

## RESEARCH ARTICLE

# Characteristics and sources of atmospheric pollutants in typical inland cities in arid regions of central Asia: A case study of Urumqi city

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## Abstract

The arid zone of central Asia secluded inland and has the typical features of the atmosphere. Human activities have had a significant impact on the air quality in this region. Urumqi is a key city in the core area of the Silk Road and an important economic center in Northwestern China. The urban environment is playing an increasingly important role in regional development. To study the characteristics and influencing factors of the main atmospheric pollutants in Urumqi, this study selected Urumqi's daily air quality index (AQI) data and observation data of six major pollutants including fine particulate matter (PM<sub>2.5</sub>), breathable particulate matter (PM<sub>10</sub>), sulfur dioxide (SO<sub>2</sub>), nitrogen dioxide (NO<sub>2</sub>), carbon monoxide (CO), and ozone (O<sub>3\_8h</sub>) from 2014 to 2018 in conjunction with meteorological data to use a backward trajectory analysis method to study the main characteristics of atmospheric pollutants and their sources in Urumqi from 2014 to 2018. The results showed that: (1) From 2014 to 2018, the annual average of PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub> and CO concentrations showed a downward trend, and O<sub>3\_8h</sub> concentrations first increased, then decreased, and then increased, reaching the highest value in 2018 (82.15 μg·m<sup>-3</sup>); The seasonal changes of PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub> and CO concentrations were characterized by low values in summer and fall seasons and high values in winter and spring seasons. The concentration of O<sub>3\_8h</sub>, however, was in the opposite trend, showing the high values in summer and fall seasons, and low values in winter and spring seasons. From 2014 to 2018, with the exception of O<sub>3\_8h</sub>, the concentration changes of the other five major air pollutants were high in December, January, and February, and low in May, June, and July; the daily changes showed a "U-shaped" change during the year. The high-value areas of the "U-shaped" mode formed around the 50th day and the 350th day. (2) The high-value area of AQI was from the end of fall (November) to the beginning of the following spring (March), and the low-value area was from April to October. It showed a U-shaped change trend during the year and the value was mainly distributed between 50 and 100. (3) The concentrations of major air pollutants in Urumqi were significantly negatively correlated with precipitation, temperature, and humidity

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( $P < 0.01$ ), and had the highest correlation coefficients with temperature. (4) Based on the above analysis results, this study analyzed two severe pollution events from late November to early December. Analysis showed that the  $PM_{2.5}/PM_{10}$  ratio in two events remained at about 0.1 when the pollution occurred, but was higher before and after the pollution (up to 1.46). It was shown that the pollution was a simple sandstorm process. Backward trajectory analysis clustered the airflow trajectories reaching Urumqi into 4 categories, and the trajectories from central Asia contributed the maximum values of average  $PM_{2.5}$  and  $PM_{10}$  concentrations.

## 1 Introduction

Since the political and economic restructuring and opening up, China's urbanization level has increased from 18% in the early stages (1980s) to the current of 57.40% [1,2], and the number of urban populations has increased from 173 million in 1979 to 813 million in 2017 [3]. The development of China's urbanization started late [1] and was mainly followed a partial and short-term targeted developing pattern in the early stage [4]. The ecological protection and environmental governance work left behind, which has caused a series of ecological and environmental problems [5,6]. Among these environmental problems, the air pollution in urban areas is one of the most serious ones. Since the appearance of the air pollution, there have been a large number of related studies try to explain the internal mechanism of air quality changes and provide feasible guidance and related suggestions for improving urban air quality [7–10].

In China, urban air pollution in different urban areas has different properties, but the coarse particulate matter ( $PM_{10}$ ) and fine particulate matter ( $PM_{2.5}$ ) are the two pollutants that have been paid more attention than others (such as sulfur dioxide and nitrogen dioxide) by scientists. It has been proved that the air quality in Chinese cities was generally improved, and the proportion of cities that meet air quality standards is rising [11]. The Beijing-Tianjin-Hebei region (three provinces around Beijing, also named Jingjinji region) was classified as the poorest air quality region in China and corresponding research have shown that high  $PM_{2.5}$  concentration of the region was due to coal consumption, high population density and construction [12]. Based on analysis of the source of air pollution, two transmission paths of pollutants has been identified by Gao, Wang [13] and they also found that the presence of a pressure equalization field, a low-level inversion layer, and the southern warm and humid airflow provided favorable conditions for the formation of  $PM_{2.5}$  in the region. The central plain region and Yangtze River Delta area of China are also facing the problem of air pollution mainly contribute by  $PM_{10}$  and  $PM_{2.5}$ , as has been reported elsewhere [14,15].

The current studies corresponding to air quality of urban area in China mainly focused  $PM_{10}$  and  $PM_{2.5}$  and conducted in eastern coastal areas or central plain region of China, little attention has been paid in urban area of northwestern part (also inland part) of the country. It should be noted that the urbanization in inland regions also witnessed severe air pollution [16], a systematic analysis on the component, source and transmission path of urban air pollution is helpful to provide decision support for the government to make reasonable and effective air pollution control measures in these regions. This study took the Urumqi city (in recent years, the air pollution has become increasingly prominent in urban areas of the city) in northwest part of China as an example to analyze the spatial-temporal characteristics of air pollution of typical inland city in northwest China. Specifically, we aimed to understanding the component, source and transmission path of the pollutant of the city and quantify the relationship

between the Air Quality Index (AQI) with meteorological data based on air quality observation data collected from 2014 to 2018. The backward trajectory analysis was used to study the temporal variation of two severe and typical air pollution events that occurred from late November to early December in 2018.

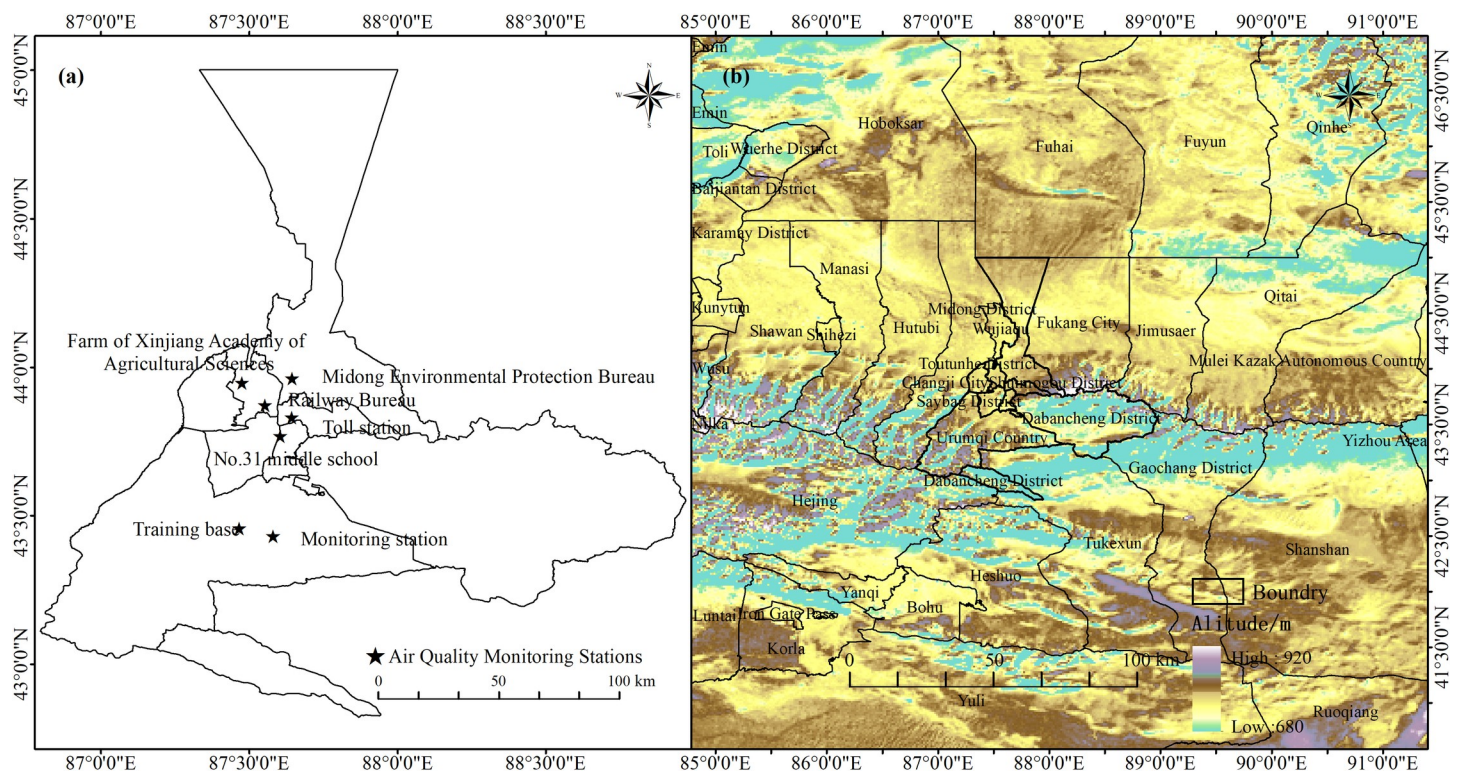
## 2. Materials and methods

### 2.1 Study area

Urumqi ( $42^{\circ}45'32''-44^{\circ}08'00''\text{N}$ ,  $86^{\circ}37'33''-88^{\circ}58'24''\text{E}$ , total area of  $14,216\text{ km}^2$ ), locating in the hinterland of Eurasia and is one of the most important cities of the Silk Road (Fig 1A). The city is experiencing rapid economic development and population growth at the present. The GDP reached 309.98 billion RMB, an increase of 7.8% over the previous year in 2018. According to statistics yearbook 2018, Urumqi reached 90.2% of urbanization rate and a build-up area of  $365.88\text{ km}^2$ . Urumqi is surrounded by mountains in the east, south and west, with high terrain in the southeast and low in the northwest (Fig 1B). The elevation of the city ranges from 680 to 920 meters with an average of 800 meters in the urban area. It has a temperate continental arid climate. The mean temperature is  $7.3^{\circ}\text{C}$  and reached the highest in July and August with an average of  $25.7^{\circ}\text{C}$ , and lowest in January with an average of  $-15.2^{\circ}\text{C}$ . The annual average precipitation is 236 mm, and the 64% falls in Spring and Summer.

### 2.2 Data sources

The AQI is a dimensionless indicator that quantitatively describes the air quality [17–20]. According to new ambient air quality standard of China (GB395-2012), the index is the highest concentrations of  $\text{PM}_{2.5}$ ,  $\text{PM}_{10}$ , sulfur dioxide ( $\text{SO}_2$ ), nitrogen dioxide ( $\text{NO}_2$ ), carbon



**Fig 1.** Map of study area ((a)Location and distribution of monitoring stations; (b)Orographic distribution in Urumqi and its surrounding areas).

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monoxide (CO), and maximum 8-hour average ozone ( $O_3_{8h}$ ). There are also different definition of AQI, such as fuzzy-based AQI (FAQI), which weighted the concentration of  $PM_{10}$ ,  $SO_2$ ,  $NO_2$ , CO and  $O_3_{8h}$ , based on fuzzy algorithm [21] and air pollution index (API), which do not cover the concentration of  $PM_{2.5}$  [22]. The present study considered the AQI and the level of classification are shown in Table 1. AQI is widely used in China due to its simple definition [23,24].

The air quality monitoring data (concentrations of  $PM_{2.5}$ ,  $PM_{10}$ ,  $SO_2$ ,  $NO_2$ , CO and  $O_3_{8h}$ ) used to calculate the AQI in this study was obtained from the China Air Quality Monitoring and Analysis Platform (<https://www.aqistudy.cn/>). Seven automatic air quality monitoring stations (Training base, Toll station, Monitoring station, No.31 middle school, Railway Bureau, Farm of Xinjiang Academy of Agricultural Sciences, Midong Environmental Protection Bureau) were selected. These seven air quality monitoring stations were classified according to functional areas, representing clean areas (Training base and Farm of Xinjiang Academy of Agricultural Sciences), residential areas (Toll station, Monitoring station and Railway Bureau), industrial areas (Midong District Environmental Protection Bureau), and cultural and educational areas (No.31 Middle School). When calculating the AQI, we used the daily average of these seven monitoring stations in Urumqi. The data was collected from January 1, 2014 to December 31, 2018. Meteorological data for backward trajectory analysis were downloaded from synchronized Global Data Assimilation System (GDAS) maintained by the National Center for Environmental Forecasting (NCEP, <ftp://arlftp.arlhq.noaa.gov/pub/archives/gdas1>). The climate data were obtained from China Meteorological Data Network (<http://data.cma.cn/>).

China Air Quality Monitoring and Analysis Platform is an automatic air quality monitoring system consists of monitoring stations, quality control laboratories and system support laboratories. The quality control laboratories are responsible for the standardization, calibration and audit of monitoring equipment, ensuring the accurate transmission and storage of data. In addition, these laboratories also responsible for correcting and eliminating of the abnormal data (missing data and errors caused by non-human factors such as instrument failure and power failure). With the help of the laboratory, the system could provide accurate, reliable, continuous and timely environmental monitoring data including AQI,  $PM_{2.5}$ ,  $PM_{10}$ ,  $SO_2$ ,  $NO_2$ ,  $O_3$ , CO, temperature, humidity, wind force scale, wind direction, satellite cloud image and other monitoring items. In the system, all data are updated hourly and automatically.

### 2.3 Statistical analysis

In this study, the Pearson's correlation coefficient was used to analyze the relationship between AQI and the meteorological factors, the simple liner regression was used to quantify the

**Table 1. Standard of air quality index [24,25] (GB3095-2012 environmental air quality standard).**

AQI	Air Quality Level	Air Quality Grade	Health Effects
0–50	Level 1	Excellent	The body feels comfortable
51–100	Level 2	Good	A very small number of very sensitive people feel lightly ill
101–150	Level 3	Light pollution	People susceptible to increased discomfort, healthy people appear stimulating symptoms
151–200	Level 4	Moderate pollution	Exacerbates symptoms in susceptible populations and causes respiratory distress in healthy populations
201–300	Level 5	Heavy pollution	Patients with heart and lung disease are uncomfortable and exercise tolerance is reduced
>300	Level 6	Severe pollution	Exercise tolerance is reduced and early symptoms of some diseases appear among healthy people

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response of AQI to meteorological factors, the corresponding calculations were performed using the Statistics Package for Social Science software v.22.0 (SPSS Inc, Chicago, IL). In addition, the MeteoInfo [26] software was used to analyze the backward trajectory and related statistical analysis.

TrajStat trajectory model, a plug-in of MeteoInfo, is used for backward trajectory analysis, which is a professional model for calculating air mass trajectory and was widely used in the study of air pollutant transport and diffusion. In this paper, we analyzed the backward trajectory of the air flow that affected two air pollution events in Urumqi (the first one appeared in November 25<sup>th</sup> 2018 and the second one started at 1<sup>st</sup> and ended at 2<sup>nd</sup> in December of 2018). The corresponding analysis was simulated at height of 500 m from the ground and the center of Urumqi (43°37'N, 87°50'E) were setup as starting point. In addition, the trajectory estimation time is 24h. Based on the simulation results, the air flow trajectories gathered in Urumqi were classified into four categories. Furthermore, the different transport trajectories of PM<sub>2.5</sub> and PM<sub>10</sub> are obtained by overlying the concentrations of PM<sub>2.5</sub> and PM<sub>10</sub> to the classified categories.

### 3. Results

#### 3.1 Temporal variation of major atmospheric pollutants

From 2014 to 2018, the annual average concentrations of PM<sub>2.5</sub> and CO followed an increase-decrease trend and reached the highest values (72.77  $\mu\text{g}\cdot\text{m}^{-3}$  and 1.46  $\text{mg}\cdot\text{m}^{-3}$ , respectively) in 2016 and the lowest values (52.64  $\mu\text{g}\cdot\text{m}^{-3}$  and 1.24  $\text{mg}\cdot\text{m}^{-3}$ , respectively) in 2018 (Fig 2). The concentrations of PM<sub>10</sub> and SO<sub>2</sub> showed a downward trend and reached the lowest values (110.15  $\mu\text{g}\cdot\text{m}^{-3}$  and 10.64  $\mu\text{g}\cdot\text{m}^{-3}$ , respectively) in 2018. The inter-annual variation of NO<sub>2</sub> concentration fluctuated from 2014 to 2018 and the highest (54.99  $\mu\text{g}\cdot\text{m}^{-3}$ ) and the lowest

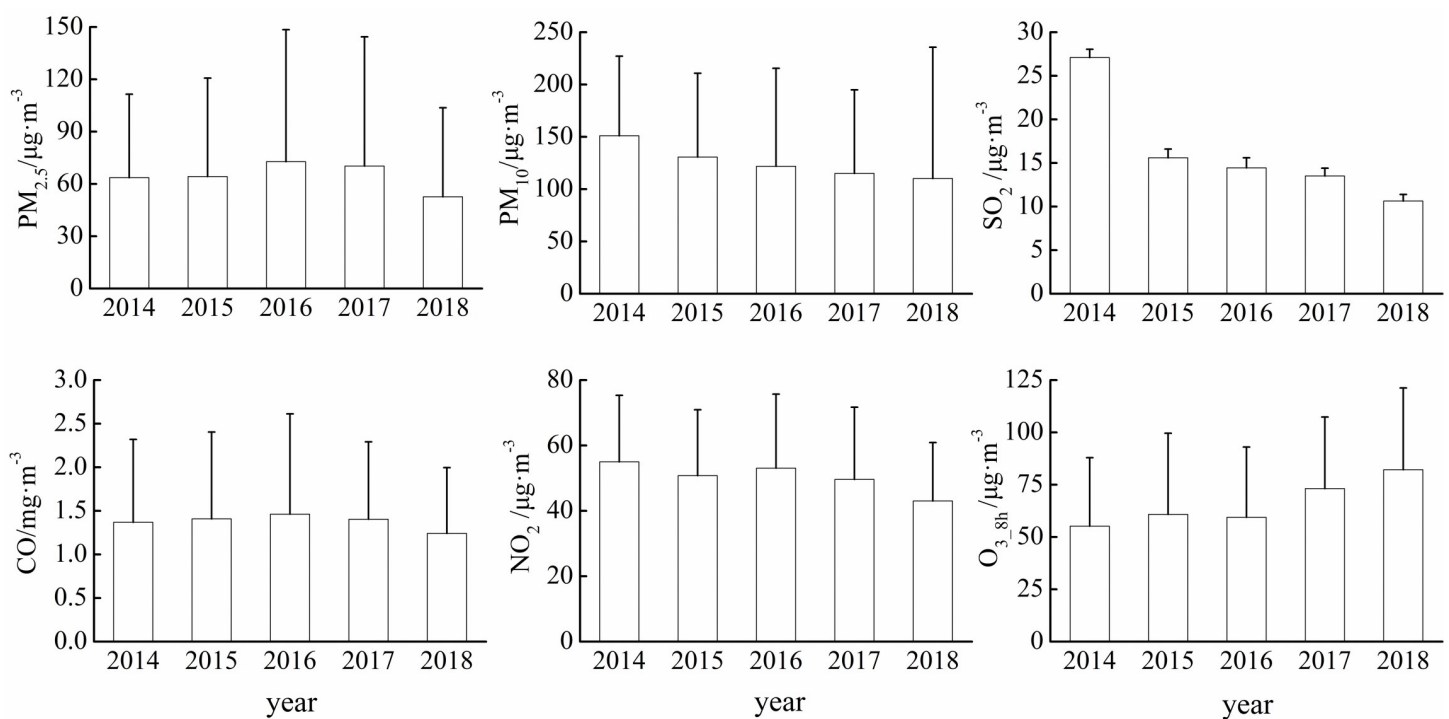


Fig 2. Inter-annual variations of major air pollutants in Urumqi.

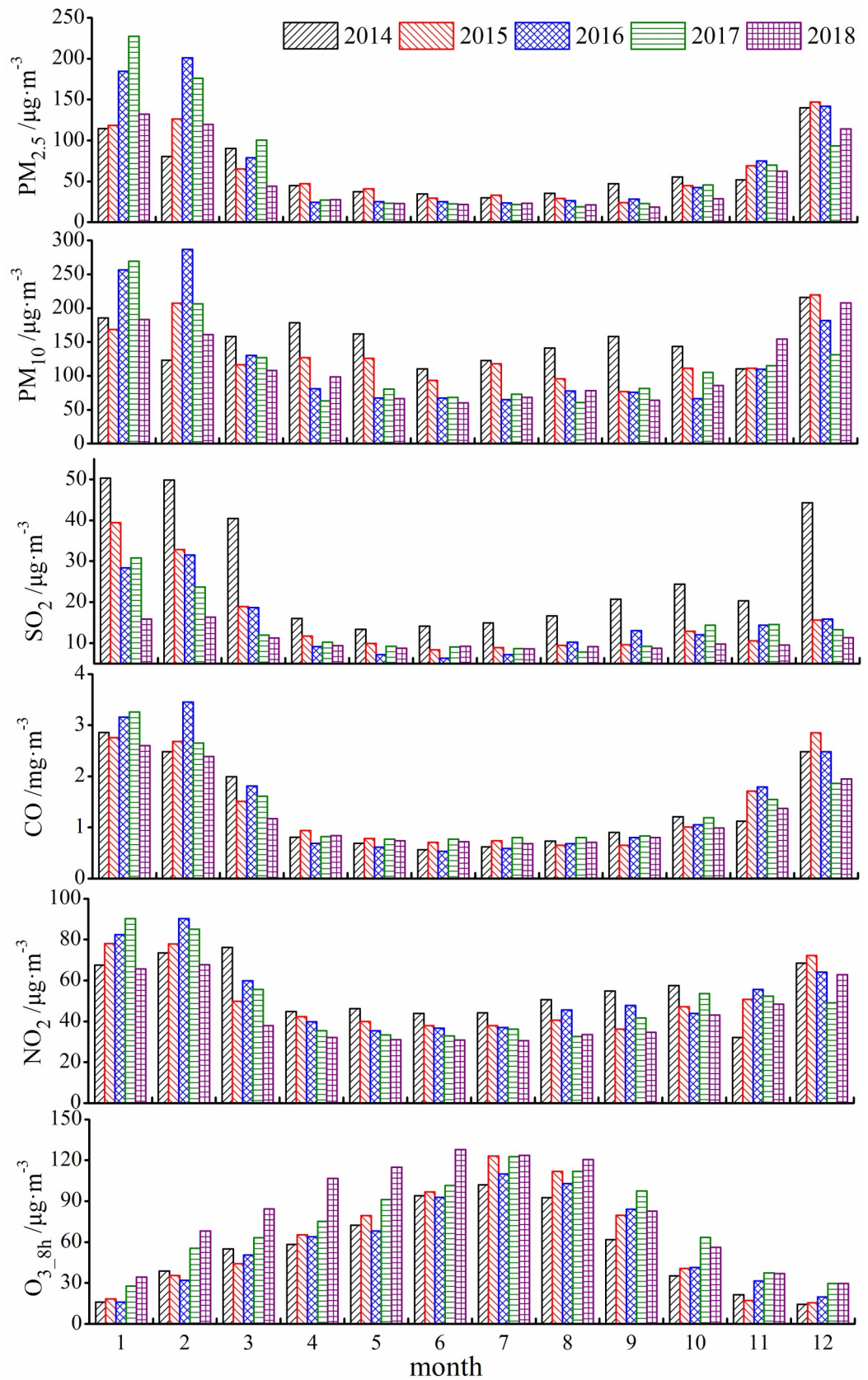
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( $43.50 \mu\text{g}\cdot\text{m}^{-3}$ ) concentrations of  $\text{NO}_2$  appeared in 2014 and 2018, respectively. The concentration of  $\text{O}_3_{8\text{h}}$  showed an upward trend, which declined slightly in 2016 and reached the highest value ( $82.15 \mu\text{g}\cdot\text{m}^{-3}$ ) in 2018.

The inter-annual variations of  $\text{PM}_{2.5}$ ,  $\text{PM}_{10}$ ,  $\text{SO}_2$ ,  $\text{NO}_2$  and CO were similar with lowest and highest value appeared in summer and winter, respectively (Fig 3). In contrast, the concentration changes of  $\text{O}_3_{8\text{h}}$  reached highest in summer and lowest in winter. The Urumqi is the remotest city from any ocean in the world and the feature of climate is short warm summers and long cold winters [27]. It determined that Urumqi has very high demands and consumption rates of fossil fuels for wintertime heating. Moreover, lots of industries and people were attracted to Urumqi from all parts of China with the strongly developing economy, enhanced the increasing consumption of energy by fossil fuels and the steady growing fleet of motor vehicles. It is likely to explain the characteristics of pollutants ( $\text{PM}_{2.5}$ ,  $\text{PM}_{10}$ ,  $\text{SO}_2$ ,  $\text{NO}_2$  and CO) mentioned above in this area. The variations concentration of  $\text{O}_3_{8\text{h}}$  rather different from other pollutants, as well, it can be explained by the increasing of solar radiation [28].

The monthly average concentrations of  $\text{PM}_{2.5}$  and CO were low from April to September and started to rise in October, after reached the annual highest in December and January, they gradually decreased from February to March (Fig 3). During the study period (from January, 2014 to December, 2018), monthly concentration of  $\text{PM}_{2.5}$  reached its highest value ( $227.52 \mu\text{g}\cdot\text{m}^{-3}$ ) in January, 2017. For CO, the highest concentration value ( $3.45 \text{mg}\cdot\text{m}^{-3}$ ) were reached in February 2016 (Fig 3). In February 2016, the concentrations of  $\text{PM}_{10}$  reached its monthly highest value in five years with  $286.69 \mu\text{g}\cdot\text{m}^{-3}$ . The monthly average concentration of  $\text{SO}_2$  was obviously higher in 2014 than in other years (especially in December).  $\text{NO}_2$  concentration reached its highest value ( $90.19 \mu\text{g}\cdot\text{m}^{-3}$ ) in January 2017. The monthly average concentration of  $\text{O}_3_{8\text{h}}$  followed a unimodal pattern in all years and reached the highest value in July. Compare with other years, the  $\text{O}_3_{8\text{h}}$  concentration was higher in 2018.

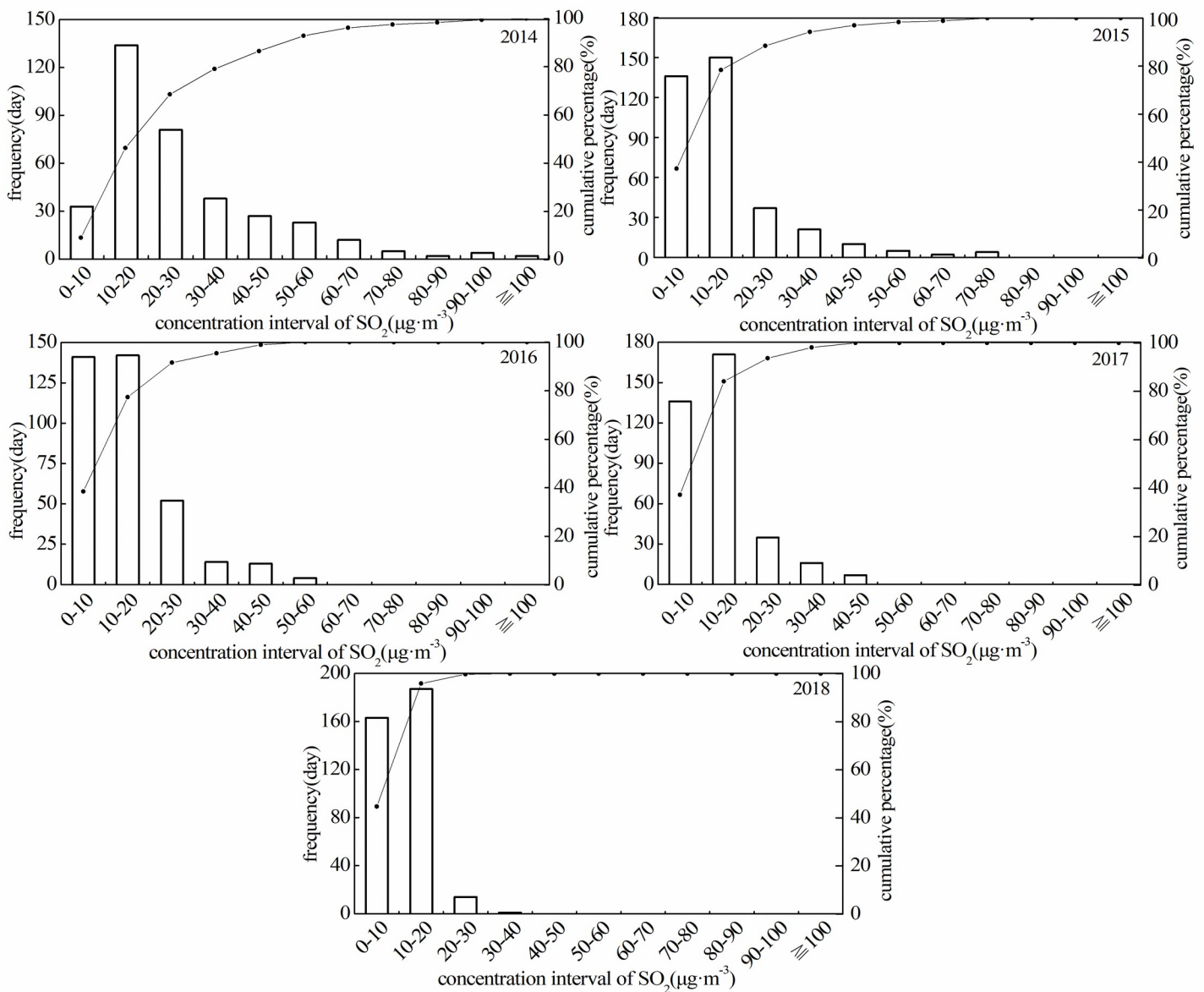
During our study period, the interannual and seasonal variations of  $\text{PM}_{2.5}$ ,  $\text{PM}_{10}$ ,  $\text{SO}_2$ , CO, and  $\text{NO}_2$  concentrations showed a decreasing trend, and the variation of  $\text{SO}_2$  concentrations was the most obviously (Figs 2 and 3). Additionally, the histogram (Fig 4) demonstrated that the variation range of  $\text{SO}_2$  concentration is shrinking, decreasing yearly and the concentration tends to be stable (CV from 70.53% in 2014 to 36.96% in 2018, Table 2). It should be noted that coal combustion is one of the main sources of urban air pollution in most of cities at western of China [29,30]. In 2014, aimed to reduce the degree of air pollution and improve the quality of human settlements, Urumqi launched 22 projects including energy conservation, air pollution prevention, and water pollution prevention, etc. In terms of air pollution control, the corresponding projects including coal-to-gas (a measure to change the energy structure), grid-connected boiler network (compare to independent heating unit) and relocation of polluting enterprises [31]. Due to these effective measures (along with air pollution prevention measures implemented in the steel industries and enhancement of transformation and obsolescence of high-emission equipment), the concentration of  $\text{SO}_2$  significantly declined from 2014 to 2018, as shows in the present study and elsewhere [32]. Additionally, the meteorological condition also contributed to the high concentration of  $\text{SO}_2$ , since when temperature is low, an inversion layer is easy to generate for a city located in valley, and the layer could hinder the diffusion of pollutants [33]. It is important to note although the concentrations of  $\text{PM}_{2.5}$ ,  $\text{PM}_{10}$ , CO,  $\text{NO}_2$  also decreased, the reduction is less effective compare with  $\text{SO}_2$ . This might be due to the measures (primarily aimed at reducing pollutants emitted by industry and energy use) that currently the local government enacted [32]. Since the air pollution is dominated by particulate pollutants ( $\text{PM}_{2.5}$  and  $\text{PM}_{10}$ ), more effective prevention and control measures for particulate pollutants would fundamentally improve air quality in the region.



**Fig 3. Monthly changes of major air pollutants in Urumqi.**

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The daily concentration of PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub>, CO and O<sub>3\_8h</sub> were similar with their intra-annual variation (Fig 5). Among these six pollutants, PM<sub>2.5</sub> and CO varied synchronously throughout the entire study period. Interestingly, for concentrations of PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, and CO, the lower (higher) the concentrations itself, the lower (higher) the standard deviation of the concentrations. In contrast, the NO<sub>2</sub> and O<sub>3\_8h</sub> did not displayed such pattern. In addition, there exist two extreme concentrations of PM<sub>10</sub> in late November and early December of 2018 (1274 μg·m<sup>-3</sup> and 1700 μg·m<sup>-3</sup>), indicating that there was a serious air pollution event in Urumqi during the period.



**Fig 4. Histogram of daily SO<sub>2</sub> concentration distribution in Urumqi.**

<https://doi.org/10.1371/journal.pone.0249563.g004>



Table 2. The Statistical characteristics of variation of main pollutant concentrations in Urumqi.

	PM <sub>2.5</sub> /μg·m <sup>-3</sup>			PM <sub>10</sub> /μg·m <sup>-3</sup>			SO <sub>2</sub> /μg·m <sup>-3</sup>		
	Mean	Standard Deviation	Variable coefficient (CV/%)	Mean	Standard Deviation	Variable coefficient (CV/%)	Mean	Standard Deviation	Variable coefficient (CV/%)
2014	63.62	47.87	75.25	151.00	76.14	50.42	27.10	19.11	70.53
2015	64.26	56.45	87.86	130.65	80.14	61.34	15.60	12.56	80.55
2016	72.77	75.69	104.02	121.88	93.57	76.77	14.45	10.26	70.98
2017	70.31	74.09	105.38	115.05	79.83	69.39	13.52	7.97	58.98
2018	52.64	51.02	96.93	110.15	125.48	113.92	10.64	3.93	36.96
	CO/mg·m <sup>-3</sup>			NO <sub>2</sub> /μg·m <sup>-3</sup>			O <sub>3_8h</sub> /μg·m <sup>-3</sup>		
	Mean	Standard Deviation	Variable coefficient (CV/%)	Mean	Standard Deviation	Variable coefficient (CV/%)	Mean	Standard Deviation	Variable coefficient (CV/%)
2014	1.37	0.95	69.29	54.99	20.34	36.98	55.19	32.68	59.22
2015	1.41	1.00	70.72	50.76	20.16	39.71	60.80	38.78	63.79
2016	1.46	1.15	78.55	53.08	22.63	42.62	59.43	33.59	56.53
2017	1.40	0.89	63.19	49.64	22.06	44.45	73.13	34.20	46.76
2018	1.24	0.75	60.78	43.05	17.85	41.45	82.15	39.10	47.59

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### 3.2 variation of in AQI

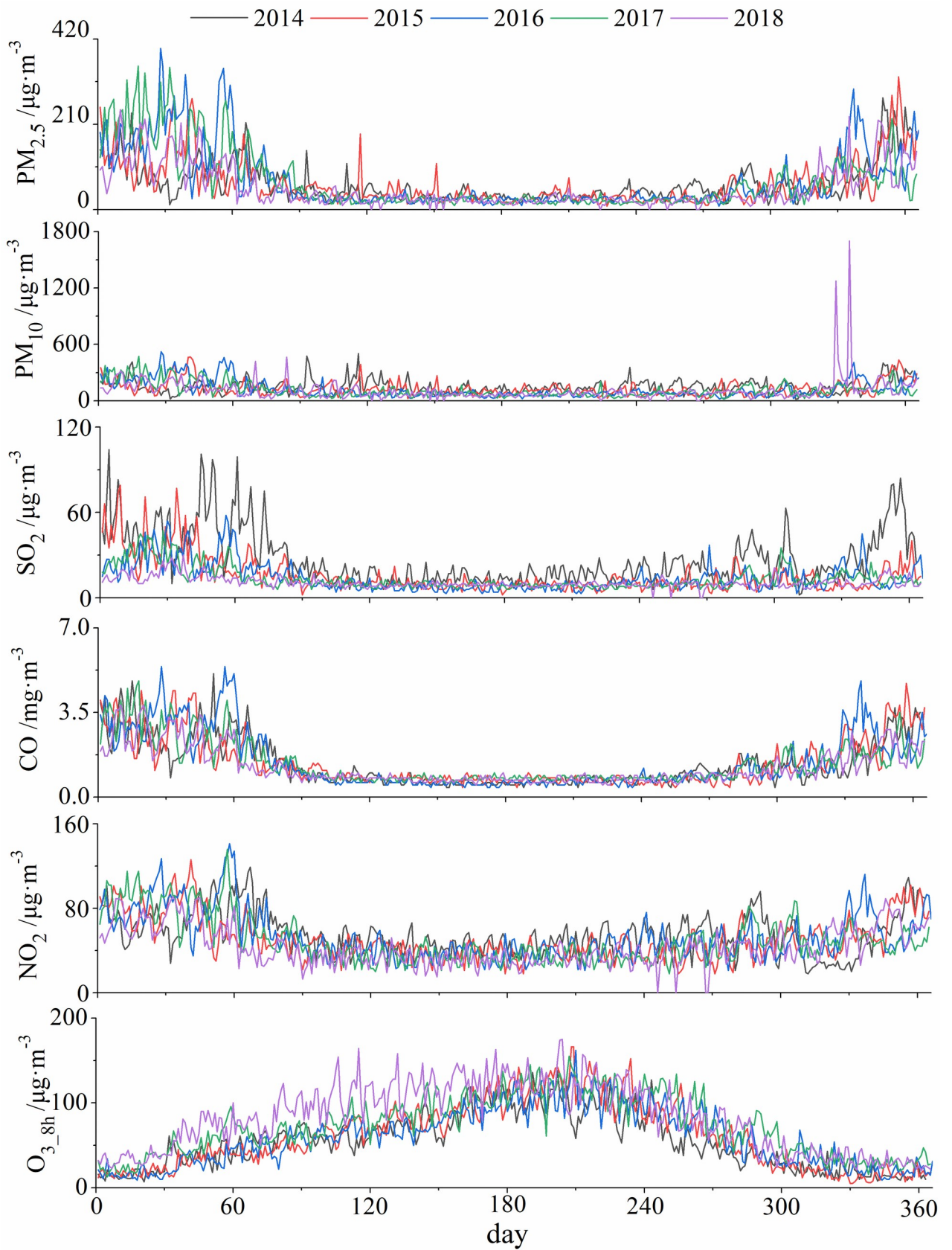
AQI was relatively higher from November of first year to March of the following year and lower from April to October with a unimodal trend. Majority of AQI value ranged between 51–100, which denoted good air quality (Fig 6). The proportion of days with good air quality (AQI value ranged between 51–100) was 45.73% in 2014 and increased to 60.55% in 2018 (Fig 7). Overall, number of days with excellent air quality (AQI value less than 51) increased from 2014 to 2018. Moreover, number of days with severe air quality (AQI value higher than 300) was fewest among six air quality levels. The light, moderate and heavy pollution (the range of AQI value are 101–150, 151–200 and 201–300, respectively) days may appear in any time of the year, but the possibility was relatively higher from November to May of next year. Summer (May to August) was the optimal season with the best air quality and winter (December to February) was the worst one.

### 3.3 Analysis of meteorological factors

As displays in Table 3, among six meteorological factors, the temperature is the only factor that have relatively higher correlation coefficients with six pollutants. Specifically, it significantly and negatively correlated with PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, CO, and NO<sub>2</sub> ( $p < 0.01$ ), and significantly and positively correlated with O<sub>3\_8h</sub> ( $p < 0.01$ ). The correlation coefficient between humidity, air pressure and wind speed with six pollutants are less than 0.3, which indicated that they had little effect on the concentration of pollutants.

### 3.4 Analysis of two severe pollution events in late November and early December of 2018

From November 25<sup>th</sup> to December 2<sup>nd</sup>, 2018, Urumqi suffered two separate severe air pollution events (Fig 8). During the two events, the AQI is maintained at the highest value of 500 and the pollution level is 6, which are serious pollution events. The first one began at 9:00 on November 25<sup>th</sup>, 2018 and ended at 20:00 on the same day. Within 12 hours of the event, the concentrations of PM<sub>2.5</sub> and PM<sub>10</sub> exceeded 100 μg·m<sup>-3</sup> and 1000 μg·m<sup>-3</sup>, respectively, and reached their highest value with 254 μg·m<sup>-3</sup> and 4061 μg·m<sup>-3</sup>. The concentrations of PM<sub>2.5</sub> and

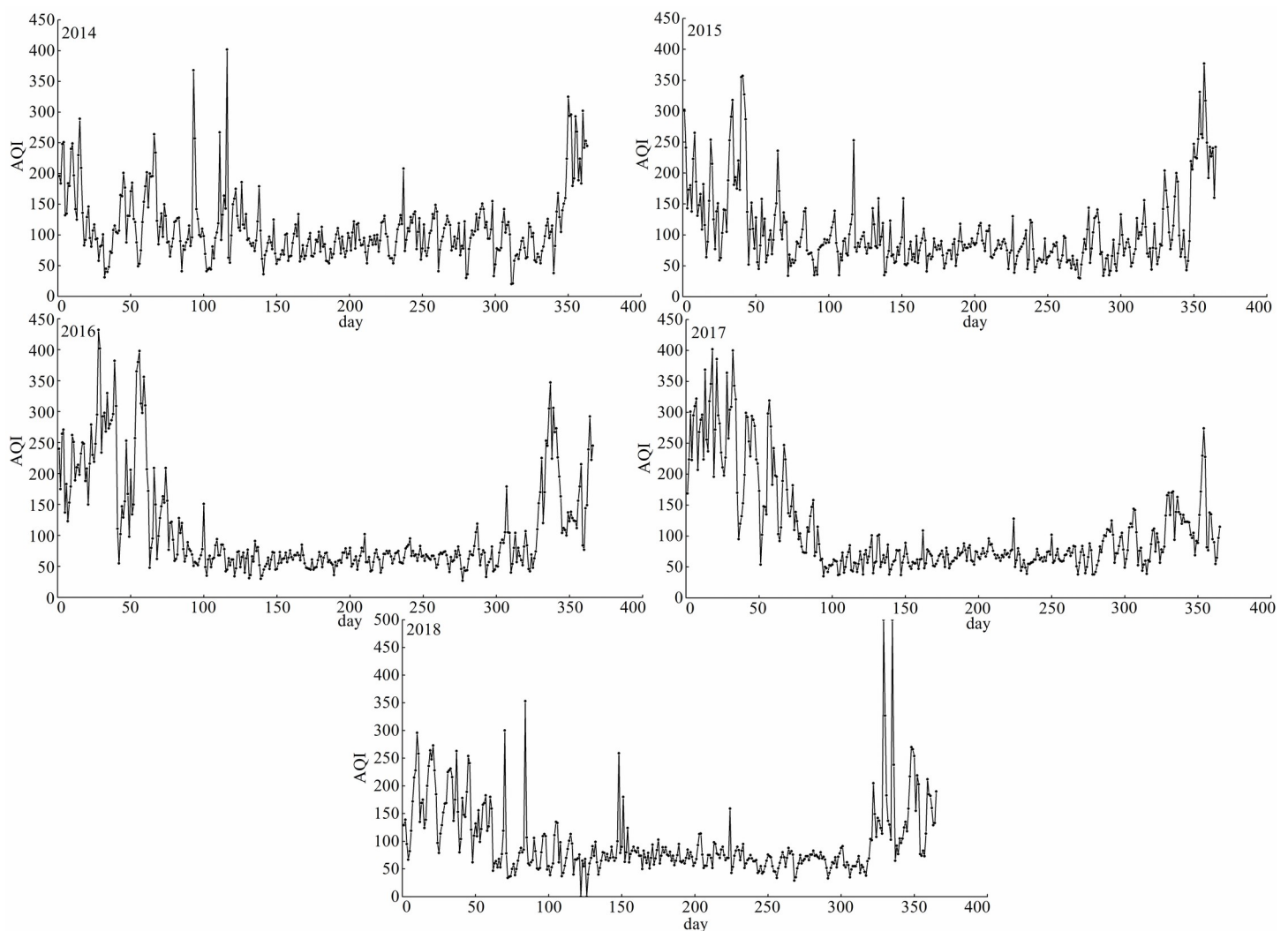


**Fig 5. Daily changes of atmospheric pollutants in Urumqi.**

<https://doi.org/10.1371/journal.pone.0249563.g005>

PM<sub>10</sub> decreased rapidly to normal values in 6 hours after the event ended. The second event lasted 21 hours and the concentrations of PM<sub>2.5</sub> and PM<sub>10</sub> reached 120  $\mu\text{g}\cdot\text{m}^{-3}$  and 1028  $\mu\text{g}\cdot\text{m}^{-3}$ , respectively, at 0:00 of December 1<sup>st</sup>, and then returned to normal values. The pollution restarted at 7:00 and concentrations of PM<sub>2.5</sub> and PM<sub>10</sub> rose to the highest values with 351  $\mu\text{g}\cdot\text{m}^{-3}$  and 4578  $\mu\text{g}\cdot\text{m}^{-3}$ , respectively, at 16:00. The PM<sub>2.5</sub>/PM<sub>10</sub> ratio maintained around 0.1 during the two events and increased to about 1.46 after the second event, indicated that the pollutants of the two events dominated by sandstorm.

The two air pollution events mainly occurred on November 25<sup>th</sup>, and from December 1<sup>st</sup> to 2<sup>nd</sup>, 2018, the characteristics of the local weather was air humidity increased significantly, accompanied by precipitation and strong winds. The meteorological conditions accompanied pollution events mainly show the following characteristics (Fig 9): Firstly, the relative air humidity during the events was apparently higher than the humidity before and after

**Fig 6. Daily changes of AQI in Urumqi.**

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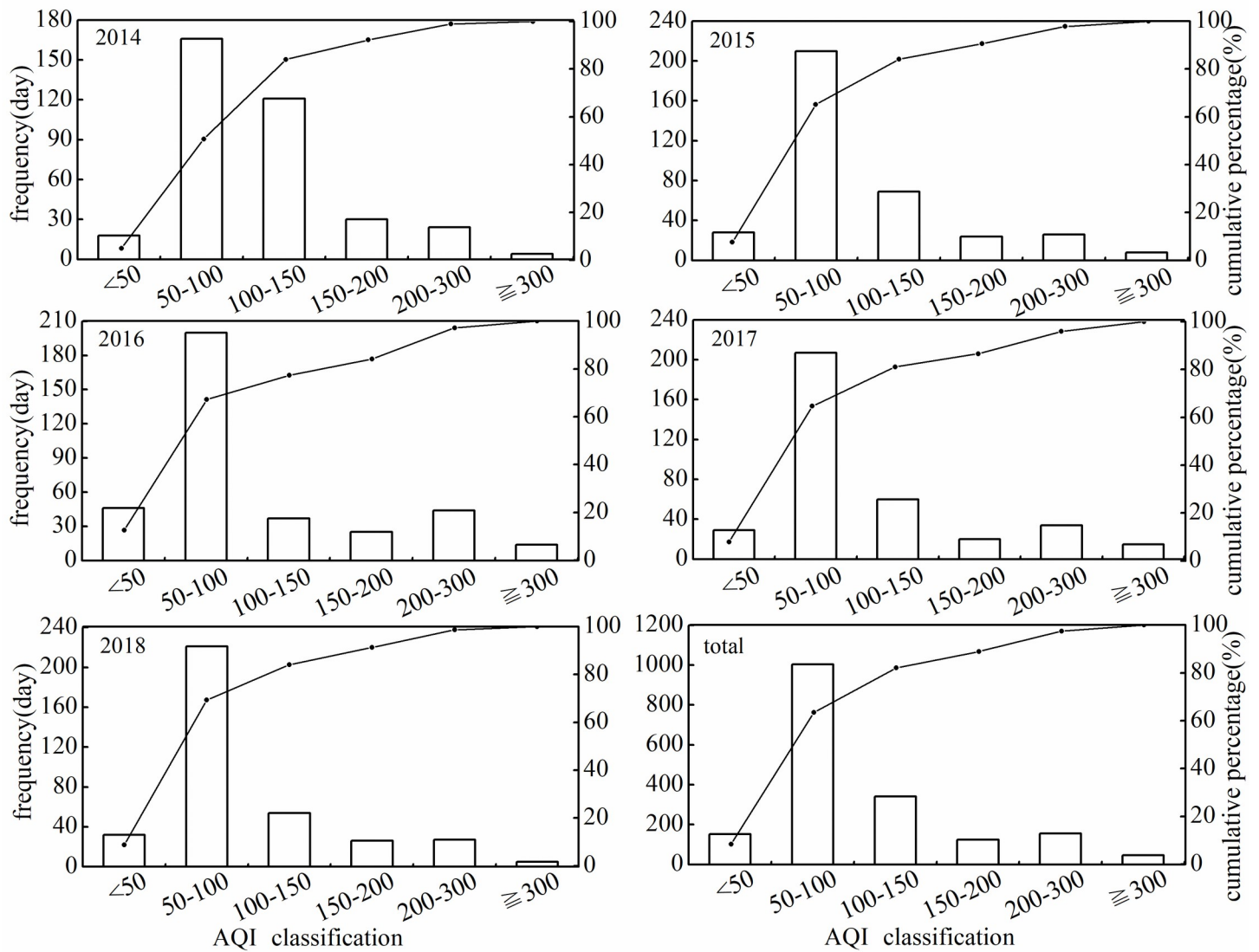


Fig 7. Frequency of pollution levels and cumulative percentage changes.

<https://doi.org/10.1371/journal.pone.0249563.g007>

Table 3. Correlation coefficients between main pollutants and meteorological factors in Urumqi.

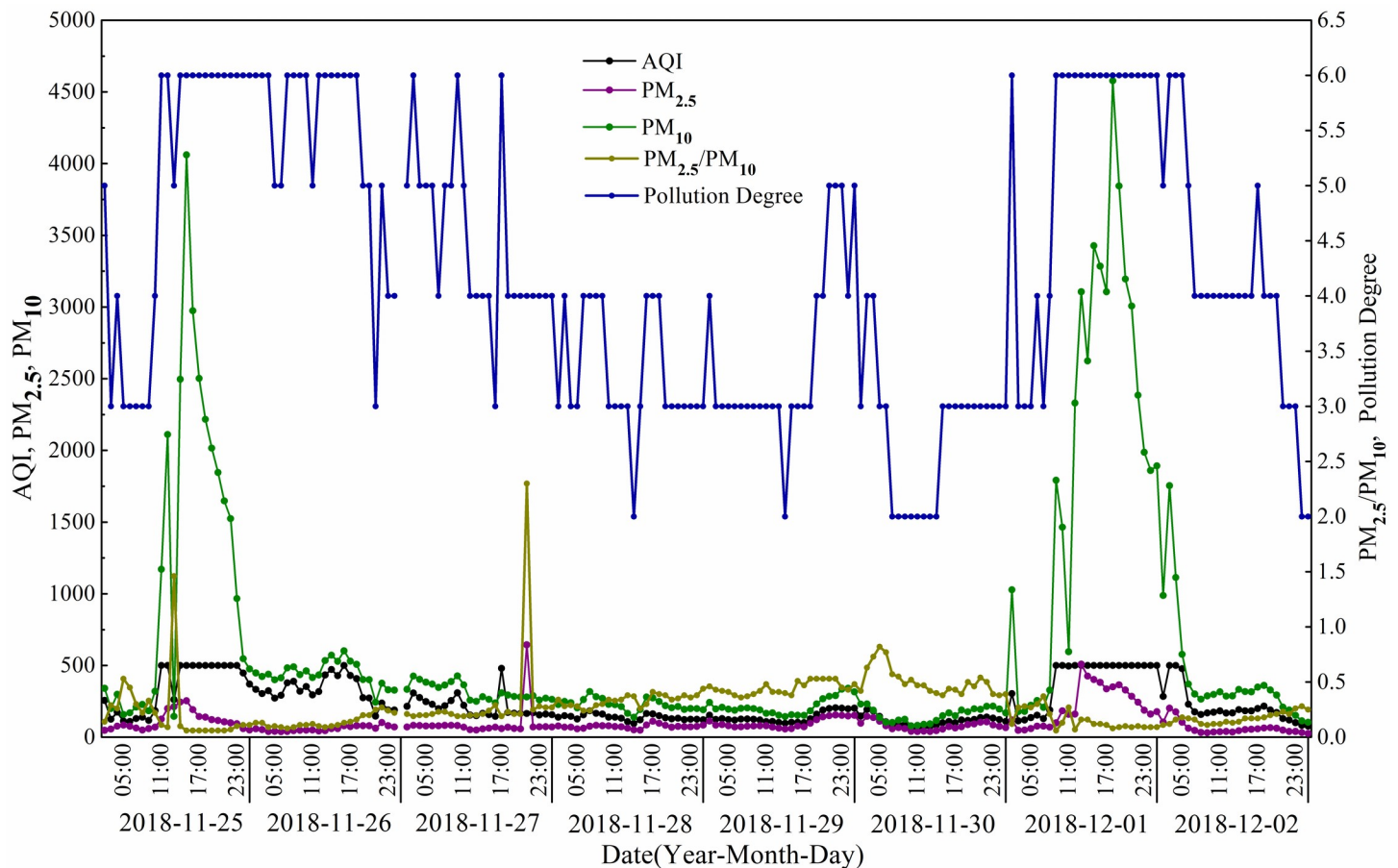
Pollutants	Precipitation	Temperature	Humidity	Air Pressure	Wind Speed
PM <sub>2.5</sub>	-0.212**	-0.555**	-0.170**	-0.003	-0.025
PM <sub>10</sub>	-0.201**	-0.299**	-0.300**	-0.011	0.094**
SO <sub>2</sub>	-0.193**	-0.430**	-0.183**	-0.032	-0.038
CO	-0.216**	-0.659**	-0.116**	0.030	-0.067**
NO <sub>2</sub>	-0.291**	-0.489**	-0.261**	0.035	0.016
O <sub>3_8h</sub>	0.101**	0.812**	-0.221**	-0.187**	0.209**

\*\*significant with  $p < 0.01$

\* significant with  $p < 0.05$ .

<https://doi.org/10.1371/journal.pone.0249563.t003>





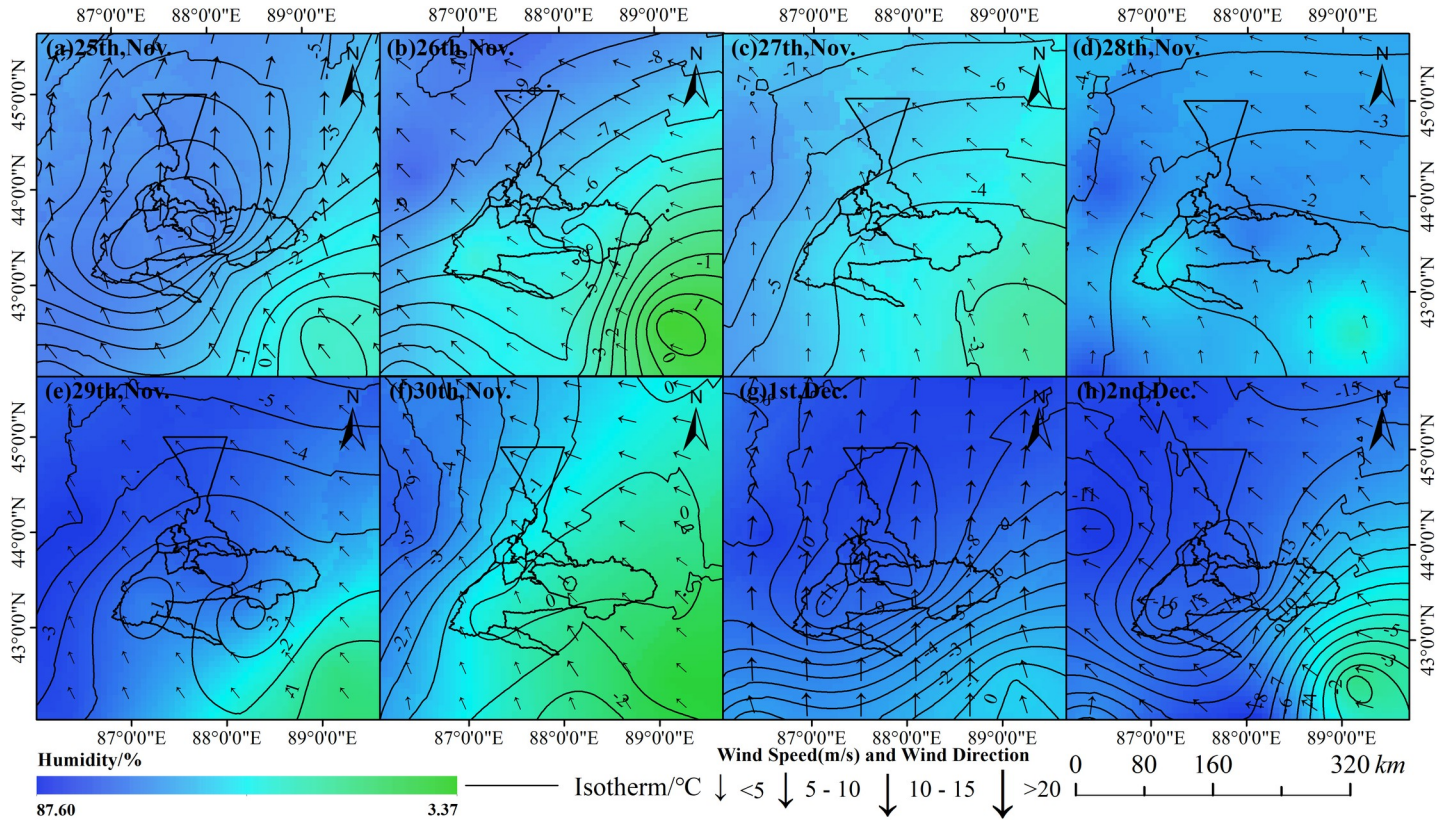
**Fig 8. Variations of AQI and concentrations of  $PM_{2.5}$  and  $PM_{10}$  during the two air pollution events emerged from November 25<sup>th</sup> to December 2<sup>nd</sup>, 2018.**

<https://doi.org/10.1371/journal.pone.0249563.g008>

pollution, and the spatial distribution of relative humidity is higher in northwest than in south-east portion of the study area. Secondly, the isotherm is more concentrated and curved during the events (Fig 9A, 9G and 9H) compare with other time (Fig 9B–9F), indicating the drastic variation of air temperature. Finally, during the events, the wind speed is faster (the maximum wind speed is more than 10m/s), and the wind direction was also changed (dominated by south and southeast wind).

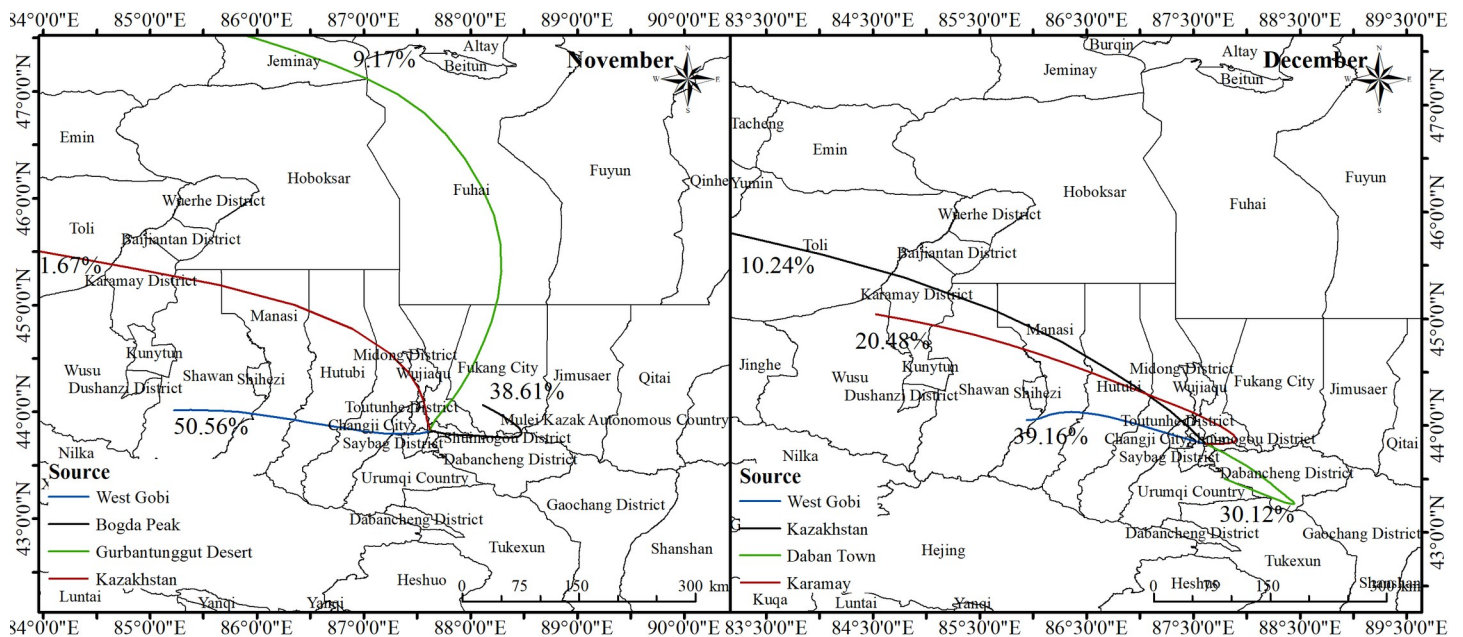
Backward trajectories analysis revealed that air mass trajectories that arrived in Urumqi could be classified to four categories with the major source originated from west and southeast, and the transmission paths is consistent with the wind direction (Fig 10). The trajectory with the largest number of airflows originated from the Gobi Desert in the west of Urumqi. The transmission distance was short and the overall airflow of the Gobi Desert accounting for 50.56% and 39.16% of the total flows during the two events, respectively. The trajectory with the longest transmission distance came from Central Asia (Kazakhstan) and brought sandy pollutants from the Gurban Tungut Desert. This trajectory accounted for 1.67% and 10.24% of the total airflow in the first and second event, respectively.

The distribution of various trajectories, the concentrations of corresponding pollutants, and terrain around Urumqi demonstrated that (Table 4 and Fig 1B): (1) The trajectory from Central Asia Kazakhstan had a long transmission distance and came from the sand source area. There were enough sand and dust pollutants and active atmospheric reaction conditions.



**Fig 9. Weather conditions in Urumqi and its surrounding areas from November 25<sup>th</sup> to December 2<sup>nd</sup>, 2018.**

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**Fig 10. Backward trajectory cluster distribution.**

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**Table 4. Distribution characteristics and average pollutant concentration of all backward trajectory.**

Time	Clustering	Source	Path Area	Probability%	PM <sub>2.5</sub> Concentration (µg/m <sup>3</sup> )	PM <sub>10</sub> Concentration (µg/m <sup>3</sup> )
Nov.	1	West of China	West Gobi, Changji	50.56	80.98	347.54
	2	East of China	Bogda Peak, Daban Town	38.61	48.24	337.05
	3	North of China	Kazakhstan, Ulungur Lake, Gurbantunggut Desert, Fukang,	9.17	20.12	23.15
	4	West of China	Kazakhstan, Tacheng area, Karamay, Gurbantunggut Desert	1.67	142.58	1384.47
Dec.	1	West of China	West Gobi, South of Shihezi, Hutubi, Urumqi	39.16	55.02	147.66
	2	Northwest of China	Kazakhstan, Tacheng area, Karamay, Gurbantunggut Desert	10.24	237.65	2031.76
	3	Southeast of China	urumqi county, Daban Town	30.12	73.10	105.66
	4	Northwest of China	Karamay, North of Shihezi, Wujiaqu, Midong District	20.43	81.82	555.50

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The concentrations of PM<sub>2.5</sub> and PM<sub>10</sub> were the highest. (2) The trajectories from the West Gobi and Daban Town were easy to accumulate pollutants due to the short transmission distance and relatively stable meteorological conditions. (3) Compared with the air mass trajectory in December, the trajectory from the Karamay was close to and intersected with the trajectory from Central Asia in December. The concentrations of PM<sub>2.5</sub> and PM<sub>10</sub> were reached 81.82 µg·m<sup>-3</sup> and 555.50 µg·m<sup>-3</sup>, respectively. (4) The minimums of PM<sub>2.5</sub> and PM<sub>10</sub> occurred in November in the Central Asian airflow trajectory because of the long transmission distance and obstruction of the Bogda Peak, reached 20.12 µg·m<sup>-3</sup> and 23.15 µg·m<sup>-3</sup>, respectively.

#### 4. Discussion

Several studies have shown that meteorological conditions have great influence on the concentration of atmospheric pollutants in cities with similar climate like Urumqi [34–36]. In general, air pollution in these cities was prone to occur during heating systems are operating (beginning at November until the end of March) and high dust generation periods (usually the Spring), rather than periods with strong ultraviolet radiation [18]. This is confirmed by the high negative correlations between the concentrations of six major air pollutants and air temperature (Table 3). The intensively developed continental anticyclone control the climate of Urumqi, downward flows strengthened the accumulation of pollutants, and snow-covered surface results in strong cooling of the adjacent air, which in turn is prone for the build-up of surface inversions. As mentioned previously, an inversion layer which could hinder the diffusion of pollutants is easy to generate for a city located in a valley [33]. In contrast, the air quality in coastal cities is mainly affected by other climatic conditions such as the monsoon [37,38], humidity [39,40] or atmospheric pressure [41,42]. It is interesting that in our study, the correlation relationship between wind speed and concentrations of six major pollutants in Urumqi is weak (Table 3). This is different from Trabzon city [43], Florence, Milan and Vicence [44] where diffusion effect of wind usually results in a relatively high correlation coefficient between the concentration of atmospheric pollutant and wind speed. We speculate the weak relationship in Urumqi might be due to the balance of two mechanisms. Firstly, although in summer the wind speed is high, which is favorable condition to sand storm, whereas abundant rainfall in the same season restricted the floating of dust [45]. Secondly, a relatively lower wind



speed in winter is accompanied by lower air temperature, which usually results an inversion layer that prevents the diffusion of pollutants [33].

The topographical situation played an important role in air pollution of the city according to the analysis of PM concentrations in Xinjiang [46,47], which contains China's two large deserts, the Taklimakan and Gurbantunggut desert. Moreover, the west, east, and the south of Urumqi are bordered by Tianshan mountains, the north is resisted by northwest wind from Siberia carrying abundant dust (the northwest wind comes through the Gurbantunggut desert) [48]. Relevant studies have shown that the dust storms in Xinjiang usually occurred in spring and summer [49,50], and affect a large region downwind, such as the glaciers in the Tibetan Plateau [51], Eastern Asia [52], and across the Pacific, even influencing the air quality over North America [53]. The highest PM concentrations were recorded in cities surrounding the Taklimakan Desert during the spring season and the highest PM<sub>2.5</sub>/PM<sub>10</sub> ratio was recorded during the winter, indicating the influence of anthropogenic emissions in winter [50]. However, the PM<sub>2.5</sub>/PM<sub>10</sub> ratio maintained around 0.1 during the two events occurred from November 25<sup>th</sup> to December 2<sup>nd</sup> (Fig 8), combined with the characteristics of the weather at the time of the pollution events that the local air humidity increased significantly, accompanied by precipitation and strong winds, indicated that the pollutants of the two events dominated by sandstorm. Backward trajectory analysis showed that the airflow from Central Asia Kazakhstan carries a large amount of dust pollutants (the concentration of PM<sub>10</sub> reached 2031.76  $\mu\text{g}\cdot\text{m}^{-3}$ , Table 4) when passing through Gurbantunggut Desert, which is blocked by Tianshan Mountains bordered in the east, west, and the south of Urumqi, and forms sandstorms under the influence of winter downward flow, strengthened the concentrations of the PM<sub>10</sub>, reached 4578  $\mu\text{g}\cdot\text{m}^{-3}$ .

Examining the source of atmospheric pollutants is help for a better understanding of the formation mechanisms of pollution events [54–56]. Relevant studies have shown that air masses from different regions contribute differently to air pollution [57–59]. An analysis of the trajectories of PM<sub>2.5</sub> transport in the Yangtze River Delta found that most of the heavily polluted days belonged to shorter trajectory groups and were controlled by high-pressure [60]. In the Lanzhou city, which is locates in Northwest inland region of China, backward trajectory analysis revealed that airflow from sand sources was prone to cause PM<sub>10</sub> pollution events [61]. The present study found that although the long-distance airflow trajectories from Central Asia was fewer, the sand and dust carried by these long-distance airflow trajectories had a higher contribution to air pollution in Urumqi. In contrast, although there were more short-distance trajectories, due to relatively lower of concentrations of pollutants, they contributed lesser than long-distance ones.

## 5 Conclusions

At present the issue of air quality has become a focus for city with the rapid development of economic and urbanization. In this paper, Urumqi city in Xinjiang province of China is selected as a study area to analysis the characteristics and sources of atmospheric pollutants (PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub>, CO and O<sub>3\_8h</sub>) based on conventional statistical methods and backward trajectory analysis. The results showed: (1) The annual average concentrations of major atmospheric pollutants showed different decline trend during the period from 2014 to 2018; the intra-annual variations of PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, CO and NO<sub>2</sub> were similar with lowest and highest value appeared in summer and winter, respectively; diametrically, the concentrations change of O<sub>3\_8h</sub> reached highest in summer and lowest in winter. (2) AQI is higher from November to March and lower from April to October with values varies between 50 and 100. (3) The temperature obtained relatively higher correlation relationship with air quality



compare with humidity, air pressure, and wind speed. (4) Airflows arrived to Urumqi can be clustered into four categories using backward trajectory analysis, and the one from Central Asia contributed the mostly to the concentration of PM<sub>2.5</sub> and PM<sub>10</sub>.

## Supporting information

### S1 Dataset.

(ZIP)

## Author Contributions

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## References

1. Weiqi Z., *Urbanization Process and Its Eco-environmental Effects in Typical Regions of China*. Beijing: Science Press, 2017.
2. Tan Y., Xu H., and Zhang X., *Sustainable urbanization in China: A comprehensive literature review*. Cities, 2016. 55: p. 82–93.
3. Guan X., et al., *Assessment on the urbanization strategy in China: Achievements, challenges and reflections*. Habitat International, 2018. 71: p. 97–109.
4. Guilin W. and Wei Z., *Effects of Urban Expansion and Changes in Urban Characteristics on PM<sub>2.5</sub> Pollution in China*. Environmental Science, 2019. 40(08): p. 3447–3456.
5. Bongaarts J., *United nations department of economic and social affairs, population division world mortality report 2005*. Population and Development Review, 2006. 32(3): p. 594–596.
6. Han L., et al., *Urbanization strategy and environmental changes: An insight with relationship between population change and fine particulate pollution*. Science of the Total Environment, 2018. 642: p. 789–799. <https://doi.org/10.1016/j.scitotenv.2018.06.094> PMID: 29920465
7. Becken S., et al., *Urban air pollution in China: destination image and risk perceptions*. Journal of Sustainable Tourism, 2017: p. 1–18.
8. Chen X., et al., *Long-term exposure to urban air pollution and lung cancer mortality: A 12-year cohort study in Northern China*. Science of the Total Environment, 2016: p. S0048969716315108. <https://doi.org/10.1016/j.scitotenv.2016.07.064> PMID: 27425436
9. Fan H., Zhao C., and Yang Y., *A comprehensive analysis of the spatio-temporal variation of urban air pollution in China during 2014–2018*. Atmospheric Environment, 2020. 220: p. 117066.
10. He K., Huo H., and Qiang Z., *URBAN AIR POLLUTION IN CHINA: Current Status, Characteristics, and Progress*. 2011. 27(1): p. 397–431.
11. Lin B. and Zhu J., *Changes in urban air quality during urbanization in China*. Journal of Cleaner Production, 2018. 188: p. 312–321.

12. Hu M., et al., *Insight into characteristics and sources of PM<sub>2.5</sub> in the Beijing–Tianjin–Hebei region, China*. National Science Review, 2015. 2(3): p. 257–258.
13. Gao J., et al., *Temporal-spatial characteristics and source apportionment of PM<sub>2.5</sub> as well as its associated chemical species in the Beijing–Tianjin–Hebei region of China*. Environmental pollution, 2018. 233: p. 714–724. <https://doi.org/10.1016/j.envpol.2017.10.123> PMID: 29126093
14. Hu J., et al., *Spatial and temporal variability of PM<sub>2.5</sub> and PM<sub>10</sub> over the North China Plain and the Yangtze River Delta, China*. Atmospheric Environment, 2015. 95(oct.): p. 598–609.
15. Shen F., et al., *Air pollution characteristics and health risks in Henan Province, China*. Environmental Research, 2017. 156: p. 625–634. <https://doi.org/10.1016/j.envres.2017.04.026> PMID: 28454015
16. Zhu C.-S., et al., *Indoor and outdoor chemical components of PM<sub>2.5</sub> in the rural areas of Northwestern China*. Aerosol Air Qual. Res, 2012. 12: p. 1157–1165.
17. Farong D., et al., *Analysis of air pollution episodes over different cities in the Yangtze River Delta*. China Environmental Science, 2018. 38(02): p. 401–411.
18. Lina X., et al., *The study of characteristics of air pollution and meteorological factors in Hohhot*. Journal of Arid Land Resources and Environmental, 2019. 33(06): p. 150–157.
19. Xiang H., et al., *Analysis of the driving factors of PM<sub>2.5</sub> in Jiangsu province based on grey correlation model*. Acta Geographica Sinica, 2016. 71(07): p. 1119–1129.
20. Huan Z., *Analysis of Beijing Air Quality Based on Time Series*. Science and Technology Innovation, 2019(09): p. 1–3.
21. Sowlat M.H., et al., *A novel, fuzzy-based air quality index (FAQI) for air quality assessment*. Atmospheric Environment, 2011. 45(12): p. 2050–2059.
22. Cairncross E.K., John J., and Zunckel M., *A novel air pollution index based on the relative risk of daily mortality associated with short-term exposure to common air pollutants*. Atmospheric environment, 2007. 41(38): p. 8442–8454.
23. Zhan D., et al., *The driving factors of air quality index in China*. Journal of Cleaner Production, 2018. 197: p. 1342–1351.
24. Zhan D., et al., *The driving factors of air quality index in China*. Journal of Cleaner Production, 2018: p. S095965261831758X-.
25. Yan W., et al., *Spatial Distributions of Atmospheric Contaminations and Effect of Surface Wind in Urumqi*. Journal of Desert Research, 2008(05): p. 986–991.
26. Wang Y.Q., *MeteoInfo: GIS software for meteorological data visualization and analysis*. Meteorological Applications, 2014. 21(2).
27. Domrös M. and Peng G., *The climate of China*. 2012: Springer Science & Business Media.
28. Costabile F., et al., *A preliminary assessment of major air pollutants in the city of Suzhou, China*. Atmospheric Environment, 2006. 40(33): p. 6380–6395.
29. Dongjie Y., et al., *Inventory of SO<sub>2</sub> emission in Xi'an and its contribution to the PM<sub>2.5</sub> of the municipality area*. Journal of Safety and Environment, 2018. 18(05): p. 2002–2007.
30. Zaili L., *Impacts of Relocation, Development, and West-East Transmission of Energy and Resource Industry in Western China on SO<sub>2</sub> Emission and Contamination in Northwestern, China*. 2018, Lanzhou University.
31. Meng X., et al., *Seasonal Characteristics and Particle-size Distributions of Particulate Air Pollutants in Urumqi*. International journal of environmental research and public health, 2019. 16(3): p. 396. <https://doi.org/10.3390/ijerph16030396> PMID: 30708935
32. Gou H., et al., *Assessment of microbial communities in PM<sub>1</sub> and PM<sub>10</sub> of Urumqi during winter*. Environmental Pollution, 2016. 214: p. 202–210. <https://doi.org/10.1016/j.envpol.2016.03.073> PMID: 27086076
33. Zhou J., et al., *Implications of the 11th and 12th Five-Year Plans for energy conservation and CO<sub>2</sub> and air pollutants reduction: a case study from the city of Urumqi, China*. Journal of Cleaner Production, 2016. 112: p. 1767–1777.
34. Wang Y., *Change Trend on Environmental Air Quality in Shihezi City [J]*. Arid Environmental Monitoring, 2004. 18(3): p. 169–171.
35. Tolkacheva G., *Problems of Air Quality in Tashkent City, in Environmental Simulation Chambers: Application to Atmospheric Chemical Processes*. 2006, Springer. p. 379–392.
36. Carlsen L., et al., *Assessment of the air quality of Almaty. Focussing on the traffic component*. International Journal of Biology and Chemistry, 2013. 5(1): p. 49–69.

37. Hien P., Loc P., and Dao N., *Air pollution episodes associated with East Asian winter monsoons*. Science of the total environment, 2011. 409(23): p. 5063–5068. <https://doi.org/10.1016/j.scitotenv.2011.08.049> PMID: 21925714
38. Chin M., et al., *Connection Between East Asian Air Pollution and Monsoon System, in Air Pollution in Eastern Asia: An Integrated Perspective*. 2017, Springer. p. 87–103.
39. Fang L., Clausen G., and Fanger P.O., *Impact of temperature and humidity on the perception of indoor air quality*. Indoor air, 1998. 8(2): p. 80–90.
40. Cheng Y., et al., *Humidity plays an important role in the PM<sub>2.5</sub> pollution in Beijing*. Environmental pollution, 2015. 197: p. 68–75. <https://doi.org/10.1016/j.envpol.2014.11.028> PMID: 25497308
41. Li L., et al., *Spatial and temporal analysis of Air Pollution Index and its timescale-dependent relationship with meteorological factors in Guangzhou, China, 2001–2011*. Environmental Pollution, 2014. 190: p. 75–81. <https://doi.org/10.1016/j.envpol.2014.03.020> PMID: 24732883
42. Ilten N. and Selici A.T., *Investigating the impacts of some meteorological parameters on air pollution in Balikesir, Turkey*. Environmental monitoring and assessment, 2008. 140(1–3): p. 267–277. <https://doi.org/10.1007/s10661-007-9865-1> PMID: 17955338
43. Cuhadaroglu B. and Demirci E., *Influence of some meteorological factors on air pollution in Trabzon city*. Energy and buildings, 1997. 25(3): p. 179–184.
44. Cogliani E., *Air pollution forecast in cities by an air pollution index highly correlated with meteorological variables*. Atmospheric Environment, 2001. 35(16): p. 2871–2877.
45. Mamtimin B. and Meixner F.X., *Air pollution and meteorological processes in the growing dryland city of Urumqi (Xinjiang, China)*. Science of the Total Environment, 2011. 409(7): p. 1277–1290.
46. Zhang X., et al., *Ambient TSP concentration and dustfall variation in Urumqi, China*. Journal of Arid Land, 2014. 6(6): p. 668–677.
47. Zhang X.-X., et al., *Dust deposition and ambient PM<sub>10</sub> concentration in north-west China: spatial and temporal variability*. Atmospheric Chemistry and Physics, 2017. 17(3): p. 1699–1711. <https://doi.org/10.1021/es504005q> PMID: 25470755
48. Li J., et al., *Characteristics of air pollution events over Hotan Prefecture at the southwestern edge of Taklimakan Desert, China*. Journal of Arid Land, 2018. 10(5): p. 686–700.
49. Mamtimin B. and Meixner F.X., *Air pollution and meteorological processes in the growing dryland city of Urumqi (Xinjiang, China)*. Sci Total Environ, 2011. 409(7): p. 1277–90. <https://doi.org/10.1016/j.scitotenv.2010.12.010> PMID: 21239046
50. Rupakheti D., et al., *Spatio-temporal characteristics of air pollutants over Xinjiang, northwestern China*. Environ Pollut, 2021. 268(Pt A): p. 115907. <https://doi.org/10.1016/j.envpol.2020.115907> PMID: 33120351
51. Kang S., et al., *Linking atmospheric pollution to cryospheric change in the Third Pole region: current progress and future prospects*. National Science Review, 2019. 6(4): p. 796–809.
52. Uno I., et al., *Asian dust transported one full circuit around the globe*. Nature Geoscience, 2009. 2(8): p. 557–560.
53. Husar R.B., et al., *Asian dust events of April 1998*. Journal of Geophysical Research: Atmospheres, 2001. 106(D16): p. 18317–18330.
54. Wang Y., Zhang X., and Draxler R.R., *TrajStat: GIS-based software that uses various trajectory statistical analysis methods to identify potential sources from long-term air pollution measurement data*. Environmental Modelling & Software, 2009. 24(8): p. 938–939.
55. Voutsas D., et al., *Ionic composition of PM<sub>2.5</sub> at urban sites of northern Greece: secondary inorganic aerosol formation*. Environmental Science and Pollution Research, 2014. 21(7): p. 4995–5006. <https://doi.org/10.1007/s11356-013-2445-8> PMID: 24363054
56. Ma L., et al., *Comparative analysis of chemical composition and sources of aerosol particles in urban Beijing during clear, hazy, and dusty days using single particle aerosol mass spectrometry*. Journal of Cleaner Production, 2016. 112: p. 1319–1329.
57. Pongkiatkul P. and Oanh N.T.K., *Assessment of potential long-range transport of particulate air pollution using trajectory modeling and monitoring data*. Atmospheric Research, 2007. 85(1): p. 3–17.
58. Shie R.-H. and Chan C.-C., *Tracking hazardous air pollutants from a refinery fire by applying on-line and off-line air monitoring and back trajectory modeling*. Journal of hazardous materials, 2013. 261: p. 72–82. <https://doi.org/10.1016/j.jhazmat.2013.07.017> PMID: 23912073
59. Wu X., et al., *The characteristics of air pollution induced by the quasi-stationary front: Formation processes and influencing factors*. Science of The Total Environment, 2020. 707: p. 136194. <https://doi.org/10.1016/j.scitotenv.2019.136194> PMID: 31972916

60. Fu Q., et al., *Mechanism of formation of the heaviest pollution episode ever recorded in the Yangtze River Delta, China*. *Atmospheric Environment*, 2008. 42(9): p. 2023–2036.
61. Jian Q., et al., *Analysis on Seasonal Differences of Pollution Characteristics and Potential Sources of PM<sub>2.5</sub> of Zhenjiang*. *Environmental Engineering*, 2019. 37(06): p. 123–130.