

International Journal of Environmental Research and Public Health



Article Predicted Future Mortality Attributed to Increases in Temperature and PM₁₀ Concentration under Representative Concentration Pathway Scenarios

Jiyun Jung¹, Jae Young Lee², Hyewon Lee^{2,3} and Ho Kim^{1,2,*}

- ¹ Graduate School of Public Health, Seoul National University, 1 Gwanak-Ro, Gwanak-Gu, Seoul 08826, Korea; bestjudy@hanmail.net
- ² Institute of Health and Environment, Seoul National University, 1 Gwanak-Ro, Gwanak-Gu, Seoul 08826, Korea; jaeyoung.lee@alumni.stanford.edu (J.Y.L.); woniggo@gmail.com (H.L.)
- ³ Department of Neuropsychiatry, Seoul National University Bundang Hospital, Seongnam 13620, Korea
- * Correspondence: hokim@snu.ac.kr; Tel.: +(82)-2-880-2701-3; Fax: +82-(2)-745-9104

Received: 1 February 2020; Accepted: 7 April 2020; Published: 10 April 2020



Abstract: As climate change progresses, understanding the impact on human health associated with the temperature and air pollutants has been paramount. However, the predicted effect on temperature associated with particulate matter (PM_{10}) is not well understood due to the difficulty in predicting the local and regional PM_{10} . We compared temperature-attributable mortality for the baseline (2003–2012), 2030s (2026–2035), 2050s (2046–2055), and 2080s (2076–2085) based on a distributed lag non-linear model by simultaneously considering assumed levels of PM_{10} on historical and projected temperatures under representative concentration pathway (RCP) scenarios. The considered projected PM_{10} concentrations of 35, 50, 65, 80, and 95 $\mu g/m^3$ were based on historical concentration quantiles. Our findings confirmed greater temperature-attributable risks at PM_{10} concentrations above 65 $\mu g/m^3$ due to the modification effect of the pollutants on temperature. In addition, this association between temperature and PM_{10} was higher under RCP8.5 than RCP4.5. We also confirmed regional heterogeneity in temperature-attributable deaths by considering PM_{10} concentrations in South Korea with higher risks in heavily populated areas. These results demonstrated that the modification association of air pollutants on health burdens attributable to increasing temperatures should be considered by researchers and policy makers.

Keywords: modification effect; high-temperature; inhalable particulate matter; regional variation

1. Introduction

In recent years, health risks have increased due to climate change associated with advancing industrialization. Studies based on historical data have shown the impact of temperature on human health, including the total and cause-specific mortality [1] and morbidity [2]. In particular, increased mortality was associated with extremely cold or hot days, usually defined as percentile and value of temperature [1,3,4].

Due to adverse effect of climate change on human health, the projection of future mortality rates under representative concentration pathway (RCP) scenarios has become an important area of study. RCP is based on different energy use, economic, and demographic assumptions, which is the typical trajectory of gas emissions recommended by the 5th Intergovernmental Panel on Climate Change (IPCC) assessment report. RCP8.5 is supposed to be the current trend of GHG emissions as Business As Usual (BAU), while RCP4.5 is a mitigation scenario. In the 51 largest cities of China, the projected heat-related excess mortality in 2041–2061 relative to 1970–2000 was 37,800 and 31,700 per year under RCP8.5 and RCP4.5, respectively [5]. A study of metropolitan areas in the United States also estimated heat-related deaths under RCP8.5 and RCP4.5 during 2085–2095 compared to 1992–2002, highlighting the risk of high temperatures [6]. Likewise, the projected heat-related risks on health have been studied in England [7], India [8], and South Korea [9]. However, few studies have considered the modification effect of air pollution on temperature-related mortality.

The complex interactions between temperature and air pollutants have been investigated based on historical data and have shown an increased risk of mortality due to the modification effect [10–12]. Two methods have been used to investigate the interaction of air pollution and temperature: (1) the modification effect of air pollution on temperature-related mortality, and (2) the modification effect of temperature on air-pollutant-related mortality. Using the former method, we estimated the modification effect of air pollution on temperature-related mortality. The effect of air pollution on temperature-related mortality has been identified in several studies [13–15]. However, few projection studies have examined effects on health by considering both particulate matter (PM) and temperature because of the difficulties in projecting future PM levels. Predicting the effect of climate change on PM is challenging because of unreliable deterministic predictive variables such as precipitation, frequency, and mixing depth [16].

Therefore, to better understand the modification effect of air pollution on temperature-attributable mortality, our research analyzed the effects on health by simultaneously considering not only the assumed levels of PM_{10} , but also the projected temperature under RCP scenarios. Unlike the data for $PM_{2.5}$ concentrations, PM_{10} was measured across the country. In this study, the assumed PM_{10} level was expressed in a simple scenario based on historical PM_{10} concentration quantiles. The current and anticipated health effects of temperature were considered by including the modification effect of PM_{10} . The results of our study could be used to provide guidelines for policy makers on hazardous air pollution levels under all climate change scenarios.

2. Methods

2.1. Data Collection

We obtained historical data on the baseline (2003–2012) and projected data for the projection period (2030s, 2050s, and 2080s) from 229 districts in South Korea. For baseline, we collected non-accidental mortality from the Korean Statistical Information Service (KOSIS) and historical climate data, such as daily mean temperature, the ambient daily concentration of PM₁₀, and humidity, from the Korean Meteorological Administration (KMA) for 229 districts in South Korea. We obtained historical PM_{10} measured from 274 stations through the national atmospheric monitoring network. We assigned the average of hourly basis into daily mean concentration. For the projection period, future daily mean temperature under representative concentration pathways (RCPs) scenarios were obtained from the Climate Change Information Center. RCP scenarios were simulated for up to 2100 by the Hadley Centre Global Environmental Model version 3 (HadGEM3-RA, Korea) regional climate model and the Modified Korean-Parameter-elevation Regressions an Independent Slopes Model (MK-PRISM, Korea) in Korea. Our study employed RCP4.5 and RCP8.5 for the 2030s (2026–2035), 2050s (2046–2055), and 2080s (2076–2085) from 230 districts. Because the future climate data were provided for 230 districts, the average temperature of two districts in the predicted period, namely Chungcheongbuk-do Cheongwon-gun and Chungcheongnam-do Yeonggi-gun, was used for the Sejong-si district in historical data to match the administrative district.

2.2. Temperature-Mortality Relationship Considering Interaction between PM₁₀ and Temperature

City-specific exposure-response curves of seven major cities in Korea were extracted using a distributed lag non-linear model (DLNM) with the temperature, day of week, time, and humidity. In addition, the interactions between temperature and PM_{10} were included or not included, depending on the analysis. The DLNM method adjusted for the non-linear relationship and the lag effect between temperature and non-accidental mortality [17]. A quadratic B-spline function with

knots placed at the 10th, 75th, and 90th percentiles was selected for assessing the temperature-mortality relationship. We placed lag knots at intervals of the log scale. Maximum lags of 21 were used with three degrees of freedom because cold-related mortality has the highest association with longer lag [18,19], and we used time adjustment by using a natural spline (ns) with eight degrees of freedom based on QAIC. The complete regression model used in Equation (1) is given by

$$log(E(Y)) = \alpha + cb(temp) + cb(temp) * cPM_{10} + PM_{10} + Factor(day of week) +ns(date, df = 8 * years) + humidity$$
(1)

where cPM₁₀ is the centered PM₁₀ at 35, 50, 65, 80, or 95 μ g/m³. *PM*₁₀ concentration at baseline was used in the model without removing outliers. We used humidity as the covariate because increased humidity was associated with decreased temperature as well as formation of secondary particles. *cPM*₁₀ indicated the modeled *PM*₁₀ concentration, which was centered at *PM*₁₀ for 35, 50, 65, 80, or 95 μ g/m³. In addition, Y, α , and cb(temp) indicated the death counts, the intercept, and the cross-basis object for temperature, respectively.

The estimated association between temperature and mortality in seven major cities was then pooled through multivariate meta-regression models using the coefficients of our basic model. We included the difference in the temperature range and average temperature of seven cities as meta-regression predictors to consider district heterogeneity [20] and to obtain stable relative risk. The pooled estimate from the seven cities was used for all other districts. The relative risks (RR) were based on 99th versus 95th percentile temperatures for 229 districts and 4 periods (baseline and projected periods) in order to investigate extreme heat effects, assuming that adaptation has occurred in the projected period [21]. Therefore, we obtained different RRs by district and period because the 95th and 99th percentile temperatures varied by district and period, even though we derived one curve for the relationship between mortality and temperature in Korea.

2.3. Historical Mortality at Baseline and Estimated Mortality of Projected Period

The number of deaths at baseline was presented as the sum of non-accident-related mortality between 2003 and 2012. Future mortality was estimated based on the United Nations (UN) projection from the World Population Prospects (WPP 2017) medium-variant scenario. The medium-variant scenario was selected for this work because our research focused only on the modification effect of PM_{10} on temperature rather than on the trends of deaths that depend on various mortality scenarios. Thus, our use of the medium-variant scenario reduced the uncertainty in our estimation of the number of future attributable deaths. Estimates from the WPP model are given in five-year increments, so the estimated deaths for the predicted period in the 2030s, 2050s, and 2080s were average estimates between 2025–2030 and 2030–2035, 2045–2050 and 2050–2055, and 2075–2080 and 2080–2085, respectively.

Because the UN mortality data provided only the total number of deaths in Korea, the mortality rate of the baseline period (the number of deaths per district divided by the total number of deaths) was used to predict the mortality in 229 districts. Therefore, the ten-year average deaths of the 229 regions in the projected period were calculated in Equation (2) using the mortality rate for the ten-year average deaths for those regions in baseline:

$$Deaths_{ii} = predicted \ deaths_i \times mortality \ rate \ of \ baseline_i$$
(2)

$$(i = 1, 2, 3 \text{ and } j = 1, \dots, 229)$$

Attributable deaths_{ij} = $\frac{RR_{ij} - 1}{RR_{ii}} \times Deaths_{ij}$ (3)

Here, *i* and *j* indicate the projected period (2030s, 2050, and 2080s) and the district number for the 229 considered districts, respectively. Finally, Equation (3) estimates attributable deaths. All analyses were conducted by R version 3.1.0 (R Foundation for Statistical Computing, Vienna, Austria).

3. Results

Table 1 and Figure S1 show the 95th and 99th percentile temperature during baseline and the projection period (2030s, 2050s and 2080s). The 95th and 99th percentile temperatures under RCP8.5 were both higher than those under RCP4.5. Both percentiles in the 2080s were higher than those in the 2030s.

Table 1. The 95th and 99th percentile temperatures for the 229 districts in South Korea at baseline and in the projected periods (2030s, 2050s, and 2080s).

Democratile	RCP Scenario	Period				
Percentile		Baseline	2030s	2050s	2080s	
95%	4.5 8.5	26.61 ± 1.14	26.74 ± 1.17 26.75 ± 1.11	27.27 ± 1.25 28.52 ± 1.29	28.02 ± 1.20 29.83 ± 1.21	
99%	4.5 8.5	28.46 ± 1.10	28.96 ± 1.25 28.75 ± 1.15	29.49 ± 1.15 30.80 ± 1.29	30.34 ± 1.16 31.90 ± 1.25	

Figure 1 shows the mean PM_{10} concentration based on historical data from 2003 to 2012. The mean and standard deviation of the baseline were 53.61 µg/m³ and 28.48 µg/m³, respectively. Outliers, which were defined as values outside the lower limit (Q1 – 1.5 × interquartile range (IQR: Q3–Q1)) and upper limit (Q3 + 1.5 × IQR), were removed to identify the trends over a decade. The mean and standard deviation after removing outliers were 50.5 µg/m³ and 20.03 µg/m³, respectively. Because PM_{10} concentration trends in the baseline were not evident, the 25th, 50th, and 75th percentiles of the outlier-removed data were maintained in the projected period. Therefore, we utilized the assumed future PM_{10} concentrations of 35, 50, and 65. The concentrations of 80 and 95 µg/m³ were additionally analyzed based on the interval of 25th, 50th, and 75th percentiles to verify the effect on relatively high concentrations.

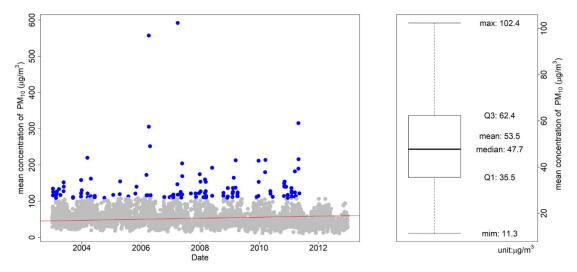


Figure 1. Mean PM₁₀ concentration distribution (grey dot) with outlier (blue dot) and its trend (red line) during the baseline period (2003–2012) (**left**), and the quartile concentration of the baseline period without outliers, which ranged from Q1 – $1.5 \times$ IQR to Q3 + $1.5 \times$ IQR (**right**).

The relationships between temperature and mortality were non-linear for all considered PM_{10} concentration values (Figure 2). Similar associations were found in seven major cities before meta-analysis (Figure S2). The relative risk shown here is the temperature-attributable risk for various PM_{10} levels. As shown in Figure 2, the relative risk for PM_{10} concentrations of 35 and 50 µg/m³ was lower than that not considering the PM_{10} concentration (0 µg/m³). However, when the concentration of PM_{10} varied from 65 µg/m³ to 95 µg/m³, the relative risk increased. Thus, as PM_{10}

concentration increased above 50 μ g/m³, the relative risks of high temperature on mortality also increased. In addition, we found the tendency to decrease relative risk as the lag period get longer (Figure S3).

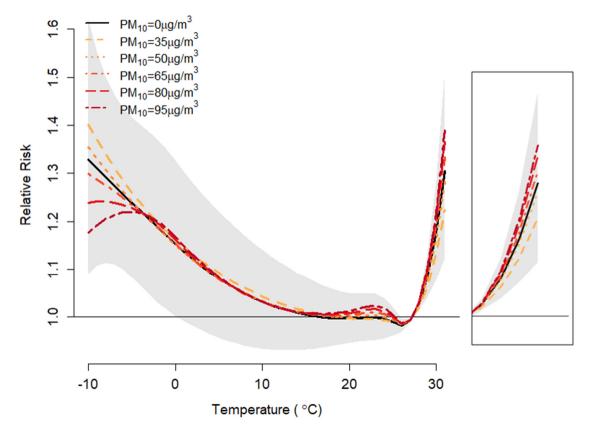


Figure 2. Estimated temperature-mortality curves with various PM_{10} concentrations (0, 35, 50, 65, 80, and 95 µg/m³) in South Korea (**left**) with 95% confidence interval (gray shaded region), and enlarged high-temperature region (**right**).

Figure 3 shows boxplots for the relative risk in the baseline, 2030s, 2050s, and 2080s when PM_{10} concentration increases. Table S1 also shows the mean temperature-attributable relative risks and 95% confidence interval with various PM_{10} levels in the 229 districts in South Korea. In the baseline period, the average relative risk without considering PM₁₀ concentrations was equal to or lower than those between 50 μ g/m³ and 65 μ g/m³. Interestingly, risks at relatively low concentrations such as 35 μ g/m³ and 50 μ g/m³ were lower than those not considering the PM₁₀ interaction effect. However, the relative risk of temperature, considering the association of temperature and air pollution under all scenarios and PM_{10} concentrations, tended to increase with concentrations of PM_{10} at higher temperature ranges. Table 2 and Figure S4 show the mean of the attributable death counts for the baseline and projected period by applying the relative risk and estimated mortality for 229 districts. Generally, attributable deaths increased with the period and scenario. When the PM_{10} concentration was assumed to be $95 \,\mu g/m^3$, the temperature-attributable mortality increased from 74.89 to 552.62 under RCP4.5, and to 673.91 under RCP8.5. The results indicated that PM₁₀ concentrations affect temperature and contribute to increased temperature-attributable deaths. Since RCP8.5 assumes greater temperature increases than RCP4.5, the number of attributable deaths was greater under RCP8.5 for both the 2050s and 2080s. However, this tendency was not observed in the 2030s because the projected temperatures under RCP4.5 and 8.5 were similar.

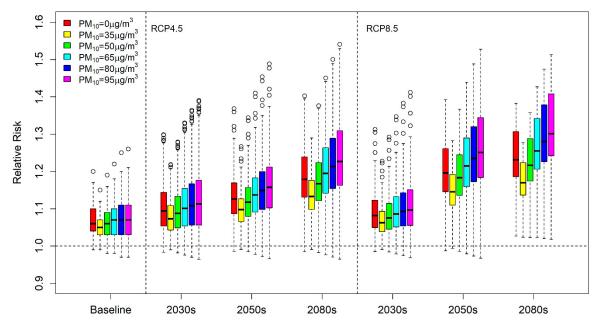


Figure 3. Relative risk boxplots of the 229 districts for the baseline and projection periods (2030s (2026–2035), 2050s (2046–2055), and 2080s (2076–2085)) under predicted temperature scenarios (RCP4.5 and 8.5) and the medium-variant population scenario. Red, yellow, green, light blue, navy, and pink colors correspond to PM_{10} levels of 0, 35, 50, 65, 80, and 95 µg/m³, respectively.

Table 2. Ten-year-average attributable deaths and the 95% confidence interval of the baseline and	nd		
projected periods (2030s, 2050s, and 2080s) by considering the relative risks of 229 districts.			

Period/Scenario		Levels of PM_{10} (in $\mu g/m^3$)							
		0	35	50	65	80	95		
Ba	seline	65.68 ± 62.7	51.53 ± 48.29	61.92 ± 59.88	68.72 ± 67.75	72.69 ± 73.19	74.89 ± 77.18		
4.5	2030s 2050s 2080s	$\begin{array}{c} 153.37 \pm 160.84 \\ 366.2 \pm 323.02 \\ 453.61 \pm 383.98 \end{array}$	$\begin{array}{c} 120.34 \pm 125.26 \\ 283.65 \pm 247.46 \\ 351.81 \pm 298.45 \end{array}$	$\begin{array}{c} 143.27 \pm 151.94 \\ 342.62 \pm 303.27 \\ 428.68 \pm 365.06 \end{array}$	$\begin{array}{c} 161.9 \pm 173.51 \\ 389.85 \pm 346.93 \\ 488.63 \pm 417.48 \end{array}$	$\begin{array}{c} 172.92 \pm 186.53 \\ 417.03 \pm 374.11 \\ 525.34 \pm 449.94 \end{array}$	$\begin{array}{r} 179.97 \pm 196.53 \\ 437.54 \pm 395.21 \\ 552.62 \pm 474.18 \end{array}$		
8.5	2030s 2050s 2080s	$\begin{array}{c} 127.59 \pm 124.56 \\ 540.54 \pm 462.33 \\ 545.09 \pm 428.93 \end{array}$	$\begin{array}{c} 100.07 \pm 96.51 \\ 419.15 \pm 357.01 \\ 422.34 \pm 333.63 \end{array}$	$\begin{array}{c} 119.67 \pm 118.3 \\ 512.19 \pm 439.14 \\ 517.94 \pm 408.48 \end{array}$	$\begin{array}{c} 135.05 \pm 135.29 \\ 584.02 \pm 501.57 \\ 591.23 \pm 464.97 \end{array}$	$\begin{array}{c} 143.44 \pm 145.29 \\ 628.69 \pm 540.72 \\ 637.42 \pm 501.01 \end{array}$	$\begin{array}{c} 149.1 \pm 152.12 \\ 663.11 \pm 570.31 \\ 673.91 \pm 530.28 \end{array}$		

In this study, we also confirmed regional heterogeneity in temperature-attributable deaths considering the PM_{10} concentrations in South Korea because different relative risks and deaths were applied in attributable deaths. As shown in Figure 4, metropolitan cities such as Seoul and Busan with higher population densities showed higher attributable mortality than other locations. In addition, the number of attributable deaths depended on the characteristics of various districts such as temperature and estimated mortality.

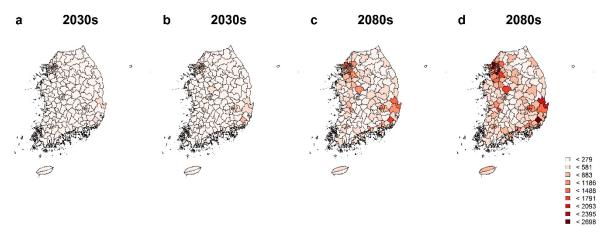


Figure 4. National distribution of temperature-attributable deaths considering simultaneous effects of PM_{10} and temperature when PM_{10} was assumed to be 35 µg/m³, for (**a**) and (**c**), and 95 µg/m³, for (**b**) and (**d**), under RCP8.5 in the 2030s and 2080s.

4. Discussion

This study examined how future temperature-related mortality changes under different assumed PM_{10} concentrations. We observed a modification effect due to PM_{10} in the relationship between temperature and mortality over the predicted periods, and we identified higher mortality risks at ranges of PM_{10} concentration over 65 µg/m³. Our results that confirmed the modification effect at the baseline agreed with previous studies conducted by Chen et al. (2018) [22] and Li et al. (2015) [23].

In this study, the attributable deaths in both RCP scenarios increased along the period because of temperature increase. In addition, the estimated temperature-attributable deaths at PM₁₀ concentrations of 65 and 95 μ g/m³ were higher than those at concentrations of 0, 35, and 50 μ g/m³. This result suggests that the interaction effect between temperature and PM_{10} may depend on the PM_{10} concentration. Surface and atmospheric cooling occurs when aerosols are reflected, while atmospheric warming was observed when absorbing aerosols. However, various properties of aerosol and characteristics of surfaces could make the effects of atmospheric aerosols on temperature different [24]. Although the underlying mechanism of how air pollution modifies the temperature effect is unclear, several hypotheses have been proposed. First, there may be a synergistic effect between the ambient temperature and air pollution on mortality [10]. Correlation between those two variables is generally high in many places [25], and the PM concentration is highest during the warm season [26]. Second, air pollutant exposure may increase because people tend to open doors and windows or go outside on a warmer day [22]. Moreover, air pollutants are more readily absorbed by the body when temperature is high due to increased skin permeability and increased respiratory rate [23]. Third, a biological mechanism supporting the modification effect of PM_{10} on temperature is plausible, especially in the case of cardiorespiratory disease, which can lead to mortality through inhalation [14], and modulation of the automatic nervous system [27]. Further, temperature could aggravate preexisting disease because of physiological and psychological stress [27]. Therefore, the results of our study suggest the importance of ambient PM_{10} emission control because of the greater temperature-attributable risks at concentrations above 65 μ g/m³ due to the modification effect.

Our results also show regional variations, which demonstrated greater health risks in metropolitan areas. Local variation of health risk caused by temperature can be manifested by various sources. Region-specific temperatures may be strongly affected by orographic precipitation, coastlines, and other local climate patterns [28]. In addition, health risks depend on social, economic, and demographic characteristics [29]. In this study, the attributable risk was estimated to be greater in metropolitan areas with higher mortality; however, determinants of health risks between rural and metropolitan vary [30]. The heat island effect caused by poor ventilation, increased thermal storage in the urban environment, and heat generated from air conditioning and vehicles may increase risks in

metropolitan areas [31]. Moreover, people in high density settlements are more vulnerable to higher temperatures and thermal discomfort [32].

There were several limitations on our study. First, we assumed that the 229 district-specific mortality rates in the projected period did not change from the current regional mortality rates because of difficulty in estimating the changing rates. We assumed current association between PM_{10} and temperature is maintained in future periods, not considering adaptation and mitigation of climate change and local particulate matter. However, regional mortality differences could be caused by a variety of reasons [33], and rural regions can be burdened by diseases that are not prevalent in metropolitan areas [34]. Second, cold-related mortality with PM₁₀ modification was not addressed in this study. Based on Figure 2, the tendency of cold temperature-attributable deaths for various PM_{10} levels will differ from that of high temperature. Therefore, further studies are needed to investigate the biological mechanisms and the predicted attributable mortality for cold temperatures. Third, we used assumed levels of PM₁₀ rather than climate predicting of PM₁₀ levels through modeling. The prediction of PM levels has been performed by artificial neural networks [35], time-varying statistical models [36], and support vector machines [37] to overcome the difficulties in predicting the precipitation, wind, temperature, and relative humidity that affects PM_{10} . In addition, we did not estimate the PM_{10} concentration where the relative risk considering only temperature is as much as the relative risk considering the interaction effect. However, our assumption is simple and easy to interpret. Also, the identical temperature-mortality relationship of the baseline in the projected periods was used, although different thresholds of each period were applied. Finally, this study included the limitation of ecological studies such as uncontrolled confounding and measurement error.

Overall, we found an increased modification on association between particulate matter and temperature-attributable deaths and identified higher risks for PM_{10} concentrations over 65 µg/m³. Our findings also showed substantial geographical variation in heavily populated areas, which is a point of increasing interest in worldwide integrated research. Therefore, this research provides an essential foundation for researchers and policy makers in understanding heat-attributable health burdens considering PM_{10} .

Supplementary Materials: The following are available online at http://www.mdpi.com/1660-4601/17/7/2600/s1, Figure S1: The 95th and 99th percentile temperatures (°C) in the 2030s, 2050s, and 2080s, Figure S2: Association between temperature and mortality modified by air pollutants in 7 major cities in Korea, Figure S3: Association between lag on temperature and relative risk in considering various PM10, Figure S4: Ten-year-average attributable deaths and the 95% confidence interval for each period by considering the relative risks of 229 districts, Table S1: Mean relative risk and 95% confidence interval with various PM10 levels in the 229 considered districts.

Author Contributions: Supervision, H.K.; Validation, J.Y.L. and H.L.; Writing—original draft, J.J.; All authors have read and agreed to the published version of the manuscript.

Funding: This work was supported by the Korea Ministry of Environment via the "Climate Change Correspondence Program" (project number: 2014001310007).

Conflicts of Interest: The authors declare no conflict of interest.

References

- 1. Gasparrini, A.; Guo, Y.; Hashizume, M.; Lavigne, E.; Zanobetti, A.; Schwartz, J.; Tobias, A.; Tong, S.; Rocklöv, J.; Forsberg, B. Mortality risk attributable to high and low ambient temperature: A multicountry observational study. *Lancet* **2015**, *386*, 369–375.
- 2. Ye, X.; Wolff, R.; Yu, W.; Vaneckova, P.; Pan, X.; Tong, S. Ambient temperature and morbidity: A review of epidemiological evidence. *Environ. Health Perspect.* **2011**, *120*, 19–28. [PubMed]
- 3. Han, J.; Liu, S.; Zhang, J.; Zhou, L.; Fang, Q.; Zhang, J.; Zhang, Y. The impact of temperature extremes on mortality: A time-series study in Jinan, China. *BMJ Open* **2017**, *7*, e014741. [PubMed]
- Mills, D.; Schwartz, J.; Lee, M.; Sarofim, M.; Jones, R.; Lawson, M.; Duckworth, M.; Deck, L. Climate change impacts on extreme temperature mortality in select metropolitan areas in the United States. *Clim. Chang.* 2015, 131, 83–95.

- 5. Li, Y.; Ren, T.; Kinney, P.L.; Joyner, A.; Zhang, W. Projecting future climate change impacts on heat-related mortality in large urban areas in China. *Environ. Res.* **2018**, *163*, 171–185.
- Weinberger, K.R.; Haykin, L.; Eliot, M.N.; Schwartz, J.D.; Gasparrini, A.; Wellenius, G.A. Projected temperature-related deaths in ten large US metropolitan areas under different climate change scenarios. *Environ. Int.* 2017, 107, 196–204.
- 7. Weinberger, K.R.; Kirwa, K.; Eliot, M.N.; Gold, J.; Suh, H.H.; Wellenius, G.A. Projected changes in temperature-related morbidity and mortality in Southern New England. *Epidemiology* **2018**, *29*, 473–481.
- 8. Murari, K.K.; Ghosh, S.; Patwardhan, A.; Daly, E.; Salvi, K. Intensification of future severe heat waves in India and their effect on heat stress and mortality. *Reg. Environ. Chang.* **2015**, *15*, 569–579.
- 9. Kim, D.-W.; Deo, R.C.; Chung, J.-H.; Lee, J.-S. Projection of heat wave mortality related to climate change in Korea. *Nat. Hazards* **2016**, *80*, 623–637.
- 10. Katsouyanni, K.; Pantazopoulou, A.; Touloumi, G.; Tselepidaki, I.; Moustris, K.; Asimakopoulos, D.; Poulopoulou, G.; Trichopoulos, D. Evidence for interaction between air pollution and high temperature in the causation of excess mortality. *Arch. Environ. Health* **1993**, *48*, 235–242.
- 11. Kinney, P.L. Interactions of Climate Change, Air Pollution, and Human Health. *Curr. Environ. Health Rep.* **2018**, *5*, 179–186. [CrossRef] [PubMed]
- 12. Roberts, S. Interactions between particulate air pollution and temperature in air pollution mortality time series studies. *Environ. Res.* **2004**, *96*, 328–337. [CrossRef] [PubMed]
- Ren, C.; Williams, G.M.; Morawska, L.; Mengersen, K.; Tong, S. Ozone modifies associations between temperature and cardiovascular mortality: Analysis of the NMMAPS data. *Occup. Environ. Med.* 2008, 65, 255–260. [CrossRef] [PubMed]
- 14. Ren, C.; Williams, G.M.; Tong, S. Does particulate matter modify the association between temperature and cardiorespiratory diseases? *Environ. Health Perspect.* **2006**, *114*, 1690–1696. [CrossRef]
- Nawrot, T.; Torfs, R.; Fierens, F.; De Henauw, S.; Hoet, P.; Van Kersschaever, G.; De Backer, G.; Nemery, B. Stronger associations between daily mortality and fine particulate air pollution in summer than in winter: Evidence from a heavily polluted region in western Europe. *J. Epidemiol. Community Health* 2007, *61*, 146–149. [CrossRef]
- 16. Jacob, D.J.; Winner, D.A. Effect of climate change on air quality. Atmos. Environ. 2009, 43, 51–63. [CrossRef]
- 17. Gasparrini, A.; Armstrong, B.; Kenward, M.G. Distributed lag non-linear models. *Stat. Med.* **2010**, *29*, 2224–2234. [CrossRef]
- 18. Anderson, B.G.; Bell, M.L. Weather-related mortality: How heat, cold, and heat waves affect mortality in the United States. *Epidemiology* **2009**, *20*, 205. [CrossRef]
- 19. Díaz, J.; García, R.; López, C.; Linares, C.; Tobías, A.; Prieto, L. Mortality impact of extreme winter temperatures. *Int. J. Biometeorol.* 2005, *49*, 179–183. [CrossRef]
- Gasparrini, A.; Guo, Y.; Hashizume, M.; Kinney, P.L.; Petkova, E.P.; Lavigne, E.; Zanobetti, A.; Schwartz, J.D.; Tobias, A.; Leone, M. Temporal variation in heat–mortality associations: A multicountry study. *Environ. Health Perspect.* 2015, 123, 1200–1207. [CrossRef]
- 21. Marí-Dell'Olmo, M.; Tobías, A.; Gómez-Gutiérrez, A.; Rodríguez-Sanz, M.; de Olalla, P.G.; Camprubí, E.; Gasparrini, A.; Borrell, C. Social inequalities in the association between temperature and mortality in a South European context. *Int. J. Public Health* **2019**, *64*, 27–37. [CrossRef] [PubMed]
- Chen, K.; Wolf, K.; Breitner, S.; Gasparrini, A.; Stafoggia, M.; Samoli, E.; Andersen, Z.J.; Bero-Bedada, G.; Bellander, T.; Hennig, F. Two-way effect modifications of air pollution and air temperature on total natural and cardiovascular mortality in eight European urban areas. *Environ. Int.* 2018, *116*, 186–196. [CrossRef] [PubMed]
- Li, L.; Yang, J.; Guo, C.; Chen, P.-Y.; Ou, C.-Q.; Guo, Y. Particulate matter modifies the magnitude and time course of the non-linear temperature-mortality association. *Environ. Pollut.* 2015, *196*, 423–430. [CrossRef] [PubMed]
- 24. Mitchell, J.M., Jr. The effect of atmospheric aerosols on climate with special reference to temperature near the earth's surface. *J. Appl. Meteorol.* **1971**, *10*, 703–714. [CrossRef]
- 25. Brunekreef, B.; Holgate, S.T. Air pollution and health. Lancet 2002, 360, 1233–1242. [CrossRef]
- 26. Schwartz, J. Air pollution and daily mortality: A review and meta analysis. *Environ. Res.* **1994**, *64*, 36–52. [CrossRef]

- 27. Gordon, C.J. Role of environmental stress in the physiological response to chemical toxicants. *Environ. Res.* **2003**, *92*, 1–7. [CrossRef]
- 28. Patz, J.A.; Campbell-Lendrum, D.; Holloway, T.; Foley, J.A. Impact of regional climate change on human health. *Nature* **2005**, *438*, 310–317. [CrossRef]
- 29. McMichael, A.J.; Woodruff, R.E.; Hales, S. Climate change and human health: Present and future risks. *Lancet* **2006**, *367*, 859–869. [CrossRef]
- 30. Dixon, J.; Welch, N. Researching the rural–metropolitan health differential using the social determinants of health. *Aust. J. Rural Health* **2000**, *8*, 254–260.
- 31. Hajat, S.; Kosatky, T. Heat-related mortality: A review and exploration of heterogeneity. *J. Epidemiol. Community Health* **2010**, *64*, 753–760. [CrossRef] [PubMed]
- 32. Harlan, S.L.; Brazel, A.J.; Prashad, L.; Stefanov, W.L.; Larsen, L. Neighborhood microclimates and vulnerability to heat stress. *Soc. Sci. Med.* **2006**, *63*, 2847–2863. [CrossRef] [PubMed]
- 33. Lozano, R.; Naghavi, M.; Foreman, K.; Lim, S.; Shibuya, K.; Aboyans, V.; Abraham, J.; Adair, T.; Aggarwal, R.; Ahn, S.Y. Global and regional mortality from 235 causes of death for 20 age groups in 1990 and 2010: A systematic analysis for the Global Burden of Disease Study 2010. *Lancet* 2012, *380*, 2095–2128. [CrossRef]
- 34. Song, H.-N.; Go, S.-I.; Lee, W.S.; Kim, Y.; Choi, H.J.; Lee, U.S.; Kang, M.H.; Lee, G.-W.; Kim, H.-G.; Kang, J.H. Population-based regional cancer incidence in Korea: Comparison between urban and rural areas. *Cancer Res Treat* **2016**, *48*, 789. [CrossRef] [PubMed]
- 35. Grivas, G.; Chaloulakou, A. Artificial neural network models for prediction of PM10 hourly concentrations, in the Greater Area of Athens, Greece. *Atmos. Environ.* **2006**, *40*, 1216–1229. [CrossRef]
- Hoi, K.; Yuen, K.; Mok, K. Prediction of daily averaged PM10 concentrations by statistical time-varying model. *Atmos. Environ.* 2009, 43, 2579–2581. [CrossRef]
- 37. Taheri Shahraiyni, H.; Sodoudi, S. Statistical modeling approaches for PM10 prediction in urban areas; A review of 21st-century studies. *Atmosphere* **2016**, *7*, 15. [CrossRef]



© 2020 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (http://creativecommons.org/licenses/by/4.0/).