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# *Article* **Vertical Distribution and Controlling Factors Exploration of Sc, V, Co, Ni, Mo and Ba in Six Soil Profiles of The Mun River Basin, Northeast Thailand**

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**Abstract:** Exploring the enrichment and controlling factors of heavy metals in soils is essential because heavy metals can cause severe soil contamination and threaten human health when they are excessively enriched in soils. Soil samples (total 103) from six soil profiles (T1 to T6) in the Mun River Basin, Northeast Thailand, were collected for the analyses of the content of heavy metals, including Sc, V, Co, Ni, Mo, Ba. The average contents of soil heavy metals decrease in the following order: Ba, V, Ni, Sc, Co, and Mo (T1, T3, T4 and T5); Ni, V, Ba, Co, Sc, Mo, and Ba (T2); Ba, V, Sc, Ni, Mo, and Co (T6). An enrichment factor (EF) and geoaccumulation index were calculated to assess the degree of heavy metal contamination in the soils. The EFs of these heavy metals in most samples range from 0 to 1.5, which reveals that most heavy metals are slightly enriched. Geoaccumulation indexes show that only the topsoil of T1 and T2 is slightly contaminated by Ba, Sc, Ni, and V. Soil organic carbon (SOC), soil pH and soil texture are significantly positively correlated with most heavy metals, except for a negative correlation between soil pH and Mo content. In conclusion, the influence of heavy metals on soils in the study area is slight and SOC, soil pH, soil texture dominate the behavior of heavy metals.

**Keywords:** soil heavy metals; enrichment factor; geoaccumulation index; Mun River Basin; Northeast Thailand

#### **1. Introduction**

Soil is a huge sink for heavy metals derived from parent rocks and anthropogenic activities, most of which will cause serious contamination once enriched in soils [\[1\]](#page-10-0). Because air, water and rocks interface in soil environment, the heavy metals in soils are transported by various carriers, which may lead to the enrichment of heavy metals in some areas [\[2–](#page-10-1)[4\]](#page-10-2). The Mun River Basin is located in Northeast Thailand, which is one of the largest agricultural regions in Thailand [\[5\]](#page-10-3). The enrichment of toxic heavy metals will lower soil fertility and increase the input of these heavy metals to the food chain, which may be adverse to human health [\[2\]](#page-10-1). Therefore, it is highly significant to study the distribution characteristics of heavy metals in the Mun River Basin.

The heavy metals Co, Ni, Ba and V are considered as important indicators of soil toxicology, variability in soil composition and human health risks [\[6\]](#page-10-4). With the urbanization and industrialization in the Mun River Basin, these heavy metals are frequently emitted out from tire abrasion, industrial production, and fertilizer application and can easily accumulate in soils [\[6\]](#page-10-4). The Mun River Basin is an important agricultural region in Thailand and large amounts of fertilizers are used, so soils in the basin are easily polluted, making it essential to explore the distributions of these heavy metals in soil profiles. Scandium (Sc) has drawn great attention due to its potential contamination to the environment and the risks to cause various diseases with its wide application in medicine and technology [\[7\]](#page-10-5), so it is significant to explore the potential pollution of Sc emitted out from factories and hospitals across the basin. In the last three decades, considerable urbanization and industrialization have occurred in Thailand, which may cause soil contamination of these heavy metals by the combustion of fossil fuel, emission of vehicle exhaust, etc. [\[6,](#page-10-4)[8\]](#page-11-0). Although the concentrations of molybdenum (Mo) are usually quite low in soils and rocks, Mo is essential for plants, animals and humans [\[9\]](#page-11-1). Relative to soil contamination problems, Mo deficiency has attracted more attention because it will largely reduce crop yield [\[10\]](#page-11-2). Mo deficiency in the crops grown in Northeast Thailand has been reported, making it necessary to study the distribution of Mo there [\[11](#page-11-3)[,12\]](#page-11-4). Moreover, as the most important agricultural region in Thailand, heavy metal contamination or deficiency would affect human health through the food chain. Therefore, these six heavy metals (Sc, V, Co, Ni, Mo, Ba) were selected to help to assess the soil condition and provide the basis for the strategies of agricultural management.

The mobility of soil heavy metals is controlled by various factors, such as soil pH, soil organic matter (SOM) and soil texture [\[2,](#page-10-1)[13\]](#page-11-5). Generally, the more mobile heavy metals are, the more available they are for plants, which increases the risk that they enter the food chain of humans. The relationships between these heavy metals and soil properties may vary with different soil conditions, for example, the adsorption of Mo on clay is regulated by soil pH [\[9\]](#page-11-1). It is therefore necessary to explore the relationships between heavy metals and soil properties to better understand the distributions of these heavy metals. This study aims to: (1) learn about the spatial distributions of these heavy metals in six soil profiles; (2) assess the risks of soil heavy metal contamination; and (3) explore the controlling factors of soil heavy metal behaviors.

#### **2. Materials and Methods**

#### *2.1. Study Area*

This study was conducted in the Mun River Basin, the largest river basin (82,000 km<sup>2</sup>) in Northeast Thailand  $(14°00'~16°00' N, 101°30'~105°30' E)$ . The Mun River Basin, of which the elevation ranges from 100 to 1065 m (Figure [1\)](#page-2-0), is dominated by a tropical savanna climate which leads to a dry season and a rainy season all year round [\[14\]](#page-11-6). Most crops are cultivated during the rainy season (April to November) when 85%–90% of the annual precipitation falls [\[15\]](#page-11-7). Alluvial and sandy soils in the basin are heavily leached and have low organic matter contents, leading to the frequent use of fertilizers [\[15,](#page-11-7)[16\]](#page-11-8). Land cover types are various in the basin, including crops, forest, grass, water, etc., and most forest lands are distributed near the southwestern and southeastern border of the basin (Figure [1\)](#page-2-0). In the Mun River Basin, shale, siltstone, sandstone, and conglomeritic sandstone are the main rock types of the rock basement which is covered by the Quaternary alluvial sediment [\[17\]](#page-11-9).

#### *2.2. Sampling and Analysis*

Soil samples (total 103) were collected in March 2018, from six soil profiles in the Mun River Basin. These soil profiles were dug on paddy lands (T1, T3), forest lands (T2, T6), wetlands (T4) and built-up lands (T5), respectively. Soil samples were collected at intervals of 10 cm and the number of soil samples was 22, 5, 10, 20, 41, and 5 in the six soil profiles, respectively. These soil profiles are chosen because they are under the main land-use types in the Mun River Basin and are distributed across the basin (Figure [1\)](#page-2-0). The distribution of heavy metals in these profiles would reflect the general condition of heavy metals in the Mun River Basin. These soil profiles are described in detail in Table [1.](#page-2-1)

<span id="page-2-0"></span>

from the website: https://[opentopography.org](https://opentopography.org/start)/start). **Figure 1.** Map of elevation (**a**) and land cover types (**b**) in the Mun River Basin (Data of elevation come

<b>Sampling Point</b>	<b>Land Cover</b>	Depth (cm)	<b>Field Description</b>
T1	Crop	220	Relatively uniform at all depths
T <sub>2</sub>	Oak forest	50	Compact soil; brown red
T3	Crop	100	0–12 cm: Black root layer $12-100$ cm: Fine silt >100 cm: Iron manganese nodules
T4	Grass	193	0–20 cm: Root layer $20-112$ cm: Fine silt 112–180 cm: Weathering crust 180-193 cm: Gray and green sand >193 cm: Bedrock
T <sub>5</sub>	Artificial building	405	$0-105$ cm: Fine silt 105–205 cm: Iron manganese nodules 205-405 cm: Clay layer
T <sub>6</sub>	Forest	50	Fine silt

<span id="page-2-1"></span>**Table 1.** General description of the sampling information.

Soil samples were air-dried at 25  $^{\circ}$ C and unwanted materials such as stones and plant roots were removed by passing soil samples through a 2 mm sieve  $[18,19]$  $[18,19]$ . Soil samples were ball milled thoroughly to pass a 200-mesh sieve ( $<$ 75  $\mu$ m) with the Retsch MM400 (Retsch GmbH, Haan, Germany), a widely used grinding machine [\[20\]](#page-11-12). 100 mg soil powders were digested in Teflon beakers using ultra-pure HNO<sub>3</sub> (70%) and ultra-pure HF (40%) (3:1) at 120 °C for seven days [\[21\]](#page-11-13). If some organic solid residue remains undissolved, aqua regia (HNO<sub>3</sub>: HCl, 1:3) would be used to remove organic matters. When all samples were dissolved, the digestion process was repeated twice using 50 ml ultra-pure HNO<sub>3</sub> (70%) to break down potential fluorine compounds [\[21\]](#page-11-13). After these treatments, the prepared samples were diluted in ultra-pure  $HNO<sub>3</sub>$  (2%) of about 50 ml. Finally, the concentration of the heavy metals Co, Ni and Mo were measured by ICP-MS (Elan DRC-e, Perkin Elmer, Waltham, MA, USA) and the concentrations of the heavy metals Ba, Sc, V and Ti were measured by ICP-OES (Optima 5300DV, Perkin Elmer, Waltham, MA, USA).

#### *2.3. Data Treatment*

The enrichment factor (EF) index and the geoaccumulation index (*Igeo*) are two important methods to assess the enrichment and contamination of heavy metals in soils [\[22,](#page-11-14)[23\]](#page-11-15). The upper continental crust (UCC) is usually chosen as the reference material for the calculation of both indexes [\[1\]](#page-10-0). All background values were provided by Wedepohl [\[24\]](#page-11-16).

For the enrichment factor, Ti was used as a reference element due to its characteristic of low occurrence variability [\[25\]](#page-11-17). The enrichment factor was calculated according to the following formula [\[26,](#page-11-18)[27\]](#page-11-19):

$$
EF = ([M]/[Ti])_S / ([M]/[Ti])_{UCC},
$$

where M represents heavy metal concentrations and S refers to soil samples.

Based on the EF values, the enrichment of heavy metals can be categorized into five grades: minimal enrichment ( $EF \le 2$ ); moderate enrichment ( $2 < EF < 5$ ); significant enrichment ( $5 \le EF < 20$ ); high enrichment (20  $\leq$  EF  $<$  40); and extremely high enrichment (EF  $\geq$  40) [\[28\]](#page-11-20).

The geoaccumulation index was calculated as follows [\[29\]](#page-11-21):

$$
I_{geo} = \log_2(C_n/1.5B_n),
$$

where  $C_n$  represents the concentration of the heavy metal n in soil samples;  $B_n$  represents the concentration of the heavy metal n in reference material (UCC); and the factor 1.5 is used to eliminate the lithogenic effects [\[1\]](#page-10-0). According to the *Igeo* values, soil contamination can be categorized into seven classes from uncontaminated (*Igeo* < 0) to very highly contaminated (*Igeo* > 5) [\[22\]](#page-11-14).

#### *2.4. Statistical Analysis*

In this study, Pearson correlation analysis was applied to analyze the relationship between SOC, soil pH, clay and soil heavy metals. Principal component analysis (PCA) was applied to identify the sources of soil heavy metals. All mathematical analyses were done by SPSS 20 (Statistical Package for Social Science, IBM, Armonk, NY, USA). Figure [1](#page-2-0) was produced by ArcMap 10.0 software and other figures were produced by OriginPro 9.0 software (OriginLab Corporation, Northampton, MA, USA).

#### **3. Results and Discussion**

#### *3.1. Contents of Heavy Metals in Soil*

The profile distributions of six heavy metals (Sc, V, Co, Ni, Mo, Ba) in six profiles under four land-use types are shown in Figure [2.](#page-4-0) The average contents of heavy metals in samples from soil profiles T1, T3, T4 and T5 decrease in the following sequence: Ba > V > Ni > Sc > Co > Mo (Table [2\)](#page-4-1). In soil profiles T2 and T6, these contents are in the following sequences respectively:  $Ni > V > Ba > Co$  $Sc > Mo$  and  $Ba > V > Sc > Ni > Mo > Co$  (Table [2\)](#page-4-1).

<span id="page-4-0"></span>

Figure 2. Vertical distributions of heavy metals, including Sc (a), V (b), Co (c), Ni (d), Mo (e) and Ba (f). (**f**). (Some data were omitted because of their small variation ranges). (Some data were omitted because of their small variation ranges).

Soil Profile	Item	Sc	V	Co	Ni	Mo	Ba
T1	Min	13.65	84.65	11.30	24.76	0.48	259.62
	Max	21.82	204.50	16.12	31.58	1.12	1659.68
	Mean (SD)	17.95(2.27)	123.47 (39.68)	13.70(1.50)	28.58 (1.97)	0.75(0.14)	764.44 (561.23)
T <sub>2</sub>	Min	12.54	124.26	16.15	102.61	1.77	8.41
	Max	23.40	176.18	18.98	146.22	2.61	49.94
	Mean (SD)	15.72 (4.39)	146.64 (19.79)	17.47 (1.28)	128.15 (21.88)	2.05(0.36)	20.62 (16.79)
T3	Min	1.76	1.26	0.49	2.05	0.17	12.57
	Max	6.76	105.87	5.15	7.14	1.42	324.63
	Mean (SD)	3.26(1.53)	18.22 (32.50)	1.45(1.36)	4.41(1.66)	0.77(0.41)	48.62 (97.07)
T4	Min	2.63	7.64	1.55	4.10	0.22	37.43
	Max	12.46	280.59	17.23	18.00	3.04	367.81
	Mean (SD)	6.89(3.52)	59.75 (62.29)	5.78(3.89)	10.72(5.10)	1.19(0.97)	188.94 (142.42)
T5	Min	4.83	12.72	0.55	2.53	0.19	2.75
	Max	13.88	437.43	22.55	58.55	1.76	534.58
	Mean (SD)	10.45(2.67)	111.98 (107.71)	10.05(6.45)	18.35 (15.25)	0.65(0.36)	202.71 (171.92)
T6	Min	4.38	26.37	0.24	1.80	0.29	13.86
	Max	6.66	51.76	1.97	6.01	2.85	89.26
	Mean (SD)	5.19(0.89)	34.75 (10.79)	0.67(0.74)	2.98(1.81)	1.02(1.05)	65.37 (29.51)

<span id="page-4-1"></span>**Table 2.** Statistical results for heavy metal contents (mg kg<sup>-1</sup>) of soil samples from six soil profiles.

Vertically, the contents of heavy metals generally increase with depth in most soil profiles, excluding profile T1, which is quite marked in Sc, V, and Co (Figure [2\)](#page-4-0). Most rainfall in the study

area is mainly concentrated in the rainy season [\[15\]](#page-11-7), which causes a large amount of precipitation in a short period and results in severe leaching of heavy metals down these soil profiles [\[30\]](#page-12-0). The vertical distributions of heavy metal concentrations can reflect the relative mobility of these metals in soil profiles [\[3\]](#page-10-6). Based on the distribution characteristics of these heavy metals in the six profiles (Figure [2\)](#page-4-0), it can be concluded that the mobility of heavy metals Sc, V, and Co is higher than the mobility of Ni, Mo, and Ba in these soil profiles. Soil profile T1 is dug in rice fields where fertilizer is frequently used. The higher concentrations of these heavy metals in the topsoil of profile T1 are caused by the applications of pesticide and fertilizer [\[23\]](#page-11-15). Huge positive peaks of heavy metal concentrations (V, Co, Mo) in soil profile T5 were observed at approximately 165–200 cm depth. According to the field recordings, there is a layer of iron-manganese oxides at 160–205 cm depth, where these positive peaks occur (Table [1\)](#page-2-1). The phenomenon is attributed to the adsorption of iron–manganese oxides on these heavy metals [\[31](#page-12-1)[,32\]](#page-12-2). Due to the surface complex formation and the existence of specific adsorption potentials, iron–manganese oxides own a relatively strong ability to absorb and concentrate heavy metals in soils [\[31,](#page-12-1)[33\]](#page-12-3). However, heavy metals Sc, Ni and Ba do not show positive peaks of concentrations in that layer, indicating that the influence of iron–manganese oxides on these heavy metals is limited. Potential reasons behind the phenomenon need to be further explored.

#### *3.2. Assessment of Soil Heavy Metal Contamination*

#### 3.2.1. Enrichment of heavy metals

The calculated EF factor values of most samples are lower than 2 (Figure [3\)](#page-5-0), which indicates the minimal enrichment of these heavy metals in soils [\[28\]](#page-11-20). Several extremely abnormal EF values occur in soil profile T5, exactly at the depth where abnormal heavy metal concentrations appear (Figure [2\)](#page-4-0), which can be attributed to the absorption by iron–manganese oxides. The sources of heavy metals in soil profiles comprise crustal materials (e.g. parent rocks) and non-crustal materials (e.g. anthropogenic sources). Due to lithogenic effects, an EF value = 1.5 is regarded as an important threshold for assessing whether human activities influence heavy metal concentrations in soils [\[34\]](#page-12-4). Except for the V in the soil profile T5 and Mo in the soil profile T6, the average EF values of these heavy metals in different soil profiles are lower than 1.5 which indicates that these heavy metals are mainly derived from natural processes [\[23](#page-11-15)[,35\]](#page-12-5). The EF values of V (2.2) and Mo (1.8) in soil profiles T5 and T6 are higher than 1.5, indicating that non-crustal materials are also important contributors for these heavy metals [\[35\]](#page-12-5). The EF values of Mo in soil profile T1 that was dug in paddy fields are lower than 0.5 which shows that the Mo deficiency may have occurred in local agricultural lands. Molybdenum fertilizer may be a good choice to make up the Mo deficiency in local agricultural lands [\[11\]](#page-11-3). *Int. J. Environ. Res. Public Health* **2020**, *17*, x 7 of 14

<span id="page-5-0"></span>

T3 (**c**), T4 (**d**), T5 (**e**) and T6 (**f**). **Figure 3.** Box plots of heavy metal EF values in six soil profiles, including soil profiles T1 (**a**), T2 (**b**), T3 (**c**), T4 (**d**), T5 (**e**) and T6 (**f**).

#### 3.2.2. Contamination Assessment by Igeo Values in Topsoil.

Heavy metal contamination was reported to be most obvious in topsoil (0–40 cm) [\[36\]](#page-12-6), so the contamination assessment should be concentrated on topsoil. The *I*<sub>geo</sub> values of these heavy metals in topsoils (0–10, 10–20, 20–30, 30–40 cm) of six soil profiles are presented in Figure [4.](#page-6-0) These values of most metals are lower than 0 at all depths of all soil profiles, except for Sc and V in soil profiles T1 and T2 (Figure [4\)](#page-6-0). The results show that the soils of the six soil profiles are not polluted by these heavy metals in the course of human activities. The positive *I*<sub>geo</sub> values of heavy metals mainly occur in the topsoil of soil profiles T1 and T2, which are located in a paddy field and oak forest, respectively. The *I*<sub>geo</sub> values of Ba and Sc in profile T1 and Ni, Sc, and V in profile T2 are roughly in the range of 0–1, indicating that the topsoil of these two soil profiles is slightly polluted by these heavy metals. Except for the surface soil layer (0–10 cm), the *I*<sub>geo</sub> values of V are in the range of 1–2, indicating the moderate contamination of V in the topsoil of soil profile T1. Heavy metals that may be adverse for human health, such as Ni and V, have been reported to accumulate in soils accompanied by the frequent application of fertilizers [\[37\]](#page-12-7), so a potential reason for the contamination is the application of fertilizers. Chemical fertilizers cause heavy metal pollution by directly inputting various heavy metals into soils, while organic fertilizers enhance heavy metal pollution levels by forming complexes with heavy metal ions. *Int. J. Environ. Res. Public Health* **2020**, *17*, x 8 of 14

<span id="page-6-0"></span>

Figure 4. Geoaccumulation indexes of heavy metals in the surface layers of six soil profiles. 0–10 cm (a); (**a**); 10–20 cm (**b**); 20–30 cm (**c**); 30–40 cm (**d**). 10–20 cm (**b**); 20–30 cm (**c**); 30–40 cm (**d**).

*3.3. Correlation Analysis Between Soil Heavy Metals and Soil Properties. 3.3. Correlation Analysis Between Soil Heavy Metals and Soil Properties.*

## 3.3.1. Characteristics of Inter-element Relationships 3.3.1. Characteristics of Inter-element Relationships

The results of Pearson correlation analysis between soil heavy metals and the three soil The results of Pearson correlation analysis between soil heavy metals and the three soil properties (SOC, soil pH and clay co[nt](#page-7-0)ents) are presented in Table 3. The inter-element relationships show that all

heavy metals are closely correlated (*P* < 0.01, *P* < 0.05) (except for Mo–Co, Mo–Sc, Mo–V and Ba–Ni), which indicates that the heavy metals Co, Ni, Ba, Sc and V are derived from similar natural processes and have similar geochemical behavior [\[23,](#page-11-15)[38\]](#page-12-8). The reason Mo is distinct from other heavy metals may be that Mo usually forms oxyanions but not cations in soil and rarely coordinates with ligands such as OH<sup>-</sup> and Cl<sup>-</sup> like most heavy metals [\[39\]](#page-12-9).

<span id="page-7-0"></span>

	Co	Ni	Mo	Ba	<b>Sc</b>	v	<b>SOC</b>	pH	Clay
Co		$0.652**$	$-0.063$	$0.419**$	$0.763**$	$0.558**$	$0.298**$	$0.451**$	$0.739**$
Ni	$0.652**$		$0.270**$	0.083	$0.510**$	$0.248*$	$0.407**$	0.003	$0.530**$
Mo	$-0.063$	$0.270**$	1	$-0.239*$	$-0.155$	0.115	$0.329**$	$-0.223*$	$-0.064$
Ba	$0.419**$	0.083	$-0.239*$		$0.701**$	$0.321**$	$0.360**$	$0.225*$	$0.575**$
Sc	$0.763**$	$0.510**$	$-0.155$	$0.701**$	1	$0.517**$	$0.414**$	$0.378**$	$0.897**$
V	$0.558**$	$0.248*$	0.115	$0.321**$	$0.517**$	1.	0.159	$0.365**$	$0.543**$
SOC.	$0.298**$	$0.407**$	$0.329**$	$0.360**$	$0.414**$	0.159	1	$-0.298**$	$0.433**$
pH	$0.451**$	0.003	$-0.223*$	$0.225*$	$0.378**$	$0.365**$	$-0.298**$		$0.357**$
Clay	$0.739**$	$0.530**$	$-0.064$	$0.575**$	$0.897**$	$0.543**$	$0.433**$	$0.357**$	

**Table 3.** Correlation analysis between heavy metals and soil properties.

\* Correlation is significant at the 0.05 level. \*\* Correlation is significant at the 0.01 level

#### 3.3.2. Influence of Soil Organic Matter on Heavy Metals

The mobility of heavy metals is reported to be controlled by the process of adsorption and desorption in soils, which is affected by some soil properties such as soil organic matter, soil pH and clay contents [\[13](#page-11-5)[,40](#page-12-10)[,41\]](#page-12-11). Soil organic matter can be converted into SOC contents with a conversion Factor, making SOC contents an important measurement for SOM [\[42\]](#page-12-12). Except for V, these heavy metals are significantly positively correlated with SOC contents  $(P < 0.01)$ , indicating that organic matter benefits the accumulation of heavy metals (Figure [5\)](#page-7-1). The heavy metal V occurs in various oxidation states (II, III, IV, or V), and mainly exists as  $V^{5+}$  under weathering conditions [\[9\]](#page-11-1). Because SOM is equipped with numerous negatively charged groups, it owns a strong ability to absorb heavy metal divalent and trivalent [cati](#page-11-13)ons [21]. From this perspective, the weak relationship between V and SOC contents can be explained by the higher valence. However, while the main oxidation states of Mo in the lithosphere are Mo(IV) and Mo(VI), Mo is still closely correlated with SOC contents ( $P < 0.01$ ). The phenomenon may demonstrate that Mo is stabilized by organic matter in a different mechanism, such as complexing with catechol groups of tannin-like compounds in soil organic matter  $[43]$ .  $f_n$  making SOC contents an important measurement for SOM  $[12]$ . Except for  $v_n$  these heavy in as complex

<span id="page-7-1"></span>

Figure 5. Linear correlation between soil organic carbon (SOC) and the heavy metals Co (a), Ni (b), Mo Mo (**c**), Ba (**d**), Sc (**e**), V (**f**). (**c**), Ba (**d**), Sc (**e**), V (**f**).

#### 3.3.3. Influence of Soil pH on Heavy Metals

Soil pH is widely regarded as the most important factor influencing the mobility and solubility of soil heavy metals [\[13](#page-11-5)[,44\]](#page-12-14). Except for Ni, all heavy metals are significantly influenced by soil pH (*P* < 0.01 for Co, Sc, and V; *P* < 0.05 for Mo and Ba) (Figure [6\)](#page-8-0). Soil heavy metals Co, Sc, Ba and V are positively correlated with soil pH while Mo is negatively correlated with soil pH. The positive relationship between heavy metals and soil pH has been widely reported [\[9,](#page-11-1)[13,](#page-11-5)[23\]](#page-11-15). Soil pH regulates heavy metal contents by changing the available heavy metal contents [\[45,](#page-12-15)[46\]](#page-12-16): higher pH means more OH<sup>-</sup> existing in soil environment, which would increase the fixation of most heavy metals and reduce the amount of heavy metals uptake by plants, thus benefitting the accumulation of heavy metals. The existing form of Mo in soils results in the negative correlation between Mo and soil pH: Mo always forms oxyanions instead of cations like most heavy metals [\[9\]](#page-11-1). In this case, Mo is more available and mobile under higher pH conditions which corresponds to the negative relationship between Mo and soil pH. The soil pH values of these samples are in the range of 5.6–7.2 for most samples (around 75%) [\[47\]](#page-12-17). The heavy metal Ni is reported to be most sensitive to soil pH when the soil pH ranges from 5.0–5.5 [\[48\]](#page-12-18), so the weak relationship between Ni and soil pH may result from the ranges of soil pH values. *Int. J. Environ. Res. Public Health* **2020**, *17*, x 10 of 14

<span id="page-8-0"></span>

Figure 6. Linear correlation between soil pH and the heavy metals Co (a), Mo (b), Ba (c), Sc (d), V (e), Ni (**f**).

## 3.3.4. Influence of Soil Texture on Heavy Metals 3.3.4. Influence of Soil Texture on Heavy Metals

Soil texture is a term used to describe the distributions of soil particle sizes in soils [49]. Soil texture is a term used to describe the distributions of soil particle sizes in soils [\[49\]](#page-12-19). According  $\Delta$  cordination of soil particles, they can be categorized into three groups: clay, silt and can  $\Delta$  All b to the diameter of soil particles, they can be categorized into three groups: clay, silt and sand. All heavy<br>The diameter of soil particles, they can be categorized into three groups: clay, silt and sand. All heavy metals (excluding Mo) are significantly correlated ( $P < 0.01$ ) with clay contents in soils (Figure [7\)](#page-9-0), indicating the important role of clay in stabilizing soil heavy metals. Finer soil particles are reported to have a stronger ability to absorb and fix heavy metals on their surface because of their bigger surface area and various functional groups [\[1](#page-10-0)[,50–](#page-12-20)[52\]](#page-13-0). Moreover, clay is also found to stabilize soil organic matter that would also fix heavy metals [13,47].  $\blacksquare$ 

Therefore, the clay benefits the accumulation of soil heavy metals. However, relatively high soil pH would prevent the adsorption of molybdate by clay [\[53\]](#page-13-1), which provides a potential explanation for the weak relationship between Mo and clay contents.

<span id="page-9-0"></span>

Figure 7. Linear correlation between clay contents and the heavy metals Co (a), Ni (b), Ba (c), Sc (d), V(**e**), Mo (**f**). V(**e**), Mo (**f**).

#### *3.4. Principal Component Analysis for Heavy Metals 3.4. Principal Component Analysis for Heavy Metals*

<span id="page-9-1"></span>Principal component analysis (PCA) is an important multivariate analysis method applied to identify the sources of soil heavy metals [\[54\]](#page-13-2). The prerequisite of the application of PCA is to pass the Kaiser–Meyer–Olkin (KMO) and Bartlett's sphericity tests: the KMO measure should be higher than 0.5 and Bartlett's sphericity test should be significant (*P* < 0.001) [\[28](#page-11-20)[,55,](#page-13-3)[56\]](#page-13-4). The KMO value equals 0.61 and Bartlett's sphericity test is significant  $(P = 0)$  in this study, indicating that PCA is applicable to the analysis of these soil samples. The results of PCA are presented in Table [4.](#page-9-1)

Element	Component			
	PC <sub>1</sub>	PC2		
Cо	0.90	0.11		
Ni	0.64	0.56		
Mo	$-0.05$	0.85		
Вa	0.66	$-0.52$		
Sc.	0.93	$-0.17$		
V	0.68	0.13		
Eigenvalue	2.98	1.36		
Variance %	49.70	22.71		
Cumulative %	49.70	72.41		

**Table 4.** Principal component analysis (PCA) of heavy metals.

On condition that eigenvalues are higher than 1, two principal components are extracted which explain 72.41% of the total variance (PC1: 49.70%; PC2: 22.71%) (Table [4\)](#page-9-1). As shown in Table [4,](#page-9-1) PC1 is positively correlated with the heavy metals Co, Ni, Ba, Sc and V, and PC2 positively correlates with Mo. Based on the results of PCA, these heavy metals can be categorized into two groups (PC1: Co, Ni, Ba, Sc and V; PC2: Mo). The heavy metals Co, Ni, Sc and V are usually regarded as lithophile elements that are associated with soil minerals [\[3](#page-10-6)[,52\]](#page-13-0), while Mo and Ba are regarded as anthropogenic elements [\[52\]](#page-13-0). As an important agricultural region, anthropogenic sources in the Mun River Basin should be the application of fertilizers, including chemical fertilizers and organic fertilizers as mentioned before. Therefore, the first factor may refer to soil mineral compositions which are largely determined by the parent rocks and the second factor may refer to the impact of human activities. In this study, only the heavy metal Mo is dominated by the second factor, indicating that soil in the study area is little influenced by the heavy metals Co, Ni, Ba, Sc and V derived from anthropogenic activities. Although

the topsoil seems to be slightly or moderately polluted by these heavy metals due to the application of fertilizers as discussed before, the impact of anthropogenic activities is relatively limited compared with that of natural processes according to the PCA results.

### **4. Conclusions**

The soil heavy metal concentrations decrease following the sequence of Ba, V, Ni, Sc, Co and Mo in soil profiles T1, T3, T4 and T5; the sequence of Ni, V, Ba, Co, Sc, Mo in soil profile T2; and the sequence of Ba, V, Sc, Ni, Mo, Co in soil profile T6. The EF values of most samples are lower than 1.5, indicating that these heavy metals are mainly derived from natural processes. The EF values of Mo in paddy fields are lower than 0.5, which provides evidence for the existence of Mo deficiency in the agricultural lands of Northeast Thailand. The application of molybdenum fertilizer should be advocated to eliminate the adverse effect of Mo deficiency on crops. The geoaccumulation index shows that only the topsoil of the soil profiles T1 and T2 is slightly contaminated by the heavy metals Ba, Sc, Ni and V. Overall, soils in the study area are in relatively good condition. The PCA results reveals that two factors (parent rocks and human activities) are responsible for 72.4% of the total variations of heavy metals in soil profiles: Co, Ni, Ba, Sc and V are dominated by the first factor while Mo is dominated by the second factor. The correlation analysis shows that soil organic matter and clay benefit the accumulation of heavy metals. Soil pH is positively correlated with heavy metals Co, Ba, Sc and V but is negatively correlated with Mo, which may result due to Mo usually existing in the form of oxyanions instead of cations like most heavy metals.

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#### **References**

- <span id="page-10-0"></span>1. Rajmohan, N.; Prathapar, S.A.; Jayaprakash, M.; Nagarajan, R. Vertical distribution of heavy metals in soil profile in a seasonally waterlogging agriculture field in Eastern Ganges Basin. *Environ. Monit. Assess.* **2014**, *186*, 5411–5427. [\[CrossRef\]](http://dx.doi.org/10.1007/s10661-014-3790-x)
- <span id="page-10-1"></span>2. Krishna, A.K.; Govil, P.K. Soil contamination due to heavy metals from an industrial area of Surat, Gujarat, Western India. *Environ. Monit. Assess.* **2007**, *124*, 263–275. [\[CrossRef\]](http://dx.doi.org/10.1007/s10661-006-9224-7)
- <span id="page-10-6"></span>3. Li, P.Z.; Lin, C.Y.; Cheng, H.G.; Duan, X.L.; Lei, K. Contamination and health risks of soil heavy metals around a lead/zinc smelter in southwestern China. *Ecotox. Environ. Safe.* **2015**, *113*, 391–399. [\[CrossRef\]](http://dx.doi.org/10.1016/j.ecoenv.2014.12.025) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/25540851)
- <span id="page-10-2"></span>4. Tang, Y.; Han, G.L. Characteristics of major elements and heavy metals in atmospheric dust in Beijing, China. *J. Geochem. Explor.* **2017**, *176*, 114–119. [\[CrossRef\]](http://dx.doi.org/10.1016/j.gexplo.2015.12.002)
- <span id="page-10-3"></span>5. Qu, R.; Han, G.; Liu, M.; Li, X. The Mercury Behavior and Contamination in Soil Profiles in Mun River Basin, Northeast Thailand. *Int. J. Environ. Res. Public Health* **2019**, *16*, 4131. [\[CrossRef\]](http://dx.doi.org/10.3390/ijerph16214131) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/31717757)
- <span id="page-10-4"></span>6. Li, X.P.; Feng, L.N. Multivariate and geostatistical analyzes of metals in urban soil of Weinan industrial areas, Northwest of China. *Atmos. Environ.* **2012**, *47*, 58–65. [\[CrossRef\]](http://dx.doi.org/10.1016/j.atmosenv.2011.11.041)
- <span id="page-10-5"></span>7. Gomez-Aracena, J.; Martin-Moreno, J.M.; Riemersma, R.A.; Bode, P.; Gutierrez-Bedmar, M.; Gorgojo, L.; Kark, J.D.; Garcia-Rodriguez, A.; Gomez-Gracia, E.; Kardinaal, A.F.M.; et al. Association between toenail scandium levels and risk of acute myocardial infarction in European men: The EURAMIC and Heavy Metals Study. *Toxicol. Ind. Health* **2002**, *18*, 353–360. [\[CrossRef\]](http://dx.doi.org/10.1191/0748233702th156oa)
- <span id="page-11-0"></span>8. Zarcinas, B.A.; Pongsakul, P.; McLaughlin, M.J.; Cozens, G. Heavy metals in soils and crops in southeast Asia. 2. *Thailand. Environ. Geochem. Health* **2004**, *26*, 359–371. [\[CrossRef\]](http://dx.doi.org/10.1007/s10653-005-4670-7)
- <span id="page-11-1"></span>9. Alloway, B.J. *Heavy Metals in Soils: Trace Metals and Metalloids in Soils and Their Bioavailability*; Springer Science & Business Media: Berlin/Heidelberg, Germany, 2012; Volume 22.
- <span id="page-11-2"></span>10. Clarkson, D.T.; Hanson, J.B. The mineral nutrition of higher plants. *Annu. Rev. Plant Physiol.* **1980**, *31*, 239–298. [\[CrossRef\]](http://dx.doi.org/10.1146/annurev.pp.31.060180.001323)
- <span id="page-11-3"></span>11. Osotsapar, Y. *Micronutrients in crop production in Thailand*; Food and Fertilizer Technology Center: Taipei, Taiwan, 2000.
- <span id="page-11-4"></span>12. Tiaranan, N.; Pimsarn, S.; Claimon, S.; Punpruk, P. Correction of nutrient deficiencies of legumes in Thailand. *Trop. Legume Improv.* **1985**, *8*, 54–55.
- <span id="page-11-5"></span>13. Zeng, F.R.; Ali, S.; Zhang, H.T.; Ouyang, Y.B.; Qiu, B.Y.; Wu, F.B.; Zhang, G.P. The influence of pH and organic matter content in paddy soil on heavy metal availability and their uptake by rice plants. *Environ. Pollut.* **2011**, *159*, 84–91. [\[CrossRef\]](http://dx.doi.org/10.1016/j.envpol.2010.09.019) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/20952112)
- <span id="page-11-6"></span>14. Wijnhoud, J.D.; Konboon, Y.; Lefroy, R.D.B. Nutrient budgets: Sustainability assessment of rainfed lowland rice-based systems in northeast Thailand. *Agric. Ecosyst. Environ.* **2003**, *100*, 119–127. [\[CrossRef\]](http://dx.doi.org/10.1016/S0167-8809(03)00180-4)
- <span id="page-11-7"></span>15. Floch, P.; Molle, F. Irrigated agriculture and rural change in Northeast Thailand: Reflections on present developments. In *Governing the Mekong: Engaging in the Politics of Knowledge*; Daniel, R., Lebel, L., Manorom, K., Eds.; Strategic Information and Research Development Centre (SIRD): Petaling Jaya, Malaysia, 2013; pp. 185–198.
- <span id="page-11-8"></span>16. Khunthasuvon, S.; Rajastasereekul, S.; Hanviriyapant, P.; Romyen, P.; Fukai, S.; Basnayake, J.; Skulkhu, E. Lowland rice improvement in northern and northeast Thailand 1. Effects of fertiliser application and irrigation. *Field Crop. Res.* **1998**, *59*, 99–108. [\[CrossRef\]](http://dx.doi.org/10.1016/S0378-4290(98)00109-9)
- <span id="page-11-9"></span>17. Nimnate, P.; Thitimakorn, T.; Choowong, M.; Hisada, K. Imaging and locating paleo-channels using geophysical data from meandering system of the Mun River, Khorat Plateau, Northeastern Thailand. *Open Geosci.* **2017**, *9*, 675–688. [\[CrossRef\]](http://dx.doi.org/10.1515/geo-2017-0051)
- <span id="page-11-10"></span>18. Liu, M.; Han, G.; Zhang, Q. Effects of agricultural abandonment on soil aggregation, soil organic carbon storage and stabilization: Results from observation in a small karst catchment, Southwest China. *Agric. Ecosyst. Environ.* **2020**, *288*, 106719. [\[CrossRef\]](http://dx.doi.org/10.1016/j.agee.2019.106719)
- <span id="page-11-11"></span>19. Han, G.L.; Li, F.S.; Tang, Y. Variations in soil organic carbon contents and isotopic compositions under different land uses in a typical karst area in Southwest China. *Geochem. J.* **2015**, *49*, 63–71. [\[CrossRef\]](http://dx.doi.org/10.2343/geochemj.2.0331)
- <span id="page-11-12"></span>20. Liu, M.; Han, G.L.; Zhang, Q.; Song, Z.L. Variations and Indications of C-13(SOC) and N-15(SON) in Soil Profiles in Karst Critical Zone Observatory (CZO), Southwest China. *Sustainability* **2019**, *11*, 16.
- <span id="page-11-13"></span>21. Han, G.L.; Li, F.S.; Tang, Y. Organic Matter Impact on Distribution of Rare Earth Elements in Soil Under Different Land Uses. *Clean-Soil Air Water* **2017**, *45*, 9. [\[CrossRef\]](http://dx.doi.org/10.1002/clen.201600235)
- <span id="page-11-14"></span>22. Zoller, W.H.; Gladney, E.S.; Duce, R.A. Atmospheric concentrations and sources of trace metals at the South pole. *Science* **1974**, *183*, 198–200. [\[CrossRef\]](http://dx.doi.org/10.1126/science.183.4121.198)
- <span id="page-11-15"></span>23. Muller, G. Index of geoaccumulation in sediments of the Rhine River. *Geojournal* **1969**, *2*, 108–118.
- <span id="page-11-16"></span>24. Wedepohl, K.H. The composition of the continental-crust. Geochim. *Cosmochim. Acta* **1995**, *59*, 1217–1232. [\[CrossRef\]](http://dx.doi.org/10.1016/0016-7037(95)00038-2)
- <span id="page-11-17"></span>25. Lu, X.W.; Wang, L.J.; Lei, K.; Huang, J.; Zhai, Y.X. Contamination assessment of copper, lead, zinc, manganese and nickel in street dust of Baoji, NW China. *J. Hazard. Mater.* **2009**, *161*, 1058–1062. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jhazmat.2008.04.052) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/18502044)
- <span id="page-11-18"></span>26. Balls, P.; Hull, S.; Miller, B.; Pirie, J.; Proctor, W. Trace metal in Scottish estuarine and coastal sediments. *Mar. Pollut. Bull.* **1997**, *34*, 42–50. [\[CrossRef\]](http://dx.doi.org/10.1016/S0025-326X(96)00056-2)
- <span id="page-11-19"></span>27. Zeng, J.; Han, G.L.; Wu, Q.X.; Tang, Y. Heavy Metals in Suspended Particulate Matter of the Zhujiang River, Southwest China: Contents, Sources, and Health Risks. *Int. J. Environ. Res. Public Health* **2019**, *16*, 1843. [\[CrossRef\]](http://dx.doi.org/10.3390/ijerph16101843) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/31137629)
- <span id="page-11-20"></span>28. Fang, X.H.; Peng, B.; Wang, X.; Song, Z.L.; Zhou, D.X.; Wang, Q.; Qin, Z.L.; Tan, C.Y. Distribution, contamination and source identification of heavy metals in bed sediments from the lower reaches of the Xiangjiang River in Hunan province, China. *Sci. Total Environ.* **2019**, *689*, 557–570. [\[CrossRef\]](http://dx.doi.org/10.1016/j.scitotenv.2019.06.330) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/31279202)
- <span id="page-11-21"></span>29. Zhang, Q.; Han, G.L.; Liu, M.; Liang, T. Spatial distribution and controlling factors of heavy metals in soils from Puding Karst Critical Zone Observatory, southwest China. *Environ. Earth Sci.* **2019**, *78*, 13. [\[CrossRef\]](http://dx.doi.org/10.1007/s12665-019-8280-6)
- <span id="page-12-0"></span>30. Zeng, J.; Yue, F.-J.; Wang, Z.-J.; Wu, Q.; Qin, C.-Q.; Li, S.-L. Quantifying depression trapping effect on rainwater chemical composition during the rainy season in karst agricultural area, southwestern China. *Atmos. Environ.* **2019**, *218*, 116998. [\[CrossRef\]](http://dx.doi.org/10.1016/j.atmosenv.2019.116998)
- <span id="page-12-1"></span>31. Wong, S.C.; Li, X.D.; Zhang, G.; Qi, S.H.; Min, Y.S. Heavy metals in agricultural soils of the Pearl River Delta, South China. *Environ. Pollut.* **2002**, *119*, 33–44. [\[CrossRef\]](http://dx.doi.org/10.1016/S0269-7491(01)00325-6)
- <span id="page-12-2"></span>32. Kabata-Pendias, A.; Pendias, H. *Trace Elements in Soils and Plants*, 2nd ed.; CRC Press: Boca Raton, FL, USA, 1992.
- <span id="page-12-3"></span>33. Chao, T.T.; Theobald, P.K. The significance of secondary iron and manganese oxides in geochemical exploration. *Econ. Geol.* **1976**, *71*, 1560–1569. [\[CrossRef\]](http://dx.doi.org/10.2113/gsecongeo.71.8.1560)
- <span id="page-12-4"></span>34. Xu, F.J.; Hu, B.Q.; Yuan, S.Q.; Zhao, Y.F.; Dou, Y.G.; Jiang, Z.Z.; Yin, X.B. Heavy metals in surface sediments of the continental shelf of the South Yellow Sea and East China Sea: Sources, distribution and contamination. *Catena* **2018**, *160*, 194–200. [\[CrossRef\]](http://dx.doi.org/10.1016/j.catena.2017.09.022)
- <span id="page-12-5"></span>35. Zhang, J.; Liu, C.L. Riverine composition and estuarine geochemistry of particulate metals in China–Weathering features, anthropogenic impact and chemical fluxes. *Estuar. Coast. Shelf Sci.* **2002**, *54*, 1051–1070. [\[CrossRef\]](http://dx.doi.org/10.1006/ecss.2001.0879)
- <span id="page-12-6"></span>36. Wong, C.S.C.; Li, X.D.; Thornton, I. Urban environmental geochemistry of trace metals. *Environ. Pollut.* **2006**, *142*, 1–16. [\[CrossRef\]](http://dx.doi.org/10.1016/j.envpol.2005.09.004) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/16297517)
- <span id="page-12-7"></span>37. Mortvedt, J.J. Heavy metal contaminants in inorganic and organic fertilizers. *Fertil. Res.* **1996**, *43*, 55–61. [\[CrossRef\]](http://dx.doi.org/10.1007/BF00747683)
- <span id="page-12-8"></span>38. Yan, X.; Luo, X.G. Heavy Metals in Sediment from Bei Shan River: Distribution, Relationship with Soil Characteristics and Multivariate Assessment of Contamination Sources. *Bull. Environ. Contam. Toxicol.* **2015**, *95*, 56–60.
- <span id="page-12-9"></span>39. Kabata-Pendias, A.; Mukherjee, A.B. *Trace Elements From Soil to Human*; Springer Science & Business Media: Berlin/Heidelberg, Germany, 2007.
- <span id="page-12-10"></span>40. Krishnamurti, G.S.R.; Huang, P.M.; Kozak, L.M. Sorption and desorption kinetics of cadmium from soil: Influence of phosphate. *Soil Sci.* **1999**, *164*, 888–898. [\[CrossRef\]](http://dx.doi.org/10.1097/00010694-199912000-00002)
- <span id="page-12-11"></span>41. Lair, G.J.; Gerzabek, M.H.; Haberhauer, G. Sorption of heavy metals on organic and inorganic soil constituents. *Environ. Chem. Lett.* **2007**, *5*, 23–27. [\[CrossRef\]](http://dx.doi.org/10.1007/s10311-006-0059-9)
- <span id="page-12-12"></span>42. Pribyl, D.W. A critical review of the conventional SOC to SOM conversion factor. *Geoderma* **2010**, *156*, 75–83. [\[CrossRef\]](http://dx.doi.org/10.1016/j.geoderma.2010.02.003)
- <span id="page-12-13"></span>43. Wichard, T.; Mishra, B.; Myneni, S.C.B.; Bellenger, J.P.; Kraepiel, A.M.L. Storage and bioavailability of molybdenum in soils increased by organic matter complexation. *Nat. Geosci.* **2009**, *2*, 625–629. [\[CrossRef\]](http://dx.doi.org/10.1038/ngeo589)
- <span id="page-12-14"></span>44. Muhlbachova, G.; Simon, T.; Pechova, M. The availability of Cd, Pb and Zn and their relationships with soil pH and microbial biomass in soils amended by natural clinoptilolite. *Plant Soil Environ.* **2005**, *51*, 26–33. [\[CrossRef\]](http://dx.doi.org/10.17221/3552-PSE)
- <span id="page-12-15"></span>45. Chuan, M.C.; Shu, G.Y.; Liu, J.C. Solubility of heavy metals in a contaminated soil: Effects of redox potential and pH. *Water Air Soil Pollut.* **1996**, *90*, 543–556. [\[CrossRef\]](http://dx.doi.org/10.1007/BF00282668)
- <span id="page-12-16"></span>46. Zeng, J.; Han, G.; Wu, Q.; Tang, Y. Effects of agricultural alkaline substances on reducing the rainwater acidification: Insight from chemical compositions and calcium isotopes in a karst forests area. *Agric. Ecosyst. Environ.* **2020**, *290*, 106782. [\[CrossRef\]](http://dx.doi.org/10.1016/j.agee.2019.106782)
- <span id="page-12-17"></span>47. Zhou, W.X.; Han, G.L.; Liu, M.; Li, X.Q. Effects of soil pH and texture on soil carbon and nitrogen in soil profiles under different land uses in Mun River Basin, Northeast Thailand. *PeerJ* **2019**, *7*, 15. [\[CrossRef\]](http://dx.doi.org/10.7717/peerj.7880) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/31637132)
- <span id="page-12-18"></span>48. Blake, L.; Goulding, K.W.T. Effects of atmospheric deposition, soil pH and acidification on heavy metal contents in soils and vegetation of semi-natural ecosystems at Rothamsted Experimental Station, UK. *Plant Soil* **2002**, *240*, 235–251. [\[CrossRef\]](http://dx.doi.org/10.1023/A:1015731530498)
- <span id="page-12-19"></span>49. Dexter, A.R. Soil physical quality–Part I. Theory, effects of soil texture, density, and organic matter, and effects on root growth. *Geoderma* **2004**, *120*, 201–214. [\[CrossRef\]](http://dx.doi.org/10.1016/j.geoderma.2003.09.004)
- <span id="page-12-20"></span>50. Damrongsiri, S.; Vassanadumrongdee, S.; Tanwattana, P. Heavy metal contamination characteristic of soil in WEEE (waste electrical and electronic equipment) dismantling community: A case study of Bangkok, Thailand. *Environ. Sci. Pollut. Res.* **2016**, *23*, 17026–17034. [\[CrossRef\]](http://dx.doi.org/10.1007/s11356-016-6897-5)
- 51. Yin, Y.J.; Impellitteri, C.A.; You, S.J.; Allen, H.E. The importance of organic matter distribution and extract soil: Solution ratio on the desorption of heavy metals from soils. *Sci. Total Environ.* **2002**, *287*, 107–119. [\[CrossRef\]](http://dx.doi.org/10.1016/S0048-9697(01)01000-2)
- <span id="page-13-0"></span>52. Wang, X.S.; Qin, Y.; Chen, Y.K. Heavy meals in urban roadside soils, part 1: Effect of particle size fractions on heavy metals partitioning. *Environ. Geol.* **2006**, *50*, 1061–1066. [\[CrossRef\]](http://dx.doi.org/10.1007/s00254-006-0278-1)
- <span id="page-13-1"></span>53. McBride, M.B.; Richards, B.K.; Steenhuis, T.; Spiers, G. Molybdenum uptake by forage crops grown on sewage sludge-amended soils in the field and greenhouse. *J. Environ. Qual.* **2000**, *29*, 848–854. [\[CrossRef\]](http://dx.doi.org/10.2134/jeq2000.00472425002900030021x)
- <span id="page-13-2"></span>54. Gerdol, R.; Marchesini, R.; Iacumin, P.; Brancaleoni, L. Monitoring temporal trends of air pollution in an urban area using mosses and lichens as biomonitors. *Chemosphere* **2014**, *108*, 388–395. [\[CrossRef\]](http://dx.doi.org/10.1016/j.chemosphere.2014.02.035)
- <span id="page-13-3"></span>55. Dahal, H. Factor analysis for soil test data: A methodological approach in environment friendly soil fertility management. *J. Agric. Environ.* **2007**, *8*, 8–19. [\[CrossRef\]](http://dx.doi.org/10.3126/aej.v8i0.722)
- <span id="page-13-4"></span>56. Han, G.L.; Song, Z.L.; Tang, Y.; Wu, Q.X.; Wang, Z.R. Ca and Sr isotope compositions of rainwater from Guiyang city, Southwest China: Implication for the sources of atmospheric aerosols and their seasonal variations. *Atmos. Environ.* **2019**, *214*, 10. [\[CrossRef\]](http://dx.doi.org/10.1016/j.atmosenv.2019.116854)



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