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RESEARCH ARTICLE

Study on the adsorption of nitrogen and phosphorus from biogas slurry by NaCl-modified zeolite

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Abstract

A NaCl-modified zeolite was used to simultaneously remove nitrogen and phosphate from biogas slurry. The effect of pH, contact time and dosage of absorbants on the removal efficiency of nitrogen and phosphate were studied. The results showed that the highest removal efficiency of NH₄⁺-N (92.13%) and PO₄³⁻-P (90.3%) were achieved at pH 8. While the zeolite doses ranged from 0.5 to 5 g/100 ml, NH₄⁺-N and PO₄³⁻-P removal efficiencies ranged from 5.19% to 94.94% and 72.16% to 91.63% respectively. The adsorption isotherms of N and P removal with NaCl-modified zeolite were well described by Langmuir models, suggesting the homogeneous sorption mechanisms. While through intra-particle diffusion model to analyze the influence of contact time, it showed that the adsorption process of NH₄⁺-N and PO₄³⁻-P followed the second step of intra-particle diffusion model. The surface diffusion adsorption step was very fast which was finished in a short time.

Introduction

Piggery wastewater is well known for its high concentration of organic matters(OM), nitrogen (N) and phosphorus (P)[1,2]. Because of low biomass production and low treatment cost, Anaerobic digestion had been widely used as an efficiency treatment. While biogas slurry are still rich in nitrogen (NH₄⁺-N > 400 mg/L) and phosphorus (PO₄³⁻-P >100 mg/L)[3,4]. Usually biogas slurry are directly used to irrigate the land filed directly to improve soil fertility and consequently be bebeficial to crop growth [5]. Meanwhile, excessive land application will lead nutrient loss from soil to water through runoff and leaching which could cause adverse consequences to groundwater[6,7]. Therefore, an effective and economical method to removal of N and P from biogas slurry is necessary.

Instead land application, adsorption is considered to be a simple and effective technique for the removal of nutrients from the wastewater. To remove organic contaminations from



wastewater, some commonly used adsorbents are activated carbon [8,9], clay minerals [10,11], chemical amendment[12,13], biochar[14,15], and zeolites[16-18]. Among those adsorbents, zeolites are widely applied for the purpose of reusing the effluent water, reducing pollution of water resources, reducing gas emissions through modifying the physiochemical properties of manure and decreasing water consumption[19-21]. An Australian zeolite with iron-coating and without iron-coating were used to remove Pb, Cu, Cd, Cr and Zn from aqueous solutions in batch and column experiments. Results showed that with pH 6.5, the Langmuir adsorption capacities of those five heavy metals ranged from 5.0-11.2 mg/g for single metal, while 3.7-7.6 mg/g for mixed metals solution [22]. Similar results were obseved by Egashira et al. They used Mongolian natural zeolites to adsorb Cu, Zn and Mn from model aqueous wastewater (pH 3-5), and found out that the adsorption capacities of 8.32-10.24 mg/g for Cu, 9.1-54.6 mg/g for Zn, 6.05-11.05 mg/g respectively for Mn [23]. Lin et al studied the effect of natural zeolite on the removal of P and NH₄⁺-N from orthophosphate and ammonium-nitrogen laden wastewaters at pH 3-11 in batch and continuous tests. The results showed that the highest removal rate (98.9% for P and 68.9% for NH₄⁺-N) was reached at pH 9.3[24]. Chen et al. used nano-zeolites synthesized from fly ash (ZFA) was used to simultaneously remove ammonium (N) and phosphate(P) in anaerobically digested swine wastewater [25]. N and P removal efficiencies ranged from 41% to 95% and 75% to 98%, respectively with a range of ZFA doses from 0.25 to 8 g/100 mL. The adsorption capacity is related to the wastewater properties as well as the colloidal properties and negatively-charged layers of zeoloties.

By evaluation parameters such as pH, adsorbent dosage and initial concentration of zeolite, the aim of this work was to investigate the impact of NaCl-modified zeolite on the removal of nitrogen and phosphorus from biogas slurry. The adsorption isotherms were adjusted to the models of Langmuir and Freundlich. Kinetic models of adsorption were used to analyze the kinetics and the zeolite adsorption mechanisms on the adsorbents.

Materials and method

My study did not involve human participants, specimens or tissue samples, or vertebrate animals, embryos or tissues:

- 1. We state clearly that no specific permissions were required for these locations/activities, and provide details on why this is the case;
- 2. We confirm that the field studies did not involve endangered or protected species.
- 3. We confirm that the authors had received approval from the COFCO Corporation to collect samples from the pig treatment plant.

About the name and product number of the zeolite acquired from the Wuhan rhyme siphon water treatment material limited company was the natural zeolite where there were no specific product number. They sell the zeolite in bulk and we just obtained some from them to do the research.

The natural zeolite used in this research was obtained from Wuhan rhyme siphon water treatment material limited company in Wuhan, China. The chemical composition of the natural zeolite is given in Table 1. The particle size of the used zeolite was < 0.15 mm. The zeolite

Table 1. The chemical composition of zeolite (n = 3, SD).

Components	SiO ₂	Al ₂ O ₃	Fe ₂ O ₃	K₂O	CaO	MgO	Na ₂ O	other
wt%	69.58±1.30	12.20±0.50	0.87±0.08	1.13±0.05	2.59±0.20	0.13±0.02	2.59±0.10	10.91±0.13

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Table 2. The charateristic of biogas slurry (n = 3, SD).

Item	рН	COD _{Cr} (mg/L)	NH ₃ -N (mg/L)	PO ₄ ³⁻ -P (mg/L)	VFA (mg/L)	EC (μs/cm)
Mean-value	8.09±0.30	2113.33±12.25	708.43±8.35	21.62±1.50	218.95±5.21	8607±15.31

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was fully washed several times with deionized water and dried for 24 h at 100°C. Natural zeolite was mixed with the NaCl solution in the ratio of 1:50 (g zeolite:mL NaCl). A magnetic stirrer was used to mix the sample mixture solution for 24 h at 150 rpm using a horizontal shaker at 30°C. Biogas slurry was collected from a pig treatment plant built by COFCO Corporation, located in Xinzhou, Wuhan, China (114°77′44.41"E, 30°54′12.84"N). Table 2 presents the chemical composition of the biogas slurry used in the experiment. The samples of biogas slurry were stored at 4°C until utilized.

The chemical composition of the zeolites was determined by energy dispersive X-ray spectros-copy (EDS) (SEA1000A, Japan). The mineralogy of the zeolite was determined using a XRD Shimadzu S6000 (Japan) diffractometer on powder samples of the zeolite. The X-ray diffractometer was equipped with a Cu target operated at 40 kV and 30 mA with a setting of 0-80° (40 min), step time 2°/min. Scanning electron microscopy (SEM), Fourier transform infrared (FTIR) spectroscopy were also inspect the structure of zeolite. TS (total solid), NH₄⁺-N, PO₄³⁻-P and total alkalinity (titrated to pH 4.3) were tested according the Standard Methods[26]. Batch experiments were carried out to obtain the adsorption data relative to contact time, sorptive concentration, pH, and dose of zeolite. pH was adjusted by the addition drops of strong HNO₃/NaOH solution. Then kept constant during the whole adsorption experiments. Experiments used to determine the equilibrium time were performed with the contact time between adsorbent and adsorbate in the range from 10 min to 24 h, at pH 8.1, with an initial concentration of NH₄⁺-N (708.4mg/L) and PO₄³⁻-P(21.62mg/L). The influences of adsorbent dosage (0.5, 1.0, 1.5, 2.0, 3.0, 4.0 and 5.0 g of adsorbent) and of pH (6.0, 7.0, 8.0,9.0 and 10.0) on the removal of nutrients by NaCl-zeolite were studied. All adsorption experiments were performed in triplicate. The removal rate of nitrogen and phosphorus was calculated as Eq (1):

$$\eta = \frac{C_0 - C_e}{C_0} \times 100\% \tag{1}$$

Where η is the removal rate (%), C_0 is the initial concentration of nitrogen and phosphorus (mg/L), C_e is the concentration of nitrogen and phosphorus at equilibrium (mg/L).

Adsorption isotherms model

Adsorption isotherms is necessary to describe the equilibrium relationships between the amounts of ion exchanged by zeolite and its equilibrium concentration in the solution which could be helpful for the analysis and the design of the sorption systems. In this study, two adsorption isotherms were developed by Langmuir model (Eq.(2)) and Freundlich model (Eq.(3)).

$$\frac{C_e}{q_e} = \frac{1}{q_{\text{max}} k_L} + \frac{C_e}{q_{\text{max}}} \tag{2}$$

$$\log q_e = \log k_F + \frac{1}{n} \log C_e \tag{3}$$

Where k_L is the Langmuir isotherm constant (L/mg), \mathbf{q}_{max} represent the maximum



ammonium ion-exchange capacity of zeolite (mg/g). k_F is the Freundlich isotherm constant that indicates the maximum adsorption capacity(mg/g), 1/n is the Freundlich isotherm constant which is dimensionless. The n value indicates the degree of nonlinearity between solution concentration and adsorption; when n=1, n<1 and n>1 is adsorption linear, chemical and physical process, respectively. The n values of within the range of 1–10 represent good adsorption. C_e is the equilibrium concentration of the adsorbed substance in the liquid phase (mg/L) and q_e is the constant that indicates the maximum adsorbate quantity of the adsorbent (mg/g).

Statistical analysis

The data was analysed by SPSS 15.0 using regression analysis. The goodness of correlation was evaluated with the correlation coefficient \mathbb{R}^2 .

Results and discussion

Zeolite characterization

The characteristics of the nature zeolite (NZ) and modified zeolite (MZ) were displayed based on the results of XRD and XRF analyses. The XRD patterns and the elemental analysis of the NZ and MZ were shown in Fig 1 and Table 3, respectively. The XRD spectra indicated that the main mineral species of the zeolite before and after modification remained unchanged. The main composition of NZ and MZ were clinoptilolite, quartz and ferrosilite, illustrating the absence of structural degradation during modification. The content of cations like Na $^+$, K $^+$, and Ca $^{2+}$ in zeolites determines the exchange ability and adsorption. The more cations Na $^+$, K $^+$, and Ca $^{2+}$, the stronger the ion-exchange ability. While Si/Al ratio influences the thermal and physical stability. Zeolites with higher Si/Al exhibits a very high physico-chemical

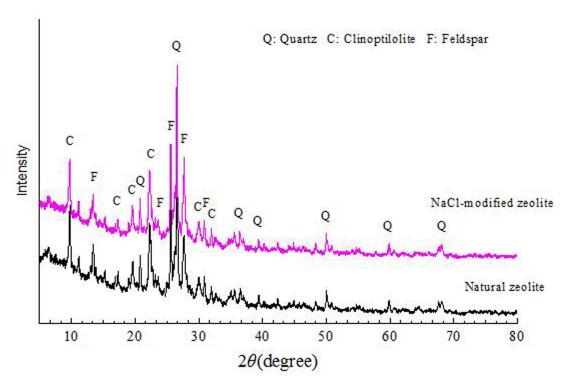


Fig 1. XRD patterns of the natural and NaCl-modified zeolites.

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Oxide	NZ	MZ
Al_2O_3	15.14	12.75
SiO ₂	72.85	77.58
K₂O	1.71	1.09
CaO	2.55	0.88
TiO ₂	0.13	0.09
Na ₂ O	3.53	6.04
MnO	0.05	0.04
Fe ₂ O ₃	0.56	0.58
others	3.48	1.13

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durability. The improvements of chemical compositions through modification of zeolites such as substitution of some (Si) and aluminium (Al) (or other metals) with cations Na⁺, K⁺, and Ca²⁺ lead to a negative charge on the framework which will increase the ion exchange ability of zeolites, especially for Na⁺ to the ion-exchange with NH₄⁺. The XRF results showed that after NaCl modification, the contents of the exchangeable cations such as Al³⁺, Ca²⁺, and K⁺ decreased, while the Si⁴⁺, Na⁺ amount increased significantly. It demonstrated that Al³⁺, K⁺, and Ca²⁺ were replaced by Na⁺ which was benefit for the removal of the other cations especially NH₄⁺ [27]. Fig 2 showed the FTIR spectra of NZ and MZ in the range of 500–4000 cm⁻¹. The typical FTIR spectra for zeolite bands was observed at the region below 1700 cm⁻¹. In this region, the bands from the zeolites backbones (Si–O–Si and Si–O–Al) which was composed of bending and stretching of Si–O–Si, Si–O and Al could be observed. Adsorption bands occurred at 3500 cm⁻¹ and 1639 cm⁻¹ were associated with–OH stretching and bending

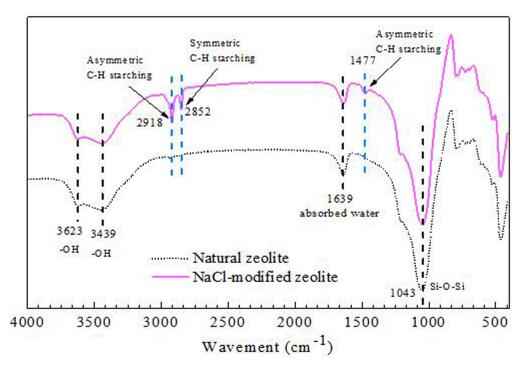


Fig 2. FI-IR spectra obtained for natural zeolite and NaCl-modified zeolite.

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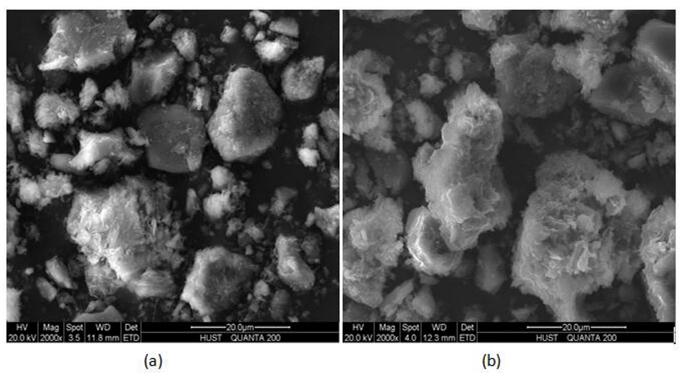


Fig 3. Scanning electron microscope (SEM) images of natural zeolite(a) and NaCl-zeolite (b).

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frequencies, respectively. A peak obtained from 2852 and 2918 cm⁻¹ was significantly increased after modification with NaCl. These results appeared that the successful grafting of Na⁺ onto the surface hydroxyl groups of zeolite which was verified by the results of XRD.

Scanning Electron Microscope (SEM) images of natural zeolite and NaCl-modified zeolite were shown in Fig 3(a) and 3(b). Before modification, natural zeolite had an small pore size. After modification, the structure of the zeolite was looser than that of the natural zeolite. There were obvious changes in its original size, shape, crumb structure, and pore structure after NaCl modified which would improve adsorption capacity.

Optimal conditions for NH₄⁺-N and PO₄³⁻-P removal by modified zeolite

pH is an important parameter in the batch adsorption study which will make influence on the adsorption process by changing the surface charge distribution of adsorbents used. While the dosage of zeolites can make influence on the adsorption process through changing unsaturation of the ion-exchange sites. A high adsorbent dosage can effectively decrease the unsaturation of the ion-exchange sites resulting in a lower ion exchange capacity[28].

Effect of pH. To determine the optimal pH for NH_4^+ -N and PO_4^{3-} -P removal, the ion-exchange performance of the MZ were investigated at different pH(from 6 to 10, (Fig 4(a)).

Result showed that MZ had the best removal efficiencies of $\mathrm{NH_4}^+$ -N(92.13%) and $\mathrm{PO_4}^{3-}$ -P (90.3%) were both occurred at pH 8. A higher or lower pH both would lead to negative results. The results were similar to Thornton and Marañón [29,30]. They proved an optimal range of pH 6–7 for $\mathrm{NH_4}^+$ -N remove by zeolite. However, Saltalı et al reported that the optimum $\mathrm{NH_4}^+$ -N remove efficiency was achieved at pH 8[31]. A favourable range of pH 4–8 for PO $_4^{3-}$ -P remove by La/Al-modified zeolite was observed by Meng, while the PO $_4^{3-}$ -P remove efficiency incressed with the decrease of pH varied from 3.0–10.0[32,33]. The inconsistency in the

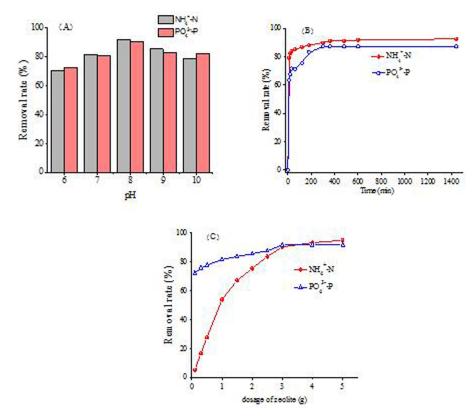


Fig 4. Optimal conditions for nutrients removal. (A)Effect of pH. (B) Effect of reaction time on removal efficiency of NH_4^+ -N and $PO_4^{3^-}$ -P by NaCl-modified zeolite (1 g/100 ml). (C) Effect of dosage on removal efficiency of NH_4^+ -N and $PO_4^{3^-}$ -P by NaCl-modified zeolite (1 g/100 ml).

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optimal pH may be attributed to the differences in the chemical compositions of the zeolites tested

In the ion-exchange process, dissociated $\mathrm{NH_4}^+$ will make the ion exchanged with Na^+ in the zeolite. While the amount of $\mathrm{NH_4}^+$ in the solution was depended on the pH of the solution. When pH is below 8, $\mathrm{NH_4}^+$ ion concentration in solution rose with decrease of pH value, while H⁺ concentration increased with the drop in the pH which would had a negative influece on the $\mathrm{NH_4}^+$ exchange [31]. When pH is over 8, $\mathrm{NH_4}^+$ was converted into $\mathrm{NH_3}$ which not be exchanged on the zeolites [34]. The same as $\mathrm{PO_4}^{3-}$ -P remove efficiency, at relatively high pH values, $\mathrm{OH^-}$ concentration increased and competed with $\mathrm{PO_4}^{3-}$ -P on the adsorbent, leading to a lower phosphorous adsorption rate.

Effect of reaction time. Fig 4(b) showed the effect of the contact time on the removal efficiency of NH_4^+ -N and PO_4^{3-} -P in biogas slurry at the dosage of 1 g zeolite/100 ml. It could be seen that both the NH_4^+ -N and PO_4^{3-} -P removal efficiency increased with the contact time.

The concentration of $\mathrm{NH_4}^+$ -N and $\mathrm{PO_4}^{3-}\mathrm{P}$ decreased rapidly during the first 0.5 h. The removal efficiency was 83.88% for $\mathrm{NH_4}^+$ -N and 71.49% for $\mathrm{PO_4}^{3-}\mathrm{P}$. Then in the following 5.5 h, corresponding $\mathrm{NH_4}^+$ -N and $\mathrm{PO_4}^{3-}$ -P removal efficiencies increased to 91.24% and 87.33%, respectively. Within 24 h, the concentration of $\mathrm{NH_4}^+$ -N decreased from 708.43 to 68.24 mg/L (92.58%) while the concentration of $\mathrm{PO_4}^{3-}$ -P decreased from 21.62 to 1.81 mg/L (87.33% removal). However, with a further increase in the contact time from 30 min to 360 min, the $\mathrm{NH_4}^+$ -N removal efficiency increased slightly and then reached equilibrium after 360 min. This behavior could be attributed to the quick utilization of the most readily available



adsorbing sites of the zeolites, leading to a fast diffusion and rapid attainment of equilibrium [35]. The increasing trend on nutrient sequestration over time (mainly within the first 6 h) is mainly related to the combined effect of dissociated Na₂O and exchangeable ion on the zeolite surface (Table 3).

In order to analyze the adsorption performance of NH_4^+ -N and PO_4^{3-} -P removal by zeolite, intra-particle diffusion model with three-linear regions was used according Eq.(4):

$$q_t = k_d t^{1/2} + C (4)$$

where k_d is the coefficient of intra-particle diffusion (mg g⁻¹ h^{-1/2}), t is the contact time(h). Table 4 showed the results of kinetic parameters for ammonium and phosphate removal using the MZ. It indicated that the process of adsorption was more than one-step which could be validated by the prior experiments. It also conclued that surface diffusion adsorption process was very fast which was finished in a short time. The adsorption process of NH_4^+ -N and PO_4^{3-} -P followed the second step of intra-particle diffusion model: surface diffusion adsorption process and particle diffusion adsorption process. Meanwhile the k_{d1} value for NH_4^+ -N was higher than k_{d2} and k_{d3} values which indicated that ammonium removal by zeolite was proposed as monolayer molecular adsorption with zeolite[36]. The k_{d1} and k_{d2} values for PO_4^{3-} -P was higher than k_{d3} value which indicated that PO_4^{3-} -P removal with zeolite was claimed as the electron exchange between phosphorus and the zeolite surface[37]. Moreover the k_d values for NH_4^+ -N were both much higher than those for PO_4^{3-} -P, indicating that adsorption phase of P was very poor in the present zeolite, PO_4^{3-} -P and removal occurred most probably in the external boundary layer film of liquor surrounding the zeolite particles [38].

Effect of zeolite dosages. Fig 4(c) showed the effect of adsorbent dosage on the NH₄⁺-N and PO₄³⁻-P removal. Both NH₄⁺-N and PO₄³⁻-P removal efficiency increased with the increase of the adsorbent dosage. The removal rates of NH₄⁺-N increased from 5% to 95% with the increase of zeolite doses from 0.1 to 5.0 g/100 ml. The PO₄³⁻-P removal rates also increased from 72% to 91% with zeolite doses ranging from 0.1 to 5.0 g/100 ml. Meanwhile when the dosage of zeolite increased from 0.1 to 3.0 g/100 ml, the NH₄⁺-N removal efficiencies increased quickly from 5% to 90.37% compared with the PO₄³⁻-P removal rates increasing from 72% to 91.63%. That was because the increasing amount of the adsorbent increased the surface area and the number of ion-exchange sites on zeolite. A negligible increase of NH₄⁺-N and PO₄³⁻-P removal efficiency occurred when the dosage was higher than 3.0 g/100 ml. This could be due to the fact that a high-adsorbent dosage can effectively decrease the unsaturation of the ion-exchange sites of the zeolites, and consequently, the number of such sites per unit mass gets reduced, resulting in comparatively lesser ion exchange at higher adsorbent amounts. Considering the nutrient removal efficiency and economic cost, the optimum zeolites dose was 3.0 g/100 ml for the kind of zeolite on the removal of NH₄⁺-N and PO₄³⁻-P.

Table 4. Kinetic parameters for ammonium and phosphate removal using MZ.

Adsorption stage	Parameters	Intra-particle diffusion model			
		Ci	k_{ip} (mg g ⁻¹ h ^{-1/2})	R _i ²	
first step	NH ₃ -N	0.2157	1.0968	0.8734	
	PO ₄ ³⁻ -P	0.0067	0.0350	0.8807	
second step	NH ₃ -N	4.4353	0.0452	0.9250	
	PO ₄ ³⁻ -P	0.1422	0.0010	0.9114	
third step	NH ₃ -N	5.2200	0.0031	0.3440	
	PO ₄ ³⁻ -P	0.1583	0.0002	0.8915	

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Table 5. Isotherms parameters of the Langmuir and Freundlich models.

Object	Langmuir parameter		Freundlich parameter			
	q _{max} (mg/g)	k _L (L/mg)	R ²	k _F (mg/g)	1/n	R ²
NH ₄ ⁺ -N	11.25	0.035	0.999	1.26	0.36	0.905
PO ₄ ³⁻ -P	6.67	0.79	0.992	3.1	0.31	0.990

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Table 6. Comparison of Langmuir sorption capacity for NH₄⁺-N and PO₄³⁻-P with various adsorbents.

Adsorbent	Adsorption c	apacity (mg/g)	Conditions	Reference	
	NH ₄ ⁺ -N	PO ₄ ³⁻ -P			
Bismuth impregnated biochar		1.48	pH 3.0, 318 k	[39]	
Thermal activated sepiolite	2.93		pH 8.0, 303k	[40]	
Raw fly ash	2.23	4.1	pH 7–8, 293k	[41]	
Natural zeolite	3.45		pH 7, 298k	[42]	
Chinese clinoptilolite	2.7–3.2		pH 6.0, 293k	[43]	
NaCl- zeolite	11.25	6.67	pH 8.0, 298k	Present study	

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Adsorption isotherms and kinetic modelling. The parameters of two models calculated from the slope and the intercept of the plots are given in Table 5. Both two models could well described the adsorption isotherms process while the Langmuir model provided a slightly more consistent t to the data (R^2 :0.999 of NH_4^+ -N and 0.992 of PO_4^{3-} -P) as compared with the Freundlich model (R^2 :0.905 of NH_4^+ -N and 0.990 of PO_4^{3-} -P).

In the present study, n values obtained between 2.78 and 3.23 which indicates a adsorption of $\mathrm{NH_4}^+$ -N, and $\mathrm{PO_4}^{3-}$ -P onto MZ. Langmuir adsorption capacity of $\mathrm{NH_4}^+$ -N, and $\mathrm{PO_4}^{3-}$ -P by MZ were compared with those of various low cost adsorbents as shown in Table 6. The maximum ion-exchange capacities of MZ at equilibrium ($\mathrm{q_{max}}$) were 11.25 and 6.67 mg/g respectively for $\mathrm{NH_4}^+$ -N and $\mathrm{PO_4}^{3-}$ -P. The differences of sorption capacity between various sorbents were caused by the difference in physico-chemical properties of adsorbents and experimental factors including the concentration range of $\mathrm{NH_4}^+$ -N, and $\mathrm{PO_4}^{3-}$ -P, pH, temperature, etc. Higher value of Langmuir constant and Freundlich adsorption capacity further reflect the improved strength and affinity of MZ for $\mathrm{NH_4}^+$ -N, and $\mathrm{PO_4}^{3-}$ -P.

Adsorption for nutrients removal is a simple and the most economical method for wastewater treatment. The price and regeneration method are the main points when chosen as adsorbent. Modified NaCl-zeolite has serval traits that can be an ideal adsorbent.1) in comparison to the conventional adsorbent, it is very efficient. 2) it can be regenerated by chlorination regeneration easily. 3) the material is inexpensive, reducing the cost of wastewater treatment. 4) it does not contaminate the wastewater.

Conclusion

In this research, simultaneous removal of $\mathrm{NH_4}^+$ -N and $\mathrm{PO_4}^{3^-}$ -P from the effluents of biogas plants was investigated using NaCl-modified zeolite as an adsorbent material. The NaCl-modified zeolite showed a good absoprtion capacity to the $\mathrm{NH_4}^+$ -N and $\mathrm{PO_4}^{3^-}$ -P which indicated that the proposed process for the treatment of digested swine wastewater is feasible. The condition of adsorption of $\mathrm{NH_4}^+$ -N and $\mathrm{PO_4}^{3^-}$ -P onto NaCl-modified zeolite was pH 8.0, 3.0 g/100 ml and 6 h contact time. The adsorption isotherm of $\mathrm{NH_4}^+$ -N and $\mathrm{PO_4}^{3^-}$ -P onto the adsorbent was well fitted to the Langmuir model and the maximum adsorption capacity (Qm) was 11.25 mg/g and 6.67 mg/g respectively.



Supporting information

S1 Fig. Fig 1 XRD patterns of the natural and modi ed zeolites. (XLSX)

S2 Fig. Fig 2 FI-IR obtained for natural zeolite and NaCl-zeolite. (XLSX)

S3 Fig. Fig 4 Optimal conditions for nutrients removal. (A)Effect of pH. (B) Effect of reaction time on removal efficiency of $\mathrm{NH_4}^+$ -N and $\mathrm{PO_4}^{3-}$ -P by NaCl-modified zeolite (1 g/100 ml). (C) Effect of dosage on removal efficiency of $\mathrm{NH_4}^+$ -N and $\mathrm{PO_4}^{3-}$ -P by NaCl-modified zeolite (1 g/100 ml). (XLSX)

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Author Contributions

Data curation: SC.

Formal analysis: YLX.

Investigation: HXL YHL.

Software: FD.

Writing - original draft: QPC.

Writing - review & editing: JFL.

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