

## RESEARCH ARTICLE

# Study on novel modified large mesoporous silica FDU-12/polymer matrix nanocomposites for adsorption of Pb(II)

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## Abstract

In this study, porous methacrylate-modified FDU-12/poly(methyl methacrylate) and amine-modified FDU-12/Nylon 6 nanocomposites were synthesized *via* a facile solution casting protocol. The physicochemical properties of the prepared materials were studied using various characterization techniques including Fourier transform-infrared spectroscopy, field emission-scanning electron microscopy, transmission electron microscopy, and nitrogen adsorption/desorption. After characterization of the materials, the prepared nanocomposites were applied as novel adsorbents for the removal of Pb(II) from aqueous media. In this regard, the effect of various parameters including solution pH, adsorbent amount, contact time, and initial concentration of Pb(II) on the adsorption process was investigated. To study the mechanism of adsorption, kinetic studies were conducted. The kinetic models of pseudo-first-order, pseudo-second-order, Elovich, and intraparticle diffusion were employed. The results revealed that the adsorption of Pb(II) onto methacrylate-modified FDU-12/poly(methyl methacrylate) and amine-modified FDU-12/Nylon 6 adsorbents followed the pseudo-second-order kinetic model. Also, different isotherms including Langmuir, Freundlich, and Dubinin-Radushkevich were applied to evaluate the equilibrium adsorption data. Langmuir isotherm provided the best fit with the equilibrium data of both adsorbents with maximum adsorption capacities of 99.0 and 94.3 mg g<sup>-1</sup> for methacrylate-modified FDU-12/poly(methyl methacrylate) and amine-modified FDU-12/Nylon 6, respectively, for the removal of Pb(II).

## OPEN ACCESS

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## 1. Introduction

Today, the discharge of environmental contaminants especially heavy metals (Pb(II), Cd(II), Hg(II), Cr(VI), etc.) in the environment caused drastic concerns about the health of individuals exposed to them. Heavy metals are an important class of environmental contaminants due to their high toxicity, carcinogenicity, nonbiodegradability, and bioaccumulative nature [1]. As an extremely toxic member of heavy metals, Pb(II) has bad effects on the kidney, liver,

reproductive system, brain functions, and basic cellular processes. It can cause mental retardation, insomnia, irritability, seizures, and renal damages. The major sources of Pb(II) in the environment are ore and metals processing, production of fertilizers, pigments, batteries, and leaded aviation fuels. Every day, health-related organizations and environmentalists express their deep concern about these contaminants. Because of these concerns, especially about human health, the US Environmental Protection Agency has announced a limit on the amount of heavy metals in soil and water. Thus, contaminant uptake, especially from aquatic media is urgent. Widespread researches have been conducted to introduce and develop decontamination methods. They are included chemical precipitation, electrochemical remediation, membrane filtration, ion exchange, solvent extraction, coagulation, adsorption, etc. Among them, the adsorption strategy is popular for effective water treatment owing to its several advantages. This technique provided a simple, relatively fast, and low-cost methodology with high efficiency and flexibility [2]. The method is also eco-friendly in large-scale manufacturing and can combine with other conventional remediation methods.

Considering intensive developments in nanotechnology especially in the fields of catalysis [3–9], extraction [10–14], carrier materials [15], and adsorption [16–21], polymer matrix nanocomposites have attracted a great deal of attention in the field of adsorption. The use of nanosized particles in the polymer matrixes provides high-performance hybrid nanocomposite materials with widespread applications. These materials combine ductility and flexibility of organic polymer matrix with the benefits of inorganic nanofiller such as high thermal and mechanical stability, and high surface area. On the other hand, the abundant functional groups (hydroxyl, carbonyl, phenyl, amine, etc.) on the surface of polymer matrix nanocomposites results in adsorbents with a good affinity toward adsorption of heavy metals [1, 22, 23]. Recently, mesoporous silica materials have attracted much attention to use as adsorbent or a part of adsorbent in polymer matrix nanocomposites due to their unique properties. They provide high surface area, functionalizable surface, and tailorable pore dimensions which make them a good candidate for adsorption processes. Over the past years, a plethora of mesoporous silica materials (examples include MCMs, SBAs, KCC-1, KITs, FDUs) with a wide range of pore geometries and particle morphologies have been introduced and used for widespread applications, especially for adsorption [24, 25]. Generally, they can be classified as 2D (e.g. SBA-15, MCM-41) and 3D (e.g. MCM-48, SBA-16, KIT-6, FDU-12) architecture based on the pore symmetry [26, 27].

As a three-dimensional large mesoporous silica, FDU-12 possesses a highly ordered structure with a superior 3D channel which makes it ideal for mass transfer and diffusion of guest molecules. It has unique properties of well-ordered pore structure, high specific surface area, ultra-large pore diameter (10–26 nm), and adjustable pore size [28, 29]. FDU-12 can be synthesized with different pore size distributions as reported in recent studies [30, 31]. To expand the application of mesoporous silica materials (e.g. FDU-12), changing hydrophobicity, and tailoring surface characteristics, the incorporation of organic functional groups onto the ordered structure of the material with uniform distribution is a widely used choice. In the case of nanocomposites, functionalization of the inorganic filler with a suitable functional group could help linking up the filler with the polymer matrix through functional groups and also more effective dispersion and penetration of the polymer chains in the porous structure of the filler. Since the surface of mesoporous silica materials is densely populated with hydroxyl groups, surface functionalization with a variety of organic moieties seems to be relatively easy [32]. In this context, several studies reported surface functionalization of mesoporous silica materials such as SBA-15, MCM-41, and KCC-1 with various organic groups [1, 2, 21, 23, 33, 34]. In comparison, there are few reports regarding the functionalization of mesoporous silica FDU-12 [35–39].

In the case of polymer matrix nanocomposites, various organic polymers have been applied for the preparation of polymer-based nanocomposites containing mesoporous materials. As a low density, transparent, thermoplastic, and rigid polymer, poly(methyl methacrylate) (PMMA) provided good flexibility and low cost. This versatile amorphous polymer has also excellent electrical and mechanical properties, and good resistance to non-polar solvents, acidic and alkaline solutions, and inorganic reagents [40, 41]. Another versatile polymer used in polymer matrix nanocomposites is Nylon. Nylon polymer is an electron-rich and polar synthetic polyamide composed of microfibrils that are interconnected forming a three-dimensional network with a porous structure [22, 42]. Among various types of commercial Nylons, Nylon 6 and Nylon 6,6 continue to be the most popular types. Nylon 6 fibers have high tensile strength and provided highly abrasion resistance. This polymer is also resistant to various types of chemicals such as acids and alkalis.

Herein, we present the synthesis and characterization of two novel modified mesoporous silica FDU-12 materials using silane coupling agents of 3-(triethoxysilyl)propyl methacrylate and  $N^1$ -(3-trimethoxysilylpropyl)diethylenetriamine. The prepared materials were then used as nanofiller for the preparation of poly(methyl methacrylate)- and Nylon 6-based nanocomposites. The nanocomposites were synthesized *via* a facile and fast method and characterized by various characterization techniques. To study the applicability of the prepared nanocomposites for adsorption purposes, the nanocomposites were used as novel adsorbents for removal of Pb(II) ions from aqueous solutions. Kinetic studies were also conducted for the two prepared materials. Our results showed the potential of methacrylate- and amine-functionalized FDU-12 for the adsorption of Pb(II) ions as a model of heavy metals from aqueous media.

## 2. Experimental

### 2.1. Materials and reagents

In this study, toluene (99%), absolute ethanol (99.5%), formic acid (99%), hydrochloric acid (37%), acetic acid (99.5%), *ortho*-phosphoric acid (85%), tetraethyl orthosilicate (TEOS), boric acid (99.5%), potassium chloride (99.5%), sodium hydroxide (97%), and lead(II) nitrate (99.5%) were obtained from Merck (Darmstadt, Germany). Poly(methyl methacrylate) (PMMA, average  $M_w \sim 120,000$ ), Nylon 6 (NY6), Pluronic F-127 ( $M_w \sim 12600$  Da), 1,3,5-trimethylbenzene (TMB, 98%), and  $N^1$ -(3-trimethoxysilylpropyl)diethylenetriamine were purchased from Sigma-Aldrich (St. Louis, MO, USA). 3-(triethoxysilyl)propyl methacrylate (98%) was obtained from TCI (Europe). Deionized water was prepared using a water purification system (Oklahoma, USA). The stock standard solution of Pb(II) ( $1000 \text{ mg L}^{-1}$ ) was prepared in water. Working standard solutions were prepared by diluting the stock solution. The Brighton-Robinson buffer system was used for the preparation of aqueous solutions with different pHs.

### 2.2. Instruments

The Fourier transform-infrared (FT-IR) spectra were recorded using a Thermo Nicolet Avatar 330 FT-IR spectrometer (USA) between  $4000$  and  $400 \text{ cm}^{-1}$  with a resolution of  $4 \text{ cm}^{-1}$ . The surface morphology of the samples was examined using a MIRA3 TESCAN-XMU field emission-scanning electron microscope (FE-SEM, Czech Republic). Transmission electron microscopy (TEM) analysis was performed on a Philips CM 120 microscope (Philips Electronics, Eindhoven, The Netherlands). For the analysis, the samples were dispersed in 2-propanol and a drop of the suspension was put on a carbon-coated nickel grid. Nitrogen adsorption/desorption analysis was performed on a Belsorp-mini II (BEL Japan Inc., Osaka, Japan) at  $77 \text{ K}$ . A flame atomic absorption spectroscopy (FAAS, Agilent, Model 240FS AA, USA) was used for

Pb quantification in the samples. In the case of samples with a low level of Pb, an inductively coupled plasma-optical emission spectrometer (ICP-OES, PerkinElmer, Optima 7300 DV, USA) was applied. A 100 W ultrasonic liquid processor (MISONIX XL-2000 Series, Raleigh, North Carolina, USA) was used for sonochemical reactions.

### 2.3. Synthesis and surface modification of large mesoporous silica FDU-12

The large mesoporous silica FDU-12 was synthesized according to the previous reports [29, 43] with some modifications. Briefly, 3.0 g of Pluronic F-127 was dissolved in 120 mL of 2.0 mol L<sup>-1</sup> hydrochloric acid. After 15 min, 7.5 g of potassium chloride was added to the solution. Then, 3.0 g of TMB was added and the mixture was stirred at 288 K for 2 h. Afterward, 12.5 g of TEOS was added to the solution and the mixture was heated to 383 K in an autoclave and remained at this temperature for 72 h. The white solid product was collected and washed with water and ethanol and oven-dried at 343 K for 24 h. The calcination of the material was performed at 823 K for 5 h.

Two silane coupling agents including 3-(triethoxysilyl)propyl methacrylate and *N*<sup>1</sup>-(3-trimethoxysilylpropyl)diethylenetriamine were used to surface modification of the prepared FDU-12 through the post-synthesis modification technique. Typically, 0.5 g of the synthesized FDU-12 and 0.5 mL of silane coupling agent were added to 40 mL of dried toluene. The mixture was refluxed at 383 K for 24 h. The resultant product was then isolated through a Büchner funnel and washed with toluene and ethanol, and oven-dried at 343 K for 24 h. The FDU-12 samples modified with 3-(triethoxysilyl)propyl methacrylate and *N*<sup>1</sup>-(3-trimethoxysilylpropyl)diethylenetriamine were denoted as FDU-12-MA and FDU-12-TA, respectively.

### 2.4. Preparation of FDU-12-MA/PMMA and FDU-12-TA/NY6

For the preparation of FDU-12-MA/PMMA nanocomposite with FDU-12-MA content of 3.0 wt%, 4.85 g of PMMA was dissolved in 40 mL of dry toluene (with the help of heating at 80°C under nitrogen atmosphere) and the mixture was sonicated for 15 min. Afterward, 0.15 g (3.0 wt%) of the prepared FDU-12-MA was added to 10 mL of dry toluene and the mixture was sonicated for 15 min. Then, the FDU-12-MA mixture was added to the polymer solution with mechanical stirring, sonicated for 15 min, and refluxed for 24 h. After cooling down to room temperature, the resultant mixture was sonicated for another 15 min, poured into a clean glass Petri dish, and dried at room temperature for 5 h. The FDU-12-MA/PMMA nanocomposite was washed with pure water, dried, and used for adsorption experiments.

In the case of FDU-12-TA/NY6, 4.85 g of NY6 was dissolved in 40 mL of formic acid (with the help of heating at 80°C under nitrogen atmosphere) and the mixture was refluxed under nitrogen atmosphere for 6 h to obtain a clear solution. Afterward, 0.15 g (3.0 wt%) of the prepared FDU-12-TA was added to 10 mL of formic acid and the mixture was sonicated for 15 min. Then, the FDU-12-TA mixture was added dropwise to the NY6 solution, sonicated for 15 min, and refluxed for 12 h. After refluxing, the mixture was sonicated for another 15 min, cast to a clean glass Petri dish, and dried at room temperature. The FDU-12-TA/NY6 nanocomposite was washed with pure water, dried, and used for adsorption experiments.

### 2.5. Adsorption experiments

Batch adsorption experiments were carried out to study the adsorption behavior of Pb(II) ions onto the FDU-12-MA/PMMA and FDU-12-TA/NY6. For each experiment, 10 mL of an aqueous standard solution of Pb(II) with the desired concentration was exposed to the accurately weighed amount of adsorbent in 20 mL polyethylene containers. The adsorption process was performed at 180 rpm and 298 K for 24 h. After that, the adsorbent was separated from the

solution and the concentration of Pb(II) in the solution was quantified by the means of FAAS or ICP-OES. The removal efficiency (*RE*) was computed using the following equation (Eq 1):

$$RE (\%) = \frac{C_i - C_e}{C_i} \times 100 \quad (1)$$

where  $C_i$  and  $C_e$  are the initial and equilibrium concentration of the metal ion in the solution ( $\text{mg L}^{-1}$ ), respectively. The adsorption capacity ( $q_e$ ,  $\text{mg g}^{-1}$ ) was calculated according to the following equation (Eq 2):

$$q_e = \left( \frac{C_i - C_e}{W} \right) \times V \quad (2)$$

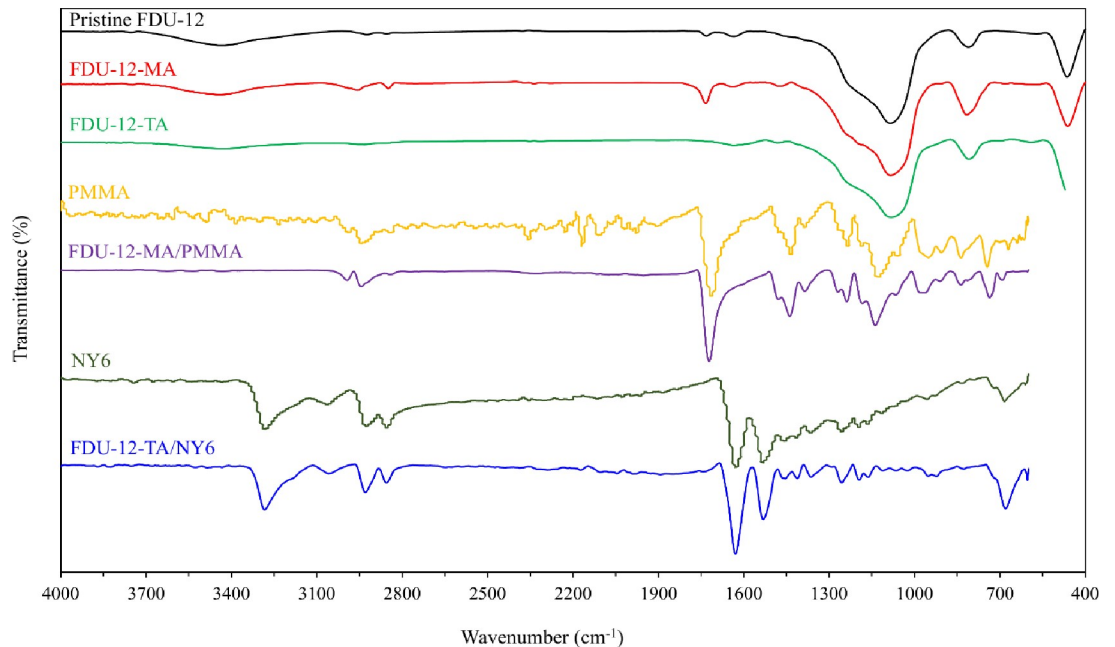
Where  $V$  is the solution volume (mL) and  $W$  refers to the adsorbent amount (mg).

### 3. Results and discussion

#### 3.1. Synthesis and characterization

Considering the relatively inert and hydrophobic nature of the pristine FDU-12, and also the tendency of nanoparticles to agglomerate, the surface functionalization is necessary to provide a homogenous dispersion of prepared mesoporous silica in the polymer matrix for the preparation of polymer nanocomposites. In the present study, the surface of the prepared pristine FDU-12 was treated with two different silane coupling agents of 3-(triethoxysilyl)propyl methacrylate and  $N^1$ -(3-trimethoxysilylpropyl)diethylenetriamine. These coupling agents improve the interfacial interaction of inorganic filler and organic polymer matrix. The methacrylate- and amine-functionalized FDU-12 materials were then applied as nanofiller for the preparation of two nanocomposites using PMMA and NY6 as organic polymers.

The chemical structure and morphology of the prepared materials were examined by FT-IR, FE-SEM, TEM, and  $N_2$  adsorption/desorption techniques. The FT-IR spectra of pristine FDU-12, FDU-12-MA, FDU-12-TA, pristine PMMA, FDU-12-MA/PMMA, pristine NY6, and FDU-12-TA/NY6 are shown in Fig 1. In the case of pristine FDU-12, the bands at 463 and 810  $\text{cm}^{-1}$  are attributed to the bending and symmetric stretching vibration of Si–O–Si. Also, the band at 1078  $\text{cm}^{-1}$  corresponds to the asymmetric stretching vibration of the Si–O–Si bond [28, 43]. The broadband around 3435  $\text{cm}^{-1}$  corresponds to O–H stretching of the surface silanol groups and the physically adsorbed water molecules. The band centered at 1633  $\text{cm}^{-1}$  is attributed to the bending vibration of H–O–H. For FDU-12-MA and FDU-12-TA, in addition to the bands observed for pure FDU-12, new bands related to the organic part of the silane coupling agent are appeared. The bands located at 2957 and 2849  $\text{cm}^{-1}$  are assigned to the C–H stretching vibrations of  $\text{CH}_2$  and the band at 1471  $\text{cm}^{-1}$  is corresponding to the bending vibration of the C–H. In the case of FDU-12-MA, the new band at 1733  $\text{cm}^{-1}$  is characteristic of the C = O bond of the silane coupling agent. For FDU-12-TA, the broadband at 3200–3500  $\text{cm}^{-1}$  indicated the presence of –OH and – $\text{NH}_2$  groups on the surface of FDU-12-TA. These data indicated successful modification of pure FDU-12 with 3-(triethoxysilyl)propyl methacrylate and  $N^1$ -(3-trimethoxysilylpropyl)diethylenetriamine. The FT-IR spectrum of FDU-12-MA/PMMA composite shows absorption bands at 2994 and 2944  $\text{cm}^{-1}$  which correspond to C–H asymmetric stretching in  $\text{CH}_3$  and  $\text{CH}_2$  groups, respectively. Also, the peak located at 2842  $\text{cm}^{-1}$  is related to the C–H symmetric stretching in  $\text{CH}_3$ . The characteristic band at 1722  $\text{cm}^{-1}$  which corresponds to the C = O bond is also observed. These bands in addition to the new bands observed in the spectrum of FDU-12-MA/PMMA correspond to different modes of  $\text{CH}_2$  and  $\text{CH}_3$  vibrational modes indicated the presence of PMMA in the structure of the prepared nanocomposite which is in accordance with other reports [23, 44].

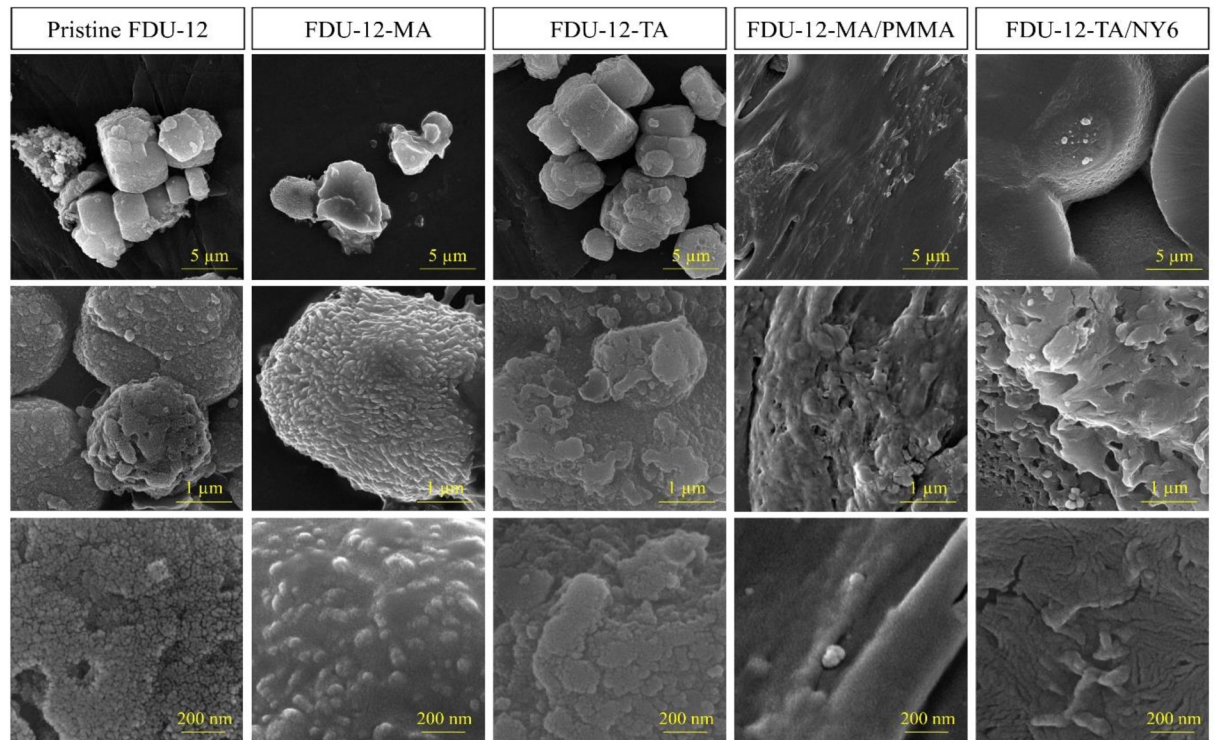


**Fig 1.** FT-IR spectra of pristine FDU-12, FDU-12-MA, FDU-12-TA, pristine PMMA, FDU-12-MA/PMMA, pristine NY6, and FDU-12-TA/NY6.

<https://doi.org/10.1371/journal.pone.0245583.g001>

Also, the similarities between the spectra of pristine PMMA and FDU-12-MA/PMMA are quite obvious. In the case of the C = O bond, the stretching vibration was observed at 1715 and 1722  $\text{cm}^{-1}$  for pristine PMMA and FDU-12-MA/PMMA, respectively. The shifts of C = O stretching vibrations may be due to the interaction of modified nanofiller with polymer matrix (hydrogen bonding interactions) which decreases the strength of the C = O bond in the polymer [45–47]. For FDU-12-TA/NY6, the characteristic peaks at 3284, 2856 & 2929  $\text{cm}^{-1}$  correspond to the stretching vibration of N-H bond and C-H stretching vibrations of the NY6. Also, the peaks located at 1631 and 1532  $\text{cm}^{-1}$  are attributed to the stretching vibration of carbonyl groups and N-H bending vibration which all are related to NY6 [1]. The similarities between the spectra of pristine NY6 and FDU-12-TA/NY6 are observable. In the case of C = O bond, the stretching vibration observed at 1628 and 1631  $\text{cm}^{-1}$  for pristine NY6 and FDU-12-TA/NY6, respectively. This shift also may be due to the hydrogen bonding interactions between nanofiller and polymer matrix which decreases the strength of the C = O bond in the polymer.

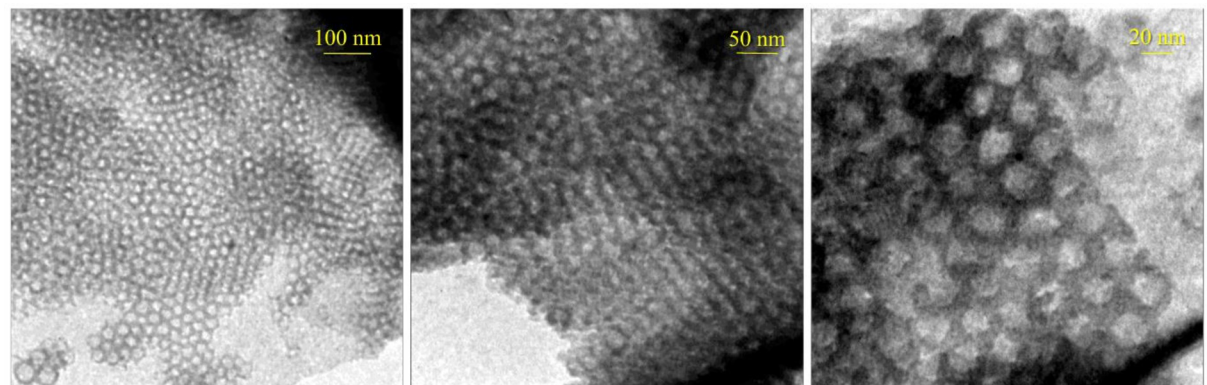
The morphology of the prepared materials was studied *via* FE-SEM and TEM techniques. The FE-SEM images of pristine FDU-12, FDU-12-MA, FDU-12-TA, FDU-12-MA/PMMA, and FDU-12-TA/NY6 are shown in Fig 2. The FE-SEM image of the pristine FDU-12 and modified samples (FDU-12-MA, FDU-12-TA) showed a relatively regular hexagonal prism morphology which is a typical characteristic morphology of FDU-12 [43, 48]. This indicates that the morphology of the FDU-12 was retained after modification with the used silane coupling agents. At higher magnifications, the growing crystals are visible. In the case of polymer composites, the modified FDU-12 particles are visible on the surface of nanocomposites suggesting that the mesoporous material is incorporated in the polymer matrix. The TEM images of pristine FDU-12 are shown in Fig 3. A well-ordered structure is seen. As reported in the literature [48–50], the FDU-12 exhibited ordered mesopore arrangement characteristics. In the images, the large cage units with regular alignment in the highly ordered lattice are obvious.



**Fig 2. FE-SEM images of pristine FDU-12, FDU-12-MA, FDU-12-TA, FDU-12-MA/PMMA, and FDU-12-TA/NY6 at three magnifications.**

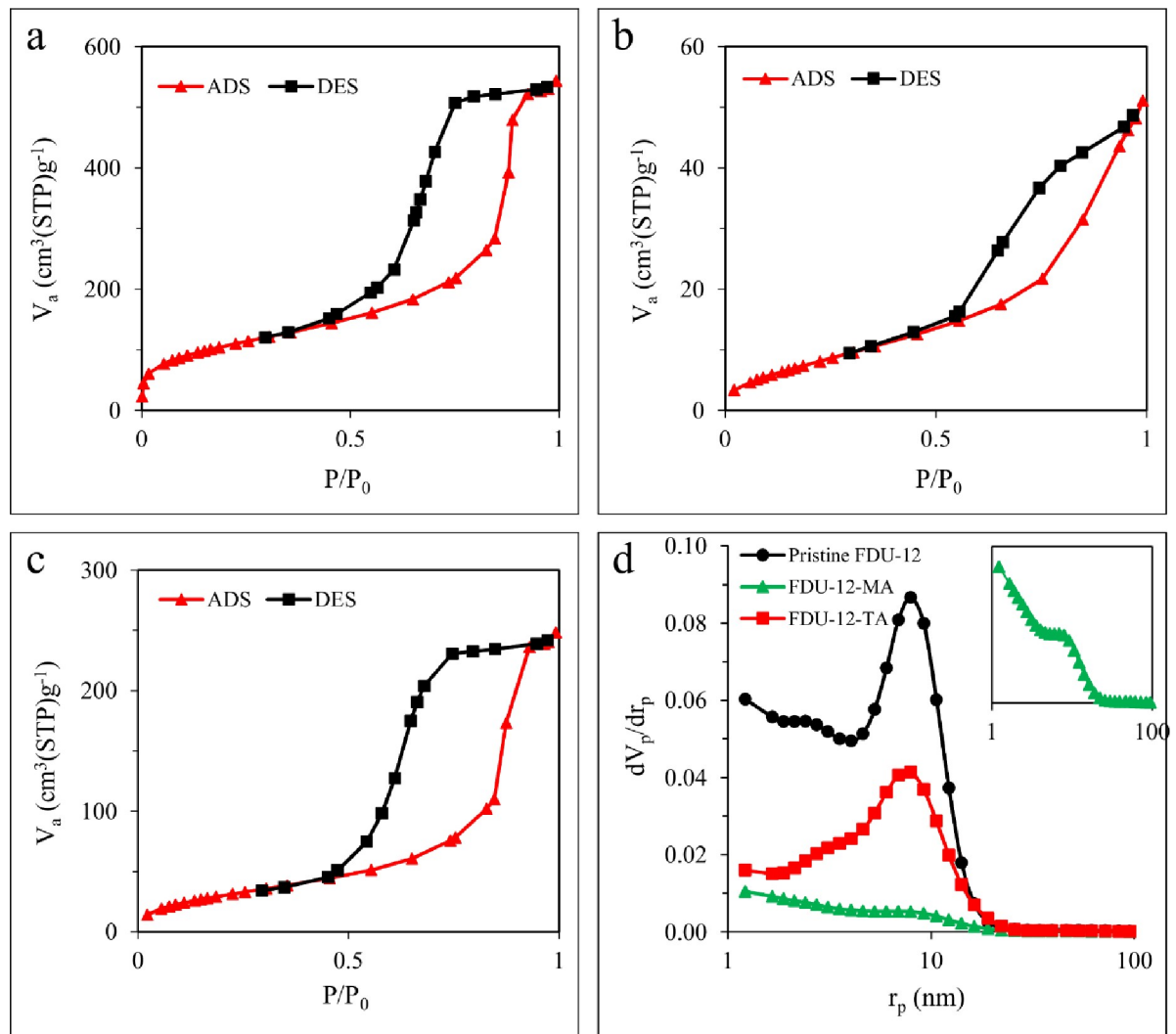
<https://doi.org/10.1371/journal.pone.0245583.g002>

The nitrogen adsorption/desorption isotherms of the pure FDU-12, FDU-12-MA, and FDU-12-TA materials are shown in Fig 4A–4C. As shown in the figure, according to the IUPAC classification, all three samples exhibited type IV isotherms with H2-type hysteresis loops. The pristine FDU-12, FDU-12-MA, and FDU-12-TA samples showed a BET surface area of 376, 29, and 115  $\text{m}^2 \text{g}^{-1}$ , respectively. According to the BJH model (Fig 4D), mean pore diameters of 9.2, 1.2, and 6.9 nm were found for pristine FDU-12, FDU-12-MA, and FDU-12-TA samples, respectively. In comparison, the textural parameters of pristine FDU-12 have reduced upon surface modification with silane coupling agents, as it was anticipatable. The obtained textural parameters of the materials are shown in Table 1.



**Fig 3. TEM images of pristine FDU-12.**

<https://doi.org/10.1371/journal.pone.0245583.g003>



**Fig 4.** N<sub>2</sub> adsorption/desorption isotherms of (a) pristine FDU-12, (b) FDU-12-MA, (c) FDU-12-TA, and (d) the BJH pore size distribution curves of the samples.

<https://doi.org/10.1371/journal.pone.0245583.g004>

**Table 1.** Textural properties of the synthesized pristine FDU-12, FDU-12-MA, and FDU-12-TA.

Sample	BET			BJH		
	SA <sup>a</sup>	PV <sup>b</sup>	PD <sup>c</sup>	SA	PV	PD
	(m <sup>2</sup> g <sup>-1</sup> )	(cm <sup>3</sup> g <sup>-1</sup> )	(nm)	(m <sup>2</sup> g <sup>-1</sup> )	(cm <sup>3</sup> g <sup>-1</sup> )	(nm)
Pristine FDU-12	376	0.8387	8.9	326	0.8004	9.2
FDU-12-MA	29	0.0787	10.9	38	0.0812	1.2
FDU-12-TA	115	0.3831	13.4	135	0.3896	6.9

<sup>a</sup> Surface area

<sup>b</sup> Total pore volume

<sup>c</sup> Mean pore diameter

<https://doi.org/10.1371/journal.pone.0245583.t001>



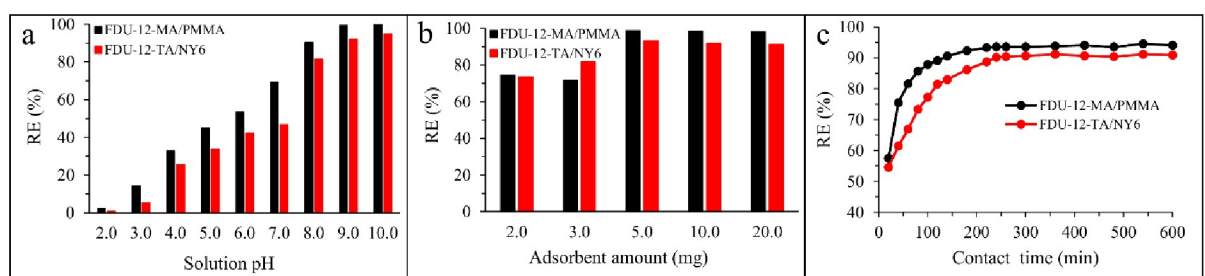
### 3.2. Adsorption studies

**3.2.1. The effect of pH on adsorption.** As an important factor affecting the adsorption process, the effect of pH on the adsorption of Pb(II) ions onto the FDU-12-MA/PMMA and FDU-12-TA/NY6 adsorbents was investigated in the range of 2.0–12.0. The experiments were performed using 10 mL of an aqueous standard solution of Pb(II) at the concentration level of 50 mg L<sup>-1</sup> with an adsorbent dosage of 10.0 mg. The adsorption was performed at 298 K for 24 h. As can be seen in Fig 5A, for both adsorbents, the removal efficiency was increased with increasing solution pH up to 9.0 and then no significant enhancement occurred. The low removal efficiency in acidic mediums could be related to the competition of H<sup>+</sup> ions with Pb(II). In other words, the H<sup>+</sup> ions in the solution occupy the attainable active sites on the adsorbent and compete with Pb(II) ions due to the electrostatic forces. This results in fewer accessible sites for Pb(II) leads to less adsorption of Pb(II) in acidic mediums. Accordingly, pH = 9.0 was selected for further experiments.

**3.2.2. The effect of adsorbent dosage.** To study the effect of adsorbent dosage on the adsorption of Pb(II) from aqueous solution, various dosages between 2.0 and 50.0 mg were tested. In these experiments, 10 mL of an aqueous standard solution of Pb(II) at the concentration level of 50 mg L<sup>-1</sup> (pH = 9.0) was used. The adsorption was performed at 298 K for 24 h. The obtained results are shown in Fig 5B. As seen in the figure, the removal efficiency for two adsorbents was enhanced with increasing adsorbent dosage from 2.0 to 5.0 mg. No surprising enhancement in removal efficiency was observed for higher amounts. With enhancing adsorbent dosage, more accessible adsorption sites are available which increases removal efficiency. So, 5.0 mg of each of the adsorbents were used for further experiments.

**3.2.3. The effect of contact time.** The effect of contact time on the adsorption process was investigated for the two adsorbents. In this step, 10 mL of an aqueous standard solution of Pb(II) at the concentration level of 50 mg L<sup>-1</sup> (pH = 9.0) with an adsorbent dosage of 5.0 mg was used. The adsorption was performed at 298 K. Fig 5C shows the effect of contact time (20 to 600 min) on Pb(II) removal by FDU-12-MA/PMMA and FDU-12-TA/NY6 adsorbents. Considerable enhancement in removal efficiency was observed when contact time was increased up to 220 and 240 min for FDU-12-MA/PMMA and FDU-12-TA/NY6, respectively. No further adsorption occurred for longer times. The obtained results showed a relatively fast adsorption process. This was mainly due to the high accessible sites on the surface of FDU-12-MA/PMMA and FDU-12-TA/NY6 adsorbents. Based on the results, contact times of 220 and 240 min (for FDU-12-MA/PMMA and FDU-12-TA/NY6, respectively) were selected for further experiments to ensure that equilibrium is reached.

**3.2.4. The adsorption kinetic studies.** To study the mechanism of adsorption, kinetic studies were conducted. Four kinetic models including pseudo-first-order (PFO), pseudo-second-order (PSO), Elovich, and intraparticle diffusion (IPD) were employed. The applied



**Fig 5.** The effect of (a) pH, (b) adsorbent amount, and (c) contact time on adsorption of Pb(II) by the prepared adsorbents.

<https://doi.org/10.1371/journal.pone.0245583.g005>

equations are expressed in the following forms:

$$\log(q_e - q_t) = \log q_e - \frac{k_1}{2.303} t \tag{3}$$

$$\frac{t}{q_t} = \frac{1}{h} + \frac{1}{q_e} t \tag{4}$$

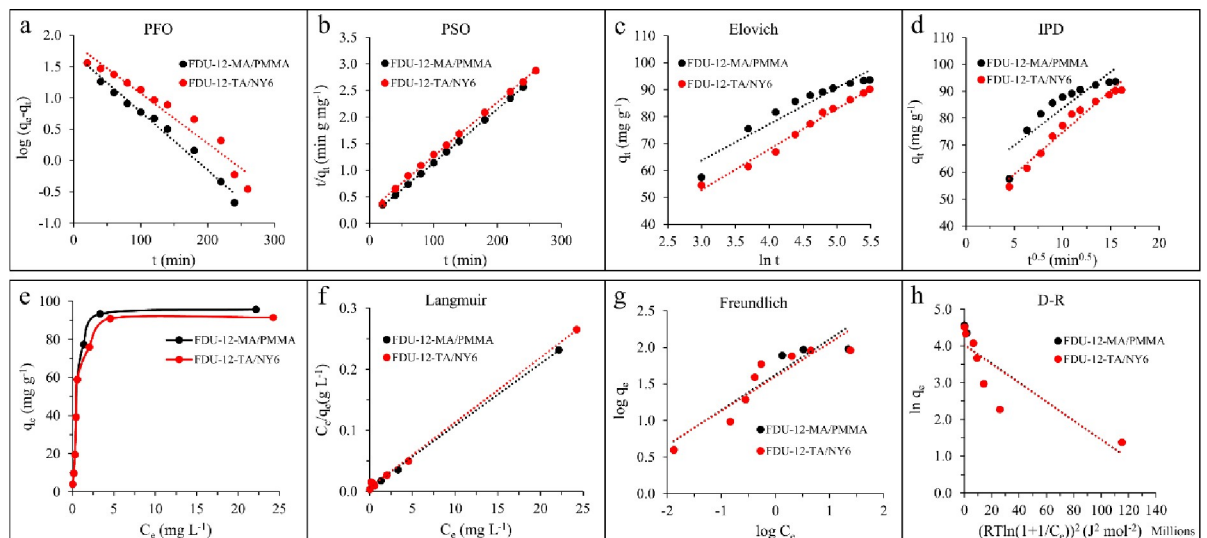
$$h = k_2 \times q_e^2 \tag{5}$$

$$q_t = \frac{\ln(\alpha\beta)}{\beta} + \frac{\ln t}{\beta} \tag{6}$$

$$q_t = k_{dif}(t)^{0.5} + C \tag{7}$$

Where  $q_e$  ( $\text{mg g}^{-1}$ ),  $q_t$  ( $\text{mg g}^{-1}$ ),  $k_1$  ( $\text{min}^{-1}$ ),  $h$  ( $\text{mg g}^{-1} \text{min}^{-1}$ ),  $k_2$  ( $\text{g mg}^{-1} \text{min}^{-1}$ ),  $\alpha$  ( $\text{mg g}^{-1} \text{min}^{-1}$ ) &  $\beta$  ( $\text{g mg}^{-1}$ ),  $k_{dif}$  ( $\text{mg g}^{-1} \text{min}^{-0.5}$ ), and  $C$  ( $\text{mg g}^{-1}$ ) are the adsorption capacity at equilibrium, the adsorption capacity at time  $t$ , PFO rate constant, the initial sorption rate in PSO model, PSO rate constant, Elovich constants, IPD rate constant, and a constant, respectively. The results of the fitting are illustrated in Fig 6A–6D and Table 2. As shown in Table 2, in the case of both adsorbents, the PSO kinetic model provided better  $R^2$  than those obtained by other models, which suggests that the chemical adsorption process can be well described with the PSO model. The presence of hydroxyl, amine, and carbonyl groups on the surface of adsorbents might be involved in the process.

**3.2.5. The effect of Pb(II) concentration and adsorption isotherm.** To study the effect of the initial concentration of Pb(II) on the adsorption behavior, a concentration range between 2.0 and 70.0  $\text{mg L}^{-1}$  (pH = 9.0) was investigated. The adsorbent dosage of 5.0 mg of each adsorbent was used for the experiments. The adsorption time was set to 220 and 240 min for FDU-12-MA/PMMA and FDU-12-TA/NY6, respectively. The obtained data are shown in



**Fig 6.** The kinetic adsorption models of (a) pseudo-first-order, (b) pseudo-second-order, (c) Elovich, (d) intra-particle diffusion; (e) the equilibrium isotherm and the isotherm models of (f) Langmuir, (g) Freundlich, and (h) Dubinin–Radushkevich for the adsorption of Pb(II) onto FDU-12-MA/PMMA and FDU-12-TA/NY6.

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**Table 2. The parameters obtained by kinetic models for the adsorption of Pb(II) onto FDU-12-MA/PMMA and FDU-12-TA/NY6.**

Model	Adsorbent	R <sup>2</sup>	Parameters <sup>a</sup>		
PFO	FDU-12-MA/PMMA	0.9846	$k_1 = 0.0214$	$q_e = 51.1$	
	FDU-12-TA/NY6	0.9469	$k_1 = 0.0184$	$q_e = 74.1$	
PSO	FDU-12-MA/PMMA	0.9999	$k_2 = 0.0008$	$q_e = 99.0$	<b><math>h = 7.6511</math></b>
	FDU-12-TA/NY6	0.9983	$k_2 = 0.0004$	$q_e = 98.0$	<b><math>h = 4.0371</math></b>
Elovich	FDU-12-MA/PMMA	0.9149	$\alpha = 75.5465$	$B = 0.0741$	
	FDU-12-TA/NY6	0.9884	$\alpha = 25.0980$	$\beta = 0.0665$	
IPD	FDU-12-MA/PMMA	0.7883	$k_{dif} = 2.7262$	$C = 56.3580$	
	FDU-12-TA/NY6	0.9596	$k_{dif} = 3.1103$	$C = 43.6304$	

<sup>a</sup> The units are as same as mentioned in Section 3.2.4.

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**Fig 6E.** The equilibrium adsorption studies were investigated by Langmuir, Freundlich, and Dubinin-Radushkevich (D-R) isotherm models. The linear forms of the applied isotherm models are expressed as follows:

$$\frac{C_e}{q_e} = \frac{1}{q_{max} \times k_L} + \frac{C_e}{q_{max}} \quad (8)$$

$$\log q_e = \frac{1}{n} \log C_e + \log k_F \quad (9)$$

$$\ln q_e = \ln q_{max} - B \left( RT \ln \left( 1 + \frac{1}{C_e} \right) \right)^2 \quad (10)$$

Where  $C_e$  ( $\text{mg L}^{-1}$ ),  $q_e$  ( $\text{mg g}^{-1}$ ),  $q_{max}$  ( $\text{mg g}^{-1}$ ),  $k_L$  ( $\text{L mg}^{-1}$ ),  $n$  &  $k_F$  ( $(\text{mg g}^{-1}) (\text{L mg}^{-1})^{1/n}$ ),  $B$  ( $\text{mol}^2 \text{kJ}^{-2}$ ),  $R$  ( $\text{J mol}^{-1} \text{K}^{-1}$ ), and  $T$  (K) are the concentration of Pb(II) at equilibrium, the adsorption capacity at equilibrium, the maximum adsorption capacity of adsorbent, the Langmuir constant, the Freundlich isotherm constants for adsorption capacity and adsorption intensity, Dubinin-Radushkevich isotherm constant, the universal gas constant, and temperature, respectively. The adsorption isotherms and the calculated parameters from the isotherms models are shown in Fig 6F–6H and Table 3, respectively. As can be seen, considering the R<sup>2</sup> value, the Langmuir model fitted the experimental data (for both adsorbent) better than the other models. The R<sup>2</sup> values of the Langmuir model were obtained 0.9970 and 0.9978 for FDU-12-MA/PMMA and FDU-12-TA/NY6 adsorbents, respectively. On the other hand, the maximum adsorption capacities of FDU-12-MA/PMMA and FDU-12-TA/NY6 obtained by

**Table 3. The parameters obtained by isotherm models for the adsorption of Pb(II) onto FDU-12-MA/PMMA and FDU-12-TA/NY6.**

Model	Adsorbent	R <sup>2</sup>	Parameters <sup>a</sup>	
Langmuir	FDU-12-MA/PMMA	0.9970	$k_L = 1.5781$	<b><math>q_{max} = 99.0</math></b>
	FDU-12-TA/NY6	0.9978	$k_L = 1.6308$	<b><math>q_{max} = 94.3</math></b>
Freundlich	FDU-12-MA/PMMA	0.8289	$k_F = 42.2474$	<b><math>n = 2.0640</math></b>
	FDU-12-TA/NY6	0.8378	$k_F = 39.4094$	<b><math>n = 2.1529</math></b>
D-R	FDU-12-MA/PMMA	0.7366	$B = 2.59 \times 10^{-7}$	<b><math>q_{max} = 56.7</math></b>
	FDU-12-TA/NY6	0.7429	$B = 2.57 \times 10^{-7}$	<b><math>q_{max} = 55.7</math></b>

<sup>a</sup> The units are as same as mentioned in Section 3.2.5.

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**Table 4. Comparison of the adsorption capacity of FDU-12-MA/PMMA and FDU-12-TA/NY6 toward Pb(II) ions with other adsorbents.**

Adsorbent	$q_{max}$ (mg g <sup>-1</sup> )	pH	Reference
Oxidized multiwalled carbon nanotubes/polypyrrole composite	26.32	6.0	[51]
SBA-15-supported Pb(II) imprinted polymer	42.55	6.0	[52]
Oil palm bio-waste/multiwalled carbon nanotubes reinforced PVA hydrogel	30.03	7.0	[53]
Triamino-functionalized KCC-1/chitosan-oleic acid nanocomposites	168	9.0	[24]
Fe <sub>3</sub> O <sub>4</sub> /cyclodextrin polymer nanocomposite	64.5	5.5	[54]
FDU-12-MA/PMMA	99.0	9.0	This work
FDU-12-TA/NY6	94.3	9.0	This work

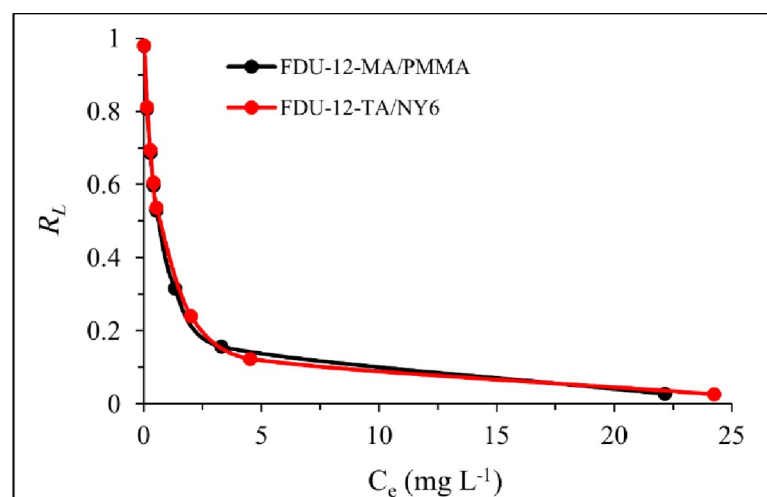
<https://doi.org/10.1371/journal.pone.0245583.t004>

the Langmuir model were found 99.0 and 94.3 mg g<sup>-1</sup>, respectively. The obtained theoretical maximum adsorption capacities were in good agreement with those obtained in experiments (93.3 and 90.2 mg g<sup>-1</sup> for FDU-12-MA/PMMA and FDU-12-TA/NY6 adsorbents, respectively). Table 4 shows data from the previous reports for the removal of Pb(II). It can be observed that the FDU-12-MA/PMMA and FDU-12-TA/NY6 adsorbents exhibited promising Pb(II) adsorption capacity when compared to other adsorbents.

Based on the Langmuir adsorption isotherm model assumption, there are a fixed number of identical adsorption sites on the saturated monolayer surface of the adsorbent and the energy of adsorption is constant. To predict the adsorption performance, the equilibrium parameter of  $R_L$  was calculated for the Langmuir model. It can be defined as follows:

$$R_L = \frac{1}{1 + (C_e k_L)} \quad (11)$$

The  $R_L$  value represents the performance of the adsorption process. In this regard,  $R_L = 0$ ,  $0 < R_L < 1$ ,  $R_L = 1$ , and  $R_L > 1$  suggests irreversible, favorable, linear, and unfavorable adsorption process. The calculated  $R_L$  values are shown in Fig 7. All the values were obtained in the range of 0 and 1 which represented favorable adsorption of Pb(II) onto the FDU-12-MA/PMMA and FDU-12-TA/NY6 adsorbents.

**Fig 7. The calculated values of  $R_L$ .**

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## 4. Conclusions

In conclusion, two types of nanocomposites including FDU-12-MA/PMMA and FDU-12-TA/NY6 were fabricated *via* the simple and fast solution polymerization technique. The three-dimensional large mesoporous silica FDU-12 was synthesized and modified with methacrylate and triamine moieties and incorporated into the organic polymers. Due to the relatively large surface area and abundant active sites of the prepared nanofillers, the newly prepared mesoporous silica-based nanocomposites showed good performance toward Pb(II) removal from aqueous solutions. After investigation of the affecting experimental parameters including sample solution pH, adsorbent dosage, contact time, and initial concentration of the metal ion, several kinetic models were studied and the best fit achieved with the pseudo-second-order model for adsorption of Pb(II) using FDU-12-MA/PMMA and FDU-12-TA/NY6. On the other hand, four isotherm models were applied to investigate the equilibrium adsorption studies. Among them, the Langmuir model showed the best fit ( $R^2$  values of 0.9970 and 0.9978 for FDU-12-MA/PMMA and FDU-12-TA/NY6, respectively). The maximum adsorption capacities of 99.0 and 94.3 mg g<sup>-1</sup> (by the Langmuir model) were obtained using FDU-12-MA/PMMA and FDU-12-TA/NY6, respectively. In conclusion, the prepared porous nanocomposites showed acceptable characteristics to be considered as effective adsorbents for heavy metals (e.g. Pb(II)) uptake from aqueous media.

## Supporting information

### S1 Graphical abstract.

(TIF)

## Author Contributions

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**Writing – review & editing:** Azam Marjani.

## References

1. Mohammadnezhad G, Soltani R, Abad S, Dinari M. A novel porous nanocomposite of aminated silica MCM-41 and nylon-6: Isotherm, kinetic, and thermodynamic studies on adsorption of Cu (II) and Cd (II). *Journal of Applied Polymer Science*. 2017; 134(40):45383.
2. Soltani R, Marjani A, Hosseini M, Shirazian S. Synthesis and characterization of novel N-methylimidazolium-functionalized KCC-1: A highly efficient anion exchanger of hexavalent chromium. *Chemosphere*. 2020; 239:124735. <https://doi.org/10.1016/j.chemosphere.2019.124735> PMID: 31499306

3. Rezaei M, Chermahini AN, Dabbagh HA, Saraji M, Shahvar A. Furfural oxidation to maleic acid with H<sub>2</sub>O<sub>2</sub> by using vanadyl pyrophosphate and zirconium pyrophosphate supported on well-ordered mesoporous KIT-6. *Journal of Environmental Chemical Engineering*. 2019; 7(1):102855.
4. Hosseini SM, Ghiaci M, Kulinich SA, Wunderlich W, Farrokhpour H, Saraji M, et al. Au-Pd@ g-C<sub>3</sub>N<sub>4</sub> as an efficient photocatalyst for visible-light oxidation of benzene to phenol: experimental and mechanistic study. *The Journal of Physical Chemistry C*. 2018; 122(48):27477–85.
5. Saini D, Kaushik J, Garg AK, Dalal C, Sonkar SK. N, S-codoped Carbon Dots for Nontoxic Cell Imaging and As a Sunlight-Active Photocatalytic Material for the Removal of Chromium. *ACS Applied Bio Materials*. 2020.
6. Shahzeydi A, Ghiaci M, Farrokhpour H, Shahvar A, Sun M, Saraji M. Facile and green synthesis of copper nanoparticles loaded on the amorphous carbon nitride for the oxidation of cyclohexane. *Chemical Engineering Journal*. 2019; 370:1310–21.
7. Babaei Z, Chermahini AN, Dinari M, Saraji M, Shahvar A. Cleaner production of 5-hydroxymethylfurfural from fructose using ultrasonic propagation. *Journal of cleaner production*. 2018; 198:381–8.
8. Chermahini AN, Hafizi H, Andisheh N, Saraji M, Shahvar A. The catalytic effect of Al-KIT-5 and KIT-5-SO<sub>3</sub>H on the conversion of fructose to 5-hydroxymethylfurfural. *Research on Chemical Intermediates*. 2017; 43(10):5507–21.
9. Babaei Z, Chermahini AN, Dinari M, Saraji M, Shahvar A. A sulfonated triazine-based covalent organic polymer supported on a mesoporous material: a new and robust material for the production of 5-hydroxymethylfurfural. *Sustainable Energy & Fuels*. 2019; 3(4):1024–32.
10. Dastkhooon M, Ghaedi M, Asfaram A, Arabi M, Ostovan A, Goudarzi A. Cu@ SnS/SnO<sub>2</sub> nanoparticles as novel sorbent for dispersive micro solid phase extraction of atorvastatin in human plasma and urine samples by high-performance liquid chromatography with UV detection: application of central composite design (CCD). *Ultrasonics sonochemistry*. 2017; 36:42–9. <https://doi.org/10.1016/j.ultsonch.2016.10.030> PMID: 28069228
11. Soltani R, Shahvar A, Dinari M, Saraji M. Environmentally-friendly and ultrasonic-assisted preparation of two-dimensional ultrathin Ni/Co-NO<sub>3</sub> layered double hydroxide nanosheet for micro solid-phase extraction of phenolic acids from fruit juices. *Ultrasonics sonochemistry*. 2018; 40:395–401. <https://doi.org/10.1016/j.ultsonch.2017.07.031> PMID: 28946438
12. Saraji M, Shahvar A. Selective micro solid-phase extraction of epinephrine, norepinephrine and dopamine from human urine and plasma using aminophenylboronic acid covalently immobilized on magnetic nanoparticles followed by high-performance liquid chromatography-fluorescence detection. *Analytical Methods*. 2016; 8(4):830–9.
13. Shahvar A, Soltani R, Saraji M, Dinari M, Alijani S. Covalent triazine-based framework for micro solid-phase extraction of parabens. *Journal of Chromatography A*. 2018; 1565:48–56. <https://doi.org/10.1016/j.chroma.2018.06.033> PMID: 29921466
14. Saraji M, Shahvar A. Metal-organic aerogel as a coating for solid-phase microextraction. *Analytica Chimica Acta*. 2017; 973:51–8. <https://doi.org/10.1016/j.aca.2017.04.029> PMID: 28502427
15. Hubbell JA, Chilkoti A. Nanomaterials for drug delivery. *Science*. 2012; 337(6092):303–5. <https://doi.org/10.1126/science.1219657> PMID: 22822138
16. Awad AM, Jalab R, Benamor A, Naser MS, Ba-Abbad MM, El-Naas M, et al. Adsorption of organic pollutants by nanomaterial-based adsorbents: An overview. *Journal of Molecular Liquids*. 2020:112335.
17. Kegl T, Kořak A, Lobnik A, Novak Z, Kralj AK, Ban I. Adsorption of rare earth metals from wastewater by nanomaterials: A review. *Journal of hazardous materials*. 2020; 386:121632. <https://doi.org/10.1016/j.jhazmat.2019.121632> PMID: 31753662
18. Soltani R, Marjani A, Shirazian S. A hierarchical LDH/MOF nanocomposite: single, simultaneous and consecutive adsorption of a reactive dye and Cr(vi). *Dalton Transactions*. 2020; 49(16):5323–35. <https://doi.org/10.1039/d0dt00680g> PMID: 32248208
19. Abu-Danso E, Peräniemi S, Leiviskä T, Kim T, Tripathi KM, Bhatnagar A. Synthesis of clay-cellulose biocomposite for the removal of toxic metal ions from aqueous medium. *Journal of hazardous materials*. 2020; 381:120871. <https://doi.org/10.1016/j.jhazmat.2019.120871> PMID: 31374372
20. Bajpai VK, Shukla S, Khan I, Kang S-M, Haldorai Y, Tripathi KM, et al. A sustainable graphene aerogel capable of the adsorptive elimination of biogenic amines and bacteria from soy sauce and highly efficient cell proliferation. *ACS Applied Materials & Interfaces*. 2019; 11(47):43949–63. <https://doi.org/10.1021/acsami.9b16989> PMID: 31684721
21. Soltani R, Shahvar A, Gordan H, Dinari M, Saraji M. Covalent triazine framework-decorated phenyl-functionalised SBA-15: its synthesis and application as a novel nanoporous adsorbent. *New Journal of Chemistry*. 2019; 43(33):13058–67.

22. Morales-Luckie RA, Sánchez-Mendieta V, Olea-Mejia O, Vilchis-Nestor AR, López-Téllez G, Varela-Guerrero V, et al. Facile solventless synthesis of a nylon-6, 6/silver nanoparticles composite and its XPS study. *International Journal of Polymer Science*. 2013; 2013.
23. Mohammadnezhad G, Abad S, Soltani R, Dinari M. Study on thermal, mechanical and adsorption properties of amine-functionalized MCM-41/PMMA and MCM-41/PS nanocomposites prepared by ultrasonic irradiation. *Ultrasonics sonochemistry*. 2017; 39:765–73. <https://doi.org/10.1016/j.ultsonch.2017.06.001> PMID: 28733004
24. Zarei F, Marjani A, Soltani R. Novel and green nanocomposite-based adsorbents from functionalised mesoporous KCC-1 and chitosan-oleic acid for adsorption of Pb (II). *European Polymer Journal*. 2019; 119:400–9.
25. Soltani R, Marjani A, Moguei MRS, Rostami B, Shirazian S. Novel diamino-functionalized fibrous silica submicro-spheres with a bimodal-micro-mesoporous network: Ultrasonic-assisted fabrication, characterization, and their application for superior uptake of Congo red. *Journal of Molecular Liquids*. 2019; 294:111617.
26. Soltani R, Marjani A, Hosseini M, Shirazian S. Meso-architected siliceous hollow quasi-capsule. *Journal of Colloid and Interface Science*. 2020. <https://doi.org/10.1016/j.jcis.2020.03.003> PMID: 32182479
27. Soltani R, Marjani A, Hosseini M, Shirazian S. Mesostructured Hollow Siliceous Spheres for Adsorption of Dyes. *Chemical Engineering & Technology*. 2020; 43(3):392–402.
28. Wu Q, Li Y, Hou Z, Xin J, Meng Q, Han L, et al. Synthesis and characterization of Beta-FDU-12 and the hydrodesulfurization performance of FCC gasoline and diesel. *Fuel processing technology*. 2018; 172:55–64.
29. Fan J, Yu C, Gao F, Lei J, Tian B, Wang L, et al. Cubic mesoporous silica with large controllable entrance sizes and advanced adsorption properties. *Angewandte Chemie International Edition*. 2003; 42(27):3146–50. <https://doi.org/10.1002/anie.200351027> PMID: 12866103
30. Fan J, Yu C, Lei J, Zhang Q, Li T, Tu B, et al. Low-temperature strategy to synthesize highly ordered mesoporous silicas with very large pores. *Journal of the American Chemical Society*. 2005; 127(31):10794–5. <https://doi.org/10.1021/ja052619c> PMID: 16076161
31. Yu T, Zhang H, Yan X, Chen Z, Zou X, Oleynikov P, et al. Pore structures of ordered large cage-type mesoporous silica FDU-12s. *The Journal of Physical Chemistry B*. 2006; 110(43):21467–72. <https://doi.org/10.1021/jp064534j> PMID: 17064096
32. Carmona D, Balas F, Santamaria J. Pore ordering and surface properties of FDU-12 and SBA-15 mesoporous materials and their relation to drug loading and release in aqueous environments. *Materials Research Bulletin*. 2014; 59:311–22.
33. Wei Q, Nie Z, Hao Y, Chen Z, Zou J, Wang W. Direct synthesis of thiol-ligands-functionalized SBA-15: Effect of 3-mercaptopropyltrimethoxysilane concentration on pore structure. *Materials Letters*. 2005; 59(28):3611–5.
34. Zou B, Hu Y, Jiang L, Jia R, Huang H. Mesoporous material SBA-15 modified by amino acid ionic liquid to immobilize lipase via ionic bonding and cross-linking method. *Industrial & Engineering Chemistry Research*. 2013; 52(8):2844–51.
35. Kao H-M, Chang P-C, Wu J-D, Chiang AS, Lee C-H. Direct synthesis, characterization and solid-state NMR spectroscopy of large-pore vinyl-functionalized cubic mesoporous silica FDU-12. *Microporous and mesoporous materials*. 2006; 97(1–3):9–20.
36. Sarvi MN, Bee TB, Gooi CK, Woonton BW, Gee ML, O'Connor AJ. Development of functionalized mesoporous silica for adsorption and separation of dairy proteins. *Chemical Engineering Journal*. 2014; 235:244–51.
37. Deka JR, Saikia D, Lai Y-S, Tsai C-H, Chang W-C, Kao H-M. Roles of nanostructures and carboxylic acid functionalization of ordered cubic mesoporous silicas in lysozyme immobilization. *Microporous and Mesoporous Materials*. 2015; 213:150–60.
38. Hodgkins RP, Garcia-Bennett AE, Wright PA. Structure and morphology of propylthiol-functionalised mesoporous silicas templated by non-ionic triblock copolymers. *Microporous and mesoporous materials*. 2005; 79(1–3):241–52.
39. Cui H-Z, Li Y-L, Liu S, Zhang J-F, Zhou Q, Zhong R, et al. Novel Pb (II) ion-imprinted materials based on bis-pyrazolyl functionalized mesoporous silica for the selective removal of Pb (II) in water samples. *Microporous and Mesoporous Materials*. 2017; 241:165–77.
40. Molla-Abbasi P, Ghaffarian SR, Danesh E. Porous carbon nanotube/PMMA conductive composites as a sensitive layer in vapor sensors. *Smart materials and structures*. 2011; 20(10):105012.
41. Moller K, Bein T, Fischer RX. Entrapment of PMMA polymer strands in micro- and mesoporous materials. *Chemistry of materials*. 1998; 10(7):1841–52.

42. Beecroft LL, Ober CK. Nanocomposite materials for optical applications. *Chemistry of materials*. 1997; 9(6):1302–17.
43. Meng Q, Du P, Wang B, Duan A, Xu C, Zhao Z, et al. Synthesis of zirconium modified FDU-12 by different methods and its application in dibenzothiophene hydrodesulfurization. *RSC advances*. 2018; 8(48):27565–73.
44. Dinari M, Mohammadnezhad G, Soltani R. Fabrication of poly (methyl methacrylate)/silica KIT-6 nanocomposites via in situ polymerization approach and their application for removal of Cu 2+ from aqueous solution. *RSC advances*. 2016; 6(14):11419–29.
45. Patel JP, Xiang ZG, Hsu SL, Schoch AB, Carleen SA, Matsumoto D. Path to achieving molecular dispersion in a dense reactive mixture. *Journal of Polymer Science Part B: Polymer Physics*. 2015; 53(21):1519–26.
46. Patel JP, Deshmukh S, Zhao C, Wamuo O, Hsu SL, Schoch AB, et al. An analysis of the role of nonreactive plasticizers in the crosslinking reactions of a rigid resin. *Journal of Polymer Science Part B: Polymer Physics*. 2017; 55(2):206–13.
47. Patel JP, Xiang ZG, Hsu SL, Schoch AB, Carleen SA, Matsumoto D. Characterization of the crosslinking reaction in high performance adhesives. *International Journal of Adhesion and Adhesives*. 2017; 78:256–62.
48. Liu Q, Dong H. In-situ Immobilizing Ni Nanoparticles to FDU-12 via Trehalose with Fine Size and Location Control for CO<sub>2</sub> Methanation. *ACS Sustainable Chemistry & Engineering*. 2020.
49. Huang L, Yan X, Kruk M. Synthesis of ultralarge-pore FDU-12 silica with face-centered cubic structure. *Langmuir*. 2010; 26(18):14871–8. <https://doi.org/10.1021/la102228u> PMID: 20726611
50. Lawrence G, Anand C, Strounina E, Vinu A. Biomolecule Encapsulation Over Mesoporous Silica with Ultra-Large Tuneable Porous Structure Prepared by High Temperature Microwave Process. *Science of Advanced Materials*. 2014; 6(7):1481–8.
51. Nyairo WN, Eker YR, Kowenje C, Akin I, Bingol H, Tor A, et al. Efficient adsorption of lead (II) and copper (II) from aqueous phase using oxidized multiwalled carbon nanotubes/polypyrrole composite. *Separation Science and Technology*. 2018; 53(10):1498–510.
52. Liu Y, Liu Z, Gao J, Dai J, Han J, Wang Y, et al. Selective adsorption behavior of Pb (II) by mesoporous silica SBA-15-supported Pb (II)-imprinted polymer based on surface molecularly imprinting technique. *Journal of hazardous materials*. 2011; 186(1):197–205. <https://doi.org/10.1016/j.jhazmat.2010.10.105> PMID: 21109351
53. Zulfiqar M, Lee SY, Mafize AA, Kahar NAMA, Johari K, Rabat NE. Efficient Removal of Pb (II) from Aqueous Solutions by Using Oil Palm Bio-Waste/MWCNTs Reinforced PVA Hydrogel Composites: Kinetic, Isotherm and Thermodynamic Modeling. *Polymers*. 2020; 12(2):430. <https://doi.org/10.3390/polym12020430> PMID: 32059376
54. Badruddoza AZM, Shawon ZBZ, Tay WJD, Hidajat K, Uddin MS. Fe<sub>3</sub>O<sub>4</sub>/cyclodextrin polymer nanocomposites for selective heavy metals removal from industrial wastewater. *Carbohydrate polymers*. 2013; 91(1):322–32. <https://doi.org/10.1016/j.carbpol.2012.08.030> PMID: 23044139