measured from maternal urine samples collected during the first half (<20 wk) and second half (≥ 20 wk) of gestation	Birth size: birth weight (g), body length (cm), and HC (cm)	Cohort (386)	Higher prenatal TCPY was associated with lower birth weight (-129.6 ; 95% CI: $-255.8, -3.5$) and smaller HC (-0.61 ; 95% CI: $-1.05, -0.17$ cm). 2,4-D was associated with lower birth weight (β per 10-fold-increase = -125.1 ; 95% CI: $-228.8, -21.5$), smaller HC ($\beta = -0.41$; 95% CI: $-0.78, -0.03$), and shorter body length ($\beta = -0.58$; 95% CI: $-1.09, -0.07$). ETU showed a nonlinear association with HC during the second half of pregnancy ^f Association of chlordecone exposure and higher BMI in boys at 3 months of age ($\beta = 0.9$; 95% CI: $0.0, 1.8$) and in girls at 8 months of age ($\beta = 0.7$; 95% CI: $0.0, 1.5$) ^g No statistically significant findings ^h
Costet et al., 2015 ⁵⁴	Guadeloupe	Chlordecone from cord blood sample collected at birth	Growth of children up to 18 months of age	Cohort (222)	
Hervé et al., 2016 ⁵³	Guadeloupe	Chlordecone concentrations in cord blood collected immediately after delivery	Birth weight	Cohort (593)	
Costet et al., 2022 ⁶⁰	Guadeloupe	Chlordecone levels from cord blood sample collected after delivery	Adiposity in 7-y-old children	Cohort (575)	Boys ($\beta = 0.99$; 95% CI: $0.38, 1.61$) and girls ($\beta = 0.48$; 95% CI: $0.03, 0.92$) had higher adiposity scores in the third quartile relative to the first quartile ⁱ No statistically significant findings
Salazar-García et al., 2004 ⁹⁰	Mexico	DDT occupational exposure during pregnancy	Sex ratio	Cross-sectional (9,187)	

Table 3. (Continued.)

Author, year	Country	Exposure	Outcome	Study design (n)	Key findings
Longnecker et al., 2007 ⁷⁵	Mexico	Maternal serum DDE and DDT levels from samples collected during the postpartum period	Anogenital and penile dimensions in newborn males	Cross-sectional (781)	No statistically significant findings ^d
Torres-Sánchez et al., 2008 ⁶⁷	Mexico	DDE concentrations from maternal blood samples collected before pregnancy and during each trimester	Anogenital distance in male and female infants	Cross-sectional (71)	Reduction of anal position index among boys ($\beta = -0.02$; 95% CI: $-0.03, -0.003$, $p = 0.02$), but not among girls, only during the first trimester of pregnancy ^k No statistically significant findings ^l
Cupul-Uicab et al., 2010 ⁷⁷	Mexico	DDE and DDT levels measured from maternal serum collected within a day of delivery	Longitudinal measurements of height and BMI up to 40 months of age	Cohort (788)	No statistically significant findings ^m
Garced et al., 2012 ⁷⁰	Mexico	DDE and DDT levels from maternal blood samples collected before pregnancy and during each trimester	Child growth parameters at birth and during the first year	Cohort (253)	No statistically significant findings ^m
Loreto-Gómez et al., 2018 ¹¹⁴	Mexico	Persistent organic pollutants from maternal blood samples collected in the third trimester	Anogenital distance	Cohort (156)	Association between PCB 28 ($\beta = -0.005$; 95% CI: $-0.008, -0.001$), PCB 74 ($\beta = -0.003$; 95% CI: $-0.006, -0.0007$), and PCB 170 ($\beta = -0.005$; 95% CI: $-0.008, -0.001$) levels and anogenital distance index among boys ⁿ
Torres-Sánchez et al., 2023 ⁹³	Mexico	<i>p,p'</i> -DDE levels measured from maternal blood samples collected before 18 wk of pregnancy	Birth weight and birth length	Cohort (170)	No statistically significant findings ^o

Note: 2,4-D, 2,4-dichlorophenoxyacetic acid; 3-PBA, 3-phenoxybenzoic acid; BMI, body mass index; CI, confidence interval; DCCA, 3-(2,2-dichlorovinyl)-2,2-dimethylcyclopropanecarboxylic acid; DDE, dichlorodiphenylchloroethylene; DDT, dichlorodiphenyltrichloroethane; ETU, ethylenethiourea; HC, head circumference; NA, data not available; OCPs, organochlorine pesticides; PCB, polychlorinated biphenyl; TCPY, 3,5,6-trichloro-2-pyridinol.

^aAdjusted for sex, gestational age, and parity.

^bAdjusted for sex and gestational age.

^cAdjusted for gestational weight change, maternal age, parity, smoking habit, gestational age, and maternal BMI.

^dThis study adopted a multinational design conducted across Brazil, Kenya, Pakistan, South Africa, Thailand, and the UK.

^eAdjusted for maternal age and fetal sex.

^fAdjusted for newborn's sex, maternal age, parity, smoking during pregnancy, prepregnancy BMI, and country of birth.

^gAdjusted for duration of gestation, maternal place of birth, maternal age, parity, maternal prepregnancy height and weight, maternal weight gain during pregnancy, education level, marital status, smoking and alcohol consumption during pregnancy, and cord total lipid concentrations.

^hAdjusted for gestational age, sex of the newborn, maternal height, maternal BMI, enrollment site, place of birth, family situation, education level, parity, gestational hypertension, gestational diabetes.

ⁱAdjusted for age at measurement, maternal place of birth, maternal BMI before pregnancy, maternal level of education, duration of breastfeeding, time spent exercising per week, time spent watching TV or playing videogames per week, obesogenic dietary habits at 7 years of age, and cord blood total lipids (g/L) and concentrations of PCB153 and *p,p'*-DDE.

^jAdjusted for birth weight, gestational age, urban vs. rural, hospital, and research nurse.

^kAdjusted for birth weight and age at evaluation.

^lAdjusted for smoking, hospital, rural residence, gestational age at delivery, maternal education, and maternal height.

^mAdjusted for age at evaluation, maternal age, height, BMI, parity, tobacco, breastfeeding, iron, zinc, and caloric intake during the first trimester of pregnancy.

ⁿAdjusted for the boy's age, age squared, and height.

^oAdjusted for newborn sex, gestational age at birth, number of previous pregnancies, tobacco consumption history, and BMI.

and *p,p'*-DDE⁹³ did not find an association with birth weight. Conversely, in one study, *p,p'*-DDE showed a statistically significant positive linear relationship with birth weight ($\beta = 0.012$; 95% CI: 0.003, 0.021), whereas *o,p'*-DDT was negatively associated with this outcome ($\beta = -0.014$; 95% CI: -0.028 , -0.0010).⁸¹ In the cohort from Guadeloupe, chlordecone in cord blood was not associated with birth weight. However, a U-shaped association between cord blood chlordecone concentration and birth weight was observed, particularly among pregnant women who had excessive gestational weight gain.⁵³ In a subsequent time-delayed study on the same cohort, chlordecone in cord blood was associated with a higher body mass index (BMI) in boys at 3 months of age and in girls at 8 and 18 months of age.⁵⁴ At 7 years of age, placement in the third quartile of prenatal exposure was associated with significantly increased adiposity, more markedly in boys.⁶⁰ DDE levels in the maternal serum taken at childbirth were not associated with differences in BMI or height in measurements taken from newborns during their first 38 months.⁷⁶ Similarly, prenatal DDE exposure was not associated with child growth parameters at birth and at 1, 3, 6, and 12 months of age.⁷⁰

A multinational study, which included data from Brazil, assessed prenatal exposure based on untargeted metabolic signature in maternal and umbilical samples.¹¹³ Exposure to hydroxy-chlorothalonil, a metabolite of the fungicide chlorothalonil, was identified in both types of samples. Levels of the metabolite in maternal blood showed associations with two extreme abdominal circumference fetal phenotypes: *a*) a positive association between exposure to chlorothalonil and early accelerating growth, and *b*) a negative association between exposure to the fungicide and faltering growth. In a Costa Rican cohort study aiming to analyze the association between prenatal exposure to pesticides and birth size parameters, TCPY and 2,4-D were associated with lower birth weight and smaller head circumference.⁴⁴ Furthermore, 2,4-D was associated with shorter body length, and ETU presented a nonlinear association with head circumference. In another study, no significant associations were found between levels of *p,p'*-DDE, chlorpyrifos, and organochlorine pesticides and neonatal anthropometric parameters.⁸⁰

In a cohort study, pregnant women who lived in a rural area and reported spraying pesticides at home had newborns with a greater head circumference than their counterparts ($p < 0.05$ for both cases).⁷⁸ Another similar study did not find a difference in newborns' head circumference between mothers from rural and urban regions. However, a higher proportion of newborns with head circumference below the fifth percentile of the standard curve, after correction for sex and gestational age, was observed in the rural group ($p < 0.05$).⁷⁹ Other studies that included head circumference as an outcome did not find significant results.^{81,111}

Neurodevelopmental and Other Outcomes throughout Childhood

We identified 29 publications on childhood outcomes and their association with prenatal pesticide exposure (Table 4). Among these, 22 studies investigated cognitive, sensory, or neurological development, 4 examined childhood cancers, and 3 focused on respiratory diseases in infancy.

Neurodevelopment. Five studies on neurodevelopment originated from the same Mexican prospective cohort, which recruited 1,585 women of reproductive age from 2001–2005 in four municipalities in the State of Morelos.^{65,66,68,69,71} In this region, DDT was used to eliminate the malaria mosquito until 1998. Pregnant women had DDT and DDE measured in blood during the three trimesters of pregnancy and their children were evaluated up to 60 months of age. The number of children evaluated

for neurobehavioral outcomes across the five studies varied from 167 to 270. The authors found a negative association between DDE maternal levels during the first trimester of pregnancy and the Psychomotor Development Index of the Bayley Scale II version throughout the first year of life ($\beta = -0.52$; 95% CI: -0.96 , -0.075).⁶⁶ In addition, assessments conducted between 3.5 and 5 years of age revealed negative associations between maternal DDE levels from the third trimester and general cognitive index ($\beta = -2.01$; 95% CI: -3.64 , -0.38), as well as the memory ($\beta = -1.26$; 95% CI: -2.20 , -0.32) component of the McCarthy Scale.⁶⁵ However, other investigations conducted during different periods of the neurodevelopment yielded null results.^{68,69,71} Three other studies from Mexico also originated from another cohort, which recruited pregnant women and their offspring from Mexico City between 1997 and 2001.^{72–74} Prenatal pesticide exposure was assessed by quantifying 3-PBA and TCPY levels in maternal urine samples collected during the third trimester of pregnancy. The sample size ranged from 137 to 187 mother–child pairs across the three studies. The outcomes assessed included neurodevelopment, sleep disorders, and attention deficit hyperactivity disorder (ADHD)–related behavior. No significant results were found, except for an association between prenatal TCPY levels and sleep duration ($\beta = 36.2$; 95% CI: 5.2, 67.3).⁷⁴

Seven studies came from a birth cohort from Guadeloupe archipelago.^{55,57–59,61,62,64} This cohort, named the TIMOUN Birth Cohort, included 1,068 pregnant women during their second or third trimester recruited between 2004 and 2007 in Guadeloupe. The number of children enrolled in the six studies varied from 111 to 410. They focused on prenatal, perinatal, and childhood exposure to chlordecone, a highly persistent organochlorine insecticide that was intensively used in banana fields. Concentrations of chlordecone and other environmental contaminants were measured in the umbilical cord blood from newborns in the cohort, and subsequently from blood samples provided by these children at various ages. The investigators reported negative associations with fine motor functions, visual recognition memory, novelty preference, and visual processing speed of pre- and postnatal exposures in 7- and 18-month-old children.^{57,58,61} In assessments of children at ~7 years of age, associations were observed between cord chlordecone levels and adverse effects on visual and neuromotor function, with differing effects noted between girls and boys concerning cognitive function and behavior problems.^{55,59,64} In addition, one study found no sex-typed toy preference in relation to prenatal exposure to chlordecone.⁶²

Two studies analyzed data from the community-based prospective birth cohort of the Costa Rica Caribbean, Infants' Environmental Health (ISA) study. In one study, the authors investigated the association between prenatal exposure to excess manganese (found in fungicides) and neurodevelopment at 1 year of age.⁴¹ It was observed that among boys, higher manganese levels were associated with lower social-emotional scores (β per 10-fold increase = -4.6 ; 95% CI: -8.5 , -0.8). In another study, the authors examined 2,4-D, pyrethroids, pyrimethanil, and chlorpyrifos metabolites.⁴⁵ An association was found between prenatal 2,4-D concentration and language (β per 10-fold increase = -2.0 ; 95% CI: -3.5 , -0.5) and motor (β per 10-fold increase = -2.2 ; 95% CI: -4.2 , -0.1) abilities composite scores for all children.

Three studies came from Ecuador, assessing parental occupation in the agricultural sector as the exposure.^{46,48,49} Prenatal maternal pesticide exposure was associated with deficits in the motor tests, visuospatial performance, and visual memory in one study,⁴⁶ and with lower fine motor skills and poorer visual acuity in another study.⁴⁹ In the study by Grandjean et al.,⁴⁸ prenatal

Table 4. Latin American and Caribbean studies on prenatal pesticide exposure and neurodevelopmental and other outcomes throughout childhood published between 2000 and 2024 (*n* = 29).

Author, year	Country	Exposure	Outcome	Study design (<i>n</i>)	Key findings
Neurodevelopment					
Benavides-Piracón et al., 2022 ¹¹⁵	Colombia	Environmental pesticide exposure during pregnancy	School-age children's cognitive ability	Cross-sectional (232)	No statistically significant findings ^a
Mora et al., 2018 ⁴¹	Costa Rica	ETU from maternal urine samples, Mn from maternal hair and blood samples collected across second and third trimesters	Neurodevelopmental outcomes in 1-y-old infants	Cohort (355)	Association between hair Mn levels and lower social-emotional scores among boys ($\beta = -4.6$; 95% CI: $-8.5, -0.8$) ^b
Conejo-Bolaños et al., 2024 ⁴⁵	Costa Rica	OH-PYR, 5-OH-TBZ, TCPY, DCCA, 3-PBA, and 2,4-D measured from maternal urine samples during the second and third trimesters	Children's neurodevelopmental status at 1 year of age	Cohort (355)	Prenatal 2,4-D concentrations were associated with lower language (β per 10-fold increase = -2.0 ; 95% CI: $-3.5, -0.5$) and motor (β per 10-fold increase = -2.2 ; 95% CI: $-4.2, -0.1$) abilities composite scores across all children ^c
Grandjean et al., 2006 ⁴⁸	Ecuador	Report of pesticide exposure during pregnancy	Blood pressure, neurological function, and neurobehavioral development	Cross-sectional (72)	Association between prenatal pesticide exposure and higher blood pressure ($\beta = 4.57$, SE = 1.89, $p = 0.018$), and lower Stanford-Binet copying score ($\beta = -1.01$, SE = 0.37, $p = 0.009$) ^d
Handal et al., 2008 ⁴⁶	Ecuador	Maternal occupation in the flower industry	Neurobehavioral development of infants and toddlers 3–23 months of age	Cross-sectional (121)	Children whose mothers worked in the flower industry during pregnancy scored lower in fine motor skills ($\beta = -8.04$; 95% CI: $-12.98, -3.10$) and had a higher likelihood of having poor visual acuity (aOR = 4.7; 95% CI: 1.1, 20) ^e
Harari et al., 2010 ⁴⁹	Ecuador	Paternal and maternal occupational pesticide exposure during pregnancy	Neurobehavior of schoolchildren 6–8 months of age	Cross-sectional (84)	Prenatal maternal exposure was associated with motor speed ($\beta = -7.1$; 95% CI: $-12.5, -1.6$), motor coordination (OR = 5.32; 95% CI: 1.03, 27.62), visuospatial performance ($\beta = 0.5$; 95% CI: 0.2, 1.0), and visual memory (OR = 6.62; 95% CI: 1.02, 42.93) ^f
Dallaire et al., 2012 ⁵⁷	Guadeloupe	Chlordecone concentrations from cord blood samples collected at birth	Cognitive, visual, and motor development of 7-month-old infants	Cohort (153)	Association with reduced novelty preference ($\beta = -0.19$; 95% CI: $-0.35, -0.03$) and low scores on the fine motor scale (OR = 1.26; 95% CI: 1.09, 1.47) ^g
Boucher et al., 2013 ⁵⁸	Guadeloupe	Chlordecone concentrations from cord blood samples collected at birth	Development in 18-month-old infants	Cohort (141)	Association with poorer fine motor scores ($\beta = -0.24$, $p = 0.03$) ^h
Cordier et al., 2015 ⁶¹	Guadeloupe	Chlordecone concentrations from cord blood samples collected at birth	Neurodevelopment at 18 months of age	Cohort (111)	The association with fine motor score remain significant among boys ($\beta = -10.1$; 95% CI: $-19.4, -0.8$) ⁱ
Cordier et al., 2020 ⁶²	Guadeloupe	Chlordecone concentrations from cord blood samples collected at birth	Sex-typed play behavior of 7-y-old children	Cohort (116)	No statistically significant findings
Saint-Amour et al., 2020 ⁵⁵	Guadeloupe	Chlordecone concentrations from cord blood samples collected at birth	Visual functions in school-age children (~ 7 years of age)	Cohort (285)	Association with lower contrast sensitivity ($\beta = -0.068$; 95% CI: $-0.133, -0.002$) ^j
Desrochers-Couture et al., 2022 ⁵⁹	Guadeloupe	Chlordecone concentrations from cord blood samples collected at birth	Visuospatial processing and neuro-motor function in 7-y-old children	Cohort (410)	Association with a regular frequency pattern of subtle hand tremors in dominant ($\beta = 0.02$; 95% CI: 0.002, 0.03) and nondominant hands ($\beta = 0.02$; 95% CI: 0.01, 0.04) ^k
Oulhote et al., 2023 ⁶⁴	Guadeloupe	Chlordecone concentrations from cord blood samples collected at delivery	Cognitive function and behavior problems in 7-y-old children	Cohort (391)	Association with higher processing speed scores in boys ($\beta = 0.84$; 95% CI: 0.08, 1.60) and higher internalizing problems scores in girls (aMR = 1.03; 95% CI: 1.00, 1.06) ^l
Torres-Sánchez et al., 2007 ⁶⁶	Mexico	Levels of p,p' -DDE in maternal serum samples collected during each trimester	Psychomotor and mental development during the first year	Cohort (244)	Levels of p,p' -DDE from the first trimester were associated with a reduction in the Psychomotor Development Index ($\beta = -0.52$; 95% CI: $-0.96, -0.075$) ^m
Torres-Sánchez et al., 2009 ⁶⁹	Mexico	Levels of p,p' -DDE and p,p' -DDT in maternal serum samples collected during each trimester	Psychomotor and mental development until 30 months of age	Cohort (270)	No statistically significant findings

Table 4. (Continued.)

Author, year	Country	Exposure	Outcome	Study design (n)	Key findings
Bahena-Medina et al., 2011 ⁶⁸	Mexico	Levels of DDE and <i>p,p'</i> -DDT in maternal serum samples collected during each trimester	Neonatal neurodevelopment at 1 month of age	Cohort (265)	No statistically significant findings
Torres-Sánchez et al., 2013 ⁶⁵	Mexico	Levels of DDE and DDT in maternal serum samples collected during each trimester	Neurodevelopment at 42–60 months of age	Cohort (203)	Association between third-trimester DDE levels and the general cognitive index ($\beta = -2.01$; 95% CI: -3.64 , -0.38) and the memory ($\beta = -1.26$; 95% CI: -2.20 , -0.32) components of the McCarthy scale ^a No statistically significant findings
Fortenberry et al., 2014 ⁷²	Mexico	Urinary TCPY levels from maternal third-trimester samples	Child neurodevelopment, specifically ADHD-related characteristics	Cohort (187)	No statistically significant findings
Osorio-Valencia et al., 2015 ⁷¹	Mexico	<i>p,p'</i> -DDE levels in maternal serum samples collected during the first, second, and third trimesters	Development of lateralization and spatial orientation in children 5 years of age	Cohort (167)	No statistically significant findings
Watkins et al., 2016 ⁷³	Mexico	3-PBA concentrations in maternal urine samples collected during the third trimester	Child neurodevelopment at 24 and 36 months of age	Cohort (187)	No statistically significant findings
Zamora et al., 2022 ⁷⁴	Mexico	Concentrations of 3-PBA and TCPY in maternal urinary samples collected during the third trimester	Sleep duration, timing, and fragmentation during adolescence	Cohort (137)	Higher tertile of TCPY levels was associated with longer sleep duration ($\beta = 36.2$; 95% CI: 5.2 , 67.3) ^c
Jenkins et al., 2024 ¹¹⁶	Puerto Rico	Levels of glyphosate and its derivative AMPA quantified from maternal urine samples collected across the second and the third trimester of pregnancy	Children's neurodevelopmental status at 6, 12, and 24 months of age	Cohort (143)	At 12 months of age, associations were found between AMPA concentration and communication score (% change = 5.32 ; 95% CI: 9.04 , 1.61), as well as the receptive (% change = 13.20 ; 95% CI: 21.97 , 4.43) communication subdomain. At 24 months of age, there was negative association between AMPA concentration and communication score (% change = 7.00 ; 95% CI: 11.75 , 2.26) and related subdomain scaled scores receptive (% change = 14.03 ; 95% CI: 24.82 , 3.25) and expressive (% change = 14.94 ; 95% CI: 26.78 , 3.11). Negative association between AMPA concentration and cognitive score (% change = 4.02 ; 95% CI: 6.72 , 1.32) and associated subdomains attention and memory (% change = 9.54 ; 95% CI: 17.21 , 1.86) and perception and concepts (% change = 8.10 ; 95% CI: 14.32 , 1.87)
Leukemia					
Ferreira et al., 2012 ⁵⁰	Brazil	Exposure to chemicals (including pesticides) during pregnancy	Childhood leukemia	Case-control (833)	Association between childhood leukemia development and maternal chemical exposure during pregnancy (OR = 1.35 ; 95% CI: 1.16 , 2.58)
Ferreira et al., 2013 ⁵¹	Brazil	Pesticide exposure during the 3 months before pregnancy (periconceptual period) or throughout each trimester of pregnancy	Childhood leukemia	Case-control (675)	Pesticide use at any time during pregnancy was associated with ALL (aOR = 2.10 ; 95% CI: 1.14 , 3.86) and AML (aOR = 5.01 ; 95% CI: 1.97 , 12.7) in children 0–11 months of age, and with ALL (aOR = 1.88 ; 95% CI: 1.05 , 5.23) in children 12–23 months of age ^d
Monge et al., 2007 ⁵⁹	Costa Rica	Occupational, environmental, and home pesticide exposures of both parents	Childhood leukemia	Case-control (879)	Maternal pesticide exposure during the first trimester or second trimester was associated with childhood leukemia (aOR = 22.0 ; 95% CI: 2.8 , 171.5 and aOR = 4.5 ; 95% CI: 1.4 , 14.7 , respectively). Paternal pesticide exposure during the second trimester also showed an association (aOR = 1.5 ; 95% CI: 1.0 , 2.3) ^d

Table 4. (Continued.)

Author, year	Country	Exposure	Outcome	Study design (n)	Key findings
Hyland et al., 2018 ⁴⁰	Costa Rica	Parental occupational and residential exposure to pesticides	Childhood ALL	Case-control (828)	Reported pregnancy exposure to insecticides was associated with ALL among boys (OR = 1.75; 95% CI: 1.13, 2.73) ^c
Respiratory infections Mora et al., 2020 ⁴²	Costa Rica	Concentrations of ETU, TCPY, 3-PBA, 2,4-D, DCCA, OH-PYR, and 5-OH-TBZ in maternal urine samples collected during each trimester of pregnancy	Respiratory health in the first year of life	Cohort (355)	Association between high ETU concentrations and wheezing (OR = 0.50; 95% CI: 0.26, 0.96)
Islam et al., 2022 ⁴³	Costa Rica	Chlorpyrifos, pyrethroids, mancozeb, pyrimethanil, thiamethoxam and 2, 4-D measured from maternal urine samples collected during pregnancy	Respiratory and allergic outcomes among 5-y-old children	Cohort (303)	No statistically significant findings
Cupul-Uicab et al., 2014 ⁷⁷	Mexico	<i>p,p'</i> -DDE and <i>p,p'</i> -DDT levels from maternal serum samples collected within a day of delivery	Recurrent lower respiratory tract infections during childhood among boys	Cohort (747)	No statistically significant findings

Note: % change, percentage change; 2,4-D, 2,4-dichlorophenoxyacetic acid; 3-PBA, 3-phenoxybenzoic acid; 5-OH-TBZ, hydroxythiabendazole; ADHD, attention deficit hyperactivity disorder; ALL, acute myeloid leukemia; AMPA, aminomethylphosphonic acid; aOR, adjusted odds ratio; BMI, body mass index; CI, confidence interval; DCCA, 3-(2,2-dichlorovinyl)-2,2-dimethylcyclopropanecarboxylic acid; DDE, dichlorodiphenyldichloroethylene; DDT, dichlorodiphenyltrichloroethane; ETU, ethylene thiourea; HOME, Home Observation for Measurement of the Environment; IQ, intelligence quotient; Mn, manganese; OH-PYR, hydroxypyrimethanil; OR, odds ratio; PCB, polychlorinated biphenyl; SE, standard error; TCPY, 3,5,6-trichloro-2-pyridinol.

^aAdjusted for child's sex, maternal educational attainment, and low height for age.

^bAdjusted for maternal education, parity, sex, gestational age at birth, child age, HOME score, and location of neurodevelopmental assessment.

^cAdjusted for age, sex, and body weight.

^dAdjusted for anemia, stimulation at home, housing construction.

^eAdjusted for the child's sex, age, BMI, number of daily meals, stunting, hematoctrit, school grade, having repeated one grade, maternal education level, family living in a traditional house, drinking-water supply, and paternal education and employment.

^fAdjusted for prepregnancy maternal employment status, cord plasma lipid concentrations, infant vitamin supplementation, postnatal exposure to maternal tobacco smoke, number of adults living with the infant, and child examiner and for plasma lipid concentrations, age of infant at testing, birthplace of mother, and domestic violence.

^gAdjusted for sex, maternal education, and child age.

^hAdjusted for maternal educational level and child age.

ⁱAdjusted for child sex and age, maternal IQ, maternal education, maternal alcohol intake during pregnancy, and plasma lipids.

^jAdjusted for child sex and age, as well as maternal Raven score and prenatal PCB-153 levels.

^kAdjusted for child's age and sex, maternal age, parity, Raven score, education, marital status, monthly household income, and alcohol and smoking during pregnancy.

^lAdjusted for birth weight, age at evaluation, breastfeeding, and HOME score at 6 months of age.

^mAdjusted for child's age at examination, mother's education, marital status when child was 4 years of age, and HOME score, sex of child, and attendance at a child care center.

ⁿAdjusted for maternal characteristics: education, parity, marital status, specific gravity, and adolescent characteristics: sex and age.

^oAdjusted for use of oral contraceptives during pregnancy, maternal age and education, child's birth weight and skin color.

^pAdjusted for residence (urban or rural).

^qAdjusted year of birth and socioeconomic status.

pesticide exposure was associated with higher blood pressure and lower visuospatial parameter scores.

In a cross-sectional study conducted in Bogotá, prenatal exposure to pesticides was not associated with cognitive abilities in children.¹¹⁵ In a prospective cohort study from Puerto Rico, the relationship between prenatal exposure to glyphosate and children's neurobehavioral development at 6, 12, and 24 months of age was assessed.¹¹⁶ No statistically significant associations were found between glyphosate or AMPA concentrations and the neurobehavioral scores at 6 months of age. However, at 12 months of age, associations were observed between AMPA concentration and the communication score, whereas at 24 months of age, associations included motor, communication, and cognitive domain scores.

Childhood leukemia. Four articles investigated the association between prenatal pesticide exposure and childhood leukemia. In a case-control study, an association was found between maternal exposure to chemicals (including pesticides) during pregnancy and childhood leukemia (OR = 1.36; 95% CI: 1.16, 1.59).⁵⁰ In another article, children born to women who had at least one exposure to pesticides during pregnancy had a greater likelihood of developing acute lymphoid leukemia (OR = 2.10; 95% CI: 1.14, 3.86) and acute myeloid leukemia (OR = 5.01; 95% CI: 1.97, 12.70) in children 0–11 months of age, and acute lymphoid leukemia (OR = 1.88; 95% CI: 1.05, 5.23) in children 12–23 months of age.⁵¹ In a different investigation, maternal exposures to any pesticides at any time (OR = 2.2; 95% CI: 1.0, 4.8), during the year before conception (OR = 2.4; 95% CI: 1.0, 5.9), and during the first (OR = 22.0; 95% CI: 2.8, 171.5) and second (OR = 4.5; 95% CI: 1.4, 14.7) trimesters were associated with an increased risk of having a child with leukemia.³⁹ In addition, an association was found for paternal exposures to any pesticides during the second trimester (OR = 2.2; 95% CI: 1.0, 4.8). In the study by Hyland et al.,⁴⁰ maternal insecticide use during pregnancy, but not herbicide use, was associated with increased odds of acute lymphoid leukemia among boys (OR = 1.75; 95% CI: 1.13, 2.73). This association appeared to be dose dependent ($p_{\text{trend}} = 0.03$).

Respiratory infections. Three studies explored the association between prenatal pesticide exposure and respiratory infections in early life. Cupul-Uicab et al.⁷⁷ did not find an association between prenatal exposure to *p,p'*-DDE or *p,p'*-DDT and lower respiratory tract infections (LRTI) in infants up to 2 years of age. In the cohort study by Mora et al.,⁴² fungicides and organophosphate (OP) insecticides were not statistically associated with LRTI, but high levels of ETU (a metabolite of mancozeb) in the second half of pregnancy were associated with reduced odds of wheezing during the first year of life (OR = 0.50; 95% CI: 0.26, 0.96). Subsequent investigation on the same cohort at 5 years of age also did not reveal an association between metabolites of pesticides and LRTI.⁴³

Molecular Effects

We found six observational human studies examining the molecular effects of prenatal pesticide exposure (Table 5). A study from Argentina aimed to determine the presence of environmental xenobiotics in the maternal blood of mothers of children with or without congenital anomalies.⁹⁵ In addition, the authors investigated the correlation between the xenobiotics present in maternal blood and genotoxicity parameters using the comet assay and the cytome assay. In the comet assay, β -HCH was positively correlated with the damage index. Furthermore, in the cytome assay, positive correlations between genotoxicity parameters were observed with levels of hexachlorobenzene (HCB), δ -HCH, lindane, exo-heptachlor epoxide, and methyl chlorpyrifos. In another study conducted in Argentina, placental

parameters, besides target enzymes, erythrocyte osmotic fragility and DNA damage from umbilical blood cells were assessed.¹¹⁷ The study aimed to investigate whether newborns of mothers living in a rural area with intense use of OPs had alterations in placental or biomarkers of oxidative stress. The researchers observed that the placental weight and placental index were higher in the group exposed to pesticides compared with the control group. In addition, higher erythrocyte osmotic fragility, higher DNA damage index, and lower superoxide dismutase (SOD) activity were observed in the rural group from the spray season compared with the rural group from the nonspray season. Acetylcholinesterase and SOD activities were found to be inversely correlated with the DNA damage index. In a Brazilian study comparing chromosomal damage in peripheral blood lymphocytes and oral epithelial cells between pregnant women from rural and urban areas, no statistical difference was found in the results of the micronucleus test.¹¹⁸

Three studies have investigated the effects of pesticide exposure on thyroid hormone levels. In a study from the TIMOUN cohort, umbilical cord chlordecone levels were associated with increased thyroid-stimulating hormone (TSH) levels at 3 months of age in boys.⁶¹ In another study conducted with the same cohort, the association between *in utero* exposure to chlordecone and levels of thyroid, metabolic, and sex-steroid hormones at 7 years of age was investigated.⁶³ Higher levels of chlordecone were associated with higher levels of TSH in girls, as well with the hormones dehydroepiandrosterone (DHEA), total testosterone (TT) and dihydrotestosterone (DHT) in both sexes. In a birth cohort study from Bolivia, newborn TSH levels were not significantly associated with *p,p'*-DDE or *o,p'*-DDT quantified in serum from cord blood.⁸¹

Discussion

Latin America constitutes a geographical region with pronounced reliance on agricultural activities as a source of income.^{2,119} The global rise in agricultural productivity over the past decades has co-occurred with an increase in the production and consumption of formulated pesticides.^{3–5,120–122} A large amount of pesticides are formulated and commercialized by enterprises situated outside of LAC.^{123–125} However, due to rigorous application of regulations within their countries of production, the use of some of these compounds is exclusively permissible in middle- and low-income nations, and not in the countries where they are produced.^{126,127} Previous literature reviews have considered the accumulating evidence regarding health risks associated with environmental contamination resulting from pesticide use.^{33,128} Compared with the review by Zúñiga-Venegas et al.,³³ our review exclusively focused on prenatal pesticide exposure and broadened the search to include six databases, as well as extending the publication date range to a larger and more recent time frame. Of the total 80 studies included in our review, 44 were not included on the previous review. The present scoping review identified a substantial body of literature from LAC concerning the potential risks associated with pesticide exposure during pregnancy. The findings underscore the importance of allocating resources from the generated agricultural wealth toward the investigation and mitigation of the repercussions linked to pesticide use in LAC agriculture.

Some studies included in the present review indicated an association between prenatal pesticide exposure and preterm birth. In animal studies, it was observed that exposure of pregnant rabbits to DDT increased the incidence of prematurity.¹²⁹ Given that DDT and its metabolites possess the capability to interact with progesterone receptors in an antagonistic manner,¹³⁰ exposure to DDT during pregnancy could potentially impact the duration of gestation. When human endometrial endothelial cells were

Table 5. Latin American and Caribbean studies on prenatal pesticide exposure and molecular adverse effects published between 2000 and 2024 (*n* = 6).

Author, year	Country	Exposure	Outcome/biological sample	Study design (<i>n</i>)	Key findings
Quintana et al., 2017 ¹¹⁷	Argentina	Living in an area with intensive pesticide application	Red blood cell osmotic fragility and cell viability	Cross-sectional (151)	Higher erythrocyte osmotic fragility and a higher DNA damage index were observed in the rural group from the spray season compared with the rural group from the nonspray season
Colussi et al., 2024 ⁹⁵	Argentina	OCPs, OPs, and PYR quantified from maternal blood samples collected after delivery	Genotoxicity assessed using the comet assay and the cytochrome assay	Case-control (290)	β-HCH was positively correlated with the damage index comet assay ($\rho = 0.354$, $p = 0.003$). HCB, δ-HCH, lindane, exo-heptachlor epoxide, and methyl chlorpyrifos were positively correlated with genotoxicity parameters
Arrebola et al., 2016 ⁸¹	Bolivia	<i>o,p'</i> -DDT and <i>p,p'</i> -DDE levels quantified in serum from cord blood samples collected after delivery	Newborn TSH levels	Cohort (200)	No statistically significant findings
Silva et al., 2019 ¹¹⁸	Brazil	Environmental and occupational exposure to agrochemicals	Chromosomal damage in peripheral blood lymphocytes and from oral epithelium samples	Cross-sectional (23)	No statistically significant findings
Cordier et al., 2015 ⁶¹	Guadeloupe	Chlordecone concentrations from cord blood samples collected at birth	TSH levels at 3 months of age	Cohort (111)	Boys from the higher exposure group showed 61.3% (95% CI: 2.0, 155.0) change in TSH levels ^a
Ayhan et al., 2021 ⁶³	Guadeloupe	Organochlorine levels measured from umbilical cord blood samples collected at birth	Levels of TSH, FT3, FT4, IGF-1, adiponectin, leptin, DHEA, TT, and E2 hormones in children at 7 years of age	Cohort (438)	The third quartile of <i>in utero</i> chlordecone exposure relative to the lowest quartile was associated with higher TSH ($\beta = 0.22$; 95% CI: 0.01, 0.44) levels in girls and higher DHEA ($\beta = 0.54$; 95% CI: 0.08, 1.01, for boys; $\beta = 0.36$; 95% CI: 0.02, 0.71, for girls), TT (aOR = 3.22; 95% CI: 1.08, 9.6, for boys; aOR = 3.28; 95% CI: 1.32, 8.17, for girls), and DHT (aOR = 3.70; 95% CI: 1.29, 10.6, for boys; aOR = 3.20; 95% CI: 1.01, 10.2, for girls) levels ^b

Note: aOR, adjusted odds ratio; β-HCH, beta-hexachlorocyclohexane; BMI, body mass index; CI, confidence interval; δ-HCH, delta-hexachlorocyclohexane; DDE, dichlorodiphenyldichloroethylene; DDT, dichlorodiphenyltrichloroethane; DHEA, dehydroepiandrosterone; DHT, dihydrotestosterone; E2, estradiol; FT3, tri-iodothyronine; FT4, thyroxine; HCB, hexachlorobenzene; IGF-1, insulin growth-factor 1; OCPs, organochlorine pesticides; OPs, organophosphates; PCB, polychlorinated biphenyl; PYR, pyrethroids; TSH, thyroid-stimulating hormone; TT, total testosterone.

^aAdjusted for hemolysis status, sample storage duration, level of education, ln PCB-153 levels, ln *p,p'*-DDE levels, and total lipids.

^bAdjusted for mothers' age at delivery, smoking during pregnancy, alcohol during pregnancy, mothers' BMI in early pregnancy, child z-score BMI, geographical origin, and breastfeeding depending on the outcome.

exposed to DDT, there was a decrease in the proportion of viable cells and an increase in the proportion of necrotic cells.¹³¹ Furthermore, administering DDT to rabbit embryos prior to the implantation stage resulted in intrauterine growth restriction,¹³² and carbamate pre- and postimplantation in pregnant rats reduced fetal weight.¹³³ Exposure to a mixture of pesticides can lead to a decrease in birth weight, even at dose levels considerably lower than the no observed adverse effect levels of a single pesticide within the mixture.¹³⁴

A possible mechanism involved in adverse obstetric outcomes is the functional impairment of the placenta due to oxidative stress.^{117,135} Oxygen is a fundamental component for cellular metabolism, facilitating various biochemical reactions, and it is particularly crucial during the initial stages of development, where it influences key processes, such as cell differentiation, tissue growth, and organ formation.¹³⁶ Pesticides increase oxidative stress and induce the formation of oxidized lipids.^{137,138} Exposure to OP pesticides leads to oxidative stress owing to decreased glutathione and increased

reactive oxygen species.¹³⁹ Moreover, organochlorine pesticide residues found in the placental tissue of individuals with preterm delivery were associated with oxidative stress.¹⁴⁰ Further studies are needed to accurately understand the molecular mechanisms involved in oxidative stress triggered by pesticide exposure.

Other pathophysiological mechanisms should be taken into consideration. Some pesticides act as androgen-receptor antagonists, inhibitors of steroidogenic enzymes, or both, thus impairing endocrine signal pathways.^{141–147} As a consequence, exposure during critical periods of development can trigger alterations in signal pathways in the tissues of the male reproductive system, causing anomalies of the external genitalia.^{148–150} Anogenital distance has also been an androgen-dependent developmental biomarker applied to assess the effects of endocrine-disrupting pesticides.¹⁵¹ Although outlying measurements of birth weight, birth length, and head circumference are not directly associated with a specific dysfunction or disease, these parameters are conventionally used to evaluate and

predict intrauterine development and possible alterations in growth.¹⁵² In addition, anogenital distance is a well-established marker of androgen exposure in humans.¹⁵³ In the present review, we identified a set of studies evaluating anthropometrics parameters, including anogenital distance, penile dimensions, and the proportion of male newborns. Nonetheless, a subset of the investigations failed to incorporate standardized anthropometric measurements, such as z-scores for weight and length stratified by both sex and gestational age, which would facilitate a more robust interstudy comparison.

Among the studies assessing congenital anomalies, particular attention was given to male genital anomalies such as hypospadias and cryptorchidism. With the exception of one study that focused on hypospadias and had a small number of participants, the overall results indicate an association between prenatal exposure to pesticides and anomalies of the genitourinary system. Other congenital anomalies, such as anomalies in the circulatory and digestive systems, and cleft lip and palate, as well as infant mortality caused by congenital anomalies in the CNS and CVS, were also associated with prenatal exposure to pesticides. Some previous works have estimated the association by using exposure to organochlorines as a measure, which, as mentioned above, have an androgenic antagonist action.¹⁴⁴ However, other studies relied on indirect measures of exposure, such as parental occupation and place of residence, or employed ecological approaches. In these cases, findings are more susceptible to confounding factors and should be interpreted with caution.

Epidemiological studies have demonstrated an association between pesticide exposure and respiratory outcomes among farmworkers.^{154–156} Among the studies included in the present review, we observed limited evidence concerning the association between prenatal exposure and respiratory outcomes in children.

We observed consistency among the results from studies in which the outcomes were childhood leukemia, which corresponds to the results of other studies conducted outside of LAC countries.^{15,157} Despite the fact that the majority of childhood leukemia cases do not have a clear etiology,¹⁵⁸ it has been proposed that certain chemicals, including pesticides, exhibit toxic mechanisms that act on early hematopoietic stem and progenitor cells. These mechanisms involve topoisomerase II inhibition properties and the excessive generation of free radicals, which can lead to DNA single- and double-strand breaks.^{159,160} The relationship between leukemia and pesticide exposure is still a relatively underexplored field in LAC. Given the morbidity and mortality associated with this disease, future studies aimed at understanding this relationship could potentially contribute to improving the childhood health indexes across countries from LAC.

We found a substantial body of literature evaluating the relationship between prenatal pesticide exposure and neurodevelopment. Notably, three Latin American cohorts were assessed for neurodevelopment parameters across different time points of childhood. Among the results of these studies, a positive association was reported between pesticide exposure and deficits in fine motor skills. In a similar way, tremors were reported among rat pups exposed to chlordecone,¹⁶¹ as well as among human adults experiencing acute exposure.¹⁶² According to the data available, authors have highlighted that chlordecone-induced tremors may involve multiple neurotransmitter systems.⁵⁸ Another possible explanation for these findings is the short-term visual memory measured by the novelty preference. Difficulty with recent memory was reported among the symptoms from workers intoxicated with chlorinated pesticides.¹⁶² Memory impairment was also observed in preweaning rats exposed to chlordecone.^{163,164} Specific to the findings on contrast sensitivity, researchers have proposed that contrast sensitivity may be mediated by dopaminergic-related mechanisms.⁵⁵

Considering the aims and the inclusion criteria for this review, we did not include experimental studies. However, findings from studies examining the molecular effects of pesticide exposure on pregnant women, including genotoxicity tests, were presented.^{95,117,118} In addition, other studies investigated TSH levels as potential mediators of adverse outcomes.^{61,81} Although not the focus of this review, we underscore the significance of exploring the potential hazards associated with pesticide exposure during pregnancy through experimental studies, particularly those addressing placental health and development. Enhanced comprehension of how pesticides influence the physiology of pregnant women and the development of embryo and fetus, as elucidated by molecular mechanisms in *in vitro* studies and animal models, holds considerable value for interpreting findings from observational human studies.

Taking into account the body of studies presented, we would like to highlight some challenges we observed among these studies. The first challenge is related to the complexity of directly assessing pesticide exposure. Occupational exposure, proximity to cultivation areas, self-reported pesticide application, and rural/urban residence were among the indirect measures used as exposure. A limited number of studies (55.0%) have directly assessed exposure biomarkers in umbilical cord, maternal blood, or urine samples. A second challenge for researchers on this topic is the variations in the timing of sample collection (e.g., trimester of pregnancy) across these studies could potentially yield disparate outcomes, which may account for divergences observed among the different studies' results. The third challenge is that some outcomes are rarely observed, and prospective research on larger sample populations may be necessary to draw more meaningful results.

We have identified gaps in the literature that should be explored in future research. Considering that new pesticides are released for application each year, we emphasize the need to examine the effects resulting from prenatal exposure to pesticides other than organochlorines. Particularly, attention should be given to the effects of pesticides banned in high-income countries but widely used in Latin America. It is worth mentioning that the present review identified studies from 13 of the 45 LAC countries, denoting a lack of research on prenatal pesticide exposure and its adverse outcomes in certain LAC countries (Anguilla, Antigua and Barbuda, Aruba, Bahamas, Barbados, Belize, British Virgin Islands, Cayman Islands, Cuba, Dominican, Dominican Republic, El Salvador, Falkland Islands (Malvinas), French Guiana, Grenada, Guyana, Haiti, Honduras, Jamaica, Martinique, Montserrat, Nicaragua, Panama, Saint Kitts and Nevis, Saint Lucia, Saint Vincent and the Grenadines, Suriname, Trinidad and Tobago, Turks and Caicos Islands, US Virgin Islands, Uruguay, and Venezuela). Local efforts to investigate the effects of prenatal pesticide exposure can lead to more effective policies. In addition, another aspect deserving attention is the impact of prenatal pesticide exposure on socially vulnerable groups. Many pregnant women exposed to pesticides in LAC hail from rural communities, indigenous populations, riverine communities, or labor under poorly regulated and protected conditions.¹⁶⁵ The compound effect of prenatal pesticide exposure and social vulnerability can yield even more pronounced consequences that should be studied. Furthermore, no studies covering multiple LAC countries were identified. Given the shared characteristics and unique aspects related to pesticides regularization, collaborative studies between countries can also yield new insights to enhance the health of the LAC population. Finally, there is a dearth of prospective epidemiological studies that use a high-quality design and statistical approach. In this context, we advocate for the advancement of research endeavors that directly measure exposure levels through biomarkers in maternal blood or urine samples, among other

methods. In addition, it is imperative to evaluate exposure during each trimester and its association with adverse outcomes. Prospective studies offer a viable approach to accomplishing these practices, facilitating more precise and wide collection of data, and consequently yielding more robust findings. Given the unique characteristics of LAC, gaining a deeper understanding of the effects of environmental factors on maternal and child health is an area that demands robust scientific investment. This can significantly contribute to formulating more effective public policies for this population.

Our study has some noteworthy aspects. It synthesizes the scientific output from observational human studies concerning prenatal exposure to pesticides and adverse outcomes in LAC, a region characterized by lax regulatory frameworks and consequent high pesticide consumption. Our methodology enabled us to identify and assess studies not only in English but also in Portuguese and Spanish, the primary languages of the region. This approach ensured the inclusion of articles from both international and local journals. By focusing on prenatal exposure, our study consistently captured scientific output on teratology, a field of research essential for informing regulatory decisions concerning products with potential human health hazards. Our study also has some limitations. We did not systematically evaluate the quality of the studies or summarize their results quantitatively, because this was not the purpose of the scoping review. Our search strategy was defined based on three main terms: pesticides, pregnancy, and Latin America. We did not include the names of specific pesticides, which may have limited the articles recovered in the search. However, we included MeSH terms to expand the search strategy to incorporate related terms into the searches. Furthermore, our study was restricted to observational studies involving human participants. This decision was made to assess the real-life impact of pesticide use across LAC. However, it inevitably excluded numerous experimental studies conducted within the region on this topic. We could search for and read articles only in Portuguese, Spanish, and English. However, there may be scientific output in other languages within the region that was not included in this review.

In summary, we identified studies conducted in LAC that evaluated prenatal exposure to pesticides and this exposure's potential associations with adverse obstetric and neonatal outcomes, congenital anomalies, anthropometric parameters, neurodevelopment, childhood leukemia, and respiratory infections. In addition, we identified a limited number of studies assessing molecular effects. Given that a significant portion of the LAC populations are exposed to pesticides and considering the diverse range of adverse effects reported in the scientific literature due to pesticide exposure, continued research into the risks of prenatal exposure is essential. There are substantial gaps in the existing knowledge, particularly concerning groups of pesticides beyond organochlorines. It is crucial to consider the unique characteristics of the LAC region, especially regarding investigating risks associated with pesticides that are banned in some countries but permitted in LAC.

Acknowledgments

We thank Conselho Nacional de Desenvolvimento Científico e Tecnológico–Brazil (CNPq) and Coordenação de Aperfeiçoamento de Pessoal de Nível Superior (CAPES) for supporting this research through scholarship awarded to R. Rohweder and S. S. Arteaga. We also thank D.M. Santin for assisting with developing the search strategy and J. French for editing the English text.

References

1. ECLAC (UN Economic Commission for Latin America and the Caribbean), United Nations. n.d. Country's demographic profile. <https://www.cepal.org/en/subtopics/>

- demographic-projections/latin-america-and-caribbean-population-estimates-and-projections/countrys-demographic-profile [accessed 14 November 2023].
2. OECD (Organisation for Economic Co-operation and Development), FAO (Food and Agriculture Organization of the United Nations). 2019. *OECD-FAO Agricultural Outlook 2019–2028*. https://doi.org/10.1787/agr_outlook-2019-en.
3. Trindade FJ, Fulginiti LE. 2015. Is there a slowdown in agricultural productivity growth in South America? *Agric Econ* 46(S1):69–81, <https://doi.org/10.1111/agec.12199>.
4. Carvalho FP. 2006. Agriculture, pesticides, food security and food safety. *Environ Sci Policy* 9(7–8):685–692, <https://doi.org/10.1016/j.envsci.2006.08.002>.
5. Zhang W, Jiang F, Ou J. 2011. Global pesticide consumption and pollution: with China as a focus. *Proc Int Acad Ecol Environ Sci* 1:125–144.
6. Melgarejo L, Morales González JC, Burity VTA, Prates LA, Cortes Rocha N. 2020. *Agrotóxicos na América Latina: violações contra o direito humano à alimentação e à nutrição adequadas: informe regional 2020*. <https://fianbrasil.org.br/wp-content/uploads/2021/04/Agrotoxicos-na-America-Latina-Portugues.pdf> [accessed 14 November 2023].
7. WHO (World Health Organization), FAO. 2010. *International Code of Conduct on the Distribution and Use of Pesticides: Guidelines for the Registration of Pesticides*. <https://www.who.int/publications/i/item/who-htm-ntd-whopes-2010.7> [accessed 25 September 2023].
8. Carson R. 1962. *Silent Spring*. Boston, MA: Houghton Mifflin Company.
9. Evangelou E, Ntritsos G, Chondrogiorgi M, Kavvoura FK, Hernández AF, Ntzani EE, et al. 2016. Exposure to pesticides and diabetes: a systematic review and meta-analysis. *Environ Int* 91:60–68, PMID: 26909814, <https://doi.org/10.1016/j.envint.2016.02.013>.
10. Muñoz-Quezada MT, Lucero BA, Iglesias VP, Muñoz MP, Cornejo CA, Achu E, et al. 2016. Chronic exposure to organophosphate (OP) pesticides and neuropsychological functioning in farm workers: a review. *Int J Occup Environ Health* 22(1):68–79, PMID: 27128815, <https://doi.org/10.1080/10773525.2015.1123848>.
11. Matich EK, Laryea JA, Seely KA, Stahr S, Su LJ, Hsu PC. 2021. Association between pesticide exposure and colorectal cancer risk and incidence: a systematic review. *Ecotoxicol Environ Saf* 219:112327, PMID: 34029839, <https://doi.org/10.1016/j.ecoenv.2021.112327>.
12. Wan MLY, Co VA, El-Nezami H. 2022. Endocrine disrupting chemicals and breast cancer: a systematic review of epidemiological studies. *Crit Rev Food Sci Nutr* 62(24):6549–6576, PMID: 33819127, <https://doi.org/10.1080/10408398.2021.1903382>.
13. Pourhassan B, Meysamie A, Alizadeh S, Habibian A, Beigzadeh Z. 2019. Risk of obstructive pulmonary diseases and occupational exposure to pesticides: a systematic review and meta-analysis. *Public Health* 174:31–41, PMID: 31306887, <https://doi.org/10.1016/j.puhe.2019.05.024>.
14. Rodrigues MDB, Carvalho DSD, Chong-Silva DC, Urrutia-Pereira M, Albuquerque GSCD, Cieslak F, et al. 2022. Association between exposure to pesticides and allergic diseases in children and adolescents: a systematic review with meta-analysis. *J Pediatr (Rio J)* 98(6):551–564, PMID: 34982974, <https://doi.org/10.1016/j.jped.2021.10.007>.
15. Karalexi MA, Tagkas CF, Markozannes G, Tseretopoulou X, Hernández AF, Schüz J, et al. 2021. Exposure to pesticides and childhood leukemia risk: a systematic review and meta-analysis. *Environ Pollut* 285:117376, PMID: 34380208, <https://doi.org/10.1016/j.envpol.2021.117376>.
16. Biosca-Brull J, Pérez-Fernández C, Mora S, Carrillo B, Pinos H, Conejo NM, et al. 2021. Relationship between autism spectrum disorder and pesticides: a systematic review of human and preclinical models. *Int J Environ Res Public Health* 18(10):5190, PMID: 34068255, <https://doi.org/10.3390/ijerph18105190>.
17. Pinos H, Carrillo B, Merchán A, Biosca-Brull J, Pérez-Fernández C, Colomina MT, et al. 2021. Relationship between prenatal or postnatal exposure to pesticides and obesity: a systematic review. *Int J Environ Res Public Health* 18(13):7170, PMID: 34281107, <https://doi.org/10.3390/ijerph18137170>.
18. Goodman M, Mandel JS, DeSesso JM, Scialli AR. 2014. Atrazine and pregnancy outcomes: a systematic review of epidemiologic evidence. *Birth Defects Res B Dev Reprod Toxicol* 101(3):215–236, PMID: 24797711, <https://doi.org/10.1002/bdrb.21101>.
19. Khoshhali M, Davoodi S, Ebrahimpour K, Shoshtari-Yeganeh B, Kelishadi R. 2020. The association between maternal exposure to organophosphate pesticides and neonatal anthropometric measures: a systematic review and meta-analysis. *J Res Med Sci* 25(1):79, PMID: 33088316, https://doi.org/10.4103/jrms.JRMS_919_19.
20. Addissie YA, Kruzka P, Troia A, Wong ZC, Everson JL, Kozel BA, et al. 2020. Prenatal exposure to pesticides and risk for holoprosencephaly: a case-control study. *Environ Health* 19(1):65, PMID: 32513280, <https://doi.org/10.1186/s12940-020-00611-z>.
21. Brender JD, Weyer PJ. 2016. Agricultural compounds in water and birth defects. *Curr Environ Health Rep* 3(2):144–152, PMID: 27007730, <https://doi.org/10.1007/s40572-016-0085-0>.

22. García AM, Benavides FG, Fletcher T, Orts E. 1998. Paternal exposure to pesticides and congenital malformations. *Scand J Work Environ Health* 24(6):473–480, PMID: 9988089, <https://doi.org/10.5271/sjweh.371>.
23. García AM, Fletcher T, Benavides FG, Orts E. 1999. Parental agricultural work and selected congenital malformations. *Am J Epidemiol* 149(1):64–74, PMID: 9883795, <https://doi.org/10.1093/oxfordjournals.aje.a009729>.
24. Garry VF, Harkins ME, Erickson LL, Long-Simpson LK, Holland SE, Burroughs BL. 2002. Birth defects, season of conception, and sex of children born to pesticide applicators living in the Red River Valley of Minnesota, USA. *Environ Health Perspect* 110(suppl 3):441–449, PMID: 12060842, <https://doi.org/10.1289/ehp.02110s3441>.
25. García-Rodríguez J, García-Martín M, Nogueras-Ocaña M, de Dios Luna-del-Castillo J, Espigares García M, Olea N, et al. 1996. Exposure to pesticides and cryptorchidism: geographical evidence of a possible association. *Environ Health Perspect* 104(10):1090–1095, PMID: 8930551, <https://doi.org/10.1289/ehp.104-1469503>.
26. Kristensen P, Irgens LM, Andersen A, Bye AS, Sundheim L. 1997. Birth defects among offspring of Norwegian farmers, 1967–1991. *Epidemiology* 8(5):537–544, PMID: 9270956, <https://doi.org/10.1097/00001648-199709000-00014>.
27. Loffredo CA, Silbergeld EK, Ferencz C, Zhang J. 2001. Association of transposition of the great arteries in infants with maternal exposures to herbicides and rodenticides. *Am J Epidemiol* 153(6):529–536, PMID: 11257060, <https://doi.org/10.1093/aje/153.6.529>.
28. Nurminen T, Rantala K, Kurppa K, Holmberg PC. 1995. Agricultural work during pregnancy and selected structural malformations in Finland. *Epidemiology* 6(1):23–30, PMID: 7888440, <https://doi.org/10.1097/00001648-199501000-00006>.
29. Rappazzo KM, Warren JL, Meyer RE, Herring AH, Sanders AP, Brownstein NC, et al. 2016. Maternal residential exposure to agricultural pesticides and birth defects in a 2003 to 2005 North Carolina birth cohort. *Birth Defects Res A Clin Mol Teratol* 106(4):240–249, PMID: 26970546, <https://doi.org/10.1002/bdra.23479>.
30. Thulstrup AM, Bonde JP. 2006. Maternal occupational exposure and risk of specific birth defects. *Occup Med (Lond)* 56(8):532–543, PMID: 17151389, <https://doi.org/10.1093/occmed/kql115>.
31. Weidner IS, Møller H, Jensen TK, Skakkebaek NE. 1998. Cryptorchidism and hypospadias in sons of gardeners and farmers. *Environ Health Perspect* 106(12):793–796, PMID: 9831539, <https://doi.org/10.1289/ehp.98106793>.
32. Kalliora C, Mamoulakis C, Vasilopoulos E, Stamatiades GA, Kalafati L, Barouni R, et al. 2018. Association of pesticide exposure with human congenital abnormalities. *Toxicol Appl Pharmacol* 346:58–75, PMID: 29596925, <https://doi.org/10.1016/j.taap.2018.03.025>.
33. Zúñiga-Venegas LA, Hyland C, Muñoz-Quezada MT, Quirós-Alcalá L, Butinof M, Buralli R, et al. 2022. Health effects of pesticide exposure in Latin American and the Caribbean populations: a scoping review. *Environ Health Perspect* 130(9):096002, PMID: 36173136, <https://doi.org/10.1289/EHP9934>.
34. Peters M, Godfrey C, McInerney P, Munn Z, Tricco A, Khalil H. 2020. Chapter 11: Scoping Reviews (2020 version). In: *JBI Manual for Evidence Synthesis*. Aromataris E, Munn Z, eds. Adelaide, Australia: JBI. <https://doi.org/10.46658/jbimes-20-12>.
35. Rohweder R, Salcedo S, da Silva Gomes VL, Schulze PAC, Schüler-Faccini L. 2022. Pesticides and pregnancy: a scoping review protocol. *OSF* 10 April 2023 <https://doi.org/10.17605/osf.io/d4tuv>.
36. Peters MDJ, Marnie C, Tricco AC, Pollock D, Munn Z, Alexander L, et al. 2020. Updated methodological guidance for the conduct of scoping reviews. *JBI Evid Synth* 18(10):2119–2126, PMID: 33038124, <https://doi.org/10.1112/JBIES-20-00167>.
37. Ouzzani M, Hammady H, Fedorowicz Z, Elmagarmid A. 2016. Rayyan—a web and mobile app for systematic review. *Syst Rev* 5(1):210, PMID: 27919275, <https://doi.org/10.1186/s13643-016-0384-4>.
38. Lo Russo G, Spolveri F, Ciancio F, Mori A. 2013. Mendeley: an easy way to manage, share, and synchronize papers and citations. *PLoS One* 8(12):e7466, PMID: 23714842, <https://doi.org/10.1097/PRS.0b013e31828bd400>.
39. Monge P, Wesseling C, Guardado J, Lundberg I, Ahlbom A, Cantor KP, et al. 2007. Parental occupational exposure to pesticides and the risk of childhood leukemia in Costa Rica. *Scand J Work Environ Health* 33(4):293–303, PMID: 17717622, <https://doi.org/10.5271/sjweh.1146>.
40. Hyland C, Gunier RB, Metayer C, Bates MN, Wesseling C, Mora AM. 2018. Maternal residential pesticide use and risk of childhood leukemia in Costa Rica. *Int J Cancer* 143(6):1295–1304, PMID: 29658108, <https://doi.org/10.1002/ijc.31522>.
41. Mora AM, Córdoba L, Cano JC, Hernández-Bonilla D, Pardo L, Schnaas L, et al. 2018. Prenatal mancozeb exposure, excess manganese, and neurodevelopment at 1 year of age in the Infants' Environmental Health (ISA) study. *Environ Health Perspect* 126(5):057007, PMID: 29847083, <https://doi.org/10.1289/EHP1955>.
42. Mora AM, Hoppin JA, Córdoba L, Cano JC, Soto-Martínez M, Eskenazi B, et al. 2020. Prenatal pesticide exposure and respiratory health outcomes in the first year of life: results from the Infants' Environmental Health (ISA) study. *Int J Hyg Environ Health* 225:113474, PMID: 32066110, <https://doi.org/10.1016/j.ijheh.2020.113474>.
43. Islam JY, Hoppin J, Mora AM, Soto-Martínez ME, Gamboa LC, Castañeda JEP, et al. 2023. Respiratory and allergic outcomes among 5-year-old children exposed to pesticides. *Thorax* 78(1):41–49, PMID: 35210357, <https://doi.org/10.1136/thoraxjnl-2021-218068>.
44. van Wendel de Joode B, Peñaloza-Castañeda J, Mora AM, Corrales-Vargas A, Eskenazi B, Hoppin JA, et al. 2024. Pesticide exposure, birth size, and gestational age in the ISA birth cohort, Costa Rica. *Environ Epidemiol* 8(2):e290, PMID: 38617432, <https://doi.org/10.1097/EE9.0000000000000290>.
45. Conejo-Bolaños LD, Mora AM, Hernández-Bonilla D, Cano JC, Menezes-Filho JA, Eskenazi B, et al. 2024. Prenatal current-use pesticide exposure and children's neurodevelopment at one year of age in the Infants' Environmental Health (ISA) birth cohort, Costa Rica. *Environ Res* 249:118222, PMID: 38272290, <https://doi.org/10.1016/j.envres.2024.118222>.
46. Handal AJ, Harlow SD, Breilh J, Lozoff B. 2008. Occupational exposure to pesticides during pregnancy and neurobehavioral development of infants and toddlers. *Epidemiology* 19(6):851–859, PMID: 18813021, <https://doi.org/10.1097/EDE.0b013e318187cc5d>.
47. Handal AJ, Harlow SD. 2009. Employment in the Ecuadorian cut-flower industry and the risk of spontaneous abortion. *BMC Int Health Hum Rights* 9(1):25, PMID: 19814818, <https://doi.org/10.1186/1472-698X-9-25>.
48. Grandjean P, Harari R, Barr DB, Debes F. 2006. Pesticide exposure and stunting as independent predictors of neurobehavioral deficits in Ecuadorian school children. *Pediatrics* 117(3):e546–e556, PMID: 16510633, <https://doi.org/10.1542/peds.2005-1781>.
49. Harari R, Julvez J, Murata K, Barr D, Bellinger DC, Debes F, et al. 2010. Neurobehavioral deficits and increased blood pressure in school-age children prenatally exposed to pesticides. *Environ Health Perspect* 118(6):890–896, PMID: 20185383, <https://doi.org/10.1289/ehp.0901582>.
50. Ferreira JD, Couto AC, Alves LC, Pombo-de-Oliveira MS, Koifman S. 2012. Exposições ambientais e leucemias na infância no Brasil: uma análise exploratória de sua associação. *Rev Bras Estud Popul* 29(2):477–492, <https://doi.org/10.1590/S0102-30982012000200014>.
51. Ferreira JD, Couto AC, Pombo-de-Oliveira MS, Koifman S, Brazilian Collaborative Study Group of Infant Acute Leukemia. 2013. *In utero* pesticide exposure and leukemia in Brazilian children < 2 years of age. *Environ Health Perspect* 121(2):269–275, PMID: 23092909, <https://doi.org/10.1289/ehp.1103942>.
52. Kadhel P, Monfort C, Costet N, Rouget F, Thomé J-P, Multigner L, et al. 2014. Chlordecone exposure, length of gestation, and risk of preterm birth. *Am J Epidemiol* 179(5):536–544, PMID: 24401561, <https://doi.org/10.1093/aje/kwt313>.
53. Hervé D, Costet N, Kadhel P, Rouget F, Monfort C, Thomé J-P, et al. 2016. Prenatal exposure to chlordecone, gestational weight gain, and birth weight in a Guadeloupean birth cohort. *Environ Res* 151:436–444, PMID: 27560981, <https://doi.org/10.1016/j.envres.2016.08.004>.
54. Costet N, Pelé F, Comets E, Rouget F, Monfort C, Bodeau-Livinec F, et al. 2015. Perinatal exposure to chlordecone and infant growth. *Environ Res* 142:123–134, PMID: 26133809, <https://doi.org/10.1016/j.envres.2015.06.023>.
55. Saint-Amour D, Muckle G, Gagnon-Chauvin A, Rouget F, Monfort C, Michineau L, et al. 2020. Visual contrast sensitivity in school-age Guadeloupean children exposed to chlordecone. *Neurotoxicology* 78:195–201, PMID: 32217184, <https://doi.org/10.1016/j.neuro.2020.02.012>.
56. Rouget F, Kadhel P, Monfort C, Viel JF, Thome JP, Cordier S, et al. 2020. Chlordecone exposure and risk of congenital anomalies: the Timoun Mother–Child Cohort Study in Guadeloupe (French West Indies). *Environ Sci Pollut Res Int* 27(33):40992–40998, PMID: 31376129, <https://doi.org/10.1007/s11356-019-06031-y>.
57. Dallaire R, Muckle G, Rouget F, Kadhel P, Bataille H, Guldner L, et al. 2012. Cognitive, visual, and motor development of 7-month-old Guadeloupean infants exposed to chlordecone. *Environ Res* 118:79–85, PMID: 22910562, <https://doi.org/10.1016/j.envres.2012.07.006>.
58. Boucher O, Simard M-N, Muckle G, Rouget F, Kadhel P, Bataille H, et al. 2013. Exposure to an organochlorine pesticide (chlordecone) and development of 18-month-old infants. *Neurotoxicology* 35:162–168, PMID: 23376090, <https://doi.org/10.1016/j.neuro.2013.01.007>.
59. Desrochers-Couture M, Cordier S, Rouget F, Michineau L, Monfort C, Thomé J-P, et al. 2022. Visuospatial processing and fine motor function among 7-years old Guadeloupe children pre- and postnatally exposed to the organochlorine pesticide chlordecone. *Neurotoxicology* 88:208–215, PMID: 34890633, <https://doi.org/10.1016/j.neuro.2021.12.003>.
60. Costet N, Lafontaine A, Rouget F, Michineau L, Monfort C, Thomé J-P, et al. 2022. Prenatal and childhood exposure to chlordecone and adiposity of seven-year-old children in the Timoun Mother–Child Cohort Study in Guadeloupe (French West Indies). *Environ Health* 21(1):42, PMID: 35439992, <https://doi.org/10.1186/s12940-022-00850-2>.

61. Cordier S, Bouquet E, Warembourg C, Massart C, Rouget F, Kadhel P, et al. 2015. Perinatal exposure to chlordecone, thyroid hormone status and neurodevelopment in infants: the Timoun cohort study in Guadeloupe (French West Indies). *Environ Res* 138:271–278, PMID: 25747818, <https://doi.org/10.1016/j.envres.2015.02.021>.
62. Cordier S, Forget-Dubois N, Desrochers-Couture M, Rouget F, Michineau L, Monfort C, et al. 2020. Prenatal and childhood exposure to chlordecone and sex-typed toy preference of 7-year-old Guadeloupean children. *Environ Sci Pollut Res Int* 27(33):40971–40979, PMID: 31264154, <https://doi.org/10.1007/s11356-019-05686-x>.
63. Ayhan G, Rouget F, Giton F, Costet N, Michineau L, Monfort C, et al. 2021. *In utero* chlordecone exposure and thyroid, metabolic, and sex-steroid hormones at the age of seven years: a study From the TIMOUN Mother–Child Cohort in Guadeloupe. *Front Endocrinol (Lausanne)* 12:771641, PMID: 34880833, <https://doi.org/10.3389/fendo.2021.771641>.
64. Oulhote Y, Rouget F, Michineau L, Monfort C, Desrochers-Couture M, Thomé J-P, et al. 2023. Prenatal and childhood chlordecone exposure, cognitive abilities and problem behaviors in 7-year-old children: the TIMOUN mother–child cohort in Guadeloupe. *Environ Health* 22(1):21, PMID: 36843015, <https://doi.org/10.1186/s12940-023-00970-3>.
65. Torres-Sánchez L, Schnaas L, Rothenberg SJ, Cebrián ME, Osorio-Valencia E, Hernández MdelC, et al. 2013. Prenatal *p,p'*-DDE exposure and neurodevelopment among children 3.5–5 years of age. *Environ Health Perspect* 121(2):263–268, PMID: 23151722, <https://doi.org/10.1289/ehp.1205034>.
66. Torres-Sánchez L, Rothenberg SJ, Schnaas L, Cebrián ME, Osorio E, del Carmen Hernández M, et al. 2007. *In utero p,p'*-DDE exposure and infant neurodevelopment: a perinatal cohort in Mexico. *Environ Health Perspect* 115(3):435–439, PMID: 17431495, <https://doi.org/10.1289/ehp.9566>.
67. Torres-Sánchez L, Zepeda M, Cebrián ME, Belkind-Gerson J, García-Hernandez RM, Belkind-Valdovinos U, et al. 2008. Dichlorodiphenyldichloroethylene exposure during the first trimester of pregnancy alters the anal position in male infants. *Ann NY Acad Sci* 1140(1):155–162, PMID: 18991914, <https://doi.org/10.1196/annals.1454.004>.
68. Bahena-Medina LA, Torres-Sánchez L, Schnaas L, Cebrián ME, Chávez CH, Osorio-Valencia E, et al. 2011. Neonatal neurodevelopment and prenatal exposure to dichlorodiphenyldichloroethylene (DDE): a cohort study in Mexico. *J Expo Sci Environ Epidemiol* 21(6):609–614, PMID: 21750576, <https://doi.org/10.1038/jes.2011.25>.
69. Torres-Sánchez L, Schnaas L, Cebrián ME, Hernández MdelC, Valencia EO, García Hernández RM, et al. 2009. Prenatal dichlorodiphenyldichloroethylene (DDE) exposure and neurodevelopment: a follow-up from 12 to 30 months of age. *Neurotoxicology* 30(6):1162–1165, PMID: 19733589, <https://doi.org/10.1016/j.neuro.2009.08.010>.
70. Garced S, Torres-Sánchez L, Cebrián ME, Claudio L, López-Carrillo L. 2012. Prenatal dichlorodiphenyldichloroethylene (DDE) exposure and child growth during the first year of life. *Environ Res* 113:58–62, PMID: 22244494, <https://doi.org/10.1016/j.envres.2011.12.002>.
71. Osorio-Valencia E, Torres-Sánchez L, López-Carrillo L, Cebrián ME, Rothenberg SJ, del Hernández Chávez Mdel C, et al. 2015. Prenatal *p,p'*-DDE exposure and establishment of lateralization and spatial orientation in Mexican preschooler. *Neurotoxicology* 47:1–7, PMID: 25572880, <https://doi.org/10.1016/j.neuro.2014.12.011>.
72. Fortenberry GZ, Meeker JD, Sánchez BN, Barr DB, Panuwet P, Bellinger D, et al. 2014. Urinary 3,5,6-trichloro-2-pyridinol (TCPY) in pregnant women from Mexico City: distribution, temporal variability, and relationship with child attention and hyperactivity. *Int J Hyg Environ Health* 217(2–3):405–412, PMID: 24001412, <https://doi.org/10.1016/j.ijheh.2013.07.018>.
73. Watkins DJ, Fortenberry GZ, Sánchez BN, Barr DB, Panuwet P, Schnaas L, et al. 2016. Urinary 3-phenoxybenzoic acid (3-PBA) levels among pregnant women in Mexico City: distribution and relationships with child neurodevelopment. *Environ Res* 147:307–313, PMID: 26922411, <https://doi.org/10.1016/j.envres.2016.02.025>.
74. Zamora AN, Watkins DJ, Peterson KE, Téllez-Rojo MM, Hu H, Meeker JD, et al. 2022. Prenatal maternal pesticide exposure in relation to sleep health of offspring during adolescence. *Environ Res* 204(pt A):111977, PMID: 34469742, <https://doi.org/10.1016/j.envres.2021.111977>.
75. Longnecker MP, Gladen BC, Cupul-Uicab LA, Romano-Riquer SP, Weber J-P, Chapin RE, et al. 2007. *In utero* exposure to the antiandrogen 1,1-dichloro-2,2-bis(*p*-chlorophenyl)ethylene (DDE) in relation to anogenital distance in male newborns from Chiapas, Mexico. *Am J Epidemiol* 165(9):1015–1022, PMID: 17272288, <https://doi.org/10.1093/aje/kwk109>.
76. Cupul-Uicab LA, Hernández-Ávila M, Terrazas-Medina EA, Pennell ML, Longnecker MP. 2010. Prenatal exposure to the major DDT metabolite 1,1-dichloro-2,2-bis(*p*-chlorophenyl)ethylene (DDE) and growth in boys from Mexico. *Environ Res* 110(6):595–603, PMID: 20566194, <https://doi.org/10.1016/j.envres.2010.06.001>.
77. Cupul-Uicab LA, Terrazas-Medina EA, Hernández-Ávila M, Longnecker MP. 2014. Prenatal exposure to *p,p'*-DDE and *p,p'*-DDT in relation to lower respiratory tract infections in boys from a highly exposed area of Mexico. *Environ Res* 132:19–23, PMID: 24742723, <https://doi.org/10.1016/j.envres.2014.03.017>.
78. Souza MS, Rovedatti MG, Sanchez S, Santa Cruz S, Pechén de D'Angelo AM, Magnarelli G. 2006. Biomarcadores para el monitoreo de la exposición intrauterina a plaguicidas en comunidades rurales. *Acta Toxicol Argent* 14(suppl):69–72.
79. Cecchi A, Alvarez G, Quidel N, Bertone MC, Anderle S, Sabino G, et al. 2021. Residential proximity to pesticide applications in Argentine Patagonia: impact on pregnancy and newborn parameters. *Environ Sci Pollut Res Int* 28(40):56565–56579, PMID: 34060016, <https://doi.org/10.1007/s11356-021-14574-2>.
80. Rodriguez PM, Oндarza PM, Miglioranza KSB, Ramirez CL, Vera B, Muntaner C, et al. 2023. Pesticides exposure in pregnant Argentinian women: potential relations with the residence areas and the anthropometric neonate parameters. *Chemosphere* 332:138790, PMID: 37142107, <https://doi.org/10.1016/j.chemosphere.2023.138790>.
81. Arrebola JP, Cuellar M, Bonde JP, González-Alzaga B, Mercado LA. 2016. Associations of maternal *o,p'*-DDT and *p,p'*-DDE levels with birth outcomes in a Bolivian cohort. *Environ Res* 151:469–477, PMID: 27567351, <https://doi.org/10.1016/j.envres.2016.08.008>.
82. Siqueira MT, Braga C, Cabral-Filho JE, Augusto LGDS, Figueiroa JN, Souza AI. 2010. Correlation Between pesticide use in agriculture and adverse birth outcomes in Brazil: an ecological study. *Bull Environ Contam Toxicol* 84(6):647–651, PMID: 20473603, <https://doi.org/10.1007/s00128-010-0027-8>.
83. Cremonese C, Freire C, Meyer A, Koifman S. 2012. Exposição a agrotóxicos e eventos adversos na gravidez no Sul do Brasil, 1996–2000. *Cad Saúde Pública* 28(7):1263–1272, PMID: 22729257, <https://doi.org/10.1590/S0102-311X2012000700005>.
84. Mendonça Guimarães R, Bueno PC, Apgáua G, Lima G, Moreira EM, Luvizotto MJ. 2014. The impact of agricultural pesticide use on the prevalence of perinatal outcomes in Brazil. *Bol Malarol Salud Ambient* 54(1):88–94.
85. de Fátima Alvim Braga I, Cozendey-Silva EN, Ertler LZ, Dos Santos Martins TG, da Silva Santos S, Silva BADFE, et al. 2023. Early abortions and congenital malformations: a comparison between agricultural and nonagricultural areas in the State of São Paulo/Brazil. *J Occup Environ Med* 65(10):820–825, PMID: 37264527, <https://doi.org/10.1097/JOM.0000000000002896>.
86. Contreras-Levicoy J, Astorga E, Castro R, Yentzen G, Cumsille M. 2005. Abortos espontáneos en hospital de Ilay y su relación con labores agrícolas de la madre. *Rev Chil Salud Pública* 9(1):7–11, <https://doi.org/10.5354/0717-3652.2005.19954>.
87. Camacho A, Mejía D. 2017. The health consequences of aerial spraying illicit crops: the case of Colombia. *J Health Econ* 54:147–160, PMID: 28570914, <https://doi.org/10.1016/j.jhealeco.2017.04.005>.
88. Torres-Arreola L, Berkowitz G, Torres-Sánchez L, López-Cervantes M, Cebrián ME, Uribe M, et al. 2003. Preterm birth in relation to maternal organochlorine serum levels. *Ann Epidemiol* 13(3):158–162, PMID: 12604158, [https://doi.org/10.1016/s1047-2797\(02\)00424-6](https://doi.org/10.1016/s1047-2797(02)00424-6).
89. Levario-Carrillo M, Amato D, Ostrosky-Wegman P, González-Horta C, Corona Y, Sanin LH. 2004. Relation between pesticide exposure and intrauterine growth retardation. *Chemosphere* 55(10):1421–1427, PMID: 15081785, <https://doi.org/10.1016/j.chemosphere.2003.11.027>.
90. Salazar-García F, Gallardo-Díaz E, Cerón-Mireles P, Loomis D, Borja-Aburto VH. 2004. Reproductive effects of occupational DDT exposure among male malaria control workers. *Environ Health Perspect* 112(5):542–547, PMID: 15064158, <https://doi.org/10.1289/ehp.112-1241918>.
91. Acosta-Maldonado B, Sánchez-Ramírez B, Reza-López S, Levario-Carrillo M. 2009. Effects of exposure to pesticides during pregnancy on placental maturity and weight of newborns: a cross-sectional pilot study in women from the Chihuahua State, Mexico. *Hum Exp Toxicol* 28(8):451–459, PMID: 19744971, <https://doi.org/10.1177/0960327109107045>.
92. Blanco-Muñoz J, Aguilar-Garduño C, Gamboa-Avila R, Rodríguez-Barranco M, Pérez-Méndez O, Huesca-Gómez C, et al. 2013. Association between PON1 genetic polymorphisms and miscarriage in Mexican women exposed to pesticides. *Sci Total Environ* 449:302–308, PMID: 23435062, <https://doi.org/10.1016/j.scitotenv.2013.01.034>.
93. Torres-Sánchez L, Escamilla-Núñez MC, Cebrian ME, Ancira-Moreno M, Rivera-Pasquel M, Lope V, et al. 2023. Exposure to *p,p'*-DDE during early pregnancy, anthropometry, and gestational age at birth, in a flower-growing region of Mexico. *J Dev Orig Health Dis* 14(1):15–23, PMID: 35593419, <https://doi.org/10.1017/S2040174422000277>.
94. Silver MK, Fernandez J, Tang J, McDade A, Sabino J, Rosario Z, et al. 2021. Prenatal exposure to glyphosate and its environmental degradate, aminomethylphosphonic acid (AMPA), and preterm birth: a nested case–control study in the PROTECT cohort (Puerto Rico). *Environ Health Perspect* 129(5):057011, PMID: 34009015, <https://doi.org/10.1289/EHP7295>.
95. Colussi CL, Martínez G, Bellenger J-P, Poletta GL, Simoniello MF. 2024. Prenatal exposure to pesticide mixture in Argentina: a pilot study in puerperal

- women from Santa Fe province. *Birth Defects Res* 116(2):e2307, <https://doi.org/10.1002/bdr2.2307>.
96. Gonçalves e Silva SR, Martins JL, Seixas S, da Silva DCG, Lemos SPP, Lemos PVB. 2011. Defeitos congênitos e exposição a agrotóxicos no Vale do São Francisco. *Rev Bras Ginecol Obstet* 33(1):20–26, PMID: 21625789.
 97. Gaspari L, Sampaio DR, Paris F, Audran F, Orsini M, Neto JB, et al. 2012. High prevalence of micropenis in 2710 male newborns from an intensive-use pesticide area of Northeastern Brazil. *Int J Androl* 35(3):253–264, PMID: 22372605, <https://doi.org/10.1111/j.1365-2605.2011.01241.x>.
 98. Cremonese C, Freire C, De Camargo AM, De Lima JS, Koifman S, Meyer A. 2014. Pesticide consumption, central nervous system and cardiovascular congenital malformations in the South and Southeast region of Brazil. *Int J Occup Med Environ Health* 27(3):474–486, PMID: 24847732, <https://doi.org/10.2478/s13382-014-0269-5>.
 99. Dutra LS, Ferreira AP. 2017. Malformações congênitas em regiões de monocultivo no estado de Minas Gerais, Brasil. *Medicina (Ribeirão Preto)* 50(5):285–296, <https://doi.org/10.11606/issn.2176-7262.v50i5p285-296>.
 100. Silvestre CMR, Silva AMC, Ferreira da Silva RCG, Bittencourt WS, Borba AM, Fernandes V, et al. 2022. Environmental factors at the periconceptional period and the occurrence of cleft lip and palate in a Midwest Brazil population: a case-control study. *J Occup Environ Med* 64(11):e751–e756, PMID: 36069817, <https://doi.org/10.1097/JOM.0000000000002689>.
 101. Rojas A, Ojeda ME, Barraza X. 2000. Malformaciones congénitas y exposición a pesticidas. *Rev Med Chil* 128(4):399–404, PMID: 10962857, <https://doi.org/10.4067/S0034-98872000000400006>.
 102. Card EB, Morales CE, Kimia R, Ramirez JM, Billingslea M, Marroquín A, et al. 2023. A retrospective and prospective cohort study comparing pediatric patients with cleft lip and palate from the United States and Guatemala. *J Craniofac Surg* 34(7):1978–1984, PMID: 37449578, <https://doi.org/10.1097/SCS.00000000000009539>.
 103. Flores-Luévano S, Farías P, Hernández M, Romano-Riquer P, Weber JP, Dewailly E, et al. 2003. Concentraciones de DDT/DDE y riesgo de hipospadias. Un estudio piloto de casos y controles. *Salud Pública México* 45(6):431–438, PMID: 14974286, <https://doi.org/10.1590/S0036-36342003000600002>.
 104. Waliszewski SM, Infanzon RM, Arroyo SG, Pietrini RV, Carvajal O, Trujillo P, et al. 2005. Persistent organochlorine pesticides levels in blood serum lipids in women bearing babies with undescended testis. *Bull Environ Contam Toxicol* 75(5):952–959, PMID: 16400584, <https://doi.org/10.1007/s00128-005-0842-5>.
 105. Lacasaña M, Vázquez-Grameix H, Borja-Aburto VH, Blanco-Muñoz J, Romieu I, Aguilar-Garduño C, et al. 2006. Maternal and paternal occupational exposure to agricultural work and the risk of anencephaly. *Occup Environ Med* 63(10):649–656, PMID: 16873458, <https://doi.org/10.1136/oem.2005.023333>.
 106. Bustamante Montes LP, Waliszewski S, Hernández-Valero M, Sanín-Aguirre L, Infanzón-Ruiz RM, Jañas AG. 2010. Exposición prenatal a los plaguicidas organoclorados y criptorquidia. *Cien Saude Colet* 15(suppl 1):1169–1174, PMID: 20640275, <https://doi.org/10.1590/s1413-81232010000700025>.
 107. Ibarra-Lopez JJ, Duarte P, Antonio-Vejar V, Calderon-Aranda ES, Huerta-Beristain G, Flores-Alfaro E, et al. 2013. Maternal C677T MTHFR polymorphism and environmental factors are associated with cleft lip and palate in a Mexican population. *J Investig Med* 61(6):1030–1035, PMID: 23787444, <https://doi.org/10.2310/JIM.0b013e31829a7e7e>.
 108. Enríquez-Vera N, Ruiz-Balbuena F, Megchún-Hernández M, Briones-Aranda A. 2023. Incidence of congenital malformations of the central nervous system in newborns in Chiapas, Mexico, and associated factors. *Rev Mex Neurocienc* 24(1):9902, <https://doi.org/10.24875/rmn.22000067>.
 109. Ojeda LC, Benítez Leite S. 2018. Factores de riesgo prenatales y su asociación a malformaciones congénitas en un hospital universitario de referencia. *Pediatr (Asunción)* 45(1):8–16, <https://doi.org/10.31698/ped.45012018002>.
 110. Tipiana IRG, Huamán AGR, Huamán ADCS. 2019. Prevalencia y riesgo de malformación congénita en mujeres gestantes expuestas a plaguicidas. en el hospital regional de Ica, Perú. *Rev Médica Panacea* 5(2), <https://doi.org/10.35563/rmp.v5i2.63>.
 111. Silvia SC, Magnarelli G, Rovedatti MG. 2020. Evaluation of endocrine disruption and gestational disorders in women residing in areas with intensive pesticide application: an exploratory study. *Environ Toxicol Pharmacol* 73:103280, PMID: 31683255, <https://doi.org/10.1016/j.etap.2019.103280>.
 112. Gibson G, Koifman S. 2008. Consumo de agrotóxicos e distribuição temporal da proporção de nascimentos masculinos no Estado do Paraná, Brasil. *Rev Panam Salud Publica* 24(4):240–247, PMID: 19133172, <https://doi.org/10.1590/s1020-49892008001000003>.
 113. Villar J, Ochieng R, Gunier RB, Papageorgiou AT, Rauch S, McGready R, et al. 2022. Association between fetal abdominal growth trajectories, maternal metabolite signatures early in pregnancy, and childhood growth and adiposity: prospective observational multinational INTERBIO-21st fetal study. *Lancet Diabetes Endocrinol* 10(10):710–719, PMID: 36030799, [https://doi.org/10.1016/S2213-8587\(22\)00215-7](https://doi.org/10.1016/S2213-8587(22)00215-7).
 114. Loreto-Gómez C, Farías P, Moreno-Macías H, Guzmán C, Riojas-Rodríguez H. 2018. Prenatal exposure to persistent organic compounds and their association with anogenital distance in infants. *Reprod Biomed Online* 37(6):732–740, PMID: 30539738, <https://doi.org/10.1016/j.rbmo.2018.09.008>.
 115. Benavides-Piracón JA, Hernández-Bonilla D, Menezes-Filho JA, van Wendel de Joode B, Lozada YAV, Bahia TC, et al. 2022. Prenatal and postnatal exposure to pesticides and school-age children's cognitive ability in rural Bogotá, Colombia. *Neurotoxicology* 90:112–120, PMID: 35306101, <https://doi.org/10.1016/j.neuro.2022.03.008>.
 116. Jenkins HM, Meeker JD, Zimmerman E, Cathey A, Fernandez J, Montañez GH, et al. 2024. Gestational glyphosate exposure and early childhood neurodevelopment in a Puerto Rico birth cohort. *Environ Res* 246:118114, PMID: 38211716, <https://doi.org/10.1016/j.envres.2024.118114>.
 117. Quintana MM, Vera B, Magnarelli G, Guinázú N, Rovedatti MG. 2017. Neonatal, placental, and umbilical cord blood parameters in pregnant women residing in areas with intensive pesticide application. *Environ Sci Pollut Res Int* 24(25):20736–20746, PMID: 28718019, <https://doi.org/10.1007/s11356-017-9642-9>.
 118. Silva MIGD, Siebel AM, Busato MA, De Sá CA, Corralo VDS. 2019. Environmental/occupational exposure to pesticides of pregnant women living in a countryside municipality. *Rev Pesqui Cuid Fundam Online* 11(5):1319–1325, <https://doi.org/10.9789/2175-5361.2019.v11i5.1319-1325>.
 119. Morris M, Sebastians AR, Perego VME. 2020. *Future Foodscapes: Re-Imagining Agriculture in Latin America and the Caribbean*. <https://documents1.worldbank.org/curated/en/942381591906970569/pdf/Future-Foodscapes-Re-imagining-Agriculture-in-Latin-America-and-the-Caribbean.pdf> [accessed 14 November 2023].
 120. Ray DK, Ramankutty N, Mueller ND, West PC, Foley JA. 2012. Recent patterns of crop yield growth and stagnation. *Nat Commun* 3(1):1293, PMID: 23250423, <https://doi.org/10.1038/ncomms2296>.
 121. Alston JM, Pardey PG. 2014. Agriculture in the global economy. *J Econ Perspect* 28(1):121–146, <https://doi.org/10.1257/jep.28.1.121>.
 122. Carvalho FP. 2017. Pesticides, environment, and food safety. *Food Energy Secur* 6(2):48–60, <https://doi.org/10.1002/fes3.108>.
 123. FAO. 2020. *Pesticides Trade. Global, Regional and Country Trends: 1990–2018*. <https://www.fao.org/3/cb0488en/CB0488En.pdf> [accessed 14 November 2023].
 124. Pelaez V, Teodorovic T, Guimarães TA, da Silva LR, Moreau D, Mizukawa G. 2016. A dinâmica do comércio internacional de agrotóxicos. *Rev Polit Agric* 25(2):39–52. <https://seer.sede.embrapa.br/index.php/RPA/article/view/1116> [accessed 14 November 2023].
 125. Simoes AJG, Hidalgo CA. 2011. The Economic Complexity Observatory: an analytical tool for understanding the dynamics of economic development. In: *AAAIWS'11-17: Proceedings of the 17th AAAI Conference on Scalable Integration of Analytics and Visualization*. 1 January 2011. Menlo Park, California: AAAI Press, 39–42.
 126. Schreinemachers P, Tipraqsa P. 2012. Agricultural pesticides and land use intensification in high, middle and low income countries. *Food Policy* 37(6):616–626, <https://doi.org/10.1016/j.foodpol.2012.06.003>.
 127. Dowler C. 2020. Thousands of tonnes of banned pesticides shipped to poorer countries from British and European factories. *Unearthed*. Published 9 September 9 2020. <https://unearthed.greenpeace.org/2020/09/10/banned-pesticides-eu-export-poor-countries/> [accessed 22 September 2023].
 128. Lopes-Ferreira M, Maleski ALA, Balan-Lima L, Bernardo JTG, Hipólito LM, Seni-Silva AC, et al. 2022. Impact of pesticides on human health in the last six years in Brazil. *Int J Environ Res Public Health* 19(6):3198, PMID: 35328887, <https://doi.org/10.3390/ijerph19063198>.
 129. Hart MM, Whang-Peng J, Sieber SM, Fabro S, Adamson RH. 1972. Distribution and effects of DDT in the pregnant rabbit. *Xenobiotica* 2(6):567–574, PMID: 4662547, <https://doi.org/10.3109/0049825720911084>.
 130. Klotz DM, Ladlie BL, Vonier PM, McLachlan JA, Arnold SF. 1997. *o,p'*-DDT and its metabolites inhibit progesterone-dependent responses in yeast and human cells. *Mol Cell Endocrinol* 129(1):63–71, PMID: 9175630, [https://doi.org/10.1016/s0303-7207\(96\)04041-5](https://doi.org/10.1016/s0303-7207(96)04041-5).
 131. Bredhult C, Bäcklin B-M, Olovsson M. 2007. Effects of some endocrine disruptors on the proliferation and viability of human endometrial endothelial cells *in vitro*. *Reprod Toxicol* 23(4):550–559, PMID: 17493787, <https://doi.org/10.1016/j.reprotox.2007.03.006>.
 132. Fabro S, McLachlan JA, Dames NM. 1984. Chemical exposure of embryos during the preimplantation stages of pregnancy: mortality rate and intrauterine development. *Am J Obstet Gynecol* 148(7):929–938, PMID: 6711631, [https://doi.org/10.1016/0002-9378\(84\)90535-0](https://doi.org/10.1016/0002-9378(84)90535-0).
 133. Giavini E, Vismara C, Broccia ML. 1983. Pre- and postimplantation embryotoxic effects of zinc dimethyldithiocarbamate (Ziram) in the rat. *Ecotoxicol Environ Saf* 7(6):531–537, PMID: 6662054, [https://doi.org/10.1016/0147-6513\(83\)90011-8](https://doi.org/10.1016/0147-6513(83)90011-8).
 134. Hass U, Christiansen S, Axelstad M, Scholze M, Boberg J. 2017. Combined exposure to low doses of pesticides causes decreased birth weights in rats.

- Reprod Toxicol 72:97–105, PMID: [28526456](https://doi.org/10.1016/j.reprotox.2017.05.004), <https://doi.org/10.1016/j.reprotox.2017.05.004>.
135. Chang A, Zhang Z, Zhang L, Gao Y, Zhang L, Jia L, et al. 2013. Proteomic analysis of preterm premature rupture of membranes in placental tissue. Arch Gynecol Obstet 288(4):775–784, PMID: [23580009](https://doi.org/10.1007/s00404-013-2837-5), <https://doi.org/10.1007/s00404-013-2837-5>.
 136. Fathollahipour S, Patil PS, Leipzig ND. 2018. Oxygen regulation in development: lessons from embryogenesis towards tissue engineering. Cells Tissues Organs 205(5–6):350–371, PMID: [30273927](https://doi.org/10.1159/000493162), <https://doi.org/10.1159/000493162>.
 137. Bonvallet N, Canlet C, Blas-Y-Estrada F, Gautier R, Tremblay-Franco M, Chevolleau S, et al. 2018. Metabolome disruption of pregnant rats and their offspring resulting from repeated exposure to a pesticide mixture representative of environmental contamination in Brittany. PLoS One 13(6):e0198448, PMID: [29924815](https://doi.org/10.1371/journal.pone.0198448), <https://doi.org/10.1371/journal.pone.0198448>.
 138. Ndonwi EN, Atogho-Tiedeu B, Lontchi-Yimagou E, Shinkafi TS, Nanfa D, Balti EV, et al. 2019. Gestational exposure to pesticides induces oxidative stress and lipid peroxidation in offspring that persist at adult age in an animal model. Toxicol Res 35(3):241–248, PMID: [31341553](https://doi.org/10.5487/TR.2019.35.3.241), <https://doi.org/10.5487/TR.2019.35.3.241>.
 139. Banerjee BD, Seth V, Ahmed RS. 2001. Pesticide-induced oxidative stress: perspective and trends. Rev Environ Health 16(1):1–40, PMID: [11354540](https://doi.org/10.1515/reveh.2001.16.1.1), <https://doi.org/10.1515/reveh.2001.16.1.1>.
 140. Anand M, Agarwal P, Singh L, Taneja A. 2015. Persistent organochlorine pesticides and oxidant/antioxidant status in the placental tissue of the women with full-term and pre-term deliveries. Toxicol Res 4(2):326–332, <https://doi.org/10.1039/C4TX00094C>.
 141. Gray LE Jr, Ostby JS, Kelce WR. 1994. Developmental effects of an environmental antiandrogen: the fungicide vinclozolin alters sex differentiation of the male rat. Toxicol Appl Pharmacol 129(1):46–52, PMID: [7974495](https://doi.org/10.1006/taap.1994.1227), <https://doi.org/10.1006/taap.1994.1227>.
 142. Kelce WR, Monosson E, Gamcsik MP, Laws SC, Gray LE Jr. 1994. Environmental hormone disruptors: evidence that vinclozolin developmental toxicity is mediated by antiandrogenic metabolites. Toxicol Appl Pharmacol 126(2):276–285, PMID: [8209380](https://doi.org/10.1006/taap.1994.1117), <https://doi.org/10.1006/taap.1994.1117>.
 143. Ostby J, Kelce WR, Lambright C, Wolf CJ, Mann P, Gray LE Jr. 1999. The fungicide procymidone alters sexual differentiation in the male rat by acting as an androgen-receptor antagonist in vivo and in vitro. Toxicol Ind Health 15(1–2):80–93, PMID: [10188193](https://doi.org/10.1177/074823379901500108), <https://doi.org/10.1177/074823379901500108>.
 144. Kelce WR, Stone CR, Laws SC, Gray LE, Kemppainen JA, Wilson EM. 1995. Persistent DDT metabolite p,p'-DDE is a potent androgen receptor antagonist. Nature 375(6532):581–585, PMID: [7791873](https://doi.org/10.1038/375581a0), <https://doi.org/10.1038/375581a0>.
 145. Lambright C, Ostby J, Bobseine K, Wilson V, Hotchkiss AK, Mann PC, et al. 2000. Cellular and molecular mechanisms of action of linuron: an antiandrogenic herbicide that produces reproductive malformations in male rats. Toxicol Sci 56(2):389–399, PMID: [10910998](https://doi.org/10.1093/toxsci/56.2.389), <https://doi.org/10.1093/toxsci/56.2.389>.
 146. McIntyre BS, Barlow NJ, Wallace DG, Maness SC, Gaido KW, Foster PMD. 2000. Effects of *in utero* exposure to linuron on androgen-dependent reproductive development in the male Crl:CD(SD)BR rat. Toxicol Appl Pharmacol 167(2):87–99, PMID: [10964759](https://doi.org/10.1006/taap.2000.8998), <https://doi.org/10.1006/taap.2000.8998>.
 147. Ding H, Zheng W, Han H, Hu X, Hu B, Wang F, et al. 2017. Reproductive toxicity of linuron following gestational exposure in rats and underlying mechanisms. Toxicol Lett 266:49–55, PMID: [28007637](https://doi.org/10.1016/j.toxlet.2016.12.013), <https://doi.org/10.1016/j.toxlet.2016.12.013>.
 148. Conley JM, Lambright CS, Evans N, Cardon M, Furr J, Wilson VS, et al. 2018. Mixed “antiandrogenic” chemicals at low individual doses produce reproductive tract malformations in the male rat. Toxicol Sci 164(1):166–178, PMID: [29945228](https://doi.org/10.1093/toxsci/kfy069), <https://doi.org/10.1093/toxsci/kfy069>.
 149. Howdeshell KL, Furr J, Lambright CR, Rider CV, Wilson VS, Gray LE Jr. 2007. Cumulative effects of dibutyl phthalate and diethylhexyl phthalate on male rat reproductive tract development: altered fetal steroid hormones and genes. Toxicol Sci 99(1):190–202, PMID: [17400582](https://doi.org/10.1093/toxsci/kfm069), <https://doi.org/10.1093/toxsci/kfm069>.
 150. Kortenkamp A. 2020. Which chemicals should be grouped together for mixture risk assessments of male reproductive disorders? Mol Cell Endocrinol 499:110581, PMID: [31525431](https://doi.org/10.1016/j.mce.2019.110581), <https://doi.org/10.1016/j.mce.2019.110581>.
 151. Schwartz CL, Christiansen S, Vinggaard AM, Axelstad M, Hass U, Svingen T. 2019. Anogenital distance as a toxicological or clinical marker for fetal androgen action and risk for reproductive disorders. Arch Toxicol 93(2):253–272, PMID: [30430187](https://doi.org/10.1007/s00204-018-2350-5), <https://doi.org/10.1007/s00204-018-2350-5>.
 152. Bertino E, Milani S, Fabris C, De Curtis M. 2007. Neonatal anthropometric charts: what they are, what they are not. Arch Dis Child Fetal Neonatal Ed 92(1):F7–F10, PMID: [17185434](https://doi.org/10.1136/adc.2006.096214), <https://doi.org/10.1136/adc.2006.096214>.
 153. Thankamony A, Pasterski V, Ong KK, Acerini CL, Hughes IA. 2016. Anogenital distance as a marker of androgen exposure in humans. Andrology 4(4):616–625, PMID: [26846869](https://doi.org/10.1111/andr.12156), <https://doi.org/10.1111/andr.12156>.
 154. Ye M, Beach J, Martin J, Senthilselvan A. 2013. Occupational pesticide exposures and respiratory health. Int J Environ Res Public Health 10(12):6442–6471, PMID: [24287863](https://doi.org/10.3390/ijerph10126442), <https://doi.org/10.3390/ijerph10126442>.
 155. Hoppin JA, Umbach DM, Long S, London SJ, Henneberger PK, Blair A, et al. 2017. Pesticides are associated with allergic and non-allergic wheeze among male farmers. Environ Health Perspect 125(4):535–543, PMID: [27384423](https://doi.org/10.1289/EHP315), <https://doi.org/10.1289/EHP315>.
 156. Negatu B, Kromhout H, Mekonnen Y, Vermeulen R. 2017. Occupational pesticide exposure and respiratory health: a large-scale cross-sectional study in three commercial farming systems in Ethiopia. Thorax 72(6):498–499, PMID: [27879416](https://doi.org/10.1136/thoraxjnl-2016-208924), <https://doi.org/10.1136/thoraxjnl-2016-208924>.
 157. Turner MC, Wigle DT, Krewski D. 2011. Residential pesticides and childhood leukemia: a systematic review and meta-analysis. Cien Saude Colet 16(3):1915–1931, PMID: [21519680](https://doi.org/10.1590/s1413-81232011000300026), <https://doi.org/10.1590/s1413-81232011000300026>.
 158. Wiemels J. 2012. Perspectives on the causes of childhood leukemia. Chem Biol Interact 196(3):59–67, PMID: [22326931](https://doi.org/10.1016/j.cbi.2012.01.007), <https://doi.org/10.1016/j.cbi.2012.01.007>.
 159. Ross JA, Potter JD, Reaman GH, Pendergrass TW, Robison LL. 1996. Maternal exposure to potential inhibitors of DNA topoisomerase II and infant leukemia (United States): a report from the Children's Cancer Group. Cancer Causes Control 7(6):581–590, PMID: [8932918](https://doi.org/10.1007/BF00051700), <https://doi.org/10.1007/BF00051700>.
 160. Hernández A, Menéndez P. 2016. Linking pesticide exposure with pediatric leukemia: potential underlying mechanisms. Int J Mol Sci 17(4):461, PMID: [27043530](https://doi.org/10.3390/ijms17040461), <https://doi.org/10.3390/ijms17040461>.
 161. Mactutus CF, Unger KL, Tilson HA. 1984. Evaluation of neonatal chlordecone neurotoxicity during early development: initial characterization. Neurobehav Toxicol Teratol 6(1):67–73, PMID: [6201754](https://doi.org/10.1016/0161-8131(84)90011-1).
 162. Cannon SB, Veazey JM Jr, Jackson RS, Burse VW, Hayes C, Straub WE, et al. 1978. Epidemic kepone poisoning in chemical workers. Am J Epidemiol 107(6):529–537, PMID: [78669](https://doi.org/10.1093/oxfordjournals.aje.a112572), <https://doi.org/10.1093/oxfordjournals.aje.a112572>.
 163. Mactutus CF, Tilson HA. 1984. Neonatal chlordecone exposure impairs early learning and retention of active avoidance in the rat. Neurobehav Toxicol Teratol 6(1):75–83, PMID: [6201755](https://doi.org/10.1016/0161-8131(84)90011-1).
 164. Mactutus CF, Unger KL, Tilson HA. 1982. Neonatal chlordecone exposure impairs early learning and memory in the rat on a multiple measure passive avoidance task. Neurotoxicology 3(2):27–44, PMID: [6186963](https://doi.org/10.1016/0161-8131(82)90011-1).
 165. Maristella S. 2019. *Las fronteras del neoextractivismo en América Latina: conflictos socioambientales, giro ecoterritorial y nuevas dependencias*. Vol Primera edición. Bielefeld, Germany: Bielefeld University Press.