

Nitration-Promoted Vanadate Catalysts for Low-Temperature Selective Catalytic Reduction of NO_X with NH₃

Sun-I Kim,[§] Yeong Jun Choi,[§] Min Seong Lee, and Duck Hyun Lee*

Cite This: ACS Omega 2023, 8, 34152–34159



ACCESS More Article Recommendations Supporting Information

ABSTRACT: Vanadium-based catalysts have been commercially used in selective catalytic reduction (SCR), owing to their high catalytic activity and effectiveness across a wide temperature range; however, their catalytic efficiency decreases at lower temperatures under exposure to SO_X . This decrease is largely due to ammonium sulfate generation on the catalyst surface. To overcome this limitation, we added ammonium nitrate to the V_2O_5 -WO₃/TiO₂ catalyst, producing a V_2O_5 -WO₃/TiO₂ catalyst with nitrate functional groups. With this approach, we found that it was possible to adjust the amount of these functional groups by varying the amount of ammonium nitrate. Overall, the resultant nitrate



 V_2O_5 -WO₃/TiO₂ catalyst has large quantities of NO₃⁻ and chemisorbed oxygen, which improves the density of Brønsted and Lewis acid sites on the catalyst surface. Furthermore, the nitrated V_2O_5 -WO₃/TiO₂ catalyst has a high NO_X removal efficiency and N₂ selectivity at low temperatures (i.e., 300 °C); this is because NO₃⁻ and chemisorbed oxygen, generated by nitrate treatment, facilitated the occurrence of a fast SCR reaction. The approach outlined in this study can be applied to a wide range of SCR catalysts, allowing for the development of more, low-temperature SCR catalysts.

1. INTRODUCTION

Recently, due to anthropogenic activity, NO_X, SO_X, carbon oxides, volatile organic compounds (VOCs), and particulate matter (PM) emissions have increased.^{1,2} Specifically, nitrogen oxides (i.e., NO, NO₂, and N₂O) are extremely hazardous air pollutants, which have been linked to many environmental issues (e.g., acid rain, smog, and ozone depletion) that can directly impact human health. Thus, many nations and governments have strengthened their emission regulation policies to control nitrogen oxide emissions.³ Several processes have been proposed to either eliminate or control nitrogen oxide emissions: for example, selective catalytic reduction (SCR), selective noncatalytic reduction, and non-selective catalytic reduction.⁴⁻⁸ Among these technologies, SCR is the most advanced and efficient; it works by converting NO_x in exhaust gas into N_2 and $H_2O_{\!\!\!}$ using ammonia (NH_3) as a reducing agent. $^{9-11}$ This technology can reduce up to $60{-}90\%$ of NO_X in exhaust gas, contain secondary pollutant emission, and can operate at a relatively low temperature (~350 °C).¹² Many transition-metal catalysts have been used for SCR reactions, such as vanadium, tungsten, copper, iron, and manganese oxides. Among these, V-based catalysts (V_2O_5 - WO_3 or MoO_3/TiO_2) have been commercially adopted due to their lower ability to oxidize SO₂ to SO₃; additionally, these catalysts exhibit greater catalytic activity, compared to other transition-metal catalysts.^{13–15} However, this catalytic activity is limited to the temperature range of 300-400 °C and decreases at lower temperatures (below 300 °C).¹⁰

Currently, there are methods for enhancing V-based catalyst activity at low temperatures, the first of which entails the use of an Mn-based catalyst. Kapteijn et al. have demonstrated the high catalytic activity of MnO_X at low temperatures. Additionally, MnO₂ can achieve its highest efficiency through MnO_X catalyst synthesis by controlling the oxidation states of manganese.¹⁷ However, this is limited by Mn sulfate formation when in contact with SO₂, owing to its low sulfur resistance. The second method involves adjusting the carbon material. Zhu et al. fabricated activated carbon (AC)-supported vanadium catalysts that achieved high SCR activity and resistance to SO₂ deactivation.¹⁸ Li et al. designed carbon nanotube (CNT)-containing V_2O_5/TiO_2 catalysts, which increased the specific area, pore volume, acidity, and reducibility; the catalyst added with 10 wt % CNT exhibited excellent activity in NO SCR at 100–300 °C.¹⁹ However, the commercialization of these catalysts within the context of industrial settings is difficult. This is because exhaust gas contains large amounts of sulfur and water, which deactivates Mn and causes the carbon material to thermally decompose.²⁰

Received: July 26, 2023 Accepted: August 29, 2023 Published: September 6, 2023





Recently, many researchers have attempted to control catalyst surface functional groups to improve their activity at low temperatures.^{21–24} For example, Zhao et al. reported that V_2O_5/TiO_2 catalyst NO_X removal efficiency reached 90% at 210 °C by co-doping with S and N.²⁵ These catalysts cause the formation of O2 active sites, which drive the SCR reaction. Additionally, these catalysts have been linked to a large surface area and greater quantities of chemisorbed oxygen and NH₃. In addition, S and N inhibit TiO₂ during transformation to the rutile phase. Maqbool et al. prepared CeO2-added Sb/V2O5/ TiO₂ catalysts that were pretreated with SO₂ under oxidizing conditions.²⁶ These catalysts formed sulfate species on the surface and had high NH3-desorption and catalytic performance when containing SO₂ because of the formation of more cerium(III) sulfate. Although many researchers have studied Vbased catalysts to improve the catalytic performance at low temperatures by controlling catalyst functional groups, previous studies have not attained results that can be easily applied in a commercial context.

In this study, we introduced a simple method to control the functional groups on a vanadium-based catalyst surface. Specifically, we added ammonium nitrate (NH₄NO₃) to the V₂O₅-WO₃/TiO₂ catalyst and investigated nitrate-treated V₂O₅-WO₃/TiO₂ catalyst effectiveness under SCR conditions. The functional groups on the V₂O₅-WO₃/TiO₂ catalyst were controlled by controlling the amount of ammonium nitrate that was added. These catalytic processes were characterized through X-ray diffraction (XRD), X-ray photoelectron spectroscopy (XPS), and Fourier transform infrared spectroscopy (FTIR).

2. MATERIALS AND METHODS

2.1. Catalyst Preparation. V₂O₅-WO₃/TiO₂ catalysts with NH₄NO₃ were prepared using an impregnation method. NH_4VO_3 (0.128 g, Sigma-Aldrich, 99.99%) and $(NH_4)_6H_2W_{12}O_{40} \times H_2O$ (0.531 g, Sigma-Aldrich, 99.99%) were dissolved in 50 mL of deionized water with 0.1959 g of oxalic acid, which acted as a solubility agent. Then, NH₄NO₃ was then added to the prepared mixture at varying concentrations (i.e., 0, 0.03, 0.05, and 0.08 M, respectively); the resultant mixtures were denoted VW/TiO₂, 0.03 M NH₄NO₃, 0.05 M NH₄NO₃, and 0.08 M NH₄NO₃, respectively. TiO₂ powder (4.400 g, NANO Co., Ltd., NT-01) was added to the prepared solution, and the mixture was stirred for 2 h. The solution was evaporated at 85 °C in an oil bath and placed in an oven at 110 °C for 12 h, and the resultant powders were then calcinated at 500 °C in a furnace for 5 h under atmospheric pressure.

2.2. Catalyst Characterization. The sample surface morphologies and elemental compositions were investigated using field-emission scanning electron microscopy (model: SU8020/Hitachi, Tokyo, Japan), transmission electron microscopy (TEM, model: JEM-2100F/JEOL Ltd., Tokyo, Japan), and electron energy loss spectroscopy (EELS) at an accelerating voltage of 10.0 kV. Textural properties were analyzed using the Brunauer–Emmett–Teller method (BET, model: ASAP2020/Micromeritics Instrument Corp., Norcross, USA). Catalyst chemical compositions were determined using X-ray fluorescence spectrometry (XRF; Zetium; Malvern Panalytical, Malvern, UK). We analyzed the extent of crystallinity using X-ray diffraction (XRD, model: Ultima IV/Rigaku, Tokyo, Japan) with Cu K α ($\lambda = 0.15406$ nm) radiation in the 2θ range from 20–80° at a 1°/min scan rate. X-ray

photoelectron spectroscopy (XPS, model: K Alpha+/Thermo Scientific, Waltham, USA) was conducted with Al K α radiation to confirm the sample oxidation states, and C 1s binding energy was normalized as 284.8 eV. Catalyst reduction properties were measured by NH₃-temperature-programmed desorption (NH₃-TPD, model: AutoChem II 2920/Micromeritics Instrument Corp., Norcross, USA). The samples were pretreated at 150 °C in a current of N₂ for 4 h to remove physisorbed NH₃ species and organic matter, and NH₃ was then adsorbed with 10% NH₃/He gas at 150 °C for 1 h. The TPD experiment was conducted under 100–900 °C. An H₂-temperature-programmed reduction (H₂-TPR, model: AutoChem II 2920/Micromeritics Instrument Corp, Norcross, USA) experiment was conducted, during which the samples were immersed in a current of 10% H₂/Ar at 150–900 °C.

2.3. Catalytic Activity Evaluation. The catalyst NO_x removal efficiencies were evaluated in a fixed-bed reactor under high atmospheric pressure, with each sample placed in a stainless-steel tube and analyzed using a powder catalyst. The temperatures under which the samples were analyzed varied from 150–350 $^{\circ}$ C, and the reactive gas consisted of 300 ppm NO_{X} , 300 ppm NH_3 ($NH_3/NO_X = 1.0$), 300 ppm SO_2 , and 5 vol % of O₂, with N₂ balance under a total flow rate of 500 sccm. The powder catalyst (0.35 g) was tested, and the gas hourly space velocity was set at 60,000 h⁻¹. Commercial platetype catalysts were evaluated in a microreactor. The temperatures under which the catalysts were evaluated varied from 160–240 °C, and the reactive gas consisted of 300 ppm NO_{X} , 300 ppm NH₃ (NH₃/NO_x = 1.2), 300 ppm SO₂, and 5 vol % of O_{22} with N_2 balance under a total flow rate of 1.2 m³/h; the area velocity was set to 10 m/h. The in situ FTIR spectra of all samples were measured using an FTIR spectrometer (VERTEX 70v FTIR; Bruker, Billerica, MA, USA) under operating conditions and 16 scans with a resolution of 4 cm^{-1} in the 3900-1200 cm⁻¹ range. A mixture of NH₃ (500 ppm), NO (500 ppm), and O_2 (5 vol %) with N_2 was used for in situ FTIR, and the flow rate was 0.3 L/min.

3. RESULTS AND DISCUSSION

To analyze the nitrated surface impact on the V_2O_5 -WO₃/ TiO₂ catalysts, XRD measurements were performed, and the diffraction patterns are shown in Figure S1. All the prepared catalysts were detected only as anatase TiO₂ signals through XRD. These anatase peaks at around $2\theta = 25^{\circ}$, 38° , 48° , 54° , 55°, 63°, 69°, 70°, and 75° were matched to phase (1 0 1), (0 0 4), (2 0 0), (1 0 5), (2 1 1), (2 0 4), (1 1 6), (2 2 0), and (2 1 5) diffractions (JCPSD card no. 21-1272), respectively. However, V_2O_5 and WO_3 signals were not observed in the XRD results because small amounts of V2O5 and WO3 with low concentrations were uniformly dispersed on TiO₂.²⁷ ' The gross elemental composition (wt %) XRF analyses for the prepared catalyst are shown in Table 1. This table clearly indicates the presence of vanadium and W, as inferred from the XRD results. Moreover, ammonium nitrate did not affect V₂O₅-WO₃/TiO₂ catalyst crystallinity. This is because NH₄NO₃ is not substituted in the V₂O₅-WO₃/TiO₂ crystal but is in the ionic state on the catalyst surface.

The prepared nitrate-treated 0.03, 0.05, and 0.08 M V_2O_5 -WO₃/TiO₂ catalyst morphologies were observed via SEM, and the resultant images are shown in Figures 1a and S1a,b. All the catalysts exhibited similar morphologies with similar particle sizes and shapes, ranging from 15–50 nm. Ammonium nitrate treatment did not cause particle agglomeration and sustained

Table 1. X-ray Fluorescence Analysis Results of the V_2O_5 -WO₃/TiO₂, 0.03 M NH₄NO₃, 0.05 M NH₄NO₃, and 0.08 M NH₄NO₃ Added V_2O_5 -WO₃/TiO₂ Catalysts, Respectively

	chemical composition (wt %)				
sample	TiO ₂	WO ₃	V_2O_5	SO3	
VW/TiO ₂	87.35	10.04	2.03	0.58	
0.03 M NH ₄ NO ₃	87.63	9.83	2.02	0.52	
0.05 M NH ₄ NO ₃	87.65	9.87	1.94	0.54	
0.08 M NH ₄ NO ₃	87.31	9.98	2.03	0.67	



Figure 1. SEM image of (a) V_2O_5 - WO_3/TiO_2 and (b) 0.05 M NH₄NO₃ added V_2O_5 - WO_3/TiO_2 catalyst. TEM images of (c) V_2O_5 - WO_3/TiO_2 and (d) 0.05 M NH₄NO₃ added V_2O_5 - WO_3/TiO_2 catalyst. EELS elemental mapping of V, W, Ti, and N of (e) V_2O_5 - WO_3/TiO_2 and (f) 0.05 M NH₄NO₃ added V_2O_5 - WO_3/TiO_2 catalyst.

the uniform shape of the VW/TiO₂ catalysts. TEM and EELS elemental mapping were conducted to observe the synthesized catalysts in detail (Figure 1c,d). The catalyst particles were composed of V_2O_5 and WO_3 nanoparticles on TiO₂ supports. The nitrate-treated 0.05 M V_2O_5 -WO₃/TiO₂ catalysts were uniformly distributed on TiO₂ supports without agglomeration, which was confirmed by EELS mapping. The red and blue lines in the mapping images indicate V and W, respectively, and the elements are evenly distributed on the TiO₂ support. Consequently, the nitrate treatment process was not affected by catalyst morphology. Interestingly, N (yellow color) was only observed in the nitrate-treated 0.05 M V_2O_5 -WO₃/TiO₂ catalysts. These results confirm that ammonium nitrate was adsorbed on the surface of the V_2O_5 -WO₃/TiO₂ catalysts, as shown in the scheme (Figure S1).

The textural properties, such as specific surface areas, pore volumes, and pore sizes of the prepared catalysts, as well as the

 $V_2O_5,\,WO_3,\,and\,TiO_2$ weight fractions, are shown in Table 2. As ammonium nitrate was added to the $V_2O_5\text{-}WO_3/\text{Ti}O_2$

Table 2. Brunauer–Emmett–Teller (BET) Results of the V_2O_5 -WO₃/TiO₂, 0.03 M NH₄NO₃, 0.05 M NH₄NO₃, and 0.08 M NH₄NO₃ Added V_2O_5 -WO₃/TiO₂ Catalysts, Respectively

sample	$S_{\rm BET}~({\rm m^2/g})$	pore volume (cm^3/g)	pore size (nm)
VW/TiO ₂	66.05	0.24	13.49
0.03 M NH ₄ NO ₃	67.27	0.24	13.16
$0.05 \text{ M NH}_4 \text{NO}_3$	67.46	0.24	12.98
$0.08 \text{ M } \text{NH}_4 \text{NO}_3$	65.42	0.19	12.08

catalyst, the surface area and pore properties did not change significantly. However, the properties decreased when adding 0.08 M ammonium nitrate, and we assumed that the ions block the pores of the catalyst surface. To better scrutinize the V₂O₅-WO₃/TiO₂ catalyst nitrated surface chemical state, XPS analysis was conducted (Figure 2). Figure 2 shows the XPS survey peaks. The N 1s spectrum exhibited two main peaks, that is, NH_2 at approximately 399 eV and NO_3^- at approximately 402 eV, which were attributed to the neutral amine residual on the surface and NO_3^- (Figure 2b).²⁸ The NH₂ peak at approximately 399 eV originated from the ammonium-based precursors of V and W (ammonium metavanadate and ammonium metatungstate hydrate); therefore, the NH₂ peak was observed in all samples. However, the NO_3^- peak at approximately 402 eV originates from the $V_2O_5^-$ WO₃/TiO₂ catalyst nitrated surface. No signal was observed for the NO3⁻ peak of the V₂O₅-WO₃/TiO₂ catalysts. However, the intensity of the 0.03 M NH₄NO₃ catalyst increased slightly and 0.05 and 0.08 M NH₄NO₃ catalysts increased remarkably. These results were commonly observed for O 1s in the XPS results (Figure 2c). The O 1s peaks can be fitted to three different peaks, i.e., NO₃⁻ at around 531.2 eV, chemisorbed oxygen (O_{α}) at around 529.8 eV, and lattice oxygen (O_{β}) at around 528.4 eV.²⁹ The NO₃⁻, O_{α}, and O_{β} ratios are shown in Table 3. When NH4NO3 was added, the intensities of the NO3⁻ and chemisorbed oxygen peaks increased simultaneously, and the 0.08 M NH₄NO₃ catalyst had a higher ratio of O_{α} (61.18) and NO_{3}^{-} (22.70). O_{α} and NO_{3}^{-} are very important factors in SCR reaction. Surface chemisorbed oxygen plays a crucial role in the oxidation of NH₄⁺ in SCR because it is more movable than lattice oxygen and boosts the oxidation of NO to NO₂.³⁰⁻³³ In addition, nitrate species can react with NO to form more NO₂ according to eq $1.^{27}$

$$2NH_4NO_3 + NO \Rightarrow 3NO_2 + 2NH_3 + H_2O \tag{1}$$

Therefore, the presence of O_{α} and NO_{3}^{-} induces "fast SCR", and the concentrations of O_{α} and NO_{3-} are important for SCR reaction.³² V 2p is mainly composed of V⁵⁺ and V⁴⁺, and the two fitted peaks at approximately 516.8 and 515.3 eV are attributed to V⁵⁺ 2p3/2 and V⁴⁺ 2p3/2, respectively.³⁴ According to previous studies, V⁴⁺ can promote oxygen adsorption and form reactive oxygen species on the catalyst surface, leading to fast redox cycles and improved redox properties.³⁵ Figure 2d shows that V⁴⁺ increases as NH₄NO₃ is added, and 0.05 and 0.08 M catalysts had higher ratios of V⁴⁺ at 36.06 and 40.53, respectively. Additionally, the $NO_{3}^{-} + O_{\alpha} / (NO_{3}^{-} + O_{\alpha} + O_{\beta})$ and V⁴⁺/(V⁴⁺ + V⁵⁺) concentrations can be confirmed in Table 3.



Figure 2. XPS spectra representing (a) survey, (b) N 1s, (c) O 1s, and (d) V 2p of V_2O_5 -WO₃/TiO₂, 0.03 M NH₄NO₃, 0.05 M NH₄NO₃, and 0.08 M NH₄NO₃ added V_2O_5 -WO₃/TiO₂ catalysts, respectively.

Table 3. NO₃⁻, O_{α} V⁴⁺ Ratio of V₂O₅-WO₃/TiO₂, 0.03 M NH₄NO₃, 0.05 M NH₄NO₃, and 0.08 M NH₄NO₃ Added V₂O₅-WO₃/TiO₂ Catalysts, Respectively

sample	NO ₃₋	O _α	O_{β}	$\mathrm{NO}_{3-}/(\mathrm{NO}_{3-} + \mathrm{O}_{\alpha} + \mathrm{O}_{\beta})$	$(\mathrm{NO_3}^- + \mathrm{O}_{\alpha})/(\mathrm{NO_3}^- + \mathrm{O}_{\alpha} + \mathrm{O}_{\beta})$	$V^{4+}/(V^{4+} + V^{5+})$
VW/TiO ₂	4.93	52.76	42.31	4.93	57.69	15.81
0.03 M NH ₄ NO ₃	9.30	52.11	38.59	9.30	61.41	32.44
0.05 M NH ₄ NO ₃	15.46	49.94	34.60	15.46	65.40	36.06
0.08 M NH ₄ NO ₃	22.70	61.18	16.12	22.70	83.88	40.53



Figure 3. (a) NO_X removal efficiency; (b) N₂ selectivity; (c) NH₃-TPD profiles; (d) H₂-TPR profiles of V₂O₅-WO₃/TiO₂, 0.03 M NH₄NO₃, 0.05 M NH₄NO₃, and 0.08 M NH₄NO₃ added V₂O₅-WO₃/TiO₂ catalysts, respectively. Reaction conditions: [NO] & [NH₃] & [SO₂] = 300 ppm, [O₂] = 5 vol %, [GHSV] = 60,000 h⁻¹.



Figure 4. In situ Fourier transform infrared spectra of ammonia adsorption depending on the reaction time over (a) V_2O_5 -WO₃/TiO₂ and (b) 0.05 M NH₄NO₃. Nitrogen oxide and oxygen reacted with pre-adsorbed ammonia over (c) V_2O_5 -WO₃/TiO₂ and (d) 0.05 M NH₄NO₃ at 200 °C. Conditions: [NH₃] = 500 ppm, [NO] = 500 ppm, [O₂] = 5 vol % (when used), and N₂ as the balance.

The NO_X removal efficiency, N_2 selectivity, and N_2O concentration were measured using a fixed-bed reactor to confirm the effect of the nitrate species that were formed on the V₂O₅-WO₃/TiO₂ catalyst surface on catalytic performance (Figure 3). Figure 3a shows that catalysts with nitrate species on the surface have higher NO_X removal efficiency than a fresh catalyst at temperatures below 300 °C. In the case of 0.05 M NH_4NO_3 , NO_x removal efficiency improved from 53–90% at 200 °C. These results were primarily due to the large number of NO₃⁻ species that formed on the catalyst surface, increasing catalytic performance at temperatures below 300 °C. However, 0.08 M NH₄NO₃ had a lower NO_X removal efficiency than 0.05 M NH_4NO_3 . This is because the excess ammonium nitrate blocked the V2O5-WO3/TiO2 catalyst pores. When adding 0.08 M NH₄NO₃, the catalyst-specific surface area and pore properties decreased, resulting in a decrease in the active area of the catalyst (Table 1). Sulfur reacts directly with V_2O_5 -WO₃/TiO₂ catalysts and decreases NO_X removal efficiency at temperatures below 300 °C, producing ammonium sulfates.³⁶ However, nitrated V_2O_5 -WO₃/TiO₂ catalysts show a high NO_X removal efficiency in the presence of sulfur at temperatures below 300 °C; this high efficiency is largely because the nitrate species on the catalyst surface prevent its reaction with sulfur. Figure 3b shows N_2 selectivity; a trace amount of N_2O in the prepared catalyst was produced at 300 °C. The trend in N₂ selectivity was similar to that of the NO_X removal efficiency, and all prepared catalysts showed a similar N2 selectivity at temperatures ranging from 150–350 $^\circ C$.

We analyzed the NH₃-TPD and H₂-TPR profiles to obtain further information on the acidity and redox ability (Figure 3c,d). Figure 3c shows the NH₃-TPD results for the V₂O₅-WO₃/TiO₂, 0.03 M NH₄NO₃, 0.05 M NH₄NO₃, and 0.08 M NH₄NO₃ catalysts, respectively. The NH₃-TPD profile of all prepared catalysts significantly varied in the temperature range of 100–900 °C. The thermal conductivity detector signals at the lower and higher temperatures were considered to be Brønsted and Lewis acid sites, which are $NH_{4,ads}^+$ species and $NH_{3,ads}$ species, respectively.³⁴ The desorbed NH_3 concentrations for 0.05 M $\rm NH_4NO_3$ and 0.08 M $\rm NH_4NO_3$ were higher, indicative of a greater capacity for NH₃ adsorption on the catalyst surface. According to these results, 0.05 M NH₄NO₃ and 0.08 M NH₄NO₃ contained more Brønsted and Lewis acid sites. Also, we produced H₂-TPR profiles to investigate the V2O5-WO3/TiO2, 0.03 M NH4NO3, 0.05 M NH₄NO₃, and 0.08 M NH₄NO₃ redox properties, respectively. (Figure 4b). V₂O₅-WO₃/TiO₂ has three apparent reduction peaks centered at 418.9, 469.7, and 783.3 °C, which could be assigned to the co-reduction of V^{5+} to V^{3+} , which corresponds to the surface vanadium species, on which we observed the reduction of W^{6+} to W^{4+} and the reduction of W^{4+} to W^0 in tungsten oxide.^{37,38} However, three main reduction peaks of 0.03, 0.05, and 0.08 M NH4NO3 catalysts shifted to lower temperatures at 401.4, 441.2, and 782.7 °C, respectively, and these results suggest that the nitrated surface has a relatively higher NO_x reducibility and reducing amount.

To investigate the formation and transformation of adsorbed species on the prepared catalyst surfaces, in situ FTIR spectra were obtained at 200 $^{\circ}$ C (Figure 4). After applying NH₃ gas to the V_2O_5 -WO₃/TiO₂ catalyst for 20 min, bands (3379, 3250, 3155, 1583, and 1415 cm^{-1}) appeared, corresponding to coordinated NH₃ bound to the Lewis acid sites and ionic NH₄⁺ bound to the Bronsted acid sites. Moreover, 0.05 M NH₄NO₃treated V₂O₅-WO₃/TiO₂ catalyst ammonia was chemically absorbed within 1 min. Furthermore, the Lewis and Bronsted acid site intensities were higher than those of pristine V₂O₅- WO_3/TiO_2 . This demonstrates that the nitrate ions (NO₃⁻) on the 0.05 M NH₄NO₃-treated V₂O₅-WO₃/TiO₂ catalyst surface quickly reacted with the applied ammonia gases and absorbed more ammonia on the surface. Figure 4c,d indicates in situ FTIR spectra of NO_X and oxygen that reacted with the pre-absorbed ammonia on the catalyst surface. After inducing NO and O_2 gases on the catalyst, the number of adsorbed



Figure 5. (a) NO_X removal efficiency under the influence of H₂O. (b) SO₂ and H₂O tolerance of the 4 wt % V₂O₅-10 wt % WO₃/TiO₂ and 0.05 M NH₄NO₃ catalyst at 200 °C (inset shows the microscope image of the plate-type catalyst). Reaction conditions: [NO] & [NH₃] & [SO₂] = 300 ppm, $[O_2] = 5$ vol %, $[H_2O] = 10$ vol %, [GHSV] = 60,000 h⁻¹.

Lewis and Bronsted acid sites gradually decreased with the reduction in NO gas. The V_2O_5 -WO₃/TiO₂ catalyst required 10 min to remove the pre-absorbed ammonia, whereas the catalyst treated with ammonia nitrate quickly reduced NO gas on the surface in 5 min. This is because the nitrate ions (NO_3^-) on the surface produced by the ammonia treatment induced "fast SCR", causing fast chemisorption and reduction.

To apply the effect of ammonia nitrate treatment to a real SCR system, we fabricated plate-type catalysts and then evaluated them in a microreactor from 160-240 °C, as shown in Figure 5a. The V₂O₅-WO₃/TiO₂ catalyst showed a poor NO_X removal efficiency below 240 °C and the efficiency deteriorated with H₂O application into the reactor because the catalyst caused undesired SO₂ oxidation and NH₄HSO₄ and $(NH_4)_2SO_4$ formation. However, after ammonium nitrate treatment of the V_2O_5 -WO₃/TiO₂ catalyst, the NO_X removal efficiency was dramatically enhanced at lower temperatures. Particularly over 180 °C, the efficiency was not significantly decreased by H₂O. The nitrate ions (NO₃⁻) on the surface treated by 0.05 M NH₄NO₃ suppress ammonium sulfate formation and lead to "fast SCR", which improved NO_X removal efficiency. Figure 5b shows that SO₂ and H₂O affected the NO_X removal efficiency of the V_2O_5 -WO₃/TiO₂ catalysts. When SO₂ gas was introduced into the reactor, the NO_X removal efficiencies of both V_2O_5 -WO₃/TiO₂ catalysts and the 0.05 M NH₄NO₃-treated V₂O₅-WO₃/TiO₂ catalyst was maintained at 97 and 90%, respectively. However, H₂O was introduced into the reactor and the removal efficiencies of both of them decreased to 89 and 52%, respectively, and the catalyst without ammonium nitrate treatment deteriorated rapidly. After turning off H₂O and SO₂, the removal efficiency returned to 95 and 80%. While the 0.05 M NH_4NO_3 -treated V_2O_5 -WO₃/TiO₂ catalyst recovered its efficiency immediately, the V₂O₅-WO₃/TiO₂ catalysts did not recover quickly or fully. Because the nitrate ions (NO_3^-) on the surface prevented ammonium sulfate creation and caused "fast SCR", the catalyst after ammonium nitrate treatment exhibited high efficiency and good stability.

4. CONCLUSIONS

In this study, nitrate-treated V_2O_5 -WO₃/TiO₂ catalysts were synthesized by adding ammonium nitrate using a general impregnation method, and the functional groups were controlled by varying the amount of added ammonium nitrate. By adding NH₄NO₃ to the V₂O₅-WO₃/TiO₂ catalyst, the NO₃⁻ and chemisorbed oxygen intensities increased simultaneously, inducing "fast SCR". Nitrate-treated V₂O₅-WO₃/TiO₂ catalysts have more Brønsted and Lewis acid sites and a higher NO_X reducibility. These factors were linked to an improved catalytic performance at low temperatures (i.e., below 300 °C). However, superabundant nitrate blocked the V₂O₅-WO₃/TiO₂ catalyst pores, which caused a decrease in the catalyst active area and decreased the NO_X removal efficiency. Therefore, the proper nitrate, 0.05 M NH₄NO₃ catalyst exhibited the highest NO_X removal efficiency and N₂ selectivity. These results may contribute to the improvement of SCR catalysts and other classes of catalytic systems.

ASSOCIATED CONTENT

Supporting Information

The Supporting Information is available free of charge at https://pubs.acs.org/doi/10.1021/acsomega.3c05423.

XRD patterns; schematic of the V2O5-WO3/TiO2 catalyst synthesis; SEM images; and TEM images and SAED patterns (PDF)

AUTHOR INFORMATION

Corresponding Author

Duck Hyun Lee – Green Materials and Processes R&D Group, Korea Institute of Industrial Technology, Ulsan 44413, Republic of Korea; School of Advanced Materials & Electrical Engineering, Industrial Technology Center for Environment-Friendly Materials, Andong National University, Andong 36729, Republic of Korea; Orcid.org/ 0000-0002-7250-5302; Email: dulee@anu.ac.kr

Authors

- Sun-I Kim Green Materials and Processes R&D Group, Korea Institute of Industrial Technology, Ulsan 44413, Republic of Korea
- Yeong Jun Choi Green Materials and Processes R&D Group, Korea Institute of Industrial Technology, Ulsan 44413, Republic of Korea
- Min Seong Lee Green Materials and Processes R&D Group, Korea Institute of Industrial Technology, Ulsan 44413, Republic of Korea

Complete contact information is available at: https://pubs.acs.org/10.1021/acsomega.3c05423

Author Contributions

[§]S.-I.K. and Y.J.C. contributed equally to this study and should be considered co-first authors.

Funding

This work was supported by a grant from the 2023 Research Fund of Andong National University.

Notes

The authors declare no competing financial interest.

ABBREVIATIONS

AC:activated carbon

BET:Brunauer-Emmett-Teller

CNT:carbon nanotube

EELS:electron energy loss spectroscopy

FTIR:Fourier transform infrared spectroscopy

H₂-TPR:H₂ temperature-programmed reduction

NH₃-TPD:NH₃-temperature-programmed desorption

SCR:selective catalytic reduction

TEM:transmission electron microscopy

TPD:temperature-programmed desorption

XPS:X-ray photoelectron spectroscopy

XRD:X-ray diffraction

XRF:X-ray fluorescence spectrometry

REFERENCES

(1) Tiwary, A.; Williams, I. Air Pollution: Measurement, Modelling and Mitigation; CRC Press, 2018.

(2) Caillol, S.; Caillol, S. Fighting global warming: the potential of photocatalysis against CO2, CH4, N2O, CFCs, tropospheric O3, BC and other major contributors to climate change. *J. Photochem. Photobiol.*, C **2011**, *12*, 1–19.

(3) Godish, T.; Davis, W. T.; Fu, J. S. Air, Quality; CRC Press, 2014.
(4) Radojevic, M. Reduction of nitrogen oxides in flue gases. Environ. Pollut. 1998, 102, 685.

(5) Mahmoudi, S.; Baeyens, J.; Seville, J. P. K. NOx formation and selective non-catalytic reduction (SNCR) in a fluidized bed combustor of biomass. *Biomass Bioenergy* **2010**, *34*, 1393.

(6) Gao, F.; Tang, X.; Yi, H.; Zhao, S.; Li, C.; Li, J.; Shi, Y.; Meng, X. A review on selective catalytic reduction of NOx by NH3 over Mn–based catalysts at low temperatures: catalysts, mechanisms, kinetics and DFT calculations. *Catalysts* **2017**, *7*, 199.

(7) Mladenović, M.; Paprika, M.; Marinković, A. Denitrification techniques for biomass combustion. *Renewable Sustainable Energy Rev.* **2018**, *82*, 3350.

(8) Lee, M. S.; Kim, S. I.; Lee, M. J.; Ye, B.; Kim, T.; Kim, H. D.; Lee, J. W.; Lee, D. H. Effect of catalyst crystallinity on V-based selective catalytic reduction with ammonia. *Nanomaterials* **2021**, *11*, 1452.

(9) Koebel, M.; Elsener, M.; Kleemann, M. Urea-SCR: a promising technique to reduce NOx emissions from automotive diesel engines. *Catal. Today* **2000**, *59*, 335.

(10) Seher, D. H. E.; Reichelt, M.; Wickert, S. Control strategy for NOx – emission reduction with SCR, *SAE Technical Paper Series*, 2003, DOI: 10.4271/2003-01-3362.

(11) Lee, M. S.; Kim, S. I.; Jeong, B.; Park, J. W.; Kim, T.; Lee, J. W.; Kwon, G.; Lee, D. H. Ammonium ion enhanced V2O5-WO3/TiO2 catalysts for selective catalytic reduction with ammonia. *Nanomaterials* **2021**, *11*, 2677.

(12) Lee, M. S.; Choi, Y. J.; Bak, S. J.; Son, M.; Shin, J.; Lee, D. H. Polyol-mediated synthesis of V2O5–WO3/TiO2 catalysts for low-temperature selective catalytic reduction with ammonia. *Nanomaterials* **2022**, *12*, 3644.

(13) Peña, D. A.; Uphade, B. S.; Smirniotis, P. G. TiO2-supported metal oxide catalysts for low-temperature selective catalytic reduction of NO with NH3:I Evaluation and characterization of first row transition metals. *J. Catal.* **2004**, *221*, 421.

(14) Pasel, J.; Käßner, P.; Montanari, B.; Gazzano, M.; Vaccari, A.; Makowski, W.; Lojewski, T.; Dziembaj, R.; Papp, H. Transition metal oxides supported on active carbons as low temperature catalysts for the selective catalytic reduction (SCR) of NO with NH3. *Appl. Catal., B* **1998**, *18*, 199.

(15) Zhao, X.; Huang, L.; Li, H.; Hu, H.; Han, J.; Shi, L.; Zhang, D. Highly dispersed V2O5/TiO2 modified with transition metals (Cu, Fe, Mn, Co) as efficient catalysts for the selective reduction of NO with NH3. *Chin. J. Catal.* **2015**, *36*, 1886.

(16) Fu, M.; Li, C.; Lu, P.; Qu, L.; Zhang, M.; Zhou, Y.; Yu, M.; Fang, Y. A review on selective catalytic reduction of NOx by supported catalysts at 100–300 °C—catalysts, mechanism, kinetics. *Catal. Sci. Technol.* **2014**, *4*, 14–25.

(17) Kapteijn, F.; Singoredjo, L.; Andreini, A.; Moulijn, J. A. Activity and selectivity of pure manganese oxides in the selective catalytic reduction of nitric oxide with ammonia. *Appl. Catal., B* **1994**, *3*, 173.

(18) Zhu, Z.; Liu, Z.; Niu, H.; Liu, S.; Hu, T.; Liu, T.; Xie, Y. Mechanism of SO2 promotion for NO reduction with NH3 over activated carbon-supported vanadium oxide catalyst. *J. Catal.* **2001**, *197*, 6–16.

(19) Li, Q.; Yang, H.; Qiu, F.; Zhang, X. Promotional effects of carbon nanotubes on V2O5/TiO2 for NOx removal. *J. Hazard. Mater.* **2011**, *192*, 915.

(20) Liu, C.; Shi, J.-W.; Gao, C.; Niu, C. Manganese oxide-based catalysts for low-temperature selective catalytic reduction of NOx with NH3: a review. *Appl. Catal., A* **2016**, *522*, 54.

(21) Wang, F.; Wang, P.; Lan, T.; Shen, Y.; Ren, W.; Zhang, D. Ultralow-Temperature NO x Reduction over SmMn2O5 Mullite Catalysts Via Modulating the Superficial Dual-Functional Active Sites. *ACS Catal.* **2022**, *12*, 7622–7632.

(22) Si, Z.; Shen, Y.; He, J.; Yan, T.; Zhang, J.; Deng, J.; Zhang, D. SO2-induced alkali resistance of FeVO4/TiO2 catalysts for NO x reduction. *Environ. Sci. Technol.* **2022**, *56*, 605–613.

(23) Jia, Z.; Shen, Y.; Yan, T.; Li, H.; Deng, J.; Fang, J.; Zhang, D. Efficient NO x abatement over alkali-resistant catalysts via constructing durable dimeric VO x species. *Environ. Sci. Technol.* **2022**, *56*, 2647–2655.

(24) Hu, W.; He, J.; Liu, X.; Yu, H.; Jia, X.; Yan, T.; Han, L.; Zhang, D. SO2-and H2O-Tolerant Catalytic Reduction of NO x at a Low Temperature via Engineering Polymeric VO x Species by CeO2. *Environ. Sci. Technol.* **2022**, *56*, 5170–5178.

(25) Zhao, W.; Dou, S.; Zhang, K.; Wu, L.; Wang, Q.; Shang, D.; Zhong, Q. Promotion effect of S and N co-addition on the catalytic performance of V2O5/TiO2 for NH3-SCR of NOx. *Chem. Eng. J.* **2019**, *364*, 401.

(26) Maqbool, M. S.; Pullur, A. K.; Ha, H. P. Novel sulfation effect on low-temperature activity enhancement of CeO₂-added Sb-V2O5/ TiO2 catalyst for NH3-SCR. *Appl. Catal.*, B **2014**, 152-153, 28.

(27) Li, S.; Huang, W.; Xu, H.; Chen, T.; Ke, Y.; Qu, Z.; Yan, N. Alkali-induced deactivation mechanism of V2O5-WO3/TiO2 catalyst during selective catalytic reduction of NO by NH3 in aluminum hydrate calcining flue gas. *Appl. Catal., B* **2020**, *270*, No. 118872.

(28) Guimon, C.; Gervasini, A.; Auroux, A. XPS study of the adsorption of SO2 and NH3 over supported tin dioxide catalysts used in de-NOx catalytic reaction. *J. Phys. Chem. B* **2001**, *105*, 10316.

(29) Boningari, T.; Ettireddy, P. R.; Somogyvari, A.; Liu, Y.; Vorontsov, A.; McDonald, C. A.; Smirniotis, P. G. Influence of elevated surface texture hydrated titania on Ce-doped Mn/TiO2 catalysts for the low-temperature SCR of NOx under oxygen-rich conditions. J. Catal. 2015, 325, 145.

(30) Wu, Z.; Jin, R.; Liu, Y.; Wang, H. Ceria modified MnOx/TiO2 as a superior catalyst for NO reduction with NH3 at low-temperature. *Catal. Commun.* **2008**, *9*, 2217.

(31) Lee, K. J.; Kumar, P. A.; Maqbool, M. S.; Rao, K. N.; Song, K. H.; Ha, H. P. Ceria added Sb-V2O5/TiO2 catalysts for low temperature NH3 SCR: physico-chemical properties and catalytic activity. *Appl. Catal., B* **2013**, *142-143*, 705.

(32) Liu, C.; Malta, G.; Kubota, H.; Toyao, T.; Maeno, Z.; Shimizu, K.-I. Mechanism of NH3–selective catalytic reduction (SCR) of NO/ NO2 (fast SCR) over cu-cha zeolites studied by in situ/operando infrared spectroscopy and density functional theory. *J. Phys. Chem. C* **2021**, *125*, 21975–21987.

(33) Liu, F.; He, H. Structure-activity relationship of iron titanate catalysts in the selective catalytic reduction of NOx with NH3. *J. Phys. Chem. C* **2010**, *114*, 16929.

(34) Jiang, Y.; Gao, X.; Zhang, Y.; Wu, W.; Song, H.; Luo, Z.; Cen, K. Effects of PbCl2 on selective catalytic reduction of NO with NH3 over vanadia-based catalysts. *J. Hazard. Mater.* **2014**, *274*, 270.

(35) Chen, Z.; Wu, X.; Ni, K.; Shen, H.; Huang, Z.; Zhou, Z.; Jing, G. Molybdenum-decorated V2O5–WO3/TiO2: surface engineering toward boosting the acid cycle and redox cycle of NH3-SCR. *Catal. Sci. Technol.* **2021**, *11*, 1746.

(36) Muzio, L.; Bogseth, S.; Himes, R.; Chien, Y.-C.; Dunn-Rankin, D. Ammonium bisulfate formation and reduced load SCR operation. *Fuel* **2017**, *206*, 180.

(37) Zhang, Y.; Guo, W.; Wang, L.; Song, M.; Yang, L.; Shen, K.; Xu, H.; Zhou, C. Characterization and activity of V2O5-CeO2/TiO2-ZrO2 catalysts for NH3-selective catalytic reduction of NOx. *Chin. J. Catal.* **2015**, *36*, 1701.

(38) Wu, X.; Yu, W.; Si, Z.; Weng, D. Chemical deactivation of V2O5-WO3/TiO2 SCR catalyst by combined effect of potassium and chloride. *Front. Environ. Sci. Eng.* **2013**, *7*, 420.