

Evaluation of the radioactive contamination in fungi genus *Boletus* in the region of Europe and Yunnan Province in China

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Abstract Numerous species of wild-grown mushrooms are among the most vulnerable organisms for contamination with radiocesium released from a radioactive fallout. A comparison was made on radiocesium as well as the natural gamma ray-emitting radionuclide (⁴⁰K) activity concentrations in the fruiting bodies of several valued edible *Boletus* mushrooms collected from the region of Europe and Yunnan Province in China. Data available for the first time for *Boletus edulis* collected in Yunnan, China, showed a very weak contamination with ¹³⁷Cs. Radiocesium concentration activity of *B. edulis* samples that were collected between 2011 and 2014 in Yunnan ranged from 5.2±1.7 to 10±1 Bq kg⁻¹ dry matter for caps and from 4.7±1.3 to 5.5±1.0 Bq kg⁻¹ dry matter for stipes. The mushrooms *Boletus badius*, *B. edulis*, *Boletus impolitus*, *Boletus luridus*, *Boletus pinophilus*, and *Boletus reticulatus* collected from the European locations between 1995 and 2010 showed two to four orders of magnitude greater radioactivity from ¹³⁷Cs compared to *B. edulis* from Yunnan. The nuclide ⁴⁰K in *B. badius* was equally distributed between the caps and stipes, while for *B. edulis*, *B. impolitus*, *B. luridus*, *B. pinophilus*, and *B. reticulatus*, the caps were richer, and for each mushroom, activity concentration seemed to be more or less species-specific.

Keywords China/Europe · Forest · Organic food · Radiocesium · Wild mushrooms

Introduction

The nuclear accident in Chernobyl, which took place on 26th of April 1986 caused large-scale diffusion of radioactivity mostly in the Central and Northern Europe, but it was detected also in other southern areas in Turkey (IAEA 2005; Simsek et al. 2014). Because of that accident, the long-term residual radioactivity in the affected areas comes largely from radiocesium (¹³⁷Cs) (Bulko et al. 2014). Contamination of soils, pastures, and forests with the post-Chernobyl ¹³⁷Cs varied between the regions of Europe, and this fact highly impacted on regional appearance of ¹³⁷Cs in food, feed, mushrooms, grazing cattle, and wildlife and health risk of ¹³⁷Cs and other nuclides to human consumers (Barret et al. 1999; Battiston et al. 1989; De Cort et al. 1998; Smith et al. 1993; Strandberg and Knudsen 1994; Zarubina 2014). Some but minor (about 10 %) residual radioactivity from ¹³⁷Cs in the soils and wild-grown mushrooms still comes from the radioactive fallout which had taken place in the 1950s and 1960s because of the nuclear weapon tests in the atmosphere (Haselwandter et al. 1988; Steinhauser et al. 2013; Taira et al. 2011).

The nuclear power station in Fukushima Dai-ichi collapsed between 11th and 14th of March 2011 after a mega tsunami episode in the northeastern part of the Honshu Island—Tohoku region in Fukushima prefecture, Japan (Yasunari et al. 2011). In result of the Fukushima accident, a large-scale diffusion of radioactivity took place. The radioactivity plume was largely dispersed in the ocean and in small portion on land there (Teramage et al. 2014). At the local scale, mushrooms in the prefecture of Fukushima have been identified as

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the most relevant source of radiocesium intake among vegetables, especially after the first year of the accident (Merz et al. 2015). Foraging of mushrooms bypass, wittingly or unwittingly, the governmental food measuring campaigns which leads to higher intake of radioactive cesium than when consumers bought their products in commercial shops (Hayano et al. 2013; Normile 2013).

The airborne ^{137}Cs that is deposited on land is efficiently taken up and sequestered in fruiting bodies by many wild-grown mushrooms (Macromycetes), which differ in their species-specific capacity to sequester stable Cs as well as many other metallic, non-metallic, and metalloid elements in the flesh (Bakken and Olsen 1990; Barret et al. 1999; Battiston et al. 1989; Brzostowski et al. 2011; Byrne et al. 1979; Eckl et al. 1986; Drewnowska and Falandysz 2015; Falandysz et al. 1994, 2003a, b, 2007a, b, c, d; Gucia et al. 2012; Kirchner and Daillant 1998; Kojta et al. 2012; Vinichuk et al. 2011; Zhang et al. 2010). These chemical elements, depending on their physical and chemical forms, can further be available from soil solution and soil bedrock to fungal mycelia, and sometimes, they can be sequestered in fungal flesh (fruiting bodies) more or less in a dose-effect-related manner. Hence, an elevated content of many metallic elements, metalloids, and Se (nonmetal) can be found in fruiting bodies of exposed populations, while the effectiveness of uptake and sequestration is a function of many variables including biological features related to species of mushroom, mycorrhiza, and geochemical/environmental factors (Falandysz and Borovička 2013).

Edible wild-grown mushrooms are popular organic food and are even items of international trade, and *Boletus* spp. are especially popular in Europe (King Bolete, Bay Bolete, Pine Wood) and especially in Yunnan and several other provinces in China (Falandysz et al. 2011; Wang et al. 2014). This paper reports and compares data on the residual activity from ^{137}Cs as well as natural radionuclide from ^{40}K accumulated in certain *Boletus* mushrooms collected in Poland, Sweden, and Belarus and in Yunnan of China. The contamination of King Bolete (*Boletus edulis*) from Yunnan (a land of mushrooms) is reported internationally for the first time. A major source of the residual ^{137}Cs for Poland without doubt is the Chernobyl accident (Mietelski et al. 2010), while for Yunnan, the likely sources include radioactive fallout from nuclear weapon tests in the 1950s and 1960s and possibly also because of the Chernobyl and Fukushima accidents.

Materials and methods

The fruiting bodies of *Boletus badius*, *B. edulis*, *Boletus impolitus*, *Boletus luridus*, *Boletus pinophilus*, and *Boletus reticulatus* mushrooms were collected in 1995–2014 in Poland, Belarus, and Sweden in Europe and in Yunnan in

China. The mushrooms were collected across of Poland from different locations: Kacze Łęgi, Dziemiany, Sobieszewo, Mojusz, Kępice, Wdzydze, Parchowo, Bory Tucholskie, Olsztynek, Puszcza Piska, Puszcza Notecka, Porążyn, Włoszowa in Świętokrzyskie land, Chochołowska Valley in Tatra Mountains, and Kłodzka Dale in Sudety Mountains. Mushrooms from Belarus were collected from two spatially distant locations: Staroje Janczyna—location in central part, administrative circuit of Minsk, Borysowski Region, and Wasilewicz in Chojniki area of the Gomel Region. Samples of *B. edulis* were also collected from the town of Umeå and its outskirts in northern region of Sweden and from Davingie and Yimen in the Prefecture of Yuxi in the Yunnan Province of China. Fresh fruiting bodies' samples (separately caps and stipes) were sliced using a plastic knife, dried at 65 °C to constant weight, and further pulverized using ceramic mortar and kept in brand new sealed polyethylene bags that were packed into larger bags and kept under dry and clean condition in a laboratory room until analysis. Before determination of activity concentration of radionuclides, the individual mushroom samples were pooled (separately caps and stipes, from 6 to 34 individuals per pool) to obtain one large (many specimens) integrated sample representing each place and year (Table 1).

Activity concentrations of ^{137}Cs and ^{40}K were determined using gamma spectrometer with coaxial HPGe detector with a relative efficiency of 18 % and a resolution of 1.9 keV at 1.332 MeV (with associated electronics). The detector was coupled with an 8192-channel computer analyzer and GENIE 2000 software (Zalewska and Staniewski 2011). The equipment was calibrated using a multi-isotope standard, and the method was fully validated. The laboratory involved was subjected for routine checks to ensure the high standards of analytical quality and analytical control as well as took part in the intercomparison exercises organized by IAEA-MEL Monaco (IAEA-414, Irish and North Sea Fish) (Zalewska and Staniewski 2011) to verify the reliability and accuracy of the method. All numerical data gained were recalculated for dehydrated fungal material (at 105 °C) and exact date of the sample collection.

Results

Data on radioactivity (expressed in Bq kg^{-1} dry matter) of ^{137}Cs and ^{40}K in caps and stipes of the *Boletus* mushrooms are summarized according to species, place of origin, size of sample, and year of collection (Table 1). Samples of *B. edulis* were from Europe and China. The Chinese mushrooms were collected at altitude of 1600–1650 m above sea level in the Yuxi Prefecture of the mountainous Province of Yunnan in 2011–2014 (Fig. 1). The mushrooms such as *B. badius*, *B. edulis*, *B. impolitus*, *B. luridus*, *B. pinophilus*, and *B. reticulatus* collected in the region of Europe (Belarus,

Table 1 ^{137}Cs and ^{40}K in *Boletus* spp. (Bq kg^{-1} dry matter; activity concentration \pm an instrumental counting error)

| Place and year of collection (number of specimens, <i>n</i>) in a pool | ^{137}Cs | | ^{40}K | |
|---|--------------------|------------------|--------------------|---------------|
| | Whole fruit bodies | | Whole fruit bodies | |
| | Caps | Stipes | Caps | Stipes |
| <i>Boletus badius</i> Pers. | | | | |
| (1) ^a Poland, Bory Tucholskie, 2000 (<i>n</i> =19) | 5105 \pm 45 | 4611 \pm 48 | 1293 \pm 52 | 1083 \pm 43 |
| (2) Poland, Puszcza Notecka, Jesionna, 2008 (<i>n</i> =32) | 970 \pm 8 | 687 \pm 19 | 1060 \pm 28 | 713 \pm 52 |
| (3) Poland, Porążyn, 2008 (<i>n</i> =29) | 45 \pm 2 | | 1240 \pm 55 | |
| (4) Belarus, Borysów, Staroje Janczyna, 2010 (<i>n</i> =34) | 1430 \pm 18 | 1373 \pm 9 | 818 \pm 136 | 828 \pm 105 |
| (5) Belarus, Chojniki, Wasilewicz, 2010 (<i>n</i> =38) | 20,758 \pm 196 | 14,799 \pm 123 | 1090 \pm 175 | WD |
| <i>Boletus edulis</i> Bull. | | | | |
| (6) Sweden, Umeå, and outskirts, 1995 (<i>n</i> =15) | 1102 \pm 15 | 904 \pm 12 | 904 \pm 98 | 668 \pm 95 |
| (7) Poland, Pomerania, Mojusz, 2007 (<i>n</i> =11) | 1358 \pm 17 | | 912 \pm 126 | |
| (8) Poland, Pomerania, Parchowo, 2010 (<i>n</i> =15) | 497 \pm 9 | 265 \pm 4 | 731 \pm 107 | 319 \pm 76 |
| (9) Poland, Tatra Mountains ^b , 1999 (<i>n</i> =12) | 227 \pm 5 | | 762 \pm 111 | |
| (10) Poland, Sudety Mt's, Kłodzka Dale, 2000 (<i>n</i> =10) | 5722 \pm 5 | 3485 \pm 3 | 903 \pm 118 | 368 \pm 90 |
| (20) China, Yunnan, Yuxi, Yimen, 2011 (<i>n</i> =12) | 10 \pm 1 | 5.0 \pm 1.0 | 740 \pm 86 | 360 \pm 61 |
| (20) China, Yunnan, Yuxi, Yimen, 2012 (<i>n</i> =10) | 5.4 \pm 1.2 | 5.5 \pm 1.0 | 810 \pm 74 | 500 \pm 65 |
| (21) China, Yunnan, Yuxi, Dayingjie, 2013 (<i>n</i> =15) | 5.2 \pm 1.7 | 4.9 \pm 1.1 | 630 \pm 140 | 470 \pm 91 |
| (21) China, Yunnan, Yuxi, Dayingjie, 2014 (<i>n</i> =15) | 7.9 \pm 1.5 | 4.7 \pm 1.3 | <120 | <140 |
| <i>Boletus impolitus</i> Fr. | | | | |
| (11) Poland, Warmia land, Olsztynek, 2003 (<i>n</i> =15) | 276 \pm 6 | 150 \pm 4 | 608 \pm 106 | 936 \pm 91 |
| <i>Boletus luridus</i> Soverby | | | | |
| (12) Poland, Sobieszewo, 2000 (<i>n</i> =23) | 3533 \pm 36 | 1007 \pm 17 | 1008 \pm 126 | 309 \pm 136 |
| (13) Poland, Pomerania, Kępice, 2003 (<i>n</i> =15) | 245 \pm 8 | 72 \pm 4 | WD | WD |
| (14) Poland, Świętokrzyskie land ^c , 2007 (<i>n</i> =12) | 188 \pm 6 | 102 \pm 3 | 468 \pm 122 | 218 \pm 82 |
| <i>Boletus pinophilus</i> Pilát & Dermek | | | | |
| (15) Poland, Wdzydze Landscape Park, 1998 (<i>n</i> =14) | 970 \pm 18 | 631 \pm 9 | 631 \pm 154 | 358 \pm 87 |
| (16) Poland, Pomerania, Dziemiiany, 2000 (<i>n</i> =14) | 810 \pm 14 | 425 \pm 9 | 686 \pm 120 | WD |
| (17) Poland, Puszcza Notecka, Jesionna, 2000 (<i>n</i> =6) | 872 \pm 17 | 564 \pm 8 | 1075 \pm 116 | 465 \pm 108 |
| (18) Poland, Puszcza Piska, 2000 (<i>n</i> =14) | 1195 \pm 13 | 431 \pm 6 | 638 \pm 95 | 221 \pm 91 |
| <i>Boletus reticulatus</i> Schaeff. | | | | |
| (19) Poland, TLP, Kacze Łęgi, 2006 (<i>n</i> =20) | 1094 \pm 15 | 498 \pm 8 | 905 \pm 122 | 698 \pm 101 |
| (4) Belarus, Borysów, Staroje Janczyna, 2010 (<i>n</i> =18) | 393 \pm 5 | 363 \pm 8 | 790 \pm 79 | 715 \pm 119 |
| (5) Belarus, Chojniki, Wasilewicz, 2010 (<i>n</i> =15) | 6614 \pm 109 | 3482 \pm 30 | 687 \pm 100 | 405 \pm 102 |

WD without data, TLP Trójmiejski Landscape Park

^a See at the map (Fig. 1)

^b Chochołowska Valley

^c Outskirts of the Włoszowa town

Poland, Sweden) in 1995–2010 showed two to four orders of magnitude (depending on species and place) greater activity concentration of ^{137}Cs when compared to *B. edulis* collected in Yunnan in 2011–2014 (Table 1).

The nuclide, ^{40}K , is a normal constituent of total K which is an important nutrient and the most abundant element in the fruiting bodies of mushrooms with a symbiotic or saprophyte life cycle. In *B. edulis* from Yimen in Yunnan, the activity of ^{40}K was similar to that noted for the samples from Poland and Sweden (Table 1). Significantly lower values, less than

120 Bq kg^{-1} dm in caps and less than 140 Bq kg^{-1} dm in stipes, were found in mushrooms collected in 2014 in the Dayingjie region of Yunnan (the same was observed for ^{40}K). This may be an indication of the deficiency of this important mineral nutrient in soils in Dayingjie, and this is worthy of further investigation.

Distribution of ^{40}K between the two morphological parts of the fruiting bodies for *B. badius* (caps and stipes) was nearly equal for most of the sites. The exception to this pattern was for samples from Wasilewicz (Table 1). For *B. edulis*,

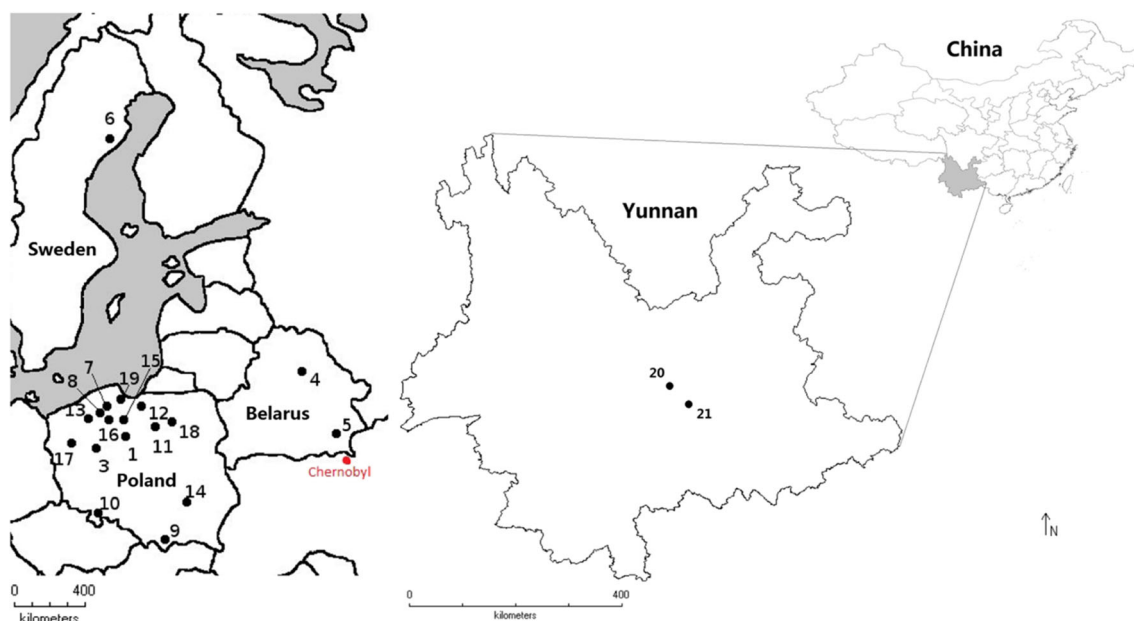


Fig. 1 Localization of the sampling sites (1–21; for details see in Table 1)

B. luridus, *B. pinophilus*, and *B. reticulatus*, the caps were frequently richer in ^{40}K than the stipes, and only in the case of *B. impolitus* was the opposite characteristic observed.

Data obtained for ^{137}Cs in *B. edulis* from Kłodzka Dale in the Sudety Mountains (southwestern Poland) showed relatively high contamination of samples with activity in caps of $5700 \pm 2 \text{ Bq kg}^{-1} \text{ dm}$. Also, samples of *B. edulis* from the region of Umeå in Sweden collected in 1995 ($1102 \pm 15 \text{ Bq kg}^{-1} \text{ dm}$ in caps and $904 \pm 12 \text{ Bq kg}^{-1} \text{ dm}$ in stipes) and Mojusz in the Pomerania land of Poland collected in 2007 ($1358 \pm 17 \text{ Bq kg}^{-1} \text{ dm}$ in whole fruiting body) were substantially contaminated with ^{137}Cs . The lowest activity concentrations of ^{137}Cs in *B. edulis* were found in samples gathered in the other Pomeranian region (Parchowo) in 2010 ($497 \pm 9 \text{ Bq kg}^{-1} \text{ dm}$ in caps and $265 \pm 4 \text{ Bq kg}^{-1} \text{ dm}$ in stipes) (Table 1). In contrast, the activity concentrations of ^{137}Cs in *B. edulis* from the Yuxi region of Yunnan were very low, i.e., from 5.2 ± 1.7 to $10 \pm 1 \text{ Bq kg}^{-1} \text{ dm}$ for caps and 4.7 ± 1.3 to $5.5 \pm 1.0 \text{ Bq kg}^{-1} \text{ dm}$ for stipes (Table 1).

High activity of ^{137}Cs were found also in other *Boletus* species: *B. luridus* collected from the forest with sandy soil bedrock at the Baltic Sea coastal place of the Sobieszewo Island near the city of Gdańsk, *B. pinophilus* from Puszcza Piska, and in *B. reticulatus* from Trójmiejski Landscape Park near the city of Gdynia (Table 1). The highest values were found in *B. luridus*; in 2000, the concentrations reached $3500 \pm 36 \text{ Bq kg}^{-1} \text{ dm}$ in caps and $1000 \pm 17 \text{ Bq kg}^{-1} \text{ dm}$ in stipes, while in *B. pinophilus* and in *B. reticulatus*, they were comparable and at the concentration of $1000 \text{ Bq kg}^{-1} \text{ dm}$ in caps and $400 \text{ Bq kg}^{-1} \text{ dm}$ in stipes. High levels of contamination with ^{137}Cs were found in *B. reticulatus* from the outskirts of Wasilewicz in the Chojniki District in the Gomel region of Belarus (Fig. 1). The activity found in this

area was $6600 \pm 109 \text{ Bq kg}^{-1} \text{ dm}$ in caps and $3500 \pm 30 \text{ Bq kg}^{-1} \text{ dm}$ in stipes, lower than the values observed in *B. badius* collected at the same place and time, indicating possible differences in the ^{137}Cs sequester capacity between these two species (Table 1).

Discussion

B. badius is well known to be susceptible to contamination with radiocesium (Malinowska et al. 2006). The samples from the post-Chernobyl polluted region of Gomel in Wasilewicz, Belarus, collected in 2010 contained high amounts of ^{137}Cs , i.e., around $21,000 \text{ Bq kg}^{-1}$ dry matter (dm) in caps and around $15,000 \text{ Bq kg}^{-1} \text{ dm}$ in stipes (Table 1). The activity of ^{137}Cs in *B. badius* from Poland varied depending on the sampling locations, and this could be roughly related to the regional differences in deposition of ^{137}Cs on and in soils because of the Chernobyl nuclear accident (Grodzinskaya and Haselwandter 2003; Haselwandter et al. 1988; Mietelski et al. 2010). The highest values of $5100 \pm 45 \text{ Bq kg}^{-1} \text{ dm}$ in caps and $4600 \pm 48 \text{ Bq kg}^{-1} \text{ dm}$ in stipes were found in samples from the Bory Tucholskie site collected in the year 2000. A slightly lower activity of ^{137}Cs , $4800 \pm 61 \text{ Bq kg}^{-1} \text{ dm}$ in caps and $2180 \pm 190 \text{ Bq kg}^{-1} \text{ dm}$ in stipes, was found in *B. badius* collected in 1995–1996 from the complex forests of the Wdzydze Landscape Park which is very close to Bory Tucholskie (Malinowska et al. 2006; Falandysz et al. 2003a, b). The mushroom *B. badius* from two other large forest complexes of the Puszcza Notecka within the outskirts of Porążyn (Fig. 1) that was sampled in 2008 showed much lower levels of contamination when compared to the corresponding values

for mushrooms sampled elsewhere in Poland in the 1990s by other researchers with $970 \pm 8 \text{ Bq kg}^{-1} \text{ dm}$ in the caps and $45 \pm 2 \text{ Bq kg}^{-1} \text{ dm}$ in the stipes (Table 1) (Malinowska et al. 2006).

The global radioactive fallout from nuclear weapon tests in the 1950s and 1960s and fallout from the Chernobyl accident has to be taken into account as a source of ^{137}Cs accumulated in *B. edulis* growing in Europe (García et al. 2015; Malinowska et al. 2006; Mietelski et al. 2010). The ^{137}Cs activity concentrations in *B. badius* and *B. reticulatus* from Belarus, *B. badius* from Poland, and *B. edulis* from the Sudety Mountains in Poland are consistent with reported ^{137}Cs general picture and “hot spot” deposition for regions of Belarus and Poland due to the Chernobyl accident (De Cort et al. 1998; Mietelski et al. 2010).

There is no information available to indicate that the most recent radioactivity release from the Fukushima accident affected Yunnan. Lack of gamma-ray radiation from ^{134}Cs in mushrooms sampled in Yunnan directly after the Fukushima accident in 2011 up to 2014 in this study (Table 1) and two other reports indicated that the contribution of Fukushima to the total radiocesium deposited there should be considered as negligible (Falandysz et al. 2015; Wang et al. 2015). On the other hand, earlier (pre-Fukushima accident period) data on the occurrence of ^{137}Cs in wild-grown mushrooms from Japan and Taiwan (in Asia) showed negligible contamination and thus can indirectly reflect depositions of small amounts of airborne ^{137}Cs after the Chernobyl accident and previous nuclear weapon tests in the atmosphere (^{134}Cs because of short life time, $t_{1/2}=2$ years, decayed until Fukushima accident) (Muramatsu et al. 1991; Tsukada et al. 1998; Wang et al. 1998). Previous data on ^{134}Cs and ^{137}Cs in *Boletus* spp. from Yunnan and other regions of China are lacking. The activity concentrations from ^{137}Cs for samples from Yimen and Dayingjie (Yuxi Prefecture) were of the same order of magnitude, and this may indicate similarities in radioactive fallout there.

Low activity concentrations of ^{137}Cs determined in fruiting bodies of *B. edulis* from Yunnan presented in this study as well as in fruiting bodies of pan-tropical mushroom *Macrocybe gigantea* (median value for dehydrated caps was 4.5 Bq kg^{-1} and 5.4 Bq kg^{-1} for stipes) and sclerotia of fungus *Wolfiporia extensa* (range from <1.4 to $7.2 \pm 1.1 \text{ Bq kg}^{-1} \text{ dm}$) (Falandysz et al. 2015; Wang et al. 2015) definitely imply that radioactive contamination, which could have resulted from both the recent (Fukushima) and earlier (Chernobyl and/or nuclear weapon tests) sources, is negligible in this region.

The content of stable Cs in fruiting bodies of mushrooms such as *B. edulis* and *B. badius* and also *Cortinarius caperatus*, *Cortinarius saturatus*, *Cortinarius traganus*, *Dermocybe semisanguinea*, *Hydnum repandum*, *Laccaria amethystina*, *Lactarius allis*, *Lactarius piperatus*, *Lactarius rufus*, *Paxillus involutus*, *Suillus luteus*, *Tricholoma album*, *Tricholoma flavovirens*, *Tricholoma fulvum*, *Tricholoma*

robustum, *Ramaria pallida*, *Sarcodon scabrum*, and *Xerocomus chrysenteron* was greater when compared to many others (Bakken and Olsen 1990; Byrne et al. 1979; Falandysz et al. 2001b, 2007a, b, c, 2008; Horyna and Řanda 1988; Karadeniz and Yaprak 2010; Tsukada et al. 1998; Yoshida et al. 2004). Nevertheless, data for ^{134}Cs , ^{137}Cs , and stable ^{133}Cs obtained for the same samples of mushrooms available from published literature is little is little (Karadeniz and Yaprak 2010; Yoshida et al. 2000).

The relative abundance of ^{137}Cs in mushroom as determined in this study for *Boletus* mushrooms from Europe can be attributed to three factors: species-specific uptake, requirement of this (stable Cs) element by the mushroom, and lastly by forest soil contamination with ^{137}Cs at the sampling sites. Of secondary importance is the soil depth where the mushroom developed its mycelia, which is species-specific (Byrne 1998; Falandysz et al. 2014a; Stijve and Poretti 1990). The bulk of radioactive fallout as well as other airborne elemental contaminants are deposited and adsorbed on the top organic layer of forest soils. Some mushrooms with shallow mycelia can accumulate them readily and in considerable concentrations (Falandysz et al. 2014b; Mietelski et al. 2010; Stijve and Poretti 1990), and they subsequently infiltrate deeper into the soil layers (and the mycelia therein). This is dependent on the element's concentration, while topography, humidity of climate, and soil structure can favor quicker vertical passage of the element under consideration into the deeper layers of the soil alongside with infiltrating rain, taking a portion of nuclides that were not readily adsorbed by litter and organic horizon of soil into deeper layers (Teramage et al. 2014).

The low concentrations of ^{137}Cs found in *B. edulis* from Yunnan are important for the inhabitants of the region. One reason is because the *Boletus* mushrooms are very popular organic foods, and numerous species are collected in Yunnan, which is well known for the mushrooms that can be found there (Wang et al. 2014; Wiekaj et al. 2014; Zhang et al. 2010). Another reason is that in Yunnan, *Boletes* mushrooms are traditionally fried with hot vegetable oil using a wok (Chinese pan) but without pre-boiling (blanching). Blanching is a common procedure when cooking or pickling *Boletus* mushrooms in many countries including Poland. It results in the leaching out of some of the mushrooms' water and water-soluble constituents (including ^{137}Cs) into the water phase, thereby reducing their content in the final mushroom dish (Barret et al. 1999).

The radioactive isotopes ^{134}Cs and ^{137}Cs can account for the “total” Cs (stable Cs) content when measured using an instrumental method that is unable to differentiate between $^{134/137}\text{Cs}$ and ^{133}Cs . This is particularly important because the content of $^{134/137}\text{Cs}$ in mushrooms is associated with ^{133}Cs (Karadeniz and Yaprak 2010; Yoshida et al. 2000).

As stated earlier, there is a dearth of data on ^{137}Cs in mushrooms from China's mainland (Marzano et al. 2001). No available information is found concerning the content of ^{134}Cs and ^{137}Cs , stable ^{133}Cs , and ^{40}K in *B. edulis* from Yunnan or other closely related species like *B. pinophilus* or *B. reticulatus*—all three are naturally rich in selenium and certain other chalcophile elements (Falandysz 2008, 2013; Falandysz et al. 2001a, 2007d, 2011; Frankowska et al. 2010; Costa-Silva et al. 2011). They and some other related species are naturally more abundant in stable ^{133}Cs than many other mushrooms (Falandysz et al. 2001b, 2007d, 2008; Horyna and Řanda 1988).

The content of potassium (K) is high in fruiting bodies of the mycorrhizal type mushrooms, e.g., at $29,000 \pm 3000 \text{ mg kg}^{-1} \text{ dm}$ in caps of *B. edulis*, from $38,000 \pm 4000$ to $55,000 \pm 2000 \text{ mg kg}^{-1} \text{ dm}$ in *Cantharellus cibarius*, and from $28,000 \pm 3000$ to $50,000 \pm 14,000 \text{ mg kg}^{-1} \text{ dm}$ in caps and from $21,000 \pm 4000$ to $35,000 \pm 4000$ in stipes of *Suillus grevillei* (Chudzyński and Falandysz 2008; Falandysz and Drewnowska 2015; Frankowska et al. 2010). Nuclide ^{40}K is a long-living isotope and is a natural part of total K, which is an essential element and undergoes a homeostatic regulation in fruiting bodies by mushrooms (Falandysz and Borovička 2013; Stijve 1996).

^{40}K is a dominant portion of the natural gamma-radioactivity contained in the flesh of the fruiting bodies of mushrooms (Karadeniz and Yaprak 2010). The activity concentration of ^{40}K had a little fluctuation and was a substantial portion of the total gamma-radioactivity contained in the flesh of the fruiting bodies of all the *Boletus* mushrooms foraged in Europe in this study, and in practice, almost a solely source in samples from Yunnan, where >100-fold exceeded activity concentration of ^{137}Cs .

In conclusion, the amount of ^{40}K found in the fruiting bodies of a particular species of *Boletus* mushrooms collected from spatially distant places was more or less species-specific and stable with respect to time. On the other hand, a spatial pattern of activity of ^{137}Cs in these mushrooms was mosaic-like, and this could be attributed to differences in the density of fallout and local soil conditions. For some areas of land located well away from the Chernobyl nuclear unit, the mosaic-like pattern of ^{137}Cs accumulated in mushrooms (*Boletus* mushrooms) can reflect possible local differences in the density of fallout and the type of nuclides available to fungi when compared to what can be deduced from the expected pattern associated with pollution by ^{137}Cs in European soils. To get a better knowledge on exposure rates and risk to consumers from ^{137}Cs and other radionuclides in wild-grown mushrooms and especially the exposure to individuals, villagers and other high-level consumers of mushrooms need to be highlighted at any locality (e.g., forest) where mushrooms are highly contaminated.

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Conflict of interest The authors declare that they have no conflict of interest.

Ethical approval This article does not contain any studies with human participants or animals performed by any of the authors.

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