Research Article

Advanced Treatment of Pesticide-Containing Wastewater Using Fenton Reagent Enhanced by Microwave Electrodeless Ultraviolet

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The photo-Fenton reaction is a promising method to treat organic contaminants in water. In this paper, a Fenton reagent enhanced by microwave electrodeless ultraviolet (MWEUV/Fenton) method was proposed for advanced treatment of nonbiodegradable organic substance in pesticide-containing biotreated wastewater. MWEUV lamp was found to be more effective for chemical oxygen demand (COD) removal than commercial mercury lamps in the Fenton process. The pseudo-first order kinetic model can well describe COD removal from pesticide-containing wastewater by MWEUV/Fenton, and the apparent rate constant (*k*) was 0.0125 min⁻¹. The optimal conditions for MWEUV/Fenton process were determined as initial pH of 5, Fe²⁺ dosage of 0.8 mmol/L, and H₂O₂ dosage of 100 mmol/L. Under the optimal conditions, the reaction exhibited high mineralization degrees of organics, where COD and dissolved organic carbon (DOC) concentration decreased from 183.2 mg/L to 36.9 mg/L and 43.5 mg/L to 27.8 mg/L, respectively. Three main pesticides in the wastewater, as Dimethoate, Triazophos, and Malathion, were completely removed by the MWEUV/Fenton process within 120 min. The high degree of pesticides decomposition and mineralization was proved by the detected inorganic anions.

1. Introduction

The widespread use of pesticides in the past decades represents serious water pollutants [1]. Various types of pesticides residues were frequently detected in surface water and aroused great public concern. The pesticides will cause potential adverse health risks even at low concentration (pg/L to ng/L) [2]. These pesticides, retained in agrochemical wastewater, are resistant to conventional biological treatment owing to their high toxicity and biological persistence. It was reported that no significant decrease in pesticide content occurred after the biological treatment and remaining recalcitrant organic carbon mainly due to pesticide molecules [3].

Advanced oxidation processes (AOPs) were expected to decompose those typically stable products into carbon dioxide, water, and inorganics or, at least, transform them into harmless compounds [4]. AOPs have been successfully applied for the removal of recalcitrant substances from organic wastewater in recent years [5-8]. Among AOPs, photo-Fenton is considered as a promising process to generate oxidizing species for pollutants degradation [9, 10]. Photo-Fenton reaction requires the presence of Fe²⁺ and hydrogen peroxide under the UV radiation to produce highly oxidative hydroxyl radicals which react with the organic pollutants and lead to the complete mineralization [11]. Photo-Fenton is effective in treating wastewater with various organic contaminants, such as dye [12], 2,4-dichlorophenol [13], and nonylphenol polyethoxylate [14]. It has been concluded that the introduction of UV light significantly promotes the degradation efficiency by photoreducing Fe^{3+} to Fe^{2+} and producing additional hydroxyl radicals. Thus, UV light source is a crucial design of photo-Fenton reactor.

Microwave electrodeless ultraviolet (MWEUV) was a new-type UV light source developed in recent years [15, 16]. It includes a mercury vapor lamp powered by microwave energy and emitted stable UV radiation. The MWEUV lamp would lead to a further development of simple and high efficiency photochemical reactor owing to its unique advantages [17-19]: high UV radiant power, long lifetime, and being immersed into reaction solution with adaptable lamp shapes. Particularly, both UV and microwave radiations are available simultaneously by using microwave energy. Many studies have reported successful applications of MWEUV lamp as light source in photolysis [20], UV/H₂O₂ [21], and heterogeneous or homogeneous photocatalysis processes [22-24] for organic wastewater treatment. It has been proved that the simultaneous application of microwave power and UV light exhibits higher efficiency in photochemical processes.

In recent years, several advanced oxidation processes for the treatment of pesticide-containing wastewater have been proposed. However, the degradation efficiency for recalcitrant pesticide substances did not meet the demand of commercial application. In this research, the MWEUV lamp was employed for enhancing Fenton process to remove nonbiodegradable organic pesticide in pesticide-containing effluent after biotreatment. Optimal parameters for COD removal by the MWEUV/Fenton process were determined. Oxidation degree and mineralization performance of residual pesticide in wastewater were also evaluated in terms of average oxidation state (AOS), carbon oxidation state (COS), DOC, and inorganic anions concentration.

2. Experimental Methodology

2.1. Materials and Reagents. Fresh wastewater was generated from pesticide container washing process in a chemical products estate (Guangdong, China). The sequential processes, coagulation, precipitation, and biodegradation, have been employed for elimination of suspended solids and biodegradable organic substances from the pesticidecontaining wastewater. Physical/chemical characteristics of fresh wastewater and the biotreated wastewater were showed in Table 1. The concentration of COD and pesticides in effluent wastewater without any advanced treatments still exceeded the discharge limits imposed by Discharge Limits of Water Pollutants (Local Standard: DB44/26-2001).

 $\rm H_2O_2$ (30%, w/w) and FeSO₄·7H₂O were prepared at a predetermined concentration as Fenton reagent. H₂SO₄ (1.0 mmol/L) and NaOH (1.0 mmol/L) were used for pH adjustment. All chemicals were of analytical grade, purchased from Sinopharm Chemical Reagent Co., Ltd (Shenzhen, China).

2.2. Experimental Setup and Methods. A U-shaped MWEUV lamp was used in this study. The lamp is made by quartz tubes filled with 1 mg mercury and 0.66 kPa argon. The external diameter is 20 mm and effective length is 250 mm. The MWEUV lamp had prominent UV emission bands at 254, 313, 365, and 405 nm. Microwave generator (Haier Co.

TABLE 1: Physical/chemical characteristics of fresh wastewater and the biotreated wastewater.

Concentration (mg/L)	Fresh wastewater	Effluent after biological treatment
pН	6.0~6.8	~7.0
TSS	618.85 ± 70.90	<15
COD	1540.34 ± 202.12	183.2 ± 9.95
BOD ₅	336.58 ± 20.86	<15
Dimethoate (C ₅ H ₁₂ NO ₃ PS ₂)	15.66 ± 3.38	14.71 ± 0.79
Triazophos (C ₁₂ H ₁₆ N ₃ O ₃ PS)	6.11 ± 0.63	5.87 ± 0.89
Malathion (C ₁₀ H ₁₉ O ₆ PS ₂)	31.65 ± 4.77	24.53 ± 4.16



FIGURE 1: Schematic diagram of experimental setup.

Ltd, China) was operated with 80 W output at frequency of 2.45 GHz. The output power of MWEUV lamp was measured by monitoring the temperature of the solution with and without the MWEUV lamp [25]. According the calculation, approximate 40 W was supplied to excite MWEUV lamp and the rest power was consumed or converted to heat.

Schematic diagram of the experimental setup is shown in Figure 1. 1000 mL pesticide-containing wastewater was added in the glass cylindrical reactor (available capacity of 1000 mL). Then, required dosage of Fenton reagent was added into the solution and mixed by a magnetic stirrer (HJ-3, Jingda Instrument Company, China). Microwave generator was applied to excited MWEUV lamp emitting UV irradiation during the reaction. Solution temperature during the reaction was kept constant at $25 \pm 0.5^{\circ}$ C by circulating solution to a cooler with a pump (DP-60, Seisun pumps Co., Ltd, China). Samples were withdrawn at predetermined time intervals, and the final pH values were adjusted to lower than 3 before analysis. All the runs were triplicated.

In this study, the experiments were carried out as follows and the results were compared: (1) Fenton oxidation run, (2) UV/Fenton run, assisted by a commercial UV mercury lamp (40 W of output, CREATOR, China), and (3) MWEUV/Fenton run, assisted by MWEUV lamp.

2.3. Analytical Methods. The pH value of solution was measured using a pH meter (PHB-5, INESA Instrument, China). Total suspended solids (TSS) were measured by gravimetry method. Chemical oxygen demand (COD) was determined according to the Standard Methods [26]. Dissolved organic carbon (DOC) was measured by TOC analyzer (Multi N/C, Jena). NO_3^- , SO_4^{2-} , and PO_4^{3-} were quantified by ion chromatography (Dionex ICS-900; column Ion Pac AS23; suppressor MMS 300). Isocratic elution was done with KOH solution, at a flow rate of 1.0 mL/min, for anions analyzed. The quantitative analysis of pesticides was performed by GC/MS (Agilent, 6890N-5975B). Samples were subsequently injected in GC, coupled with a mass spectrometer. The mass detector was operated in MRM (Multiple Reaction Monitoring) mode, selecting specific transitions for each pesticide. The final optimized method allowed the concurrent detection of three pesticides, during chromatographic runs of 20 min.

3. Results and Discussion

3.1. Evaluation of Different Treatment Processes. Photo-Fenton process produces hydroxyl radicals with powerful oxidizing ability to degrade organic pollution. The major reactions to form hydroxyl radicals are as follows [5]:

$$\operatorname{Fe}^{2+} + \operatorname{H}_2\operatorname{O}_2 \longrightarrow \operatorname{Fe}^{3+} + \bullet\operatorname{OH} + \operatorname{OH}^-$$
 (1)

$$\operatorname{Fe}^{3+} + \operatorname{H}_2 O \longrightarrow \operatorname{Fe} (OH)^{2+} + H^+$$
 (2)

$$Fe (OH)^{2+} \xrightarrow{h\nu} Fe^{2+} + \bullet OH$$
(3)

$$H_2O_2 \xrightarrow{h\nu} \bullet OH + \bullet OH$$
 (4)

In this study, treatment of pesticide-containing wastewater in different processes was evaluated. Specially, Fenton oxidation assisted by MWEUV and a commercial mercury UV lamp was compared. As shown in Figure 2, Fenton oxidation only achieved 48.7% of COD removal after 120 min, while COD removal efficiency was significantly improved by introducing UV light. The COD concentration of pesticidecontaining wastewater with UV/Fenton treatment decreased 64.0%. Such improvement was attributed to two facts: on the one hand, the reduction of Fe³⁺ (Equations (2), (3)) under UV irradiation regenerated Fe²⁺ that can further participate in the Fenton reaction and produce •OH radicals. Such recycling enhanced the •OH radical production and provides the UV/Fenton system sufficient activity for organic molecule decomposition. On the other hand, photolysis of H₂O₂ (Equation (4)) generates more OH radicals. It is also notable that MWEUV lamp as the light source is more effective in the Fenton system than a commercial mercury lamp. 72.1% of COD in wastewater was removed by MWEUV/Fenton treatment within 120 min. In this experiment, wastewater temperature during the reaction was kept constant. It means that microwave irradiation accelerates chemical reactions by nonthermal effects such as polarization, dielectric properties, nuclear spin rotation, and spin alignment [27]. It was reported that simultaneous irradiation of microwave and UV-Vis light led to effective reactant decomposition

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FIGURE 2: COD removal in different treatment processes. Experimental conditions: initial pH value of 5, H_2O_2 dosage of 40 mmol/L, and Fe²⁺ dosage of 0.6 mmol/L.

by microwave ultraviolet electrodeless lamp introduced in photolysis, photocatalysis, and UV/H₂O₂ systems [25, 27, 28].

3.2. Effect of Initial pH Value. The pH of wastewater is an important parameter for the MWEUV/Fenton process because it affects the decomposition of hydrogen peroxide [29] and the hydrolytic speciation of the ferric ion species [30]. To test the effect of initial pH, experiments were carried out at initial pH ranging from 2 to 10; the results of COD removal by MWEUV/Fenton are shown in Figure 3(a). It is clear that removal efficiency is highly dependent on initial pH, and the optimum pH is between 2 and 4. The COD removal efficiency decreased from 73.8% to 41.7% according to the increase of pH from 5 to 9. Many researchers have also reported the optimum pH as 3-5 for conventional Fenton or photo-Fenton processes [31-33]. Neamtu et al. [34] summarized various photoactive species of iron formed at different pH conditions. The dominant species at pH 1-2 is $Fe[H_2O]_6^{3+}$, 2-3 is for $Fe[OH][H_2O]_5^{2+}$, and 3-4 is for $Fe[OH]_2[H_2O]_4^{+}$. Fe(OH)²⁺ species is reported to have the highest photoreactivity in Fenton's oxidation [35]. At acidic pH, Fe³⁺ hydroxyl complexes are highly soluble and Fe(OH)²⁺ is the predominant form of the Fe^{3+} hydroxyl complexes. At higher pH, the concentration of Fe[OH]²⁺ complex ions in solution decreased with the precipitation of ferrous ion as oxyhydroxides [36]. Under the alkaline solution condition, the removal of pollutant is mainly attributed to adsorption and coagulation rather than Fenton oxidation [37].

3.3. Effect of H_2O_2 Dosage. Effect of H_2O_2 dosage on COD removal from pesticide-containing wastewater by MWEUV/Fenton is shown in Figure 3(b). As expected,



FIGURE 3: Effect of (a) initial pH value with H_2O_2 dosage of 40 mmol/L, (b) H_2O_2 dosage, and (c) Fe^{2+} dosage with H_2O_2 dosage of 100 mmol/L on COD removal. General experimental parameters: initial pH value of 5, Fe^{2+} dosage of 0.6 mmol/L, and reaction time of 120 min.

the increase of H_2O_2 dosage from 20 to 80 mmol/L accelerated COD removal. The increased degradation efficiency can be attributed to the additional OH radicals produced from H_2O_2 decomposition. However, COD removal could not be obviously improved by excessive addition of H_2O_2 (>100 mmol/L). The results showed a negligible increase from 77.5% to 78.3% in removal efficiency when H_2O_2 dosage further increased from 100 to 120 mmol/L. It can be interpreted that the excessive H_2O_2 acts as a scavenger of •OH, but the produced HO_2 • (Equations (5)–(7)) has much lower oxidation capacities [38]. Therefore, the optimum H_2O_2 dosage was found to be 100 mmol/L for advanced treatment of the pesticide-containing wastewater by the MWEUV/Fenton:

$$H_2O_2 + \bullet OH \longrightarrow HO_2 \bullet + H_2O$$
 (5)

$$HO_2 \bullet + OH \longrightarrow H_2O + O_2$$
 (6)

$$\bullet OH + \bullet OH \longrightarrow H_2O_2 \tag{7}$$

Li et al. [39] investigated catechol oxidation in nano-Fe₃O₄ catalyzing UV-Fenton process. The best operational H_2O_2 dosage was determined to obtain high efficiency of both H_2O_2 utilization and COD removal. Lucas and Peres [40] observed the decreasing dye decolorization at excessive H_2O_2 concentration (>2 mmol/L) for Reactive Black 5 in Fenton/UV-C and ferrioxalate/ H_2O_2 /solar light processes. Zhong et al. [41] also indicated that the degradation efficiency of tetrabromobisphenol A significantly decreased in heterogeneous UV/Fenton with the H_2O_2 concentration increasing to 20 mmol/L.

3.4. Effect of Fe^{2+} Dosage. To investigate effect of Fe^{2+} dosage on COD removal, experiments were carried out at Fe^{2+} dosage ranging from 0 to 1.0 mmol/L. Figure 3(c) showed a significant increase in COD removal efficiency from 44.5% to 77.9% with the increasing Fe^{2+} concentration 0 to 0.8 mmol/L. Fe²⁺ is important for formation of photoactive ferric-hydroxo complexes that absorb UV light to produce •OH [42]. However, further increases in Fe²⁺ concentration up to 0.8 mmol/L only resulted in slight increases in degradation rate. Excessive ferrous ions may act as hydroxyl radical scavenger according to the following [43]:

$$Fe^{2+} + \bullet OH \longrightarrow Fe^{3+} + OH^{-}$$
 (8)

In addition, a deep color and high turbidity at high Fe^{2+} concentration reduced the transmission of UV light in solution, which inhibited the photolysis of H_2O_2 to produce OH radicals [31]. Therefore, overdosed Fe^{2+} was inefficient for COD removal by MWEUV/Fenton process.

3.5. Kinetics of COD Removal by MWEUV/Fenton. Experiments were conducted under the optimum operating conditions to evaluate the kinetics of MWEUV/Fenton degradation for pesticide-containing wastewater. A simple pseudo-first order kinetic model was used to fit experimental data and determine the kinetic parameters:

$$\frac{-\ln C_t}{C_0} = kt,\tag{9}$$

where C_t is COD concentration (mg/L) at time t (min), C_0 is initial COD concentration (mg/L), and k is apparent rate constant (min⁻¹). A plot of $-\ln(C_t/C_0)$ versus t generates a straight line. Apparent rate constant for the COD degradation was determined from the slope of the straight line. From Figure 4, COD degradation in MWEUV/Fenton followed pseudo-first order kinetics with rate constants (k) of 0.0125 min⁻¹ and R^2 (correlation coefficient) of 0.9171.

3.6. Evaluation of Oxidation Degree and Mineralization. At present, two parameters, as AOS and COS, are defined to evaluate the oxidation degree and oxidative process efficiency, respectively [3]:

$$AOS = 4 - 1.5 \left(\frac{COD_t}{DOC_t}\right),\tag{10}$$

$$COS = 4 - 1.5 \left(\frac{COD_t}{DOC_0}\right),\tag{11}$$

where COD_t is the chemical oxygen demand (mg/L) at time t (min), DOC_t is the dissolved organic carbon (mg/L) at time t, and DOC_0 is the initial dissolved organic carbon (mg/L). AOS takes values between +4 for CO₂, the most oxidized state of C, and -4 for CH₄, the most reduced state of C.

Figures 5 and 6 present the evolution of COD, DOC, AOS, and COS with reaction time. Both the initial AOS and COS parameters were -2.32, indicating that organic compounds are at reduced state in wastewater. With the prolonging of reaction time, the concentrations of COD and DOC in wastewater gradually decreased, showing a strong oxidation of the organics. AOS and COS parameters increased to be positive, suggesting that strong mineralization occurred and highly oxidized intermediates are generated. Analyzing



FIGURE 4: Kinetics of COD decomposition by MWEUV/Fenton. Experimental parameters: initial pH value of 5, H_2O_2 dosage of 100 mmol/L, and Fe²⁺ dosage of 0.8 mmol/L.



FIGURE 5: Evolution of COD and DOC in MWEUV/Fenton. Experimental parameters: initial pH value of 5, H_2O_2 dosage of 100 mmol/L, and Fe²⁺ dosage of 0.8 mmol/L.

the different process phases, more oxidized organic intermediates were formed before 60 min of reaction without substantial mineralization, which is corroborated by rapid COD decrease and low DOC removal. After the gradual growth, the values of AOS and COS reached the plateau. Finally, the concentrations of COD and DOC decreased to 36.9 mg/L and 27.8 mg/L, respectively, with the AOS value of 2.73 and the COS value of 2.01. The results indicate that a high oxidation degree of pesticides related to the generation of some low molecular weight carbohydrates which are resistant to mineralization by hydroxyl radicals.

The concentration profiles of Dimethoate, Triazophos, and Malathion during the MWEUV/Fenton process are shown in Figure 7. They displayed similar profiles, a rapid



FIGURE 6: Evolution of AOS and COS in MWEUV/Fenton. Experimental parameters: initial pH value of 5, H_2O_2 dosage of 100 mmol/L, and Fe²⁺ dosage of 0.8 mmol/L.



FIGURE 7: Degradation of Dimethoate, Triazophos, and Malathion in MWEUV/Fenton. Experimental parameters: initial pH value of 5, H_2O_2 dosage of 100 mmol/L, and Fe²⁺ dosage of 0.8 mmol/L.

increase in pesticides concentration at the initial stage of MWEUV/Fenton process, and Dimethoate and Malathion were completely removed within 240 min. Figure 8 exhibits the concentration changes of inorganic anions (NO_3^- , SO_4^{-2-} , and PO_4^{-3-}) in the reaction. The increases of SO_4^{-2-} and PO_4^{-3-} concentration in wastewater were mainly attributed to the breakage of chemical bonds such as S=P, P-O, and S-P. The results also suggested a deep decomposition and mineralization of pesticide molecules occurring under the attack of hydroxyl radicals.



FIGURE 8: Evolution of inorganic anions in MWEUV/Fenton. Experimental parameters: initial pH value of 5, H_2O_2 dosage of 100 mmol/L, and Fe²⁺ dosage of 0.8 mmol/L.

4. Conclusions

The feasibility and superiority of the MWEUV/Fenton for advanced treatment of pesticide-containing biotreated wastewater were evaluated in this paper. In terms of COD removal, the MWEUV/Fenton process showed high efficiency comparing with the conventional Fenton process assisted by commercial mercury lamps. The optimal parameters were found at the initial pH of 5, Fe²⁺ dosage of 0.8 mmol/L, and H₂O₂ dosage of 100 mmol/L. Three main pesticides, Dimethoate, Triazophos, and Malathion, in the wastewater were completely decomposed with high oxidation and mineralization degrees.

Conflict of Interests

The authors declare that there is no conflict of interests regarding the publication of this paper.

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References

- C. Zaror, C. Segura, H. Mansilla, M. A. Mondaca, and P. Gonzalez, "Kinetic study of Imidacloprid removal by advanced oxidation based on photo-Fenton process," *Environmental Technology*, vol. 31, no. 13, pp. 1411–1416, 2010.
- [2] K. V. Plakas and A. J. Karabelas, "Removal of pesticides from water by NF and RO membranes—a review," *Desalination*, vol. 287, pp. 255–265, 2012.

- [3] V. J. P. Vilar, F. C. Moreira, A. C. C. Ferreira et al., "Biodegradability enhancement of a pesticide-containing biotreated wastewater using a solar photo-Fenton treatment step followed by a biological oxidation process," *Water Research*, vol. 46, no. 15, pp. 4599–4613, 2012.
- [4] M. M. Micó, S. Chourdaki, J. Bacardit, and C. Sans, "Comparison between ozonation and photo-fenton processes for pesticide methomyl removal in advanced greenhouses," *Ozone: Science & Engineering*, vol. 32, no. 4, pp. 259–264, 2010.
- [5] A. Babuponnusami and K. Muthukumar, "A review on Fenton and improvements to the Fenton process for wastewater treatment," *Journal of Environmental Chemical Engineering*, vol. 2, no. 1, pp. 557–572, 2014.
- [6] M. J. Benotti, B. D. Stanford, E. C. Wert, and S. A. Snyder, "Evaluation of a photocatalytic reactor membrane pilot system for the removal of pharmaceuticals and endocrine disrupting compounds from water," *Water Research*, vol. 43, no. 6, pp. 1513– 1522, 2009.
- [7] N. Klamerth, S. Malato, A. Agüera, A. Fernández-Alba, and G. Mailhot, "Treatment of municipal wastewater treatment plant effluents with modified photo-fenton as a tertiary treatment for the degradation of micro pollutants and disinfection," *Environmental Science & Technology*, vol. 46, no. 5, pp. 2885–2892, 2012.
- [8] Q.-F. Zeng, J. Fu, Y. Zhou, Y.-T. Shi, and H.-L. Zhu, "Photooxidation degradation of Reactive brilliant red K-2BP in aqueous solution by ultraviolet radiation/sodium hypochlorite," *Clean: Soil, Air, Water*, vol. 37, no. 7, pp. 574–580, 2009.
- [9] Y. Deng and J. D. Englehardt, "Treatment of landfill leachate by the Fenton process," *Water Research*, vol. 40, no. 20, pp. 3683– 3694, 2006.
- [10] C. Orbeci, I. Untea, G. Nechifor, A. E. Segneanu, and M. E. Craciun, "Effect of a modified photo-Fenton procedure on the oxidative degradation of antibiotics in aqueous solutions," *Separation and Purification Technology*, vol. 122, pp. 290–296, 2014.
- [11] M. Umar, H. A. Aziz, and M. S. Yusoff, "Trends in the use of Fenton, electro-Fenton and photo-Fenton for the treatment of landfill leachate," *Waste Management*, vol. 30, no. 11, pp. 2113– 2121, 2010.
- [12] J. X. Chen and L. Z. Zhu, "Oxalate enhanced mechanism of hydroxyl-Fe-pillared bentonite during the degradation of Orange II by UV-Fenton process," *Journal of Hazardous Materials*, vol. 185, no. 2-3, pp. 1477–1481, 2011.
- [13] A. Karci, I. Arslan-Alaton, T. Olmez-Hanci, and M. Bekbölet, "Transformation of 2,4-dichlorophenol by H₂O₂/UV-C, Fenton and photo-Fenton processes: oxidation products and toxicity evolution," *Journal of Photochemistry and Photobiology A: Chemistry*, vol. 230, no. 1, pp. 65–73, 2012.
- [14] A. Karci, I. Arslan-Alaton, M. Bekbolet, G. Ozhan, and B. Alpertunga, "H₂O₂/UV-C and Photo-Fenton treatment of a nonylphenol polyethoxylate in synthetic freshwater: followup of degradation products, acute toxicity and genotoxicity," *Chemical Engineering Journal*, vol. 241, pp. 43–51, 2014.
- [15] J. Fu, Z. Xu, Q.-S. Li et al., "Treatment of simulated wastewater containing Reactive Red 195 by zero-valent iron/activated carbon combined with microwave discharge electrodeless lamp/sodium hypochlorite," *Journal of Environmental Sciences*, vol. 22, no. 4, pp. 512–518, 2010.
- [16] J. Fu, T. Wen, Q. Wang et al., "Degradation of Active Brilliant Red X-3B by a microwave discharge electrodeless lamp in the

presence of activated carbon," *Environmental Technology*, vol. 31, no. 7, pp. 771–779, 2010.

- [17] A. I. Al-Shamma'a, I. Pandithas, and J. Lucas, "Low-pressure microwave plasma ultraviolet lamp for water purification and ozone applications," *Journal of Physics D: Applied Physics*, vol. 34, no. 18, pp. 2775–2781, 2001.
- [18] J. H. Xu, C. L. Li, P. Liu, D. He, J. F. Wang, and Q. Zhang, "Photolysis of low concentration H₂S under UV/VUV irradiation emitted from high frequency discharge electrodeless lamps," *Chemosphere*, vol. 109, pp. 202–207, 2014.
- [19] S. Horikoshi, M. Kajitani, S. Sato, and N. Serpone, "A novel environmental risk-free microwave discharge electrodeless lamp (MDEL) in advanced oxidation processes. Degradation of the 2,4-D herbicide," *Journal of Photochemistry and Photobiology A: Chemistry*, vol. 189, no. 2-3, pp. 355–363, 2007.
- [20] X. Zhang, Y. Wang, G. Li, and J. Qu, "Oxidative decomposition of azo dye C.I. Acid Orange 7 (AO7) under microwave electrodeless lamp irradiation in the presence of H₂O₂," *Journal of Hazardous Materials*, vol. 134, no. 1–3, pp. 183–189, 2006.
- [21] C. Ferrari, I. Longo, E. Tombari, and E. Bramanti, "A novel microwave photochemical reactor for the oxidative decomposition of Acid Orange 7 azo dye by MW/UV/H₂O₂ process," *Journal of Photochemistry and Photobiology A: Chemistry*, vol. 204, no. 2-3, pp. 115–121, 2009.
- [22] J. Hong, C. Sun, S.-G. Yang, and Y.-Z. Liu, "Photocatalytic degradation of methylene blue in TiO₂ aqueous suspensions using microwave powered electrodeless discharge lamps," *Journal of Hazardous Materials*, vol. 133, no. 1–3, pp. 162–166, 2006.
- [23] S. Horikoshi, M. Kajitani, N. Horikoshi, R. Dillert, and D. W. Bahnemann, "Use of microwave discharge electrodeless lamps (MDEL). II. Photodegradation of acetaldehyde over TiO₂ pellets," *Journal of Photochemistry and Photobiology A: Chemistry*, vol. 193, no. 2-3, pp. 284–287, 2008.
- [24] S. Horikoshi, H. Hidaka, and N. Serpone, "Environmental remediation by an integrated microwave/UV-illumination technique IV. Non-thermal effects in the microwave-assisted degradation of 2,4-dichlorophenoxyacetic acid in UVirradiated TiO₂/H₂O dispersions," *Journal of Photochemistry and Photobiology A: Chemistry*, vol. 159, no. 3, pp. 289–300, 2003.
- [25] X. W. Zhang, G. T. Li, Y. Z. Wang, and J. H. Qu, "Microwave electrodeless lamp photolytic degradation of acid orange 7," *Journal of Photochemistry and Photobiology A: Chemistry*, vol. 184, no. 1-2, pp. 26–33, 2006.
- [26] APHA, Standard Methods for the Examination of Water and Wastewater, American Public Health Association, Washington, DC, USA, 21st edition, 2005.
- [27] X. W. Zhang, G. T. Li, and Y. Z. Wang, "Microwave assisted photocatalytic degradation of high concentration azo dye Reactive Brilliant Red X-3B with microwave electrodeless lamp as light source," *Dyes and Pigments*, vol. 74, no. 3, pp. 536–544, 2007.
- [28] D.-H. Han, S.-Y. Cha, and H.-Y. Yang, "Improvement of oxidative decomposition of aqueous phenol by microwave irradiation in UV/H₂O₂ process and kinetic study," *Water Research*, vol. 38, no. 11, pp. 2782–2790, 2004.
- [29] Y.-Y. Zhang, C. He, V.-K. Sharma, X.-Z. Li, S.-H. Tian, and Y. Xiong, "A new reactor coupling heterogeneous Fenton-like catalytic oxidation with membrane separation for degradation of organic pollutants," *Journal of Chemical Technology and Biotechnology*, vol. 86, no. 12, pp. 1488–1494, 2011.
- [30] J.-H. Park, I.-H. Cho, and S.-W. Chang, "Comparison of Fenton and Photo-Fenton processes for livestock wastewater

treatment," Journal of Environmental Science and Health Part B Pesticides, Food Contaminants, and Agricultural Wastes, vol. 41, no. 2, pp. 109–120, 2006.

- [31] J. Rodríguez-Chueca, M. I. Polo-López, R. Mosteo, M. P. Ormad, and P. Fernández-Ibáñez, "Disinfection of real and simulated urban wastewater effluents using a mild solar photo-Fenton," *Applied Catalysis B: Environmental*, vol. 150-151, pp. 619–629, 2014.
- [32] Q. Wang, S. Tian, J. Long, and P. Ning, "Use of Fe(II)Fe(III)-LDHs prepared by co-precipitation method in a heterogeneous-Fenton process for degradation of Methylene Blue," *Catalysis Today*, vol. 224, pp. 41–48, 2014.
- [33] M. Siddique, R. Farooq, and G. J. Price, "Synergistic effects of combining ultrasound with the Fenton process in the degradation of Reactive Blue 19," *Ultrasonics Sonochemistry*, vol. 21, no. 3, pp. 1206–1212, 2014.
- [34] M. Neamtu, A. Yediler, I. Siminiceanu, and A. Kettrup, "Oxidation of commercial reactive azo dye aqueous solutions by the photo-Fenton and Fenton-like processes," *Journal of Photochemistry and Photobiology A: Chemistry*, vol. 161, no. 1, pp. 87–93, 2003.
- [35] P. K. Malik and S. K. Saha, "Oxidation of direct dyes with hydrogen peroxide using ferrous ion as catalyst," *Separation and Purification Technology*, vol. 31, no. 3, pp. 241–250, 2003.
- [36] M. I. Badawy, F. E. Gohary, M. Y. Ghaly, and M. E. M. Ali, "Enhancement of olive mill wastewater biodegradation by homogeneous and heterogeneous photocatalytic oxidation," *Journal of Hazardous Materials*, vol. 169, no. 1–3, pp. 673–679, 2009.
- [37] P. V. Nidheesh, R. Gandhimathi, and S. T. Ramesh, "Degradation of dyes from aqueous solution by Fenton processes: a review," *Environmental Science and Pollution Research*, vol. 20, no. 4, pp. 2099–2132, 2013.
- [38] W. Sabaikai, M. Sekine, M. Tokumura, and Y. Kawase, "UV light photo-Fenton degradation of polyphenols in oolong tea manufacturing wastewater," *Journal of Environmental Science and Health Part A: Toxic/Hazardous Substances and Environmental Engineering*, vol. 49, no. 2, pp. 193–202, 2014.
- [39] W. G. Li, Y. Wang, and A. Irini, "Effect of pH and H₂O₂ dosage on catechol oxidation in nano-Fe₃O₄ catalyzing UV-Fenton and identification of reactive oxygen species," *Chemical Engineering Journal*, vol. 244, pp. 1–8, 2014.
- [40] M. S. Lucas and J. A. Peres, "Degradation of reactive black 5 by Fenton/UV-C and ferrioxalate/H₂O₂/solar light processes," *Dyes and Pigments*, vol. 74, no. 3, pp. 622–629, 2007.
- [41] Y. H. Zhong, X. L. Liang, Y. Zhong et al., "Heterogeneous UV/Fenton degradation of TBBPA catalyzed by titanomagnetite: catalyst characterization, performance and degradation products," *Water Research*, vol. 46, no. 15, pp. 4633–4644, 2012.
- [42] N. Ayten, I. Arslan-Alaton, and T. Ölmez-Hanci, "Application of Photo-Fenton-like oxidation for the degradation and detoxification of commercial naphthalene sulfonates a case study with H-acid model pollutant," *Desalination and Water Treatment*, vol. 26, no. 1–3, pp. 139–144, 2011.
- [43] D. Vujević, S. Papić, N. Koprivanac, and A. L. Božić, "Decolorization and mineralization of reactive dye by UV/Fenton process," *Separation Science and Technology*, vol. 45, no. 11, pp. 1637–1643, 2010.