



Article Effects of Exposure to Indoor Fine Particulate Matter on Atopic Dermatitis in Children

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Abstract: This study aimed to investigate the short-term effect of exposure to indoor fine particulate matter (PM_{2.5}) on atopic dermatitis (AD) symptoms in children. Sixty-four children (40 boys and 24 girls) with moderate-to-severe AD, aged under 18 years were enrolled in the study. They were followed up from February 2019 through November 2020. Exposure to indoor PM2.5 in each household of the enrolled children and their AD symptoms were measured daily. The generalized linear mixed model was utilized for statistical analysis. Subdivision analysis was performed by stratifying the patients by age, sex, season, severity, the presence of family allergic diseases, sensitization, and indoor environment conditions including temperature and relative humidity. A total of 9,321 person-days of AD symptom data were collected. The average $PM_{2.5}$ concentration was 28.7 \pm 24.3 μ g/m³, with the highest value in winter ($47.1 \pm 29.6 \,\mu\text{g/m}^3$). The overall effect of PM_{2.5} on AD symptoms was not statistically significant. However, an increase of $10 \,\mu\text{g/m}^3$ in indoor PM_{2.5} concentration increased AD symptom scores by 16.5% (95% CI: 6.5, 27.5) in spring and 12.6% (95% CI: 4.3, 21.5) in winter, 6.7% (95% CI: 2.3, 11.3) at indoor temperatures of <25.5 °C, and by 15.0% (95% CI: 3.5, 27.7) with no use of an air purifier. The harmful effect of $PM_{2.5}$ in boys, in children aged ≥ 6 years, and in children with inhalant allergen sensitization was significant, showing an increase in AD symptoms of 4.9% (95% CI: 1.4, 8.6), 12.0% (95% CI: 5.3, 19.1), and 7.0% (95% CI: 1.9, 12.3) per 10 μ g/m³ of PM_{2.5}, respectively. Furthermore, children with inhalant allergen sensitization plus severe symptoms (SCORing Atopic Dermatitis, SCORAD \geq 30.7, median value) showed more harmful effects from exposure to PM_{2.5} (15.7% (95% CI: 4.5, 28.1) increase in AD symptom scores per 10 μ g/m³ of PM_{2.5} increase). Indoor exposure to PM2.5 exacerbated AD symptoms in children in spring, winter, and at indoor temperatures of < 25.5 °C. In particular, this harmful effect was prominent in children with inhalant allergen sensitization and severe symptoms. Minimizing exposure to indoor PM2.5 is needed for the proper management of AD.

Keywords: fine particulate matters; atopic dermatitis; children; indoor; season

1. Introduction

Particulate matter (PM), one of the major air pollutants, is an environmental threat to human health. The biological effects of PM on health are determined largely by the size and the composition of the PM. Long-term exposure to PM with a diameter of less than 2.5 μ m (PM_{2.5}) has been attributed to up to 4.2 million annual premature deaths, representing 7.6% of the total global mortality [1]. PM_{2.5} deposits in the respiratory bronchioles and alveoli stimulate local and systemic inflammation as well as oxidative stress [2]. PM is associated with the increased risk of allergic diseases, such as asthma and atopic dermatitis (AD). Increased exposure to ambient PM_{2.5} significantly increased the risk of AD symptoms [3,4] and the prevalence of asthma symptoms was also associated with an increase in PM_{2.5} concentrations, even at low exposure levels under standard [5,6].



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Copyright: © 2021 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). AD is a chronic inflammatory skin disease occurring mostly in early childhood and predisposes patients to food allergies and atopic march [7]. The estimated prevalence of AD in Korean children aged 18 years or younger in 2014 was about 6% [8]. One basic principle in the management of AD is avoiding aggravating factors such as microbes, allergens, emotional stress, meteorological factors, and pollutants such as PMs [3,9,10].

Most prior studies on the association between PM exposure and allergic diseases estimated the effects of ambient PM [3–5,11,12]. Eczema prevalence was significantly associated with long-term exposure to outdoor particulate matter with an aerodynamic diameter $\leq 10 \ \mu m$ (PM₁₀) [11] and PM_{2.5} [12]. Several studies also found that short-term exposure to ambient PMs exacerbated the symptoms of AD patients [3–5,13,14]. Oh et al. (2017) [15] addressed a positive association between AD symptoms and PMs including indoor PMs. However, they measured indoor PMs at two preschools during school hours (9 a.m. to 5 p.m.) and used ambient outdoor concentrations at air quality monitoring sites during non-school hours. They did not measure residential places where children spend more time. Besides, they did not include spring and winter season when the level of PMs is the highest in Korea.

The importance of indoor air quality is, however, increasingly emphasized because people spend most of their time indoors. Korean people, for example, spend approximately 88% of their time indoors, almost 60% of which is at home on weekdays [16]. A previous study [17] demonstrated that the indoor PM components were quite different from those of outdoor PMs and the effect sizes of the indoor PM components on allergic symptoms were different from those of outdoors. However, there are very limited epidemiological studies on the short-term effects of exposure to indoor PM_{2.5} on AD symptoms, mainly because of difficulties in real-time measurements indoors.

Thus, the present study aimed to examine the short-term effect of exposure to indoor $PM_{2.5}$ on AD symptoms in children by measuring real-time $PM_{2.5}$ concentrations in the patients' houses and simultaneously monitoring daily symptom scores. Subgroup analyses stratified by season, demographic status, the presence of allergen sensitization, and indoor environments were conducted to characterize the effects of indoor $PM_{2.5}$ in children with AD.

2. Materials and Methods

2.1. Study Participants and AD Symptom Monitoring

A total of 64 children (40 boys and 24 girls) aged under 18 years with AD living in the Seoul Metropolitan Areas of Korea were enrolled in the study. They were followed up from February 2019 to November 2020. The diagnosis of AD was made according to the Hanifin and Rajka criteria [18]. AD severity was assessed using the SCORing Atopic Dermatitis (SCORAD) index [19]. Only AD patients with SCORAD scores over 15 were enrolled because patients with mild AD may not be sensitive to small changes in indoor PM_{2.5} levels. Total IgE and specific IgE against common food and inhalant allergens in peripheral blood were measured using ImmunoCAP (Thermo Fisher Scientific, Waltham, MA, USA). Sensitization was defined by specific IgE levels of over 0.69 kU/L. Common allergens included egg white, cow's milk, soybean, wheat, peanut, *Dermatophagoides pteronyssinus*, and *D. farinae*.

The Atopic Dermatitis Symptom Score (ADSS) [20], a smartphone-based symptom diary, was used to monitor daily AD symptoms in the children. The parents or children were instructed to record AD symptoms daily using the symptom diary designed to record the extent of subjective symptoms (itching and sleep disturbance) and the degree of objective signs (erythema, dryness, oozing, and edema) on a scale of 0 to 4. The presence of AD symptoms was defined when each subjective symptom score was 2 or greater accompanied by at least two objective signs. This definition was used because our patients were not symptom-free at baseline. All patients were instructed to take a daily bath or shower and apply moisturizers frequently during the study period. The intermittent use of low potency topical corticosteroid (TCS) was allowed when needed. Along with symptom scores, the participants recorded the use of TCS and the presence of fever every day. Those who were allergic to inhalants or foods avoided exposure to the offending allergens.

At the time of participant enrollment, we also surveyed the presence of family allergic diseases (asthma, AD, and allergic rhinitis), the use of an air purifier, and demographic information. Written informed consent was obtained from the parent or guardian of each participating child. The study protocols were reviewed and approved by the Institutional Review Board (IRB) of Samsung Medical Center (IRB No. 2018-10-121).

2.2. Assessment of Indoor PM_{2.5} Exposure

To assess indoor PM_{2.5} exposure, an SSP100 (SENKO, Osan, Korea) laser-based lightscattering sensor was used. The SSP100 sensor is equipped with an IoT-based device, the Breeze[®] (SENKO, Osan, Korea), and automatically measures PM_{2.5} every 10 min. The Breeze[®] was placed in the living room of each participant's house. Real-time sensor data were sent to and saved in the cloud. After retrieving the 10-min PM_{2.5} data from the cloud, we averaged the 10-min data to hourly data and compared the hourly data measured by Grimm portable aerosol spectrometer (PAS) 1.108 (Grimm Aerosol Technik GmbH & Co. KG, Ainring, Germany). We stratified the hourly data into four groups based on PM_{2.5} levels of 0–10, 10–20, 30–40, and >40 µg/m³ PM_{2.5}. The regression parameters of the groups were obtained based on linear regression models for each group. The coefficient of determination of the regression model between the PM_{2.5} in the SSP100 and Grimm PAS was 0.93 (*p* < 0.0001) overall. We then calibrated the hourly indoor PM_{2.5} levels measured by the SSP100 in the households, applying the four parameters in accordance with the range of PM_{2.5} levels in each group. The details are shown in the Supplemental Materials (Supplementary Figure S1).

The average 24 h data values were calculated using the calibrated hourly $PM_{2.5}$ and matched with the daily AD symptom data for each individual.

2.3. Covariates

Breeze[®] also has temperature, relative humidity (RH), and formaldehyde sensors. Formaldehyde, indoor temperature, and indoor RH were automatically measured every 10 min. More details on indoor formaldehyde measurements are given in our previous study [21]. The average 24 h indoor temperature, indoor RH, and formaldehyde levels were calculated and used as confounding factors. For formaldehyde, the upper and lower 10% of 10 min interval data of each day were deleted and the average 24 h data value was calculated.

Ambient environment conditions such as $PM_{2.5}$, temperature, and RH are also known to be associated with the aggravation of AD symptoms in children [22]. We obtained daily outdoor $PM_{2.5}$ concentrations from the nearest air quality monitoring system (AQMS) station based on the home address of each participant. Daily outdoor temperature and RH were also obtained from the Korean Meteorological Administration (KMA). When the outdoor and indoor $PM_{2.5}$, temperature and RH were matched for each individual and each day, the outdoor $PM_{2.5}$, temperature, and RH showed strong correlations with indoor $PM_{2.5}$, temperature, and RH. The correlation between outdoor and indoor $PM_{2.5}$, temperature, and RH was 0.44 (p < 0.0001), 0.57 (p < 0.0001), 0.52 (p < 0.0001), respectively (Supplementary Figure S2). Therefore, we selected indoor $PM_{2.5}$, temperature, and RH instead of outdoor parameters to avoid the problem of collinearity between the indoor and outdoor variables, considering that the AD patients stayed indoors more than outdoors (approximately 90%) [4].

2.4. Statistical Analysis

Considering that repeated measurements of allergic symptoms could provide longitudinal data with a binomial distribution, the generalized linear mixed model (GLMM) with binomially distributed errors was utilized to estimate the effects of $PM_{2.5}$ on AD symptoms. To account for serial correlation among the repeated measurements, random effects were included in the GLMM for each subject. Potential confounding factors such as age, sex, day of the week (DOW), indoor temperature and RH, indoor formaldehyde, SCORAD at enrollment, the use of TCS, and the presence of fever as a proxy of infection were controlled for when fitting the GLMM.

We also used penalized regression curves of a generalized additive mixed model (GAMM) to examine the linearity of the relationship between the exposure to indoor $PM_{2.5}$ and AD symptoms.

For subgroup analysis, we stratified the dataset by sex, age (<6 years and \geq 6 years), season, the presence of family allergic diseases, the presence of inhalant allergen sensitization, SCORAD, household income, the use of an air purifier, and indoor environmental conditions such as temperature and RH. Indoor temperature was stratified into two categories of <25.5 °C and \geq 25.5 °C (median value). RH was stratified into three categories of <40%, 40–60%, and \geq 60%. We then fitted GLMM models to each subgroup dataset and compared the effect sizes between them. All results are expressed as percent change with 95% confidence intervals (CIs) of symptom scores per 10 µg/m³ increment of PM_{2.5}.

All analyses were conducted using R version 4.0.2 (The Comprehensive R Archive Network: http://cran.r-project.org, accessed on 22 June 2020). GLMMs were fitted using the lme4 package (version 3.1.2). All tests were two-sided. An alpha level of less than 0.05 was considered significant.

3. Results

3.1. Characteristics of AD Patients and Household Environments

Table 1 shows the characteristics of the subjects in the present study. The average age of the study subjects was 4.5 ± 3.7 years. The average SCORAD at enrollment was 33.2 ± 13.9 . Among 64 participants, the parents or siblings of 38 (61.3%) had ever been diagnosed with allergic diseases such as asthma, allergic rhinitis, or AD. Most households had air purifiers (88.7%).

| Characteristics | Total | |
|---|--------------------|--|
| Sex (boys:girls) | 40:24 | |
| Age (yrs) | 4.5 ± 3.7 | |
| BMI (kg/m^2) | 17.1 ± 2.4 | |
| SCORAD at enrollment | 33.2 ± 13.9 | |
| Total IgE (U/L) | 799.1 ± 1392.5 | |
| Sensitization, n (%) | 50/64 (78.1%) | |
| Food allergens [‡] | 47/64 (71.2%) | |
| Inhalant allergens [§] | 23/64 (46.0%) | |
| Presence of family allergic diseases, n (%) | 38/62 (61.3%) | |
| Use of air purifier, n (%) | 55/62 (88.7%) | |
| No. of record (person-days) | 9321 | |
| Presence of AD symptoms (%) | 49.3 | |
| Presence of fever (%) | 0.9 | |
| Use of TCS (%) | 53.3 | |

Table 1. Characteristics of the subjects in this study $(n = 64)^*$.

* Data are expressed as the mean ± standard deviation; [‡] Sensitized by five common food allergens, including egg white, cow's milk, soybean, wheat, or peanut; [§] Sensitized by house dust mite (*Dermatophagoides pteronyssinus*, *D. farinae*); BMI, body mass index; SCORAD, SCORing of Atopic Dermatitis; TCS, topical corticosteroid.

A total of 9321 person-days of symptom records were obtained from 64 subjects. Among them, 6410 (68.8%) and 2911 (30.2%) person-days were recorded by boys and girls, respectively. The presence of AD symptoms was 49.3% in 9321 person-days.

The average daily indoor temperature, RH, and formaldehyde level during the study period were 25.5 °C \pm 2.6 °C, 47.2% \pm 12.5%, and 13.6 ppb \pm 20.1 ppb, respectively (Supplementary Table S1). With respect to the outdoor environment during the study period, the average temperature, RH, and PM_{2.5} were 14.0 °C \pm 9.5 °C, 67.9% \pm 15.8%, and 20.0 \pm 11.8 µg/m³, respectively.

3.2. Exposure Assessment of Indoor PM_{2.5}

The average indoor PM_{2.5} concentration in 9,321 datasets was $28.7 \pm 24.3 \ \mu g/m^3$. Table 2 shows the indoor PM_{2.5} levels by season and indoor environment. The highest indoor PM_{2.5} level in four seasons was observed in winter ($47.1 \pm 29.6 \ \mu g/m^3$), followed by spring ($28.0 \pm 24.8 \ \mu g/m^3$), fall ($22.9 \pm 19.3 \ \mu g/m^3$), and summer ($20.8 \pm 15.0 \ \mu g/m^3$). With respect to the indoor environment, indoor PM_{2.5} concentrations were higher when the indoor temperature was below 25.5 °C (p < 0.0001) and the indoor RH was below 40% (p < 0.0001). Not surprisingly, the indoor PM_{2.5} level was lower when an air purifier was used in the home (p < 0.0001).

| Classification | Subgroups | Number (Person-Days) | $\begin{array}{c} Mean\pm SD\\ (\mu g/m^3) \end{array}$ | <i>p-</i> Value * |
|-----------------------|---|------------------------------|---|------------------------------------|
| All | | 9321 | 28.7 ± 24.3 | |
| Season | Spring Summer Fall Winter | 2093 2443 2781 2004 | $\begin{array}{c} 28.0 \pm 24.8 \\ 20.8 \pm 15.0 \\ 22.9 \pm 19.3 \\ 47.1 \pm 29.6 \end{array}$ | - <0.0001 <0.0001 <0.0001 |
| Indoor RH | $<\!$ | 2748 5057 1516 | 36.5 ± 28.1 26.0 ± 23.3 25.5 ± 21.7 | <0.0001 <0.0001 |
| Indoor temperature | <25.5 °C ≥25.5 °C | 4570 4751 | 35.4 ± 27.4 22.3 ± 18.7 | <0.0001 |
| Air purifier | (-) (+) | 785 8429 | $\begin{array}{c} 39.9 \pm 28.2 \\ 27.8 \pm 23.7 \end{array}$ | <0.0001 |

Table 2. Summary of indoor $PM_{2.5}$ concentration by season and indoor environments (n = 64).

* ANOVA test for season and RH and z-test for others; RH, relative humidity.

Indoor PM_{2.5} levels were compared after stratifying the patients according to their characteristics (Table 3). The indoor PM_{2.5} levels were higher in boys, children with family allergic diseases, and children with inhalant allergen sensitization than in girls (p < 0.0001), children without a family history (p < 0.0001), and children without allergen sensitizations (p = 0.0026), respectively. In children with SCORAD ≥ 40.9 , indoor PM_{2.5} levels were lower than in children with mild AD (p = 0.0003), suggesting that the more severe the AD of their child, the better indoor environment the parents were trying to make.

Table 3. Summary of indoor $PM_{2.5}$ concentration by subgroup (n = 64).

| Classification | Subgroups | Number (Person-Days) | $\begin{array}{c} Mean \pm SD \\ (\mu g/m^3) \end{array}$ | <i>p</i> -Value * |
|--|-----------------------------------|-------------------------|--|-------------------|
| All | | 9321 | 28.7 ± 24.3 | |
| Sex | Boys Girls | 6410 2911 | $\begin{array}{c} 30.7 \pm 25.2 \\ 24.4 \pm 21.6 \end{array}$ | <0.0001 |
| Age | <6 yrs ≥6 yrs | 6924 2397 | $\begin{array}{c} 28.6 \pm 24.4 \\ 29.1 \pm 23.9 \end{array}$ | 0.4362 |
| Family history of allergic diseases | (—) (+) | 3474 5740 | $\begin{array}{c} 25.8 \pm 22.2 \\ 30.7 \pm 25.4 \end{array}$ | <0.0001 |
| Inhalant allergen sensitization | (—) (+) | 3501 3908 | $\begin{array}{c} 28.2 \pm 25.4 \\ 30.0 \pm 23.7 \end{array}$ | 0.0026 |
| SCORAD at enrollment | <30.7 30.7-40.9 ≥ 40.9 | 4624 2269 2428 | $\begin{array}{c} 28.4 \pm 24.7 \\ 29.6 \pm 24.8 \\ 26.7 \pm 21.4 \end{array}$ | 0.0525 0.0003 |
| Inhalant allergen sensitization (+) | <30.7 (median) | 2379 | 31.2 ± 24.5 | < 0.0001 |
| PLUS SCORAD | \geq 30.7 (median) | 1529 | 28.0 ± 22.3 | |

* Z-test for mean differences between subgroups except for SCORAD at enrollment which was result from ANOVA test. SCORAD, SCORing of Atopic Dermatitis.

3.3. Effects of PM_{2.5} on AD Symptoms

Figure 1 shows the relationship between AD symptoms and indoor $PM_{2.5}$ levels as a result of GAMM-fitting after controlling for confounding factors including indoor formaldehyde, temperature, RH, SCORAD at enrollment, fever, DOW, and the use of TCS. The spline curves show that the AD symptoms increased monotonically in spring and winter as the level of indoor $PM_{2.5}$ increased, whereas there was no linear relationship in summer and the relationship was negative in fall.



Figure 1. Relationship of indoor $PM_{2.5}$ with atopic dermatitis (AD) symptoms. Each figure shows the spline curves (solid lines) with 95% confidence intervals (two dashed lines). The model was controlled for the severity score at the initial visit, age, sex, presence of fever, and day of week (DOW), indoor RH, temperature, and formaldehyde; OR: odds ratio.

Figures 2 and 3 show the effects of indoor $PM_{2.5}$ exposure on AD symptoms as a result of GLMM-fitting after controlling for confounding factors including indoor formaldehyde, temperature, RH, SCORAD at enrollment, fever, DOW, and the use of TCS. An increase in $PM_{2.5}$ of 10 µg/m³ was associated with a 0.9% (95% CI: -2.1, 4.0) increase in AD symptoms as a whole. The overall effect of $PM_{2.5}$ on AD symptoms was not statistically significant. However, when the patients were stratified by season, the $PM_{2.5}$ effects on AD symptoms were statistically significant in spring and winter, showing a higher effect size in spring (Figure 2). With an increase in $PM_{2.5}$ of 10 µg/m³, the presence of AD symptoms increased by 16.5% (95% CI: 6.5, 27.5) in spring and 12.6% (95% CI: 4.3, 21.5) in winter when indoor formaldehyde, temperature, and RH were controlled for. Low indoor temperature was also a risk factor to exacerbate the adverse effects of $PM_{2.5}$ on AD symptoms in children (Figure 2). When the indoor temperature was below 25.5 °C, an increase in $PM_{2.5}$ of 10 μ g/m³ increased AD symptom scores by 6.7% (95% CI: 2.3, 11.3). AD symptom scores in children living in houses without an air purifier were significantly increased by 14.9% (95% CI: 3.5, 27.7) by an increase in indoor PM_{2.5} of 10 μ g/m³, whereas there was no increase in children living in houses with an air purifier. All numbers for the effect sizes are shown in Supplementary Tables S2 and S3.



Figure 2. Percent changes of atopic dermatitis symptoms caused by $PM_{2.5}$ exposure. % change in AD symptoms per 10 µg/m³ of $PM_{2.5}$ exposure; dot or square indicate % change in AD symptoms per 10 µg/m³ of $PM_{2.5}$ exposure and lines are 95% confidence interval; RH, relative humidity.



Figure 3. Percent changes of atopic dermatitis symptoms caused by $PM_{2.5}$ exposure. % change in AD symptoms per 10 µg/m³ of $PM_{2.5}$ exposure; dot or square indicate % change in AD symptoms per 10 µg/m³ of $PM_{2.5}$ exposure and lines are 95% confidence interval; SCORAD, SCORAD at enrollment.

The harmful effect of PM_{2.5} in boys was significant, showing an increase in AD symptom scores of 4.9% (95% CI: 1.4, 8.6) per 10 μ g/m³ of PM_{2.5}. AD symptom scores in children aged 6–18 years increased significantly by 12.0% (95% CI: 5.3, 19.1) by an increase in indoor PM_{2.5} levels of 10 μ g/m³, whereas there was no increase in young children under 6 years. AD symptom scores in children with inhalant allergen sensitization increased by 7.0% (95% CI: 1.9, 12.3) per 10 μ g/m³ of PM_{2.5} increase. Notably, the harmful effect of PM_{2.5} on AD symptoms was greater in children with inhalant allergen sensitization plus severe symptoms (SCORAD at enrollment \geq 30.7), showing 15.7% (95% CI: 4.5, 28.1) per 10 μ g/m³ of PM_{2.5} (Figure 3).

4. Discussion

Although air pollution is considered to be a significant risk factor for AD [9], there is a paucity of studies investigating the effect of indoor PM on AD symptoms. In the

present study, we longitudinally followed AD patients to assess the effects of indoor $PM_{2.5}$ exposure on AD symptoms in children based on repeated measurements of indoor $PM_{2.5}$ using a real-time sensor device. To overcome the issue of measurement data accuracy, we calibrated and validated the measured data based on standard measurement devices (details are in Supplementary Figure S1). Furthermore, the effect of indoor $PM_{2.5}$ was adjusted by indoor temperature, RH, and formaldehyde, an important indoor air pollutant, in our statistical models. Finally, we obtained results on the environmental factors affecting indoor $PM_{2.5}$ levels as well as the characteristics of the patients sensitively responding to changes in indoor $PM_{2.5}$ concentrations. To the best of our knowledge, this was the first study on the association between indoor $PM_{2.5}$ and AD symptoms in children based on real-time indoor $PM_{2.5}$ measurements in residential places. Our results can be used in real clinical practice for the management of moderate-to-severe AD in children.

Indoor PM originates partly from the outside, depending upon the frequency of ventilation, whereas there are many indoor sources of PM as well [9]. The reason why the indoor PM_{2.5} levels were higher in winter and spring than in summer and fall in the present study can be related to less frequent ventilation. In summer, 39 of 62 participants (63%) ventilated their houses by opening windows every day, compared to only 24 (39%) in winter. Consequently, the correlation coefficient of indoor and outdoor PM2.5 concentrations in summer was 0.47 (p < 0.0001), whereas they were 0.32 and 0.22 in winter and spring, respectively (Supplementary Table S4). These findings imply that the higher the number of natural ventilation events, the lower the indoor PM_{2.5} levels. According to previous studies [23,24], indoor air quality in homes in urban areas was influenced by indoor human activities such as cooking, smoking, and cleaning. For example, high peak PM episodes were observed during cooking activities in Seoul [23]. Natural ventilation may be necessary to lower the levels of PM indoors even in winter and spring. In this study, indoor PM_{2.5} levels were higher when the indoor temperature and RH were lower. This can be related to relatively lower indoor temperatures and RH in winter (23.5 ± 2.1 °C, $37.6 \pm 8.8\%$) than in summer (27.3 \pm 20.0 °C, 58.6 \pm 8.9%) (p < 0.0001). Thus, when indoor PM_{2.5} concentrations increase by the influence of the season, indoor temperature, or RH, it is highly likely that AD symptoms worsen according to the exposure concentration.

In the present study, we observed that AD symptom scores were significantly increased by an increase in indoor $PM_{2.5}$ levels in specific seasons such as spring and winter, at indoor temperatures of <25.5 °C, and in houses without air purifiers. Perhaps the reason is that indoor $PM_{2.5}$ concentrations were higher under these conditions (Table 2) and implies that it is necessary to monitor whether indoor PM2.5 concentrations increase in such environmental conditions. We also found that the AD symptom scores increased, particularly in boys compared to girls, and the PM2.5 levels in households where boys lived were higher ($30.7 \pm 25.2 \ \mu g/m^3$) than in those where girls lived ($24.4 \pm 21.6 \ \mu g/m^3$, p < 0.0001). These findings explain why boys with AD were more vulnerable to PM_{2.5} exposure than girls in the present study. The reason why the association between $PM_{2.5}$ and AD flares in girls was negative is not clear. Behavioral differences might be one of the reasons. Parents having girls with severe AD may have a tendency to manage AD symptoms more strictly when the PM2.5 level is high. However, considering that there are biological differences between boys and girls in immune development, infection, and gut microbiota [25–27], it is not clear whether our results were simply due to differences in $PM_{2.5}$ concentrations. Further investigations are needed to explain the sex-specific response to indoor PM_{2.5}.

It is necessary to identify the environmental factors affecting indoor $PM_{2.5}$ levels, but it is also important to identify which patients with AD in clinical practice are more susceptible to $PM_{2.5}$ exposure, even at low concentrations. Indeed, our previous study demonstrated that responses to outdoor PM_{10} exposure varied considerably among individuals with AD [28]. In the present study, it was notable that the harmful effect of $PM_{2.5}$ in children with inhalant allergen sensitization was significant, with increases in AD symptoms of 7.0% (95% CI: 1.9, 12.3) per 10 µg/m³ of $PM_{2.5}$. This could be because house dust mite allergens are one of the components of indoor fine dust, unlike outdoor PM. Another single risk factor, severe AD (SCORAD \geq 30.7, median value), did not show statistical significance in AD symptom aggravation. However, when combined with house dust mite sensitization, AD symptom scores significantly increased by 15.7% (95% CI: 4.5, 28.1) per 10 µg/m³ of PM_{2.5}. This finding indicates that AD severity could be a risk factor in children who are sensitized to house dust mite allergen. Therefore, these host factors including house dust mite sensitization and disease severity should be considered as risk factors affecting host susceptibility to PM_{2.5} exposure in children with AD.

We observed that children aged over 6 years showed a significant increase in percent change in AD symptoms by $PM_{2.5}$ exposure, whereas those under 6 years did not. As shown in Supplementary Table S5, there was no difference in $PM_{2.5}$ levels between the two groups (p = 0.4362). When we compared the host factors between children aged ≥ 6 years and <6 years, there seemed to be a difference in AD severity, inhalant allergen sensitization, and body mass index (BMI), but was no statistical significance. It is speculated that different behaviors according to age may explain this difference. Children over age 6 years may participate in more vigorous activities indoors and spend more time outdoors where they are exposed to a variety of environmental triggers compared to children under age 6 years. Unfortunately, we did not find the exact reason to explain the age-related difference and further studies with more patients are needed.

A systematic review and meta-analysis demonstrated that PM_{2.5} was associated with the aggravation of AD symptoms at a young age (OR = 1.05; 95% CI: 0.95–1.16; $I^2 = 46\%$) [29]. The exact mechanisms through which PM can aggravate AD symptoms are not fully understood. It has been reported that the skin barrier is damaged by direct exposure to PM. A high PM_{2.5} level might generate reactive oxygen species (ROS) and induce oxidative damage to proteins in keratinocytes and eventually, lead to the disruption of the skin barrier and further exacerbation of AD [30,31]. Jin et al. [32] showed that cellular ROS production was increased by PM treatment in mice, and antioxidant N-acetyl cysteine pretreatment prevented the induction of IL-8 and matrix metalloprotease 1. A recent study demonstrated that PM_{2.5}-induced tumor necrosis factor- α downregulated filaggrin expression in human keratinocytes through the aryl hydrocarbon receptor (AhR) pathway and subsequently, caused skin barrier dysfunction [33]. It is suggested that polycyclic hydrocarbons (PAHs), a component of airborne fine particles, contributed to the activation of the AhR signaling pathway and the transactivation of neurotrophic factor artemin [34,35]. The intercellular penetration of PM was seen in barrier-disrupted skin as well [32].

One of the limitations of this study was that we did not include outdoor $PM_{2.5}$ levels in the exposure assessments. Considering that the AD patients spend over 90% of their time indoors [4] and that the AQMS-monitored $PM_{2.5}$ could have led to misclassification bias of the participants whose residences were far from the monitoring station [36], we included only indoor $PM_{2.5}$ measured in the households, with no consideration of outdoor exposure. With more spatially resolved outdoor monitoring data, $PM_{2.5}$ exposure assessment could be improved. Another limitation is that we did not consider the effect of environmental tobacco smoke (ETS) exposure as a covariate. However, we assumed that our patients were not exposed to ETS because the parents classified as current smokers reported that they did not smoke in the home. We had many missing data in daily symptom reports during the study period; the average count of daily reports was 153 ± 89 days, although children or parents were guided to report every day for a continual year. This might cause the bias of PM effect on AD symptoms. Finally, caution is needed in generalizing our results because the number of patients was small and our study was limited to a specific race, age, region, and time period.

5. Conclusions

In conclusion, indoor $PM_{2.5}$ exposure exacerbated AD symptoms in children in spring, winter, and at indoor temperatures of <25.5 °C. For the proper management of AD, indoor $PM_{2.5}$ must be considered as one of the important environmental triggers, and minimizing

the exposure to indoor PM2.5 should be individualized because host characteristics and surrounding environmental factors differ from patient to patient. Ventilation, proper use of air purifiers, and maintenance of proper indoor temperature and RH can be helpful to minimize the indoor $PM_{2.5}$ effect on AD symptoms. In particular, this harmful effect was prominent in children with inhalant allergen sensitization and severe symptoms. The control of indoor $PM_{2.5}$ should be considered in the management of AD symptoms particularly in children with inhalant allergen sensitization or with severe symptoms.

Supplementary Materials: The following are available online at https://www.mdpi.com/article/10 .3390/ijerph182111509/s1, Table S1: Summary of indoor and outdoor environments, Table S2: Percent changes of atopic dermatitis symptoms caused by PM_{2.5} exposure, Table S3: Percent changes of atopic dermatitis symptoms caused by PM_{2.5} exposure, Table S4: Indoor and outdoor PM_{2.5} levels by season during study period, Table S5: Characteristics of AD patients and indoor environments by age group, Figure S1: Comparison of the PM_{2.5} concentrations measured by SSP100 with those by Grimm PAS, Figure S2: Correlations between atopic dermatitis (AD) symptoms and environmental variables.

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References

- Cohen, A.J.; Brauer, M.; Burnett, R.; Anderson, H.R.; Frostad, J.; Estep, K.; Balakrishnan, K.; Brunekreef, B.; Dandona, L.; Dandona, R.; et al. Estimates and 25-year trends of the global burden of disease attributable to ambient air pollution: An analysis of data from the Global Burden of Diseases Study 2015. *Lancet* 2017, 389, 1907–1918. [CrossRef]
- Brook, R.D.; Rajagopalan, S.; Pope, C.A., 3rd; Brook, J.R.; Bhatnagar, A.; Diez-Roux, A.V.; Holguin, F.; Hong, Y.; Luepker, R.V.; Mittleman, M.A.; et al. Particulate matter air pollution and cardiovascular disease: An update to the scientific statement from the American Heart Association. *Circulation* 2010, 121, 2331–2378. [CrossRef]
- 3. Kim, Y.M.; Kim, J.; Han, Y.; Jeon, B.H.; Cheong, H.K.; Ahn, K. Short-term effects of weather and air pollution on atopic dermatitis symptoms in children: A panel study in Korea. *PLoS ONE* 2017, *12*, e0175229. [CrossRef] [PubMed]
- Kim, Y.M.; Kim, J.; Han, Y.; Lee, B.J.; Choi, D.C.; Cheong, H.K.; Jeon, B.H.; Oh, I.; Bae, G.N.; Lee, J.Y.; et al. Comparison of diverse estimation methods for personal exposure to air pollutants and associations with allergic symptoms: The Allergy & Gene-Environment Link (ANGEL) study. *Sci. Total Environ.* 2017, 579, 1127–1136. [PubMed]
- 5. Mirabelli, M.C.; Vaidyanathan, A.; Flanders, W.D.; Qin, X.; Garbe, P. Outdoor PM2.5, Ambient Air Temperature, and Asthma Symptoms in the Past 14 Days among Adults with Active Asthma. *Environ. Health Perspect.* **2016**, *124*, 1882–1890. [CrossRef]
- Chen, F.; Lin, Z.; Chen, R.; Norback, D.; Liu, C.; Kan, H.; Deng, Q.; Huang, C.; Hu, Y.; Zou, Z.; et al. The effects of PM2.5 on asthmatic and allergic diseases or symptoms in preschool children of six Chinese cities, based on China, Children, Homes and Health (CCHH) project. *Environ. Pollut.* 2018, 232, 329–337. [CrossRef] [PubMed]
- 7. Tham, E.H.; Leung, D.Y. Mechanisms by Which Atopic Dermatitis Predisposes to Food Allergy and the Atopic March. *Allergy Asthma Immunol. Res.* **2019**, *11*, 4–15. [CrossRef]
- 8. Lee, J.Y.; Yang, H.K.; Kim, M.; Kim, J.; Ahn, K. Is the prevalence of atopic dermatitis in Korean children decreasing?: National Database 2009-2014. *Asian Pac. J. Allergy Immunol.* **2017**, *35*, 144–149. [PubMed]
- 9. Ahn, K. The role of air pollutants in atopic dermatitis. J. Allergy Clin. Immunol. 2014, 134, 993–999. [CrossRef]
- 10. Kim, J.; Kim, B.E.; Ahn, K.; Leung, D.Y.M. Interactions Between Atopic Dermatitis and Staphylococcus aureus Infection: Clinical Implications. *Allergy Asthma Immunol. Res.* **2019**, *11*, 593–603. [CrossRef] [PubMed]
- 11. Penard-Morand, C.; Raherison, C.; Charpin, D.; Kopferschmitt, C.; Lavaud, F.; Caillaud, D.; Annesi-Maesano, I. Long-term exposure to close-proximity air pollution and asthma and allergies in urban children. *Eur. Respir. J.* 2010, *36*, 33–40. [CrossRef]

- 12. Kathuria, P.; Silverberg, J.I. Association of pollution and climate with atopic eczema in US children. *Pediatr. Allergy Immunol.* **2016**, 27, 478–485. [CrossRef] [PubMed]
- Baek, J.O.; Cho, J.; Roh, J.Y. Associations between ambient air pollution and medical care visits for atopic dermatitis. *Environ. Res.* 2021, 195, 110153. [CrossRef] [PubMed]
- 14. Song, S.; Lee, K.; Lee, Y.M.; Lee, J.H.; Lee, S.I.; Yu, S.D.; Paek, D. Acute health effects of urban fine and ultrafine particles on children with atopic dermatitis. *Environ. Res.* **2011**, *111*, 394–399. [CrossRef]
- 15. Oh, I.; Lee, J.; Ahn, K.; Kim, J.; Kim, Y.M.; Sun, S.C.; Kim, Y. Association between particulate matter concentration and symptoms of atopic dermatitis in children living in an industrial urban area of South Korea. *Environ. Res.* **2018**, *160*, 462–468. [CrossRef]
- 16. Yang, W.; Lee, K.; Park, G.; Yoon, C.; Son, B.S.; Jeon, J.M. Microenvironmental Time Activity Patterns of Weekday and Weekend on Korean. J. Korean Soc. Indoor Environ. 2009, 6, 267–274.
- Habre, R.; Moshier, E.; Castro, W.; Nath, A.; Grunin, A.; Rohr, A.; Godbold, J.; Schachter, N.; Kattan, M.; Coull, B.; et al. The effects of PM2.5 and its components from indoor and outdoor sources on cough and wheeze symptoms in asthmatic children. *J. Expo. Sci. Environ. Epidemiol.* 2014, 24, 380–387. [CrossRef]
- 18. Hanifin, J.; Rajka, G. Diagnostic features of atopic dermatitis. Acta Derm. Venereol. 1980, 92, 44-47.
- Stalder, J.F.; Taïeb, A.; Atherton, D.J.; Bieber, P.; Bonifazi, E.; Broberg, A.; Calza, A.; Coleman, R.; De Prost, Y.; Gelmetti, C.; et al. Severity scoring of atopic dermatitis: The SCORAD index: Consensus Report of the European Task Force on Atopic Dermatitis. Dermatology 1993, 186, 23–31.
- 20. Lee, J.Y.; Kim, M.; Yang, H.K.; Kim, H.M.; Cho, J.; Kim, Y.M.; Lim, I.S.; Cheong, H.K.; Kim, H.S.; Sohn, I.; et al. Reliability and validity of the Atopic Dermatitis Symptom Score (ADSS). *Pediatr. Allergy Immunol.* **2018**, *29*, 290–295. [CrossRef]
- 21. Kim, Y.M.; Kim, J.; Ha, S.C.; Ahn, K. Harmful Effect of Indoor Formaldehyde on Atopic Dermatitis in Children: A Longitudinal Study. *Allergy Asthma Immunol. Res.* 2021, 13, 468–478. [CrossRef] [PubMed]
- 22. Langan, S.M.; Bourke, J.F.; Silcocks, P.; Williams, H.C. An exploratory prospective observational study of environmental factors exacerbating atopic eczema in children. *Br. J. Dermatol.* **2006**, *154*, 979–980. [CrossRef]
- 23. Kim, J.B.; Lee, G.; Ryu, S.H.; Lee, J.Y.; Woo, S.-H.; Lee, S.-B.; Kim, K.H.; Yun, S.-T.; Bae, G.-N. Indoor air pollution of a high-rise apartment caused by combustion sources in winter. *Par. Aerosol. Res.* **2014**, *10*, 119–130. [CrossRef]
- Song, P.; Wang, L.; Hui, Y.; Li, R. PM2.5 Concentrations indoors and outdoors in heavy air pollution days in Winter. *Procedia Eng.* 2015, 121, 1902–1906. [CrossRef]
- Uekert, S.J.; Akan, G.; Evans, M.D.; Li, Z.; Roberg, K.; Tisler, C.; Dasilva, D.; Anderson, E.; Gangnon, R.; Allen, D.B.; et al. Sex-related differences in immune development and the expression of atopy in early childhood. *J. Allergy Clin. Immunol.* 2006, 118, 1375–1381. [CrossRef]
- Stone, M.R.; Faulkner, G.E. Outdoor play in children: Associations with objectively-measured physical activity, sedentary behavior and weight status. *Prev. Med.* 2014, 65, 122–127. [CrossRef]
- 27. Fish, E.N. The X-files in immunity: Sex-based differences predispose immune responses. *Nat. Rev. Immunol.* 2008, *8*, 737–744. [CrossRef] [PubMed]
- 28. Noh, S.R.; Kim, J.S.; Kim, E.H.; Jeon, B.H.; Kim, J.H.; Kim, Y.M.; Kim, J.; Han, Y.; Ahn, K.; Cheong, H.K. Spectrum of susceptibility to air quality and weather in individual children with atopic dermatitis. *Pediatr. Allergy Immunol.* **2019**, *30*, 179–187. [CrossRef]
- 29. Ngoc, L.T.N.; Park, D.; Lee, Y.; Lee, Y.C. Systematic Review and Meta-Analysis of Human Skin Diseases Due to Particulate Matter. *Int. J. Environ. Res. Public Health* **2017**, *14*, 1458. [CrossRef]
- Niwa, Y.; Sumi, H.; Kawahira, K.; Terashima, T.; Nakamura, T.; Akamatsu, H. Protein oxidative damage in the stratum corneum: Evidence for a link between environmental oxidants and the changing prevalence and nature of atopic dermatitis in Japan. *Br. J. Dermatol.* 2003, 149, 248–254. [CrossRef]
- Dijkhoff, I.M.; Drasler, B.; Karakocak, B.B.; Petri-Fink, A.; Valacchi, G.; Eeman, M.; Rothen-Rutishauser, B. Impact of airborne particulate matter on skin: A systematic review from epidemiology to in vitro studies. *Part. Fibre Toxicol.* 2020, 17, 35. [CrossRef] [PubMed]
- 32. Jin, S.P.; Li, Z.; Choi, E.K.; Lee, S.; Kim, Y.K.; Seo, E.Y.; Chung, J.H.; Cho, S. Urban particulate matter in air pollution penetrates into the barrier-disrupted skin and produces ROS-dependent cutaneous inflammatory response in vivo. *J. Dermatol. Sci.* 2018, *91*, 175–183. [CrossRef] [PubMed]
- 33. Kim, B.E.; Kim, J.; Goleva, E.; Berdyshev, E.; Lee, J.; Vang, K.A.; Lee, U.H.; Han, S.; Leung, S.; Hall, C.F.; et al. Particulate matter causes skin barrier dysfunction. *JCI Insight* 2021, *6*, e145185. [CrossRef] [PubMed]
- 34. Hidaka, T.; Ogawa, E.; Kobayashi, E.H.; Suzuki, T.; Funayama, R.; Nagashima, T.; Fujimura, T.; Aiba, S.; Nakayama, K.; Okuyama, R.; et al. The aryl hydrocarbon receptor AhR links atopic dermatitis and air pollution via induction of the neurotrophic factor artemin. *Nat. Immunol.* **2017**, *18*, 64–73. [CrossRef]
- Kabashima, K.; Otsuka, A.; Nomura, T. Linking air pollution to atopic dermatitis. *Nat. Immunol.* 2016, 18, 5–6. [CrossRef] [PubMed]
- 36. Kim, Y.M.; Oh, I.; Kim, J.; Kang, Y.H.; Ahn, K. Harmful Effects of Ambient Nitrogen Dioxide on Atopic Dermatitis: Comparison of Exposure Assessment Based on Monitored Concentrations and Modeled Estimates. *Atmosphere* **2020**, *11*, 921. [CrossRef]