



# Micro-scale H<sub>2</sub>–CO<sub>2</sub> Dynamics in a Hydrogenotrophic Methanogenic Membrane Reactor

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Biogas production is a key factor in a sustainable energy supply. It is possible to get biogas with very high methane content if the biogas reactors are supplied with exogenous hydrogen, and one of the technologies for supplying hydrogen is through gas permeable membranes. In this study the activity and stratification of hydrogen consumption above such a membrane was investigated by use of microsensors for hydrogen and pH. A hydrogenotrophic methanogenic community that was able to consume the hydrogen flux within 0.5 mm of the membrane with specific rates of up to 30 m<sup>3</sup> H<sub>2</sub> m<sup>-3</sup> day<sup>-1</sup> developed within 3 days in fresh manure and was already established at time zero when analyzing slurry from a biogas plant. The hydrogen consumption was dependent on a simultaneous carbon dioxide supply and was inhibited when carbon dioxide depletion elevated the pH to 9.2. The activity was only partially restored when the carbon dioxide supply was resumed. Bioreactors supplied with hydrogen gas should thus be carefully monitored and either have the hydrogen supply disrupted or be supplemented with carbon dioxide when the pH rises to values about 9.

**Keywords:** hydrogen, microsensor, methane, biogas, membrane reactor, methanogenesis, pH, CO<sub>2</sub>

## INTRODUCTION

The increasing concern about the rise in atmospheric CO<sub>2</sub> is driving intense research into alternative energy sources. The biological production of methane in bioreactors reduces the organic waste from several agricultural or industrial facilities and the product is a gas that may be used as fuel in combustion engines, for power generation or transport, and in the longer term as a fossil free raw material for chemical synthesis. However, the biogas produced in an anaerobic reactor is typically composed of only 50–70% of CH<sub>4</sub>, 30–50% CO<sub>2</sub>, and minor amounts of other gasses (Ghorbanian et al., 2014). The high CO<sub>2</sub> concentration reduces the quality of the raw biogas and makes it unsuitable for transport purposes. To reduce the amount of CO<sub>2</sub> in the produced gas and at the same time increase the CH<sub>4</sub> yield, an injection of H<sub>2</sub> into the reactor has been proposed (Luo and Angelidaki, 2012). The injection of H<sub>2</sub> would provide the electron donor needed to reduce more CO<sub>2</sub> to CH<sub>4</sub> by the hydrogenotrophic methanogenic community. This could upgrade biogas to a quality approaching natural gas (Luo and Angelidaki, 2013; Wang et al., 2013; Bensmann et al., 2014; Ghorbanian et al., 2014).

Renewable energy sources such as wind or sunlight are subject to weather conditions. As the supply from renewable sources and the demand fluctuates, the market price of electricity changes hourly at the established power trading platforms, and in periods with much wind the prices may be very low or even negative<sup>1</sup>. One method to efficiently utilize renewable energy in periods with low cost is to produce H<sub>2</sub> by electrolysis (Sherif et al., 2005). However, the direct use of H<sub>2</sub> as fuel or in fuel cells has several disadvantages due to complex transportation and storage (Luo et al., 2012). Hydrogen gas may, however, be used for production of methane by microbiological or chemical means (Stempien et al., 2015), as CH<sub>4</sub> has higher volumetric energy content and the infrastructure is already available for its use. The use of H<sub>2</sub>, produced by water electrolysis, for conversion to CH<sub>4</sub> enables the indirect accumulation of wind or solar power as a storable gas. When the conversion is done by microbes in a biogas plant it furthermore utilizes a readily available source of biologically formed CO<sub>2</sub>. Chemical conversion would, on the other hand, require a concentrated source of CO<sub>2</sub> obtained from for instance combustion of biomass, fossil fuels, or cement production.

Injection of H<sub>2</sub> into a working biogas reactor does not necessarily imply substantial structural modifications and the essential microbial community of hydrogenotrophic methanogens is generally well developed in biogas reactors (Demirel and Scherer, 2008). It could therefore be expected that an implementation of H<sub>2</sub> addition to biogas reactors should be straightforward. However, a number of experiments in small scale reactors have revealed several lacks in our knowledge about the complex interaction of microbial, physical, and chemical processes occurring in the reactors, resulting in an inefficient conversion of the injected H<sub>2</sub>. One critical limitation of the process is a low solubility of H<sub>2</sub>, resulting in a low rate of gas/liquid transfer by bubbling (Luo et al., 2012; Bensmann et al., 2014; Ghorbanian et al., 2014). The use of permeable membranes has been proven to be more efficient, resulting in the full consumption of the injected H<sub>2</sub> and a higher CH<sub>4</sub> content of the produced biogas (Wang et al., 2013; Díaz et al., 2015). Another limiting factor is that the consumption of CO<sub>2</sub> due to methanogenesis may cause an increase in pH to values higher than 9, resulting in the destabilization of the whole reactor (Luo et al., 2012; Xu et al., 2014; Hu et al., 2015). Addition of acidic substrates to the reactor to balance the pH has therefore been applied to maintain high CH<sub>4</sub> production rates (Luo and Angelidaki, 2013; Wang et al., 2013; Ghorbanian et al., 2014). An alternative would be to reduce the H<sub>2</sub> input to levels not resulting in almost complete CO<sub>2</sub> depletion and associated high pH.

The gas permeable membrane reactor has been shown to be an efficient design for the bio-methanization of H<sub>2</sub> (Wang et al., 2013; Díaz et al., 2015). The microbial transformation is expected to occur in a biofilm growing on the membrane, being supplied with H<sub>2</sub> by diffusion through the membrane. By use of microsensors it is possible to measure relevant chemical parameters such as H<sub>2</sub>, H<sub>2</sub>S, and CH<sub>4</sub> in such biofilms (Revsbech, 2005) with high temporal and spatial resolution. The interference of H<sub>2</sub>S on the signal from electrochemical H<sub>2</sub>-sensors could have

prevented a microscale study of H<sub>2</sub> transformations in this kind of biofilm, but recently a H<sub>2</sub> microsensor with a H<sub>2</sub>S trap has been developed (Nielsen et al., 2015).

The aim of this study was to investigate how hydrogenotrophic methanogenesis is regulated at a microscale when H<sub>2</sub> is supplied to biogas reactors. We studied an anaerobic membrane reactor where H<sub>2</sub> was supplied through a gas permeable membrane, but the results may also illustrate the microenvironment and metabolic rates around H<sub>2</sub> bubbles if H<sub>2</sub> is supplied by bubbling. Hydrogen and pH microprofiles were measured in the microbial community overlying a silicone membrane and the spatial microdistribution of processes was calculated from the profiles. The time course development of microprofiles and activity was investigated in both fresh cattle manure at 20°C and in biomass from a mesophilic biogas reactor at 38°C. The effects of various rates of CO<sub>2</sub> supply on the microdistribution of H<sub>2</sub> and pH and on the rate of H<sub>2</sub> consumption were examined in detail.

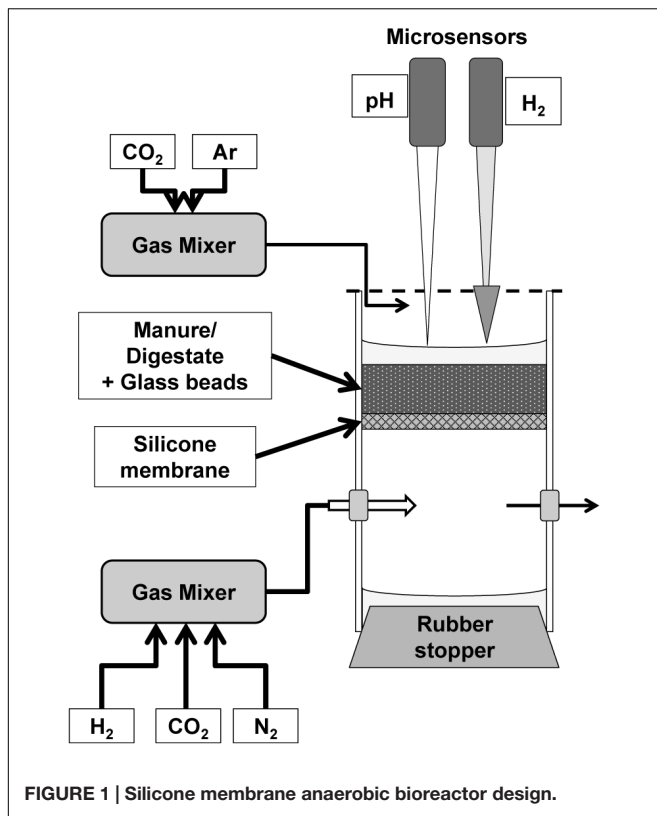
## MATERIALS AND METHODS

### Experimental Set-up: Bioreactor and Microsensor Measurements

A small scale silicone membrane reactor similar to the one described by Revsbech (1989) was used to provide a controllable environment in terms of H<sub>2</sub> supply (Figure 1). This design deviates substantially from the silicone-tubing membrane reactors used by Wang et al. (2013) and Díaz et al. (2015) as it was decided to exploit a geometry that was possible to analyze with microsensors and allowed reliable mathematical analysis. The reactor consisted of two cylindrical chambers (3 and 6 cm length) consisting of Plexiglas (i.d. 4 cm) separated by a silicone rubber membrane. The membrane was made by coating a stainless steel mesh (mesh size 1 mm) with silicone rubber (Silastic RT adhesive 732, Down coring, Midland, MI, USA), resulting in a total thickness of 1.5 mm. The membrane provided a stable base for the substrate while simultaneously being highly permeable to gases (Robb, 1968). The chamber below the membrane was used as a reservoir of gases diffusing up into the upper compartment. The bottom was closed by a rubber stopper and two holes in the sides (8 mm), closed with rubber stoppers, allowed the continuous flushing with a humidified gas flow of 100 mL/min through hypodermic needles of the chamber with a gas mixture controlled by a gas mixer (Brooks Instruments D.V., Holland). To ensure anaerobic conditions in the upper chamber, the upper opening, except for a small opening for microsensor insertion, was closed with Parafilm while it was continuously flushed with humidified Argon or an Ar:CO<sub>2</sub> mixture (~50–100 mL/min). Maintaining a higher density of the flushing gas than the surrounding air was essential to avoid turbulent mixing of atmospheric air into the chamber. The anoxic conditions were confirmed by an oxygen microsensor with the tip placed in the upper headspace.

A recently improved H<sub>2</sub> microsensor (Nielsen et al., 2015) was used to measure the distribution of H<sub>2</sub> diffusing from the silicone membrane. The new sensor has no interference from H<sub>2</sub>S, which can be present at relatively high concentrations (up to mM range) in both raw and digested cattle manure. The size

<sup>1</sup><http://www.nordpoolspot.com/historical-market-data>



of the tip (70–100  $\mu\text{m}$ ) allowed the measurement of profiles with a high spatial resolution of 100  $\mu\text{m}$ . The relatively long diffusion path from the exterior to the internal anode results in a negligible effect of variations in turbulence and diffusive properties in the analyzed medium. The H<sub>2</sub> sensors applied thus exhibited a signal difference between vigorously stirred and stagnant medium of <3% but a response time of about 30 s for 90%. pH microprofiles were also measured in some of the experiments. The pH microsensors had a pH-sensitive tip length of 150  $\mu\text{m}$ , allowing for a similar vertical resolution as the H<sub>2</sub> measurements. Hydrogen gas and pH were measured simultaneously in some of the experiments (see below) by the use of one sensor of each kind, measuring in the same reactor. The H<sub>2</sub> microsensor was connected to a pico-ammeter (PA2000, Unisense) whereas the pH microelectrode was connected to a custom-built mV-meter (Aarhus University) and to an external Ag/AgCl reference electrode positioned inside the chamber. Microsensors were mounted together in a computer-controlled micromanipulator (MC-232, Unisense) while the signals were recorded through an A/D converter (ADC-216, Unisense) using the Sensor TracePro software (Unisense).

Profiles of H<sub>2</sub> were modeled, using the numerical procedure described by Berg et al. (1998). The numerical fitting of the measured H<sub>2</sub> generate profiles with the spatial distribution of the H<sub>2</sub> consumption rates, being expressed as  $\text{nmol cm}^{-3} \text{s}^{-1}$ . The obtained data were also depth-integrated to get the total rates being expressed as  $\mu\text{mol cm}^{-2} \text{h}^{-1}$ . The diffusion coefficient in water was taken from

the diffusion coefficient tables compiled by Unisense<sup>2</sup>. The solubility of H<sub>2</sub> in water was calculated (Crozier and Yamamoto, 1974) to be 778 and 728  $\mu\text{mol L}^{-1}$  at 22 and 38°C, respectively. The effective diffusion coefficient in the sieved fresh manure was assumed to be equal to the one in water. The effective diffusion coefficient in the glass beads biofilm was calculated according to Revsbech (1989) being 0.63 times the value in the water.

## Development of an Active H<sub>2</sub> Consuming Community in Fresh Cattle Manure

The initial activity and development of an active H<sub>2</sub> consuming community was followed in fresh cow manure at room temperature (22°C). Cow manure was collected in a dairy cattle farm near Aarhus, Denmark, brought to the laboratory and sieved applying a mesh-size of 2 mm, to remove large particles that could damage the sensor. The manure was then added to the bioreactor described previously to get a layer of about 15 mm on top of the silicone membrane, ensuring the complete anoxic conditions at the bottom. The bottom chamber of the reactor was flushed with H<sub>2</sub> (100%) while the upper chamber was flushed with Ar. The system was left to equilibrate for 2 h after which H<sub>2</sub> microprofiles were measured every 5 h until a steady state was obtained as judged from several consecutive profiles.

The evolution of the H<sub>2</sub> consumption rates showed the same pattern as bacterial growth and therefore, the rates were fitted to the Gompertz equation modified by Zwietering et al. (1990), adapted to the measured rates:

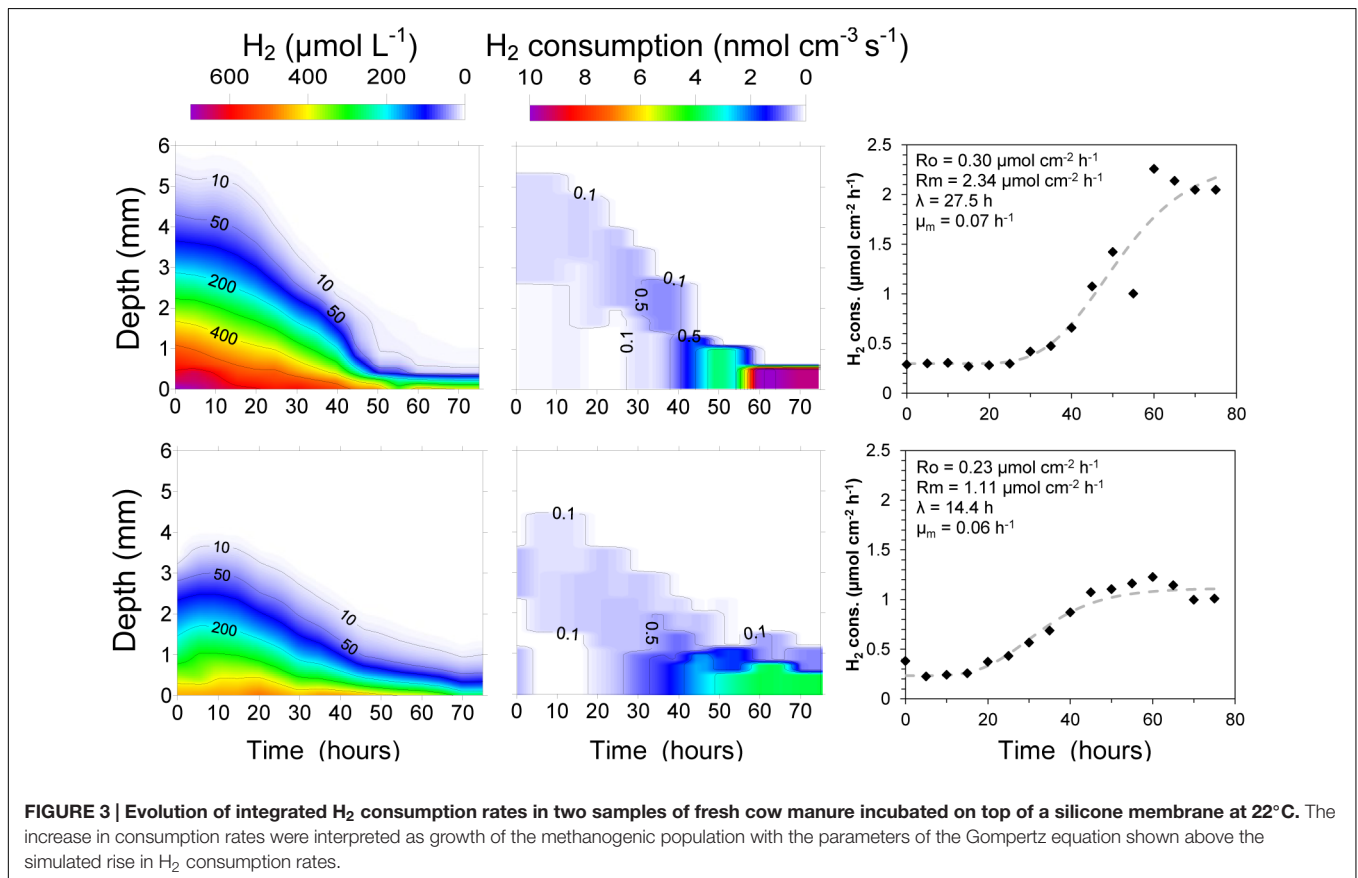
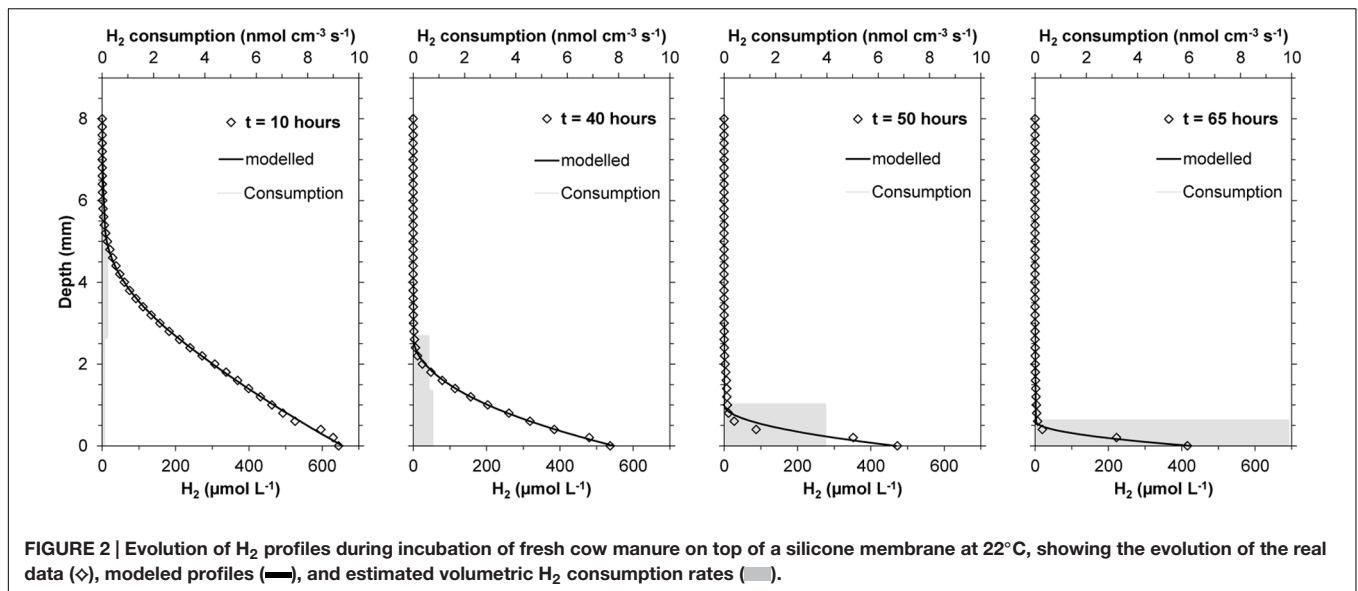
$$\ln\left(\frac{R}{R_0}\right) = R_m \times \exp\left[-\exp\left[\left(\frac{\mu_m \times e}{A}\right) \times (\lambda - t) + 1\right]\right] \quad (1)$$

Where:  $R$  is the H<sub>2</sub> consumption rate,  $R_0$  is the initial rate,  $R_m$  is the maximum rate,  $\mu_m$  is the maximum specific growth rate ( $\text{h}^{-1}$ ),  $\lambda$  is the lag time (h) and  $A$  is the asymptote, being equal to  $\ln(R_m/R_0)$ . The model was fitted to the experimental data by minimum least squares fitting, using the Solver command in Excel (Kemmer and Keller, 2010).

## Activity and CO<sub>2</sub> Dependence of H<sub>2</sub> Consumption Rates in the Digested Content of a Biogas Reactor

The development of an active H<sub>2</sub> consuming sludge matrix and the dependence on rates of CO<sub>2</sub> supply were investigated in sludge from a mesophilic anaerobic biogas reactor (hereafter referred as digestate) that was sieved through a 2 mm screen and mixed with glass beads (40–60  $\mu\text{m}$  diameter, 3 M to provide a stable matrix for profile analysis. The digested manure is a liquid and turbulent mixing due to convection would make the results difficult to interpret. The mixture of digestate and glass beads was added to the bioreactor (the bottom chamber being flushed with 100% H<sub>2</sub>) and the glass beads were allowed

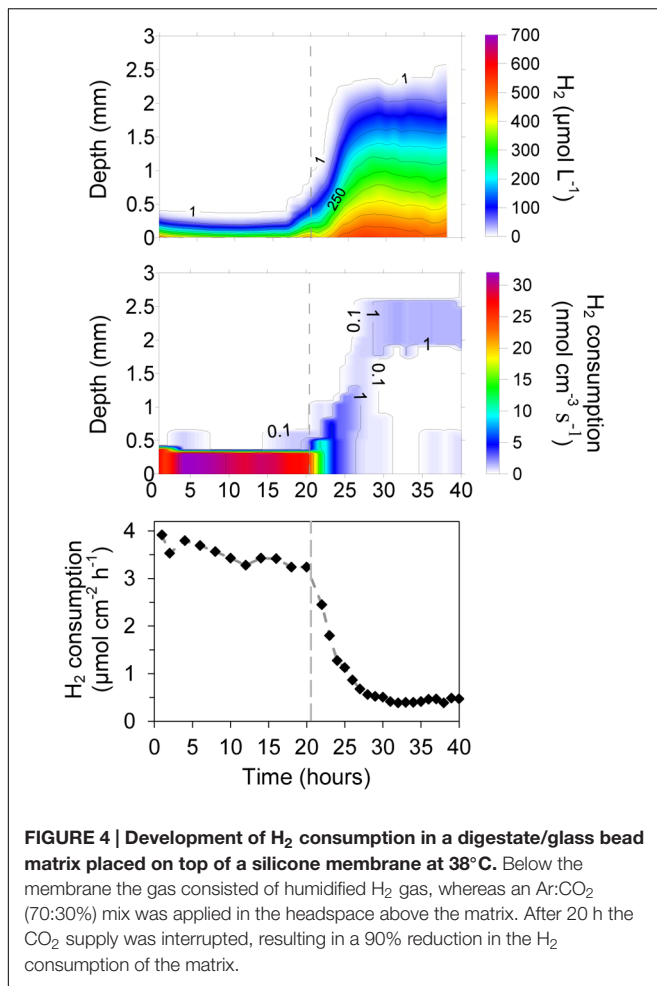
<sup>2</sup><http://www.unisense.com/files/PDF/Diverse/Seawater%20&%20Gases%20table.pdf>



to settle down for 10 min. The supernatant was then removed with a pipette to form a 4–5 mm active layer above the membrane and the upper chamber closed with Parafilm. The thickness of the layer was reduced in order to reduce the capacity for CO<sub>2</sub> production of the system and then be more dependent of the external gasses manipulations. The bioreactor

was then immersed in a temperature-controlled water bath at 38°C. A water saturated gas flow controlled by a gas mixer enabled changes in gas composition in both upper and lower chambers.

Two experiments with different types of gas supply manipulations were carried out:



### Development of the Community and Time Course after CO<sub>2</sub> Depletion

To follow the development of an active H<sub>2</sub> consuming community in the digestate, the bottom chamber was supplied with 100% H<sub>2</sub> while the upper one was supplied with a mixture of Ar:CO<sub>2</sub> (70:30%) to avoid CO<sub>2</sub> limitation. After 1 h of equilibration, H<sub>2</sub> microprofiles were measured every 2 h. After 20 h, the CO<sub>2</sub> supply was terminated while keeping the same measuring frequency for another 20 h.

### Dependence of H<sub>2</sub> Consumption of CO<sub>2</sub> Supply

The bottom compartment was supplied with a mixture of H<sub>2</sub>:N<sub>2</sub> (71:29%) while the upper one was supplied with 100% Ar during the whole experiment. After 1 h for equilibration, H<sub>2</sub> and pH microprofiles were measured hourly. Once stable conditions were reached (by measuring two consecutive profiles with apparently identical H<sub>2</sub> distributions), the CO<sub>2</sub> proportion in the gas supply of the bottom chamber was consecutively increased to the following values: 3, 6, 12, 14, 21, and 29% (reducing the N<sub>2</sub> but keeping constant the H<sub>2</sub> percentage). Finally, the flushing was returned to the initial H<sub>2</sub>:N<sub>2</sub> mix and the system was followed until steady-state profiles were approached.

## RESULTS

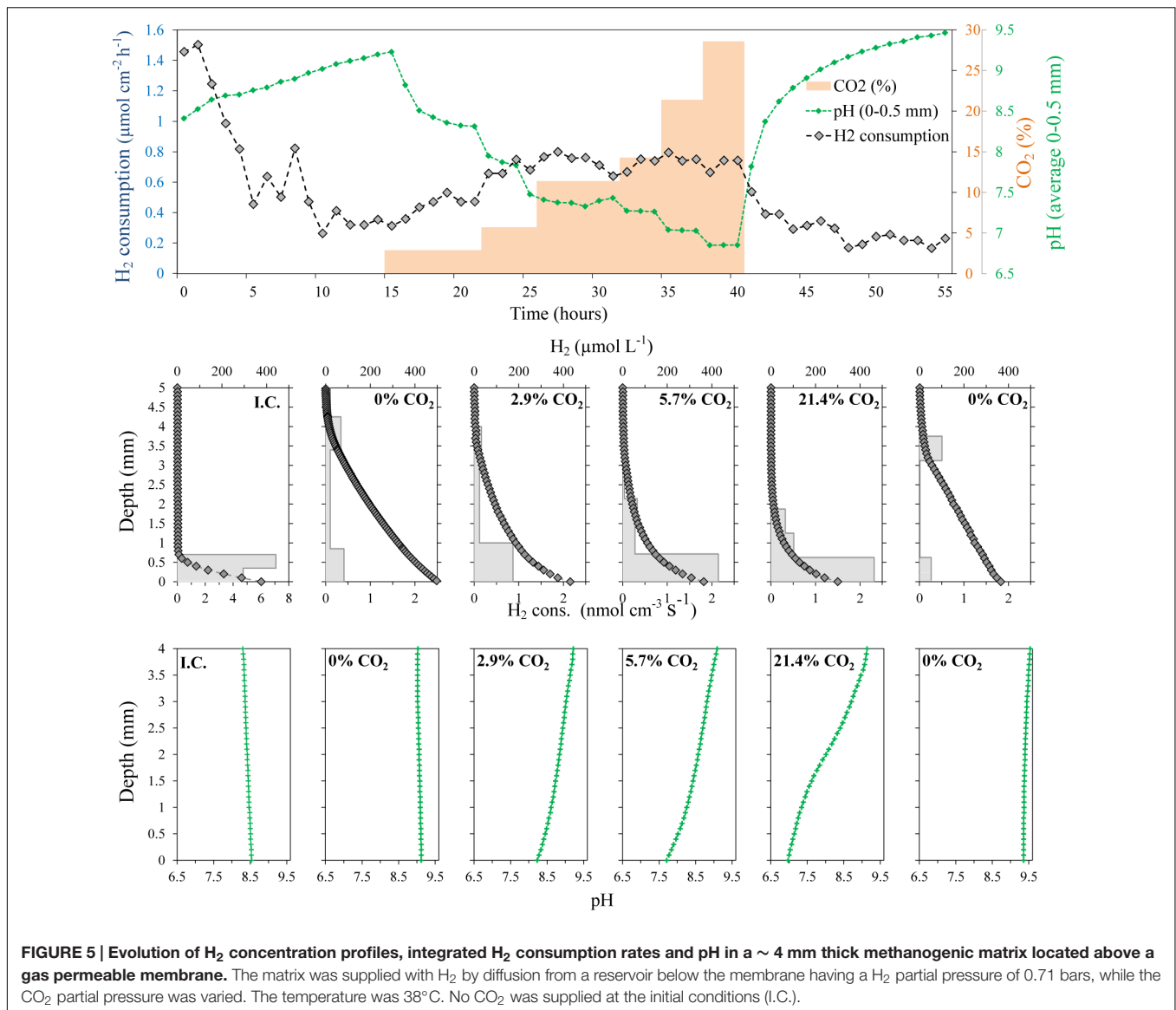
### Development of an Active H<sub>2</sub> Consuming Community in Cattle Manure

An initially large H<sub>2</sub> penetration in the cattle manure suggested relatively low activity of the microbial community already present in the manure (Figures 2 and 3). H<sub>2</sub> microprofiles were similar during the first 15–25 h in both experiments conducted (Figure 3) showing H<sub>2</sub> diffusion from the silicone membrane into the first 3–4 mm of the manure. During this period, H<sub>2</sub> consumption rates were low (<0.25 nmol cm<sup>-3</sup> s<sup>-1</sup>, equivalent to 0.55 m<sup>3</sup> H<sub>2</sub> m<sup>-3</sup> manure d<sup>-1</sup>) and extended along the 3–4 mm where H<sub>2</sub> could be measured. The integrated H<sub>2</sub> consumption rate was 0.22–0.30 μmol cm<sup>-2</sup> h<sup>-1</sup> (1.4–1.8 L m<sup>-2</sup> d<sup>-1</sup>) (Figure 3) during this initial period. After this lag period, lasting for 14.4 to 27.5 h according to the Gompertz model, the H<sub>2</sub> profiles changed quickly, resulting in a progressive decrease in H<sub>2</sub> penetration and increase in rates of consumption (Figures 2 and 3). The H<sub>2</sub> consumption was finally concentrated in a thin layer of less than 0.5 mm, with maximum rates of 5.7–9.5 nmol cm<sup>-3</sup> s<sup>-1</sup> (corresponding to 12.5–20.9 m<sup>3</sup> H<sub>2</sub> m<sup>-3</sup> manure d<sup>-1</sup>). Assuming that the H<sub>2</sub> consumption rates were proportional to the populations density of methanogens it can be calculated from the Gompertz equation that the maximum specific growth rates were 0.06–0.07 h<sup>-1</sup>. After 60 h, integrated consumption rates reached stable values of 1.11–2.05 μmol cm<sup>-2</sup> h<sup>-1</sup>, corresponding to 6.8–12.5 L m<sup>-2</sup> d<sup>-1</sup> (Figures 2 and 3).

### Activity and CO<sub>2</sub> Dependence of H<sub>2</sub> Consumption Rates in the Slurry of a Biogas Reactor

The development of an active H<sub>2</sub> consuming community was followed using digestate from a mesophilic anaerobic reactor. Contrary to the experiments with fresh cattle manure, the H<sub>2</sub> profiles showed a high rate of consumption already at the time of the first profile measurement 1 h after the set-up of the system (Figure 4). The H<sub>2</sub> diffusing out of the membrane was quickly consumed in a layer thinner than 0.5 mm with rates varying from 24.5 to 31.9 nmol cm<sup>-3</sup> s<sup>-1</sup> (53.9–70.2 m<sup>3</sup> H<sub>2</sub> m<sup>-3</sup> digestate d<sup>-1</sup>). The integrated rates were also relatively constant with values of 3.24–3.92 μmol cm<sup>-2</sup> h<sup>-1</sup> (19.8–24.0 L m<sup>-2</sup> d<sup>-1</sup>) (Figure 4).

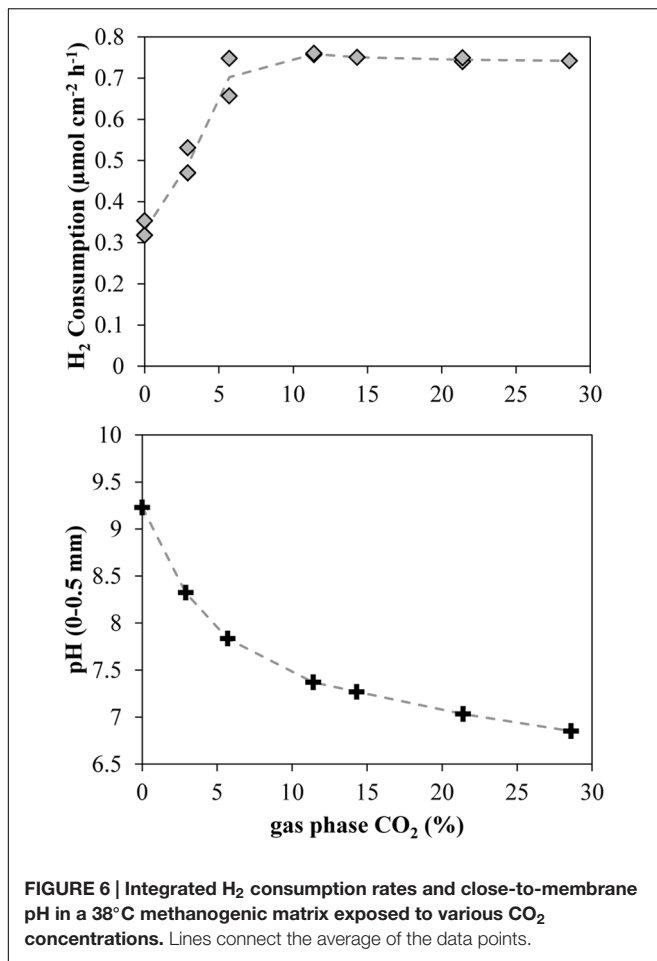
After 20 h of continuous H<sub>2</sub> supply through the silicone membrane, and CO<sub>2</sub> supply from the upper headspace (Ar:CO<sub>2</sub> 70:30%), the CO<sub>2</sub> supply was stopped, resulting in a rapid decrease in H<sub>2</sub> consumption rates of the microbial community (Figure 4). The penetration of H<sub>2</sub> into the matrix thus increased from a layer thinner than 0.5 mm to more than 2.5 mm. After 6–7 h, the H<sub>2</sub> consumption seemed to be localized in the upper part of the matrix with maximum rates around 1.6 nmol cm<sup>-3</sup> s<sup>-1</sup> (3.5 m<sup>3</sup> H<sub>2</sub> m<sup>-3</sup> digestate d<sup>-1</sup>). The integrated rates were also stable, showing low consumption values of 0.39–0.47 μmol cm<sup>-2</sup> h<sup>-1</sup> (2.4–2.9 L m<sup>-2</sup> d<sup>-1</sup>) (Figure 4). These rates may, however, be overestimates as the consumption determined from the change in profile slope in the upper part of the matrix may be



an artifact due to measurements in a meniscus at the transition from liquid to gas (Figures 4 and 5).

Due to the clear effect of CO<sub>2</sub> supply on the activity of the digestate in terms of H<sub>2</sub> consumption, the effect of the CO<sub>2</sub> concentration on the measured rates was also investigated (Figure 5). To get dependence from the external CO<sub>2</sub> supply instead of the internal CO<sub>2</sub> stock, the system was initially supplied with a gas mixture of 71% H<sub>2</sub>:29% N<sub>2</sub> to deplete the inorganic carbon in the system. Similar to the previous experiment with CO<sub>2</sub> starvation (Figure 4), the H<sub>2</sub> consumption rates decreased progressively, reaching relative stable and low values around 0.35 μmol cm<sup>-2</sup> h<sup>-1</sup> (2.1 L m<sup>-2</sup> d<sup>-1</sup>) after 5–6 h (Figure 5). The CO<sub>2</sub> starvation also resulted in substantial changes in the pH values throughout all depths. The pH thus increased progressively from about 8.5 to 9.1 throughout the digestate (Figure 5).

A progressive increase in CO<sub>2</sub> concentration in the gas supplied to the chamber below the silicone membrane (keeping H<sub>2</sub> at 71%) resulted in a progressive increase in the H<sub>2</sub> consumption rates. After a period of several hours without CO<sub>2</sub>, an increase of CO<sub>2</sub> to 2.9% produced a clear increase in the H<sub>2</sub> consumption rates, reaching a stable value of 0.5 μmol cm<sup>-2</sup> h<sup>-1</sup> (3.1 L m<sup>-2</sup> d<sup>-1</sup>) after 5 h (Figure 5). The next increase in CO<sub>2</sub> % to values of 5.7% also produced a clear increase in the H<sub>2</sub> consumption rates to a value of 0.75 μmol cm<sup>-2</sup> h<sup>-1</sup> (4.6 L m<sup>-2</sup> d<sup>-1</sup>). However, further increases did not produce any clear increase in H<sub>2</sub> consumption rates and a slight decrease could even be observed at the highest concentrations (Figures 5 and 6). The spatial distribution of the H<sub>2</sub> consumption was also modified. The H<sub>2</sub> consumption at the initial conditions (Figure 5) was mainly localized in a narrow layer of ca. 0.5 mm with a high activity of 5.9 nmol cm<sup>-3</sup> s<sup>-1</sup> (13.0 m<sup>3</sup> m<sup>-3</sup> d<sup>-1</sup>). However, the depletion and subsequent resumed CO<sub>2</sub> supply resulted in



lower rates (maximum of  $2.63 \text{ nmol cm}^{-3} \text{ s}^{-1}$  at 21.4% of CO<sub>2</sub>) distributed in a broader area of 2–3 mm (Figure 5).

An increase of the CO<sub>2</sub> supply produced a substantial pH decrease in the layers closest to the membrane (Figures 5 and 6). Just 2.9% CO<sub>2</sub> thus reduced pH close to the membrane from a value about 9.2 during CO<sub>2</sub> depletion to 8.2. At 5.7% CO<sub>2</sub> the value further decreased to 7.7 and at 21.4% it reached a pH of 7. The pH in the upper parts of the matrix stayed at a relatively high level during the increase in CO<sub>2</sub> supply from below, due to the flushing of the headspace with pure argon.

## DISCUSSION

### Vertical Micro-structure of a Hydrogenotrophic Biofilm (Conceptual Model)

In a standard methanogenic reactor, the decomposition of the various organic compounds results in acetogenesis with acetate (CH<sub>3</sub>COO<sup>-</sup>) and CO<sub>2</sub>+H<sub>2</sub> being the final products of fermentation processes. These compounds are then converted to methane by the acetoclastic and hydrogenotrophic methanogenesis, respectively (Figure 7). Hydrogenotrophic

methanogenesis is highly energetically favorable and the concentration of H<sub>2</sub> is usually maintained at very low values allowing for the syntrophic acetogenesis (Schink, 1988). However, the use of acetate is less energetically favorable and acetate is thus generally accumulating to higher concentrations. At low H<sub>2</sub> concentrations, the syntrophic acetate oxidation can also occur, converting acetate to CO<sub>2</sub>+H<sub>2</sub>, which then can be transformed into methane.

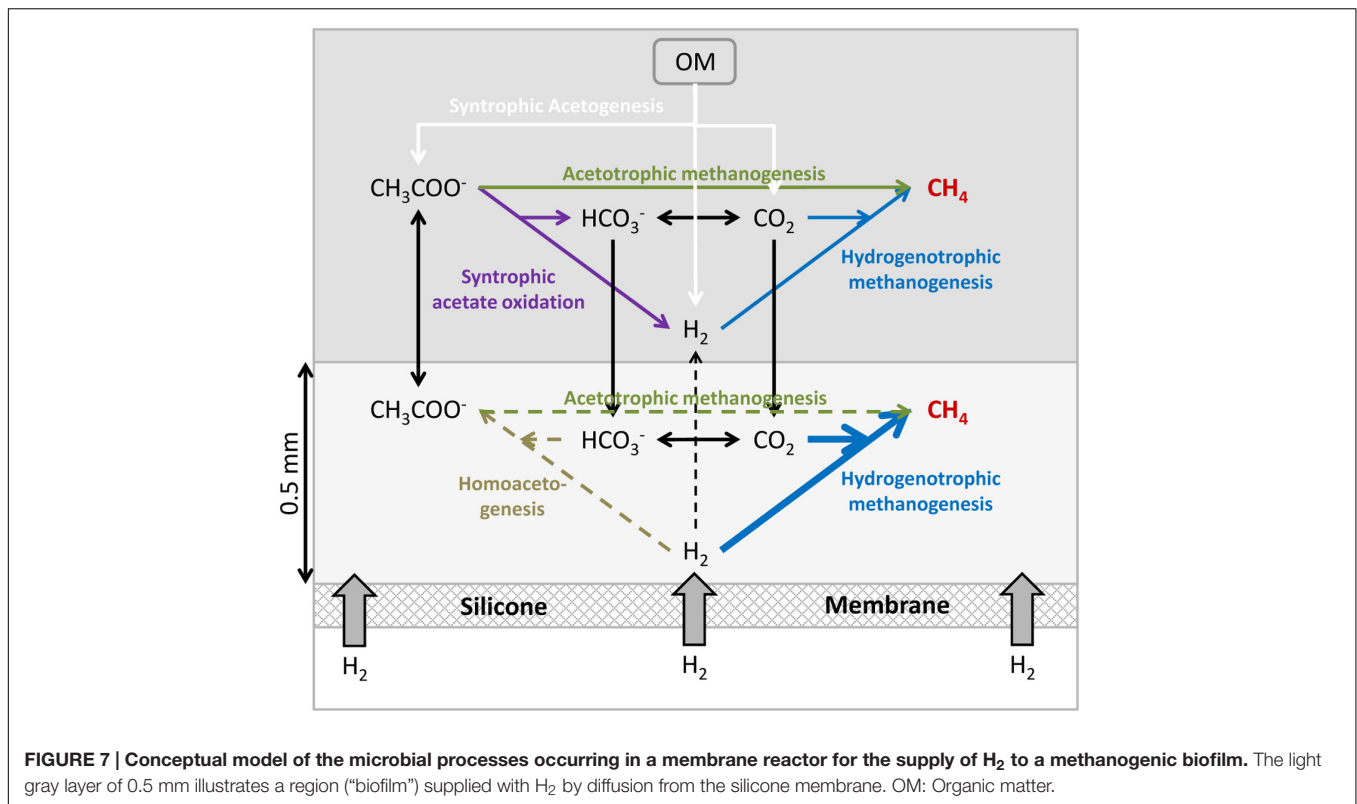
The localized supply of H<sub>2</sub> into an anaerobic reactor through a membrane is to a major extent modifying the balances and reactions taking place in the nearby microbial community. The high concentrations of H<sub>2</sub> close to the membrane inhibit syntrophic acetogenesis and syntrophic acetate oxidation, as these processes are no longer energetically favorable (Demirel and Scherer, 2008). Homoacetogenesis producing acetate from H<sub>2</sub> and CO<sub>2</sub> and hydrogenotrophic methanogenesis should then be dominant processes and consume all the H<sub>2</sub> supplied through the membrane. From Figures 2–5 it appears that complete depletion may occur over a distance as small as 0.5 mm. Among these two processes, the hydrogenotrophic methanogenesis is more energetically favorable and it has been shown to be the dominant process in reactors feed with H<sub>2</sub>, outcompeting homoacetogenesis (Díaz et al., 2015). The H<sub>2</sub> consumption processes both use CO<sub>2</sub> resulting in an increase of pH (Figure 5), which could lead to inhibition of the processes.

The processes of syntrophic acetogenesis and homoacetogenic oxidation can take place in the bulk reactor outside the zone of H<sub>2</sub> supply from the membrane. Both processes produce CO<sub>2</sub> and H<sub>2</sub> which can feed the CH<sub>4</sub> production by hydrogenotrophic methanogens. Acetogenesis and homoacetogenesis in the bulk reactor create good conditions for acetoclastic methanogenesis (Demirel and Scherer, 2008), producing CO<sub>2</sub> and thus contributing to the CO<sub>2</sub> supply of the high hydrogenotrophic activity near the membrane.

The structure and control of the activity near the membrane is thus regulated by the supply of H<sub>2</sub> and CO<sub>2</sub>. The H<sub>2</sub> supply to the zone must be balanced with the CO<sub>2</sub> supply from local production and diffusion from nearby areas. An excess of H<sub>2</sub> supply as compared to the supply of CO<sub>2</sub> leads to the increase in pH which results in the inhibition of several microbial processes and an instability of the system (Wang et al., 2013; Bensmann et al., 2014). A control of H<sub>2</sub> supply and the maintenance of moderate pH values around 8 have been demonstrated to mediate an almost complete conversion of CO<sub>2</sub> to CH<sub>4</sub> in membrane reactors (98–99% CH<sub>4</sub> in the produced biogas) with stable microbial communities, without affecting the microbial degradation of organic matter (Wang et al., 2013), inferring that most of the reactor was still characterized by very low H<sub>2</sub> concentrations allowing for syntrophic acetogenesis (Wang et al., 2013). However, an additional external CO<sub>2</sub> supply could allow for higher amounts of methane being produced.

### Development of Hydrogenotrophic Community at Low Temperature

The use of small-scale biogas reactors is an increasing practice in livestock farms from developing countries (Pham et al., 2014).



**FIGURE 7 | Conceptual model of the microbial processes occurring in a membrane reactor for the supply of H<sub>2</sub> to a methanogenic biofilm.** The light gray layer of 0.5 mm illustrates a region ("biofilm") supplied with H<sub>2</sub> by diffusion from the silicone membrane. OM: Organic matter.

The digestion of animal manure at relatively low temperatures (25°C) has been proven to be enough for a modest biogas production (<1 m<sup>3</sup> biogas/Kg volatile solids). However, the efficiency of the process is relatively low and start-up times longer than 1 month are needed (Zeeman et al., 1988; Alkaya et al., 2010; Pham et al., 2014). The injection of H<sub>2</sub> to untreated wet manure is, however, a different scenario. The methanogenic community is already present in the manure (Zeeman et al., 1988; Demirel and Scherer, 2008) and an injection of H<sub>2</sub> will boost the growth of the community. In our small reactor, we measured a H<sub>2</sub> consumption rate of 20 m<sup>3</sup> H<sub>2</sub> m<sup>-3</sup> manure day<sup>-1</sup> (Figure 3) after 3 days. The H<sub>2</sub> consumption may, however, not result in a 100% conversion to CH<sub>4</sub>. A large fraction (up to 70%) of the H<sub>2</sub> consumption has been reported to be used for growth of the methanogens and not for methane production during the start-up phase of a reactor (Díaz et al., 2015). The only about 60 h needed to obtain maximum activity in the raw manure (Figure 2) is a very short time period as compared to the start-up times needed for cattle manure reactors (Zeeman et al., 1988). Considering the integrated consumption rates as a proxy for the size of the microbial community, maximum values of 0.05–0.08 h<sup>-1</sup>, corresponding to doubling times of 9–14 h of the hydrogenotrophic microbial community, were measured (Figure 3). These values are in the range of the optimal doubling times of several species of mesophilic methanogens (Demirel and Scherer, 2008). Our data illustrate that the growth rate of the methanogenic community in raw cattle manure is high and the long start up times for methanogenic

digestors fed with cattle manure thus cannot be due to the methanogens.

The zone with measurable H<sub>2</sub> concentrations compared to the total thickness of the applied matrix (about 1.5 cm) was relatively low throughout the experiment. Even at the initial conditions, when H<sub>2</sub> penetrated more than 5 mm (Figure 3), a large fraction of the manure was net-producing CO<sub>2</sub>, and the diffusional supply from this production must have prevented CO<sub>2</sub> limitation of the process near the membrane and buffered the potential increase in pH.

### Hydrogenotrophic Community at Mesophilic Conditions: Microstructure and Control by the CO<sub>2</sub> Supply

The start-up of an anaerobic reactor may require some time (weeks) but once established the populations of the different groups of fermenters, acetogens and methanogens are active and balanced (Griffin et al., 1998). As pointed by Díaz et al. (2015), the introduction of H<sub>2</sub> into an anaerobic reactor does not require the use of specific strains, and it is possible to use for instance unspecific anaerobic sludge inoculum to produce an active hydrogenotrophic methanogenic reactor. The relatively slow growth rates of acetogens and acetoclastic methanogens result in relatively long reactor start-up times before a stable methane production can be obtained (Luo and Angelidaki, 2013; Ghorbanian et al., 2014; Díaz et al., 2015). However, the supply of H<sub>2</sub> to an established anaerobic reactor results in the immediate activation of the hydrogenotrophic community, a



fast consumption of H<sub>2</sub> and an increase in CH<sub>4</sub> production (Luo and Angelidaki, 2012; Wang et al., 2013; Xu et al., 2014; Hu et al., 2015). Our data confirm the predisposition of the microbial community present in a mesophilic anaerobic reactor to metabolize an external supply of H<sub>2</sub> (Figures 4 and 5), and also provide a relevant description of the microstructure of the process.

One of the main problems previously reported by H<sub>2</sub> addition to reactors is the limitation of H<sub>2</sub> transfer to the reactor fluid (Luo et al., 2012; Luo and Angelidaki, 2013; Bensmann et al., 2014) by direct supply of H<sub>2</sub> gas to the reactor, resulting in substantial H<sub>2</sub> concentrations in the produced biogas. The use of gas permeable membranes for the H<sub>2</sub> supply makes it possible to have localized high H<sub>2</sub> concentrations and consumption rates (Figures 3 and 4) without massive ebullition of H<sub>2</sub> to the headspace. Laboratory scale membrane reactors have thus been shown to produce biogas with only trace concentrations of H<sub>2</sub> in the produced biogas and a high quality of up to 99% CH<sub>4</sub> (Wang et al., 2013). The performance of membrane reactors is, however, limited by the amount of H<sub>2</sub> that can diffuse out of the membranes, and membrane reactors must thus contain a large surface area of permeable tubing. Hollow fiber membranes may be applied (Wang et al., 2013; Díaz et al., 2015) instead of silicone tubing, but the cost of these membranes might prohibit full scale applications.

Addition of H<sub>2</sub> to anaerobic reactors can result in high rates of methanogenesis, but as the H<sub>2</sub> supply is not coming from a catabolic process with a parallel CO<sub>2</sub> production there may be a problem with CO<sub>2</sub> depletion and rise in pH. The interruption of the CO<sub>2</sub> supply as shown in Figures 4 and 5 resulted in an immediate decrease in H<sub>2</sub> consumption rate and an increase in pH (Figures 4 and 5). Although no external CO<sub>2</sub> was supplied, the small internal CO<sub>2</sub> production and CO<sub>2</sub>/HCO<sub>3</sub><sup>-</sup> pool were supporting H<sub>2</sub> consumption rates for a few hours, dropping to low values after 10 h. During this period, the H<sub>2</sub> consumption rates decreased progressively while the CO<sub>2</sub> depletion resulted in an increase of pH to values of about 9.2 (Figure 5), similar to the values measured in other reactors exposed to high H<sub>2</sub> loadings (Wang et al., 2013; Xu et al., 2014). This value corresponds to the pK<sub>a</sub> of ammonium (9.25), which is present at high concentrations in anaerobic reactors and therefore at this pH will contribute to buffering compound in the reactor. Several microbial processes may be inhibited at such high pH, resulting in the stop of the anaerobic digestion and the destabilization of the reactor, finally leading to the dramatic decrease in H<sub>2</sub> consumption and CH<sub>4</sub> production (Wang et al., 2013). Especially the high concentration of free NH<sub>3</sub> at high pH values may be very inhibitory for both methanogens and syntrophic acetogens (Wang et al., 2015). The supply of CO<sub>2</sub> together with the H<sub>2</sub> through the silicone membrane following CO<sub>2</sub> depletion resulted in a partial recovery of the H<sub>2</sub> consumption rates (Figure 5). The optimal CO<sub>2</sub> content in the feeding gas producing the highest H<sub>2</sub> consumption rates was 5.7% (Figures 5 and 6), which seems to be far away for the theoretical

17.8% CO<sub>2</sub> needed to react with the 71% H<sub>2</sub> following a 1:4 stoichiometry for hydrogenotrophic methanogenesis. However, the CO<sub>2</sub> permeability in silicone membrane is about five times higher than for H<sub>2</sub> (Robb, 1968), resulting in higher fluxes through the membrane. Consequently, the CO<sub>2</sub> supplied by the membrane was enough to support the high biological demand from the hydrogenotrophic methanogenesis, decreasing the pH to values below 8 (Figures 5 and 6). pH values below 8 are considered as optimal for anaerobic digestion (Luo and Angelidaki, 2013; Wang et al., 2013) and allow for the highest CH<sub>4</sub> upgrade (Wang et al., 2013). Higher increases in the proportion of CO<sub>2</sub> did not increase the H<sub>2</sub> consumption rates (Figures 5 and 6), but decreased the pH of the matrix close to the membrane.

A balanced CO<sub>2</sub> supply, being external or internal, is crucial for maintaining reactor performance by addition of H<sub>2</sub> to a methanogenic reactor. The internal capacity of the reactor to produce CO<sub>2</sub>, including the extraction from bicarbonate resulting in moderate pH increases, should thus not be exceeded to maintain stable microbial processes. Addition of H<sub>2</sub> to a methanogenic reactor through a membrane or by bubbling will probably mostly be used to upgrade the methane content of the biogas by reduction of the high CO<sub>2</sub> levels present in the slurry, but it may be an advantage to have the possibility for adding more CO<sub>2</sub> along with the H<sub>2</sub> to facilitate recovery in reactors temporarily exposed to an overdose of H<sub>2</sub>. The present study is from a membrane reactor, and methanogenic membrane reactors may be difficult to implement at full scale. The results illustrate, however, that H<sub>2</sub> may be depleted at a very short distance from a rising H<sub>2</sub> bubble as the penetration of H<sub>2</sub> was ~0.5 mm from the membrane into the digestate. Apart from a gradual rise in bulk pH, a supply of H<sub>2</sub> to biogas reactors may thus only have local consequences on microbial transformations such as interspecies H<sub>2</sub> transfer.

## AUTHOR CONTRIBUTIONS

All the authors designed the study. EG-R and NR carried out the experiments. All authors interpreted the data. EG-R and NR wrote the manuscript with input from all co-authors.

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## REFERENCES

- Alkaya, E., Erguder, T. H., and Demirer, G. N. (2010). Effect of operational parameters on anaerobic co-digestion of dairy cattle manure and agricultural residues: a case study for the Kahramanmaraş region in Turkey. *Eng. Life Sci.* 10, 552–559. doi: 10.1002/elsc.201000037
- Bensmann, A., Hanke-Rauschenbach, R., Heyer, R., Kohrs, F., Benndorf, D., Reichl, U., et al. (2014). Biological methanation of hydrogen within biogas plants: a model-based feasibility study. *Appl. Energy* 134, 413–425. doi: 10.1016/j.apenergy.2014.08.047
- Berg, P., Risgaard-Petersen, N., and Rysgaard, S. (1998). Interpretation of measured concentration profiles in sediment pore water. *Limnol. Oceanogr.* 43, 1500–1510. doi: 10.4319/lo.1998.43.7.1500
- Crozier, T. E., and Yamamoto, S. (1974). Solubility of hydrogen in water, sea water, and sodium chloride solutions. *J. Chem. Eng. Data* 19, 242–244. doi: 10.1021/jc60062a007
- Demirel, B., and Scherer, P. (2008). The roles of acetotrophic and hydrogenotrophic methanogens during anaerobic conversion of biomass to methane: a review. *Rev. Environ. Sci. Biotechnol.* 7, 173–190. doi: 10.1007/s11157-008-9131-1
- Díaz, I., Pérez, C., Alfaro, N., and Fdz-Polanco, F. (2015). A feasibility study on the bioconversion of CO<sub>2</sub> and H<sub>2</sub> to biomethane by gas sparging through polymeric membranes. *Bioresour. Technol.* 185, 246–253. doi: 10.1016/j.biortech.2015.02.114
- Ghorbanian, M., Lupitsky, R. M., Satyavolu, J. V., and Berson, R. E. (2014). Impact of supplemental hydrogen on biogas enhancement and substrate removal efficiency in a two-stage expanded granular sludge bed reactor. *Environ. Eng. Sci.* 31, 253–260. doi: 10.1089/ees.2013.0496
- Griffin, M. E., McMahon, K. D., Mackie, R. I., and Raskin, L. (1998). Methanogenic population dynamics during start-up of anaerobic digesters treating municipal solid waste and biosolids. *Biotechnol. Bioeng.* 57, 342–355. doi: 10.1002/(SICI)1097-0290(19980205)57:3<342::AID-BIT11>3.0.CO;2-I
- Hu, Y., Hao, X., Zhao, D., and Fu, K. (2015). Enhancing the CH<sub>4</sub> yield of anaerobic digestion via endogenous CO<sub>2</sub> fixation by exogenous H<sub>2</sub>. *Chemosphere* 140, 34–39. doi: 10.1016/j.chemosphere.2014.10.022
- Kemmer, G., and Keller, S. (2010). Nonlinear least-squares data fitting in Excel spreadsheets. *Nat. Protoc.* 5, 267–281. doi: 10.1038/nprot.2009.182
- Luo, G., and Angelidaki, I. (2012). Integrated biogas upgrading and hydrogen utilization in an anaerobic reactor containing enriched hydrogenotrophic methanogenic culture. *Biotechnol. Bioeng.* 109, 2729–2736. doi: 10.1002/bit.24557
- Luo, G., and Angelidaki, I. (2013). Co-digestion of manure and whey for in situ biogas upgrading by the addition of H<sub>2</sub>: process performance and microbial insights. *Appl. Microbiol. Biotechnol.* 97, 1373–1381. doi: 10.1007/s00253-012-4547-5
- Luo, G., Johansson, S., Boe, K., Xie, L., Zhou, Q., and Angelidaki, I. (2012). Simultaneous hydrogen utilization and in situ biogas upgrading in an anaerobic reactor. *Biotechnol. Bioeng.* 109, 1088–1094. doi: 10.1002/bit.24360
- Nielsen, M., Larsen, L. H., Ottosen, L. D. M., and Revsbech, N. P. (2015). Hydrogen microsensors with hydrogen sulfide traps. *Sens. Actuators B Chem.* 215, 1–8. doi: 10.1016/j.snb.2015.03.035
- Pham, C. H., Triolo, J. M., and Sommer, S. G. (2014). Predicting methane production in simple and unheated biogas digesters at low temperatures. *Appl. Energy* 136, 1–6. doi: 10.1016/j.apenergy.2014.08.057
- Revsbech, N. P. (1989). Diffusion characteristics of microbial communities determined by use of oxygen microsensors. *J. Microbiol. Methods* 9, 111–122. doi: 10.1016/0167-7012(89)90061-4
- Revsbech, N. P. (2005). “Analysis of microbial communities with electrochemical microsensors and microscale biosensors,” in *Methods in Enzymology*, ed. J. R. Leadbetter (Cambridge, MA: Academic Press), 147–166.
- Robb, W. L. (1968). Thin silicone membranes—their permeation properties and some applications. *Ann. N. Y. Acad. Sci.* 146, 119–137. doi: 10.1111/j.1749-6632.1968.tb20277.x
- Shink, B. (1988). “Principles and limits of anaerobic degradation: environmental and technological aspects,” in *Biology of Anaerobic Microorganisms*, ed. A. J. B. Zehnder (New York, NY: Wiley Interscience), 771–846.
- Sherif, S. A., Barbir, F., and Veziroglu, T. N. (2005). Wind energy and the hydrogen economy—review of the technology. *Sol. Energy* 78, 647–660. doi: 10.1016/j.solener.2005.01.002
- Stempien, J. P., Ni, M., Sun, Q., and Chan, S. H. (2015). Production of sustainable methane from renewable energy and captured carbon dioxide with the use of Solid Oxide Electrolyzer: a thermodynamic assessment. *Energy* 82, 714–721. doi: 10.1016/j.energy.2015.01.081
- Wang, H., Fotidis, I. A., and Angelidaki, I. (2015). Ammonia effect on hydrogenotrophic methanogens and syntrophic acetate-oxidizing bacteria. *FEMS Microbiol. Ecol.* 91:fiv130. doi: 10.1093/femsec/fiv130
- Wang, W., Xie, L., Luo, G., Zhou, Q., and Angelidaki, I. (2013). Performance and microbial community analysis of the anaerobic reactor with coke oven gas biomethanation and in situ biogas upgrading. *Bioresour. Technol.* 146, 234–239. doi: 10.1016/j.biortech.2013.07.049
- Xu, H., Gong, S., Sun, Y., Ma, H., Zheng, M., and Wang, K. (2014). High-rate hydrogenotrophic methanogenesis for biogas upgrading: the role of anaerobic granules. *Environ. Technol.* 36, 529–537. doi: 10.1080/09593330.2014.979886
- Zeeman, G., Sutter, K., Vens, T., Koster, M., and Wellinger, A. (1988). Psychrophilic digestion of dairy cattle and pig manure: start-up procedures of batch, fed-batch and CSTR-type digesters. *Biol. Wastes* 26, 15–31. doi: 10.1016/0269-7483(88)90146-2
- Zwietering, M., Jongenburger, I., Rombouts, F., and Van't Riet, K. (1990). Modeling of the bacterial growth curve. *Appl. Environ. Microbiol.* 56, 1875–1881.

**Conflict of Interest Statement:** The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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