

SCIENTIFIC REPORTS



OPEN

Short-term effects of ambient air pollution and childhood lower respiratory diseases

Liyang Zhu^{1,2}, Xuhua Ge³, Yaoyao Chen¹, Xinying Zeng⁴, Wang Pan⁴, Xu Zhang⁴, Shuai Ben⁵, Qi Yuan¹, Junyi Xin¹, Wei Shao¹, Yuqiu Ge¹, Dongmei Wu¹, Zhong Han², Zhengdong Zhang¹, Haiyan Chu¹ & Meilin Wang¹

The association between air pollution and childhood respiratory disease is inconsistent. In the present study, we investigated a short-term effect of ambient air pollutants and daily childhood lower respiratory diseases (CLRD). Daily air pollutants, weather data, and CLRD data were collected from January 2014 to April 2015 (452 days) in Nanjing, China. Time-series regression and generalized additive models were used to assess the effects of air pollutants (PM₁₀, PM_{2.5}, NO₂, SO₂, O₃, and CO) on CLRD. We observed that an interquartile range (IQR) increase in concentrations of PM₁₀, NO₂, and SO₂ significantly increased the daily CLRD with 6 days cumulative effects (difference of estimates: 2.8%, 95% CI: 0.6–5.0%; 4.1%, 1.2–7.0%; 5.6%, 2.6–8.6%, respectively). However, no significant association was found in IQR concentrations of PM_{2.5}, O₃, and CO. Specifically, elevated PM₁₀, PM_{2.5}, NO₂, and SO₂ significantly increased the numbers of CLRD in cool season (3.6%, 1.5–5.7%; 2.4%, 0.3–4.5%; 4.9%, 2.9–7.0%; 6.3%, 3.7–9.0%, respectively). Additionally, the effect estimates of PM₁₀, NO₂, and SO₂ in female and age >27 months were more pronounced than in male and age ≤27 months. This study suggested that short-term exposure to ambient PM₁₀, NO₂, and SO₂ were associated with the increased CLRD numbers.

Air pollution is the important environment risk factor. Extensive data have indicated that air pollution is associated with a negative effect on human health¹. WHO had proposed outdoor air pollution as the 14th mortality risk factor in 2004. The ischemic heart disease and stroke were the main cause of deaths associated with air pollution. Additionally, pulmonary disorder also contributed to the disease mortality, among which acute lower respiratory diseases (LRD) was the common cause of children deaths under five years old².

It is well known that the response of individual air pollutant exposure may be determined by age, sex, genetics, lifestyle and disease. The children are vulnerable to the detrimental effects of outdoor air pollutants on health. Several previous studies demonstrated significantly harmful effects of outdoor air pollutants on childhood lung function, e.g. exposure to particulate matter less than 2.5 μm in aerodynamic diameter (PM_{2.5}), particulate matter less than 10 μm in aerodynamic diameter (PM₁₀) and nitrogen dioxide (NO₂) associated with reduced forced expiratory flows³. Recently, a few studies have found that acute and short-term exposure to outdoor air pollutants may increase the number of childhood allergic rhinitis, pneumonia, asthma symptoms^{4,5}. A recent meta-analysis including several cohort and cross-sectional studies suggested that exposure to traffic related PM₁₀, PM_{2.5}, NO_x, sulfur dioxide (SO₂), carbon monoxide (CO), ozone (O₃) had increased frequency of childhood asthma and wheeze⁶. Potential mechanisms may be air pollution inducing allergic sensitization and oxidative stress, increasing the individual susceptibility to respiratory infection⁷.

At present, few association studies between air pollutants and Childhood LRD (CLRD) have been conducted in China^{8,9}. A recent study in Changsha (China) reported that early-life exposure to air pollutants can induce

¹Department of Environmental Genomics, Jiangsu Key Laboratory of Cancer Biomarkers, Prevention and Treatment, Collaborative Innovation Center for Cancer Personalized Medicine, Nanjing Medical University, Nanjing, China.

²Department of Statistics, School of Economics, Nanjing University Of Finance & Economics, Nanjing, China.

³Department of Emergency, Children's Hospital of Nanjing Medical University, Nanjing, China. ⁴Department of Genetic Toxicology, The Key Laboratory of Modern Toxicology of Ministry of Education, School of Public Health, Nanjing Medical University, Nanjing, China. ⁵School of Public Health, Nantong University, Nantong, Jiangsu, China. Liyang Zhu, Xuhua Ge and Yaoyao Chen contributed equally to this work. Correspondence and requests for materials should be addressed to Z.Z. (email: drzdzhang@gmail.com) or H.C. (email: chy_grape@njmu.edu.cn) or M.W. (email: mwang@njmu.edu.cn)

the higher childhood asthma¹⁰. Liu *et al.* collected the daily concentrations of PM₁₀, NO₂, and SO₂ during the children life time (2006–2012) where they lived, and found that exposure to NO₂ was associated with respiratory diseases in children¹¹. Although the association studies of air pollutants and CLRD have been widely studied, the results remain inconsistent. Molter *et al.* found that outdoor air pollutants were not related with the multi-center childhood asthma in Europe¹². Studies demonstrated that different NO₂ exposure years of life were not associated with childhood asthma and asthma-related symptoms in the United States¹³.

It was worth to note that these studies had provided the effects of air pollutants on CLRD in location for public health assessments, however, the results might be affected by the local air pollutants levels, components and population susceptibility. It is important to study the effects of city air pollutants on CLRD numbers for the local city policy development¹⁴. Nanjing is a city located in the east of China, which is the provincial capital of Jiangsu province. With the development of economics, the concentrations of Nanjing's air pollutants in central monitoring sites are above the nation's levels. In the present study, we aim to conduct a short-term effect of ambient air pollutants, including PM₁₀, PM_{2.5}, NO₂, CO, and O₃, and daily CLRD in Nanjing, China.

Material and Methods

Data collection. Nanjing is a large city of Jiangsu province with a moderate subtropical climate (6597 km²). It had a population of 8.22 million in 2015. Daily CLRD data from 28 January 2014 to 24 April 2015 were obtained from Children's Hospital of Nanjing Medical University, The Second Affiliated Hospital of Nanjing Medical University, Jiangsu Women and Clinical Health Hospital, Nanjing Jiangning Hospital, Nanjing First Hospital, Nanjing Gaochun People's Hospital, Nanjing Maternity and Child Health Care Hospital. In this study, CLRD were comprised of doctor-diagnosed pneumonia (J18), bronchitis (J20), capillary bronchitis (J21), and acute asthma (J45), which were diagnosed based on the International Classification of Diseases 10th revision¹⁵.

Air pollution data including PM₁₀, PM_{2.5}, NO₂, SO₂, O₃, and CO were obtained from China Environmental Monitoring Centre. These monitoring stations are not in direct vicinity to traffic or industrial sources, housing emissions such as oil, coal, or waste burning. Thus, the daily air pollution data of monitoring stations can represent the environmental air pollutants' levels. To control the confounding effects of meteorological conditions on CLRD, we also obtained the data of daily weather data (relative humidity and mean temperature) from the Meteorological Bureau.

Statistical analysis. Spearman's correlation coefficient was applied to investigate the association between air pollutants and weather variables. We used the time-series regression model to explore the effects of each air pollutant (PM₁₀, PM_{2.5}, NO₂, SO₂, O₃, and CO) on CLRD. We also used generalized additive model (GAM) to calculate the data¹⁶. Herein, the data of the respiratory symptoms followed the over-dispersed Poisson distribution, thus, we applied the quasi-Poisson regression in the GAM¹⁷. In order to control the potential confounding effect, firstly, we included a natural cubic smooth function of calendar time with 7 degrees of freedom (*df*) per year to remove unmeasured long-term and seasonal trends¹⁸; secondly, we included the natural smooth functions of the current-day relative humidity (3 *df*) and mean temperature (3 *df*) to exclude the weather confounding effects¹⁹; thirdly, we also included the indicator variable for "day of the week (DOW)" in the model. In this study, we also calculated the data in the cool period (November through April, 4 *df*) and warm period (May to October, 4 *df*) to separately analyze the effect of air pollutants on CLRD.

Further, we used single-pollutant model to explore the air pollution's effect on CLRD with single lag days (lag0, 1, 2, 3, 4, 5, 6). To remove the potential misalignment of single day lag exposure, we considered the moving average exposure of multiple days (lag 0–1, 0–2, 0–3, 0–4, 0–5, 0–6) in additional analyses. We conducted the concentration-response relationship curves for the air pollutants concentrations with daily CLRD using a 3 *df* for the smoothing function in single-pollutant model. Herein, we performed the two-pollutants model to exam the stability of results.

The effect estimates were shown as the percentage change and 95% confidence intervals (CIs) in daily CLRD per interquartile range (IQR) increased of air pollutants' levels. In this study, all models were conducted in R software (version 3.2.1) with mgcv package. All tests were two-sided, and $P < 0.05$ were considered as statistically significant.

Results

Data description. Table 1 presents the data of descriptive variables. There were 26,423 CLRD cases (male: 16,078, female: 10,344, sex unknown: 1; mean age: 26.6 months, Supplementary Table 1) during 28 January 2014 to 24 April 2015 (452 days), among which there were 5,551 bronchitis, 19,098 pneumonia, 1,560 capillary bronchitis, and 214 asthma cases. The daily average CLRD numbers were 58.5, and the maximum was 105.0. The daily average concentrations were 117.4 µg/m³ for PM₁₀, 69.9 µg/m³ for PM_{2.5}, 50.1 µg/m³ for NO₂, 21.5 µg/m³ for SO₂, 0.94 mg/m³ for CO, and 95.8 µg/m³ for O₃. The daily average relative humidity and temperature were 72.5% and 15.1 °C, respectively.

As shown in Table 2, there were positively correlated among 6 air pollutants, except CO and O₃ ($r = -0.17$, $P = 0.001$). In addition, air pollutants were also correlated with temperature and humidity.

Main results. Single-pollutant models were applied to obtain the percentage change of CLRD associated with an IQR increase in air pollutants concentrations for single day lag (lag0, lag1, lag2, lag3, lag4, lag5, lag6) and multi-day lag (lag0–1, lag0–2, lag0–3, lag0–4, lag0–5, lag0–6) (Figure 1). An IQR increase in concentrations of PM₁₀, NO₂, SO₂ at multi-day exposure were more significant associated with estimated increase in the numbers of CLRD than single day exposure. For lag0–5, we observed the highest significant association between PM₁₀, NO₂, SO₂ and CLRD numbers (difference of estimates: 2.8%, 95% CI: 0.6–5.0%; 4.1%, 1.2–7.0%; 5.6%, 2.6–8.6%, respectively). However, no significant association was found in IQR concentrations of PM_{2.5}, O₃, and CO.

Further, we evaluated the effect estimates stratified by the season (warm and cool period). Figure 2 showed that the associations of PM₁₀, PM_{2.5}, NO₂, and SO₂ and CLRD numbers in cool period (3.6%, 1.5–5.7%; 2.4%, 0.3–4.5%; 4.9%, 2.9–7.0%; 6.3%, 3.7–9.0%, respectively) were more pronounced than in warm period. Additionally, we also explored the effect estimates of air pollutants on CLRD in the subgroup of sex, age and diseases (bronchitis,

Variables	Mean \pm SD	Minimum	25% quartile	Median	75% quartile	Maximum	Inter-quartile range
Daily CLRD numbers	58.5 \pm 13.9	23.0	48.0	58.5	68.3	105.0	20.3
Air pollution ($\mu\text{g}/\text{m}^3$)							
PM ₁₀	117.4 \pm 56.5	17.8	76.0	108.9	147.8	389.2	71.8
PM _{2.5}	69.9 \pm 36.7	13.7	45.4	63.0	87.6	247.3	42.2
NO ₂	50.1 \pm 18.8	14.2	36.2	47.5	62.3	118.1	26.1
SO ₂	21.5 \pm 11.5	3.21	12.4	19.8	28.8	71.7	16.4
O ₃	95.8 \pm 52.7	7.00	53.0	84.5	129.0	273.0	76.0
CO (mg/m^3)	0.94 \pm 0.89	0.35	0.70	0.89	1.11	2.45	0.41
Weather conditions							
Mean temperature ($^{\circ}\text{C}$)	15.1 \pm 8.2	-2.23	7.93	15.3	22.4	31.8	
Relative humidity (%)	72.5 \pm 15.5	28.0	62.2	72.9	85.4	97.2	

Table 1. The summary of descriptive statistics in this study. CLRD, Childhood lower respiratory diseases.

Variables	PM ₁₀	PM _{2.5}	NO ₂	SO ₂	CO	O ₃	Mean temperature ($^{\circ}\text{C}$)	Relative humidity (%)
PM ₁₀	—	0.87**	0.67**	0.75**	0.61**	0.24**	-0.07	-0.49**
PM _{2.5}	0.87**	—	0.53**	0.56**	0.68**	0.22**	0.03	-0.21**
NO ₂	0.67**	0.53**	—	0.75**	0.53**	0.05	-0.22**	-0.46**
SO ₂	0.75**	0.56**	0.75**	—	0.47**	0.14**	0.28**	0.71**
CO	0.61**	0.68**	0.53**	0.47**	—	-0.17**	-0.22**	-0.13**
O ₃	0.24**	0.22**	0.05	0.14**	-0.17**	—	0.58**	-0.26**
Mean temperature ($^{\circ}\text{C}$)	-0.07	0.03	-0.22**	-0.28**	-0.22**	0.58**	—	0.24**
Relative humidity (%)	-0.50**	-0.21**	-0.46**	-0.71**	-0.13**	-0.26**	0.24**	—

Table 2. Spearman correlation between air pollutants and weather variables in Nanjing. ** $P < 0.001$.

pneumonia, capillary bronchitis, and asthma cases) (Supplementary Figure 1–4). For PM₁₀, we observed that the effect estimates were pronounced in female (4.0%, 0.6–7.6%), age >27 months (13.3%, 9.0–17.6%), and pneumonia (2.7%, 0.1–5.3%). For NO₂ and SO₂, results suggested that the significant effect was found both in male (3.9%, 0.4–7.5%; 4.4%, 0.7%, 8.2%) and female (5.2%, 0.8–9.8%; 7.8%, 3.1–12.7%). When stratified by age, besides PM₁₀, an IQR increase in concentrations of PM_{2.5}, NO₂, SO₂ and O₃ exposure were significant associated with estimated increase in the numbers of CLRD (age >27 months) (12.2%, 8.5–16.0%; 13.2%, 7.8–19.0%; 9.9%, 4.5–15.6%; 12.9%, 6.2–19.9%, respectively). In the stratification of diseases, we found that PM_{2.5} was associated with positively increased risk of childhood asthma (28.2%, 3.2–59.3%), and NO₂ was associated with increased risk of pneumonia (4.0%, 0.7–7.4%) and capillary bronchitis (19.7%, 8.1–32.6%), and SO₂ was associated with increased risk of bronchitis (8.0%, 1.5–14.9%), pneumonia (4.3%, 0.9–7.8%) and capillary bronchitis (14.1%, 2.4–27.1%), and O₃ was associated with increased risk of bronchitis (14.9%, 6.5–23.8%).

Concentration-response relationships of CLRD numbers with air pollutants also were performed (Figure 3). The curves for PM₁₀, NO₂ and O₃ were similar and linearly positive and flat at higher concentrations. The effects of SO₂ tend to be linearly positive with its concentrations, and appeared to be stable when concentration was beyond 40 $\mu\text{g}/\text{m}^3$. However, the curves for PM_{2.5} and CO were linearly associated without any thresholds.

For lag 0–5, we also performed the two-pollutants model (Figure 4). Results suggested that the effect estimates of all six pollutants attenuated a little. However, when adjusted for PM_{2.5}, two-pollutants model of PM₁₀ produced the higher effect estimates for CLRD numbers than single-pollutant model (7.0%, 95% CI: 2.8–11.5%). Additionally, when adjusted for PM₁₀, an IQR increase in PM_{2.5} pollutant concentration decreased CLRD numbers (-4.1%, 95% CI: -7.6–0.6%).

Discussion

The source of air pollution is mainly manufacturing industries, traffic, agriculture, forest fires, etc., which result in increasing air pollutants level. Exposure to outdoor air pollutants such as particulate matter, NO₂, SO₂ had the adverse effects on the airway responsiveness and affected the immune competent cells^{20,21}. Epidemiological association studies between the air pollutants and childhood respiratory health have been studied widely; however, results are conflicting. For example, in a prospective birth cohort, Molter *et al.* did not find any evidence of the effect on the association between PM₁₀ and NO₂ and childhood asthma or wheeze²². Similar to their study, a study in Norway suggested that exposure to NO₂ was not associated with childhood asthma at age 9–11 years old²³. European BAMSE study also observed no significant association between PM₁₀ exposure and childhood asthma²⁴. Nevertheless, Clark *et al.* found the significant association between NO₂, PM₁₀, SO₂ exposure and Canadian childhood asthma²⁵.

Children exposure to air pollutants is vulnerable to suffering from pulmonary diseases, owing to a larger surface area of lung space than adult²⁶. Exposure to higher air pollutants levels may induce the pathologic changes in childhood airway mucous and make children prone to the prevalence of chronic inflammation in respiratory

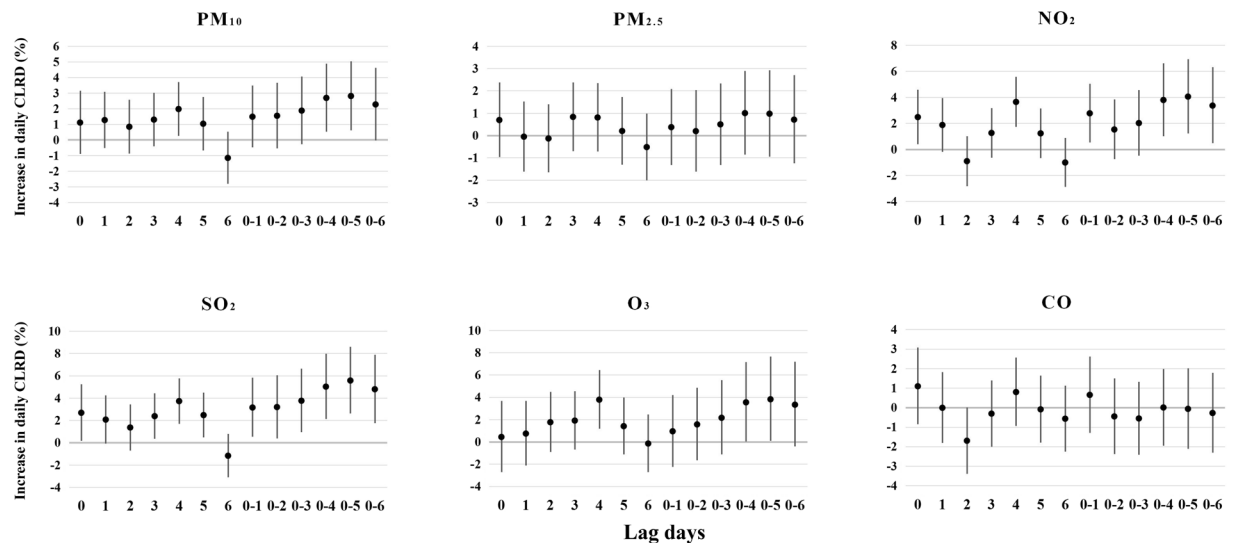


Figure 1. Estimated changes with 95% confidence intervals in daily CLRD (Childhood lower respiratory diseases) percentage deviations (%) associated with an interquartile range increase in PM₁₀, PM_{2.5}, NO₂, SO₂, O₃ or CO concentrations with different lag days in single-pollutant model.

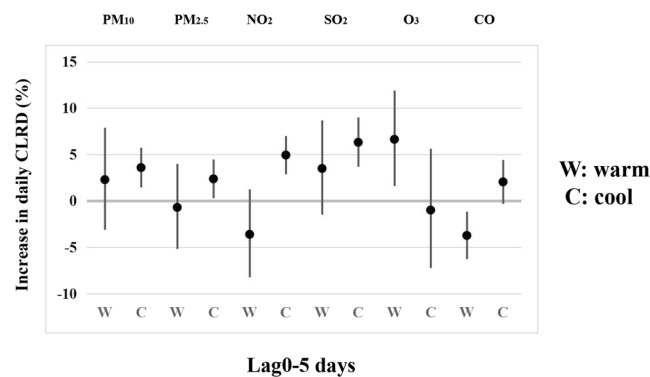


Figure 2. Estimated changes with 95% confidence intervals in daily CLRD percentage deviations (%) associated with an interquartile range increase in PM₁₀, PM_{2.5}, NO₂, SO₂, O₃ or CO concentrations (lag0–5 days) by season using the single-pollutant model.

system²⁷. Many studies have investigated the association between air pollutants and CLRD, of which the majority suggested the effects of short-term exposure to air pollutants up to 4 or 5 days^{28,29}. Similarly, we also found a cumulative effect of PM₁₀, NO₂, and SO₂ (lag0–5) on the short term effects with CLRD. Besides, we also investigated the effect estimates of PM₁₀, NO₂, and SO₂ on CLRD by stratification of diseases (bronchitis, pneumonia, capillary bronchitis, and asthma cases). Some differences observed between PM₁₀, NO₂, and SO₂ and diseases.

Exposure to PM₁₀ can induce the lung function disorder of children. In this study, we observed that an IQR increase in PM₁₀ daily concentrations increased the daily numbers of CLRD with 5, 6, or 7 days cumulative effects (lag0–4, lag0–5, lag0–6, respectively). Recently, a retrospective study also showed a positive association between the cumulative effects of 7 days (lag0–6) for PM₁₀ and CLRD¹⁵. Further, we found the significant association between PM₁₀ and pneumonia. A Jinan study revealed that PM₁₀ were associated with increased number of Childhood pneumonia³⁰. In 2010, Weinmayr and his colleagues conducted a systematic literature review to assess the effect of PM₁₀ on respiratory health in children, which had provided the clear evidence of PM₁₀ as the risk factor on childhood asthma symptom episodes³¹. However, we did not observe the similar association. It is possible that genetic background can contribute to the differences.

NO₂ exposure can damage the lung function, and induce the change of allergic airway inflammation, and responses to allergens³². Similar to PM₁₀, we found that exposure to NO₂ was significantly associated with CLRD. The estimated effect for NO₂ was consistent with many other studies^{33,34}. In Italy, Bono *et al.* found the adverse health effect of NO₂ on the risk of emergency room admissions for childhood respiratory diseases³⁴. A meta-analysis included seven countries also confirmed that exposure to NO₂ was significantly associated with the asthma and wheezing of children aged 0–18³⁵. In the present study, the significant effect estimate of NO₂ on childhood pneumonia and capillary bronchitis was more pronounced than other diseases (bronchitis, and asthma).

SO₂ is a well-known gas with strong irritant, and can induce the responses of respiratory system. A study showed that children living near a petrochemical complex with high SO₂ exposure had significantly increased incidence of bronchitis, asthma and allergic rhinitis than low exposure group³⁶. In addition, Samoli *et al.* also observed that exposure to SO₂ (10 µg/m³) significantly increased the number of childhood asthma hospital admissions³⁷. In this study, we also found that exposure to SO₂ had a positive association with the numbers of CLRD, especially with childhood bronchitis, pneumonia, and capillary bronchitis.

Season was an important factor in modifying the association between air pollutants and population health effects³⁸. In this study, we found the effects of air pollutants on CLRD numbers varied by season. Specifically, the effect estimates of PM₁₀, NO₂, and SO₂ were positive associated with CLRD numbers in cool season. Chen's findings also suggested that exposure to PM₁₀ and NO₂ had the largest effects in childhood asthma in cool season³⁹. A system review and meta-analysis also proposed that children exposure to PM₁₀, NO₂, and SO₂ have the stronger association of asthma-related hospital admission in warm season⁴⁰. The exact seasonal reasons were still not clear, and might be related with the difference in air pollution levels, surrounding environment, or races.

Several studies have found that age and sex can influence the effect estimates of air pollutants on population health. In the present study, we explored the effect estimates of air pollutants on CLRD in the subgroup of sex and age. Results showed that the effect estimates of PM₁₀, NO₂, and SO₂ in female were more pronounced than in male, which were consistent with previous study⁴¹. Similar results were observed in children with age >27 months. It was possible that children with age >27 months spent more time outdoors than age ≤27 months.

Herein, we should mention several limitations. Firstly, we obtained the data only from one city. Exposure to air pollutants effects may be influenced by person's living and other factors. Results were difficult to spread to other cities. Secondly, we did not obtain the exactly individual exposure levels of air pollutants, thus, we merely calculated the average air pollution levels of local region to evaluate the estimate effects of air pollutants. Thirdly, we collected the useful data based on one and a half years data from the seven hospitals, which may not represent the entire populations. Besides, our study is the retrospective study, we cannot obtain the biochemical indexes of CLRD cases. Further larger studies should be conducted to comprehensively clarify the effects of air pollutants on CLRD. This study also had several strengths. Firstly, it was the first time to investigate the effect estimate of air pollutants on CLRD in Nanjing, China. Secondly, we used the time-series regression and GAM analysis to evaluate the effects of air pollution. Thirdly, although we obtained the data in a city, the results were consisted with many other studies^{39, 40, 42}.

In summary, we found that the short-term exposure to ambient PM₁₀, NO₂, and SO₂ were associated with increased CLRD numbers. It was worth to note that our data provided the limited information; therefore, further studies should be conducted to validate our findings.

References

1. Beelen, R. *et al.* Effects of long-term exposure to air pollution on natural-cause mortality: an analysis of 22 European cohorts within the multicentre ESCAPE project. *Lancet*. **383**, 785–795, doi:10.1016/S0140-6736(13)62158-3 (2014).
2. Jassal, M. S. Pediatric asthma and ambient pollutant levels in industrializing nations. *International health*. **7**, 7–15, doi:10.1093/inthealth/ihu081 (2015).
3. Oftedal, B. *et al.* Residential outdoor air pollution and lung function in schoolchildren. *Epidemiology*. **19**, 129–137, doi:10.1097/EDE.0b013e31815c0827 (2008).
4. Leitte, A. M. *et al.* Size-segregated particle number concentrations and respiratory emergency room visits in Beijing, China. *Environmental health perspectives*. **119**, 508–513, doi:10.1289/ehp.1002203 (2011).
5. Sarnat, S. E. *et al.* Air pollution and acute respiratory response in a panel of asthmatic children along the U.S.-Mexico border. *Environmental health perspectives*. **120**, 437–444, doi:10.1289/ehp.1003169 (2012).
6. Gasana, J. *et al.* Motor vehicle air pollution and asthma in children: a meta-analysis. *Environmental research*. **117**, 36–45, doi:10.1016/j.envres.2012.05.001 (2012).
7. La Grutta, S. *et al.* Environmental risk factors and lung diseases in children: from guidelines to health effects. *Early human development*. **89**(Suppl 3), S59–62, doi:10.1016/j.earlhumdev.2013.07.025 (2013).
8. Zhang, J. J. *et al.* Children's respiratory morbidity prevalence in relation to air pollution in four Chinese cities. *Environmental health perspectives*. **110**, 961–967, doi:10.1289/ehp.02110961 (2002).
9. Zhao, Z. *et al.* Asthmatic symptoms among pupils in relation to winter indoor and outdoor air pollution in schools in Taiyuan, China. *Environmental health perspectives*. **116**, 90–97, doi:10.1289/ehp.10576 (2008).
10. Deng, Q. *et al.* Early life exposure to ambient air pollution and childhood asthma in China. *Environmental research*. **143**, 83–92, doi:10.1016/j.envres.2015.09.032 (2015).
11. Liu, W. *et al.* Associations of gestational and early life exposures to ambient air pollution with childhood respiratory diseases in Shanghai, China: A retrospective cohort study. *Environment international*. **92–93**, 284–293, doi:10.1016/j.envint.2016.04.019 (2016).
12. Molter, A. *et al.* A multicentre study of air pollution exposure and childhood asthma prevalence: the ESCAPE project. *The European respiratory journal*. **45**, 610–624, doi:10.1183/09031936.00083614 (2015).
13. Parker, J. D., Akinbami, L. J. & Woodruff, T. J. Air pollution and childhood respiratory allergies in the United States. *Environmental health perspectives*. **117**, 140–147, doi:10.1289/ehp.11497 (2009).
14. Lu, F. *et al.* Short-term effects of air pollution on daily mortality and years of life lost in Nanjing, China. *The Science of the total environment*. **536**, 123–129, doi:10.1016/j.scitotenv.2015.07.048 (2015).
15. Schvartzman, C., Pereira, L.A., Braga, A.L., Farhat, S.C. Seven-day cumulative effects of air pollutants increase respiratory ER visits up to threefold. *Pediatric pulmonology*. (2016).
16. Dominici, F. *et al.* Fine particulate air pollution and hospital admission for cardiovascular and respiratory diseases. *Jama*. **295**, 1127–1134, doi:10.1001/jama.295.10.1127 (2006).
17. Wang, X. *et al.* Associations between fine particle, coarse particle, black carbon and hospital visits in a Chinese city. *The Science of the total environment*. **458–460**, 1–6, doi:10.1016/j.scitotenv.2013.04.008 (2013).
18. Peng, R. D. *et al.* Seasonal analyses of air pollution and mortality in 100 US cities. *American journal of epidemiology*. **161**, 585–594, doi:10.1093/aje/kwi075 (2005).
19. Yang, Y. *et al.* The association between ambient air pollution and daily mortality in Beijing after the 2008 olympics: a time series study. *PLoS one*. **8**, e76759, doi:10.1371/journal.pone.0076759 (2013).
20. Delfino, R. J. *et al.* Repeated hospital encounters for asthma in children and exposure to traffic-related air pollution near the home. *Annals of allergy, asthma & immunology: official publication of the American College of Allergy, Asthma, & Immunology*. **102**, 138–144, doi:10.1016/S1081-1206(10)60244-X (2009).

21. Jerrett, M. *et al.* A review and evaluation of intraurban air pollution exposure models. *Journal of exposure analysis and environmental epidemiology*. **15**, 185–204, doi:10.1038/sj.jea.7500388 (2005).
22. Molter, A. *et al.* Effects of long-term exposure to PM₁₀ and NO₂ on asthma and wheeze in a prospective birth cohort. *Journal of epidemiology and community health*. **68**, 21–28, doi:10.1136/jech-2013-202681 (2014).
23. Oftedal, B., Nystad, W., Brunekreef, B. & Nafstad, P. Long-term traffic-related exposures and asthma onset in schoolchildren in oslo, norway. *Environmental health perspectives*. **117**, 839–844, doi:10.1289/ehp.11491 (2009).
24. Gruzieva, O. *et al.* Exposure to air pollution from traffic and childhood asthma until 12 years of age. *Epidemiology*. **24**, 54–61, doi:10.1097/EDE.0b013e318276c1ea (2013).
25. Clark, N. A. *et al.* Effect of early life exposure to air pollution on development of childhood asthma. *Environmental health perspectives*. **118**, 284–290, doi:10.1289/ehp.0900916 (2010).
26. Ng'ang'a, L. W. *et al.* Prevalence of exercise induced bronchospasm in Kenyan school children: an urban-rural comparison. *Thorax*. **53**, 919–926, doi:10.1136/thx.53.11.919 (1998).
27. Miller, M. D. *et al.* Differences between children and adults: implications for risk assessment at California EPA. *International journal of toxicology*. **21**, 403–418, doi:10.1080/10915810290096630 (2002).
28. Lee, S. L., Wong, W. H. & Lau, Y. L. Association between air pollution and asthma admission among children in Hong Kong. *Clinical and experimental allergy: journal of the British Society for Allergy and Clinical Immunology*. **36**, 1138–1146, doi:10.1111/j.1365-2222.2006.02555.x (2006).
29. Farhat, S. C. *et al.* Effect of air pollution on pediatric respiratory emergency room visits and hospital admissions. *Brazilian journal of medical and biological research = Revista brasileira de pesquisas medicas e biologicas/Sociedade Brasileira de Biofisica... [et al.]*. **38**, 227–235 (2005).
30. Lv, C. *et al.* The impact of airborne particulate matter on pediatric hospital admissions for pneumonia among Jinan children: a case-crossover study. *J Air Waste Manag Assoc* (2016).
31. Weinmayr, G. *et al.* Short-term effects of PM₁₀ and NO₂ on respiratory health among children with asthma or asthma-like symptoms: a systematic review and meta-analysis. *Environmental health perspectives*. **118**, 449–457, doi:10.1289/ehp.0900844 (2010).
32. Poynter, M. E. *et al.* Nitrogen dioxide enhances allergic airway inflammation and hyperresponsiveness in the mouse. *American journal of physiology. Lung cellular and molecular physiology*. **290**, L144–152, doi:10.1152/ajplung.00131.2005 (2006).
33. Hansel, N. N. *et al.* A longitudinal study of indoor nitrogen dioxide levels and respiratory symptoms in inner-city children with asthma. *Environmental health perspectives*. **116**, 1428–1432, doi:10.1289/ehp.11349 (2008).
34. Bono, R. *et al.* Air pollution, aeroallergens and admissions to pediatric emergency room for respiratory reasons in Turin, northwestern Italy. *BMC public health*. **16**, 722, doi:10.1186/s12889-016-3376-3 (2016).
35. Takenoue, Y. *et al.* Influence of outdoor NO₂ exposure on asthma in childhood: meta-analysis. *Pediatrics international: official journal of the Japan Pediatric Society*. **54**, 762–769 (2012).
36. Chiang, T. Y. *et al.* Increased incidence of allergic rhinitis, bronchitis and asthma, in children living near a petrochemical complex with SO₂ pollution. *Environment international*. **96**, 1–7, doi:10.1016/j.envint.2016.08.009 (2016).
37. Samoli, E. *et al.* Acute effects of air pollution on pediatric asthma exacerbation: evidence of association and effect modification. *Environmental research*. **111**, 418–424, doi:10.1016/j.envres.2011.01.014 (2011).
38. Yang, C. *et al.* Acute effect of ambient air pollution on heart failure in Guangzhou, China. *International journal of cardiology*. **177**, 436–441, doi:10.1016/j.ijcard.2014.09.003 (2014).
39. Chen, K. *et al.* The Effects of Air Pollution on Asthma Hospital Admissions in Adelaide, South Australia, 2003–2013: Time-Series and Case-crossover Analyses. *Clinical and experimental allergy: journal of the British Society for Allergy and Clinical Immunology*. (2016).
40. Zheng, X. Y. *et al.* Association between Air Pollutants and Asthma Emergency Room Visits and Hospital Admissions in Time Series Studies: A Systematic Review and Meta-Analysis. *PLoS one*. **10**, e0138146, doi:10.1371/journal.pone.0138146 (2015).
41. Strak, M. *et al.* Associations between lifestyle and air pollution exposure: Potential for confounding in large administrative data cohorts. *Environmental research*. **156**, 364–373, doi:10.1016/j.envres.2017.03.050 (2017).
42. Zhang, J. *et al.* Air Pollutants, Climate, and the Prevalence of Pediatric Asthma in Urban Areas of China. *BioMed research international*. **2016**, 2935163, doi:10.1155/2016/2935163 (2016).

Acknowledgements

This study was partly supported by grants from the Natural Science Foundation of Jiangsu Province (15KJB330002), Undergraduates Training Programs of Innovation and Entrepreneurship of Jiangsu Province (201510312001Z), The National Key Research and Development Program of China (2017YFC0211600), Collaborative Innovation Center for Cancer Personalized Medicine, and the Priority Academic Program Development of Jiangsu Higher Education Institutions (Public Health and Preventive Medicine).

Author Contributions

L.Z., X.G., Y.C., Z.Z., H.C., M.W. and conceived and designed the study. L.Z. and H.C. wrote the main manuscript. Y.C. and X.Z., W.P., X.Z., Q.Y., J.X., W.S., S.B. analyzed the data. D.W., Z.H., Z.Z. and H.C. analyzed and interpreted the data. All authors have read and approved the manuscript.

Additional Information

Supplementary information accompanies this paper at doi:10.1038/s41598-017-04310-7

Competing Interests: The authors declare that they have no competing interests.

Publisher's note: Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons license, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons license, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons license and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this license, visit <http://creativecommons.org/licenses/by/4.0/>.

© The Author(s) 2017