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Sonochemical degradation of surfactants with different charge types: Effect of the critical micelle concentration in the interfacial region of the cavity



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<i>Keywords:</i> Surface potential Counter-ion Coexisting effect Ion strength	Ionic surfactants tend to accumulate in the interfacial region of ultrasonic cavitation bubbles (cavities) because of their surface active properties and because they are difficult to evaporate in cavitation bubbles owing to their extremely low volatilities. Hence, sonolysis of ionic surfactants is expected to occur in the interfacial region of the cavity. In this study, we performed sonochemical degradation of surfactants with different charge types: anionic, cavitation behavior in the interfacial region of cavitation bubbles. For all of the surfactants investigated, the degradation rate increased with increasing initial bulk concentration and reached a maximum value. The initial bulk concentration to obtain the maximum degradation rate had a positive correlation with the critical micelle concentration (cmc). The initial bulk concentrations of the anionic surfactants were hower than their cmcs. These results can be explained by the negatively

charged cavity surface and the effect of the coexisting counterions of the surfactants.

1. Introduction

Ultrasonic irradiation of a liquid induces formation, growth, and collapse of gas microbubbles, cavities. During cavity collapse, local reaction zones with extremely high temperature (several thousand of degrees) and pressure (several hundred atmosphere pressure) are produced [1,2]. In aqueous solution, reactive OH radicals and hydrogen atoms are formed by sonolysis of water [3,4]. Therefore, it is considered that sonochemical degradation of organic compounds proceeds by the direct pyrolysis reaction inside cavities and by OH radicals in the interfacial region of cavities. Previous reports have shown that organic compounds with high hydrophobicity and high volatility are readily decomposed by sonochemical reactions [5] because hydrophobic compounds accumulate in the interfacial region of cavities during the bubble growth process [6]. Volatile compounds vaporize in cavity and are then thermally decomposed [7]. The interfacial region of the cavity has attracted attention since the early stages of sonochemistry [8]. Therefore, many studies have investigated this target by sonolysis of surfactants [9-24].

We previously investigated the physicochemical properties of the interfacial region of cavities by sonochemical degradation of anionic surfactants [25,26]. We found that the initial degradation rates of linear alkylbenzene sulfonates (LASs) were strongly dependent on their concentrations. However, the degradation rates of LASs with different carbon chain lengths had different maximum values at different initial bulk concentrations. In addition, there was a clear relationship between the maximum degradation rate and the critical micelle concentration (cmc) for each LAS. Furthermore, the concentrations of the LASs when the degradation rates reached their maximum values were smaller than their cmcs. We suggested that the micelles formed in the interfacial region of the cavitation bubbles could reduce the cavitation efficiency, although the detailed mechanism was unclear.

In the present study, we performed sonochemical degradation of surfactants with different charge types, such as the cationic surfactant benzylalkyldimethyl ammonium chloride (BAC) with different alkyl chain lengths. Based on the degradation rates, the effect of micelle formation on the cavitation efficiency is discussed. Furthermore, we investigated the surface potential of the cavity on the basis of the degradation behavior of surfactants with different charge types. The microbubbles generated by the aerator in the water reservoir are considered to be negatively charged because of excess adsorption of OH^- compared with H^+ at the air–water interface owing to the

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difference in the hydration energies [27]. However, the details of the surface potential of ultrasonic cavitation bubbles remain unclear. Based on the degradation rates of various surfactants with different charge types, the effect of the cavitation surface charge on the surfactant counterion, and subsequent micelle formation of the surfactant in the interfacial region of the cavity was suggested.

2. Experimental section

2.1. Materials

Benzyldimethyldodecyl ammonium chloride (BAC C₁₂: >98.0%), benzyldimethyltetradecyl ammonium chloride (BAC C₁₄: >98.0%), benzyldimethylhexadecyl ammonium chloride (BAC C16: >97.0%), benzyldimethyloctadecyl ammonium chloride (BAC C16: >97.0%), benzyldimethyloctadecyl ammonium chloride (BAC C16: >98.0%), octylphenolethoxylate (PE C₈ n = approx.10), dodecyldimethyl (3-sulfopropyl)ammonium hydroxide inner salt (SAA C₁₂: >98.0%), and octadecyldimethyl (3-sulfopropyl) ammonium hydroxide inner salt (SAA C₁₈: >97.0%) were supplied by Tokyo Chemical Industry Co., Ltd. (Tokyo, Japan). Sodium p-octylbenzene sulfonate (LAS C₈: >99.0%) and sodium p-decylbenzene sulfonate (LAS C₁₀: >99.0%) were supplied by Nacalai Tesque Inc. (Kyoto, Japan). Dodecyldimethyl glycine (Gly C₁₂: >95.0%) was supplied by Wako Pure Chemical Industries, Ltd. (Osaka, Japan). All of the surfactant solutions were prepared with Milli-Q water.

2.2. Methods

Ultrasonic irradiation was performed with a 65 mm diameter oscillator (Kaijo 4611type, MFG. No. 78E5) and an ultrasonic generator (Kaijo TA-4021type, Lot. No. S644, frequency 200 kHz), which was operated at 200 W. The diameter of the reaction vessel was 55 mm, the base of the vessel was 1 mm thick, and the distance from the oscillator was fixed at 4 mm. The details of the irradiation setup and the characteristics of the reaction vessel are described elsewhere [28]. An Arsaturated aqueous solution of each surfactant (60 mL) was sonicated in a water bath maintained at 20°C by a compact handy cooler with a temperature controller (AS ONE 102TCN).

2.3. Analysis

The concentration of each surfactant in the aqueous solution was monitored by high-performance liquid chromatography (HPLC, Shimadzu LC-20AD, SPD-20AV) with ultraviolet (UV) detection using a C18 column with a mobile phase of acetonitrile/50 mM NaClO₄ aqueous solution (80:20 v/v) flowing at 1.0 mL min⁻¹. The initial degradation rate was determined by plotting the concentration of the BAC during sonication as a function of the sonication time and then fitting the time profile to a logarithmic equation. The obtained logarithmic equation was differentiated with respect to the sonication time, and time t = 0 was substituted into the differential equation to obtain the initial degradation rate [29]. The rate was determined as the average value from several experimental runs. The critical micelle concentration (cmc) values were determined by the electrical conductivity method. The electrical conductivity measurements of the surfactant solutions were performed at room temperature (25°C). Surfactant solutions containing 3 mmol L^{-1} tert-butanol as a OH radical scavenger were also sonicated to estimate the contribution of the direct pyrolysis reaction [30]. The yields of methane, ethane, ethylene, and acetylene gases were determined by gas chromatography (Shimadzu GC-2025) with a flame ionization detector and an Alumina BOND/Na2SO4 column (RESTEK). Gas samples were taken from the headspace of the vessel for analysis after 60 min sonication.

3. Results and discussion

3.1. Effect of carbon chain length on degradation rate

The surfactant concentration change with time during sonication with an initial bulk concentration of 6 mmol L^{-1} is shown in Fig. 1. We confirmed that these results have good reproducibility. The degradation rate increased with decreasing carbon chain length in the order BAC C16 < BAC C₁₄ < BAC C₁₂. These results agree well with the order of the degradation rates of the LASs in our previous report [26]. A surfactant with shorter carbon-chain-length has lower activation energy for adsorption [31]. Consequently, the degradation rate of surfactant increased because the adsorption step on gas-liquid interface became faster. The degradation rate of each BAC as a function of the initial bulk concentration is shown in Fig. 2. For all of the BACs, the degradation rate gradually increased to a maximum value with increasing initial bulk concentration, and then decreased at different concentrations. The maximum degradation rates of the BACs depended on their carbon chain lengths. These trends are also similar to the degradation results of the LASs. The degradation rates of BAC C_{12} , C_{14} , and C_{16} reached the maximum values when their initial bulk concentrations were about 12, 10, and 6 mmol L^{-1} , respectively.

3.2. Effect of micelle formation on maximum degradation rate

Considering the cause of the local maximum value of the degradation rate, we expected involvement of micelle formation of the surfactant. It has previously been reported that micelle formation can reduce the cavitation efficiency [20,21]. In addition, based on the clear relationship between the maximum degradation rate and cmc for LASs, we suggested that micelles formed in the interfacial region of the cavity by accumulation of the anionic surfactant even when the concentration was less than the cmc and reduced the cavitation efficiency [25]. We then investigated the relation between the maximum degradation rate and the cmc for surfactants with different charge types. The data of electrical conductivity to estimate cmc values for anionic and cationic surfactants are shown in Fig. S1. The other cmc values for zwitterionic and nonionic surfactants were taken from a data book [32] because the electrical conductivity was too small to measure. The relation between the initial



Fig. 1. Sonochemical degradation of BAC C_{12} (\diamond), BAC C_{14} (\Box), and BAC C_{16} (\diamond) in aqueous solution under an Ar atmosphere. The initial bulk concentrations of the BACs were 6 mM.



Fig. 2. Relationship between the initial BAC C_n concentration and the BAC degradation rate for BAC C_{12} (\circ), BAC C_{14} (\square), BAC C_{16} (\diamond), and BAC C_{18} (\times). The error bar indicates standard deviation based on 3 times experiments.

bulk concentration for the maximum degradation rate and the cmc for each surfactant is shown in Fig. 3. For comparison, the data for LAS C_9 and C_{12} are also plotted [25]. If micelle formation of the surfactant reduced the cavitation efficiency, the degradation rate should reach the maximum value when the initial bulk concentration is the cmc (dashed line in Fig. 3). For the zwitterionic and nonionic surfactants, the degradation rates had maximum values at around the cmcs. However, the degradation rates of the anionic surfactants reached their maximum values at lower bulk concentrations than their cmcs. In contrast, the degradation rates of the cationic surfactants increased even above their cmcs. Destallats et al.[20] and Vinodgopal et al.[11] also indicated that the degradation rates decreased above cmc for non-ion surfactants, alkylphenol ethoxylate and Teric GN9, respectively. Yang et al. [22]



Int. conc. for maximum degradation rate [mmol/L]

Fig. 3. Relationship between the initial surfactant concentration when the degradation rate reaches the maximum value and the cmc. The gray dashed line represents the 1:1 relation. Charge type of the surfactant: \Box anionic, \circ cationic, \diamond zwitterionic, and \triangle nonionic.

reported that under pulse sonication the degradation rate of anion surfactants, dodecyl benzenesulfonate and octyl benzenesulfonate, decreased at the surfactant concentration over 20% of cmc. Sostaric et al. [18] showed that the radical yields from the sonication of anion surfactant peaked out at the surfactant concentration much below cmc while that for non-ion surfactant reached maximum values at around cmc. These results were well accorded with our results, indicating the effect of the electric charge of the surfactant on the relation between the maximum degradation rate and the cmc. However, adsorption or repulsion of the surfactant owing to its electrical charge cannot explain these results.

3.3. Contribution of pyrolysis

Based on the assumption that the cavity has a negatively charged surface [27], the cationic surfactant could electrostatically adsorb on the cavity surface. Surfactant adsorption on the cavity surface can be confirmed by the contribution ratios of pyrolysis in sonolysis for the BACs (Fig. S2). The pyrolysis contribution was estimated from the results of sonolysis with 3 mmol L^{-1} tert-butanol as a OH radical scavenger. The scavenging ability of tert-butanol for the OH radical has been previously reported [6,7,33,34]. The adsorbed surfactant in the interfacial region of the cavity thermally decomposed. The contribution ratios of pyrolysis in sonolysis for the BACs were constant at almost 100% irrespective of the initial bulk concentration and carbon chain length. Corresponding results of surfactant adsorption on the cavity surface have been reported in previous studies [9,12,17,18]. Nevertheless, the contribution ratio of pyrolysis was constant even above the cmc (Fig. S2). Based on these results, the further increases of the degradation rates of the BACs above their cmcs can be explained by adsorption of the cationic surfactant even in the micellar state. This is unreasonable from the viewpoint of the surface activity of the surfactant.

3.4. Ionic strength in the interfacial region of cavity

On the basis of the above results, we focused on the effect of the counterion of the surfactant. From the viewpoint of ion size, the surfactant counterions, such as the sodium and chloride ions, diffuse faster from the bulk solution to the cavity surface than the surfactant ions. For the cationic surfactants, the Cl⁻ counterion is electrostatically repulsed by the negatively charged cavity surface, as shown in Fig. 4(a). Consequently, the ion strength in the interfacial region increases and the surfactant activity decreases on the basis of Debye-Hückel theory. In general, the attracting force between ions involved in the dissociation equilibrium is shielded efficiently by dissolved solutes. Then, the actual concentration (activity) of ion decreased and the equilibrium was shifted [35]. In the interfacial region of cavity of the present study, high concentration of Cl⁻ counterion shifted the equilibrium of surfactant to undissociated state. Therefore, the decrease of the surfactant activity makes micelle formation difficult, and the cmc of the cationic surfactant increases. Conversely, the counterion of the anionic surfactants (Na⁺) is electrostatically attracted to the negatively charged cavity surface, as shown in Fig. 4(b). The increase of the surfactant activity owing to the decreased ion strength in the interfacial region results in a decrease in the cmc. These hypotheses were different from the results of previous study about the effect counterions on micelle formation in the surfactant solution with adding salt [36]. To verify the above hypotheses, we measured the cmcs of surfact ant solutions containing 10 mmol $\rm L^{-1}~\rm HCl$ (Cl⁻) under the assumption of electrostatic repulsion by the negatively charged cavity surface in the interfacial region. The electrical conductivity of each surfactant aqueous solution with addition of HCl is shown in Fig. S3. We chose the HCl concentration considering the effect of the solution pH on dissociation of the surfactants. The relation between the initial bulk concentration for the maximum degradation rate and the cmc in the surfactant solution with HCl is shown in Fig. 5. The cmc of the anionic surfactant decreased, whereas that of the cationic surfactant

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Fig. 4. Schematic diagram of the difference in the behavior between the (a) cationic surfactant and (b) anionic surfactant in the interfacial region of the cavitation bubble. The negatively charged cavity surface had the opposite effect on the counterions of the cationic and anionic surfactants (chloride and sodium ions, respectively). The adsorbed surfactant on the cavity surface and the other counterions in the interfacial region are not shown for clarity.



Fig. 5. Relationship between the initial surfactant concentration when the degradation rate reaches the maximum value and the cmc for surfactant solutions containing 10 mM HCl. The gray dashed line represents the 1:1 relation. Charge type of the surfactant: \Box anionic, \circ cationic, \diamond zwitterionic, and \triangle nonionic.

increased. These results with adding acid were different from the widely known relations between cmc and electrolyte concentration in the presence and absence of added salts. For the nonionic and zwitterionic surfactants, the cmcs remained almost the same with addition of Cl⁻. Accordingly, these plots fit on the 1:1 line, as shown in Fig. 5. While some of the cationic surfactants had values below the 1:1 line, these plots fitted the 1:1 line when the cmc was estimated for the surfactant solution with higher Cl⁻ concentration (20 mmol L⁻¹ HCl).

3.5. Critical micelle concentration in the interfacial region of cavity

From the above results, it was considered that it was clarified that formation of micelles in the interfacial region of the cavity surface inhibited surfactant degradation. This inhibition mechanism can be

explained as follows. The temperature of the hotspot and/or the number of cavities might decrease because the micelles decrease the acoustic pressure of ultrasound against the cavities by the shelter effect. Alternatively, the micelles might prevent surfactant molecules from diffusing from the bulk solution to the cavities. To investigate the mechanism, we measured the yields of C₂ hydrocarbons (ethane, ethylene, and acetylene) as the degradation products of surfactant sonolysis. In addition, based on the hydrocarbon yields, we estimated the average temperatures of the cavities by the methyl radical recombination method [1,37–39]. The total yields of C_2 hydrocarbons (ethane, ethylene, and acetylene) and average cavity temperatures estimated from sonolysis of BAC C₁₂ for different initial bulk concentrations are shown in Fig. 6. Based on comparison with the degradation results in Fig. 2, the hydrocarbon yields increased even when the initial bulk concentration was above that for the maximum degradation rate (around 12 mmol L^{-1}). Furthermore, the average cavity temperature was almost the same (about 2500 K) regardless of the initial bulk concentration. From these results, the inhibition of degradation rate by micelle formation was difficult to be explained by quenching of cavitation bubbles. Inherently,



Fig. 6. Total yields of C_2 hydrocarbons (ethane, ethylene, and acetylene) and average cavitation temperatures estimated based on the C_2 yields for sonolysis of BAC C_{12} with different initial bulk concentrations.

the multibubble situation in acoustic cavitation solution contains complexity. For example, it is known that surfactants adsorbed to the surface of bubbles inhibit bubble coalescence processes, hence affect bubble numbers. We should consider the effect of micelle formation on the conditions of caviation bubbles comprehensively. Further investigation of the number of cavities is required to clarify the mechanism.

4. Conclusion

The degradation rates of surfactants with different charge types showed maxima at different initial bulk concentrations. For the anionic and cationic surfactants, the initial bulk concentrations for the maximum degradation rates were below and above their cmcs, respectively. These results can be explained by the negative charge of the cavitation bubble surface and the effect of the coexisting counterions of the surfactants. The zwitterionic and nonionic surfactants showed little difference between the cmcs in the cavity interfacial region and bulk solution. In contrast, the anionic and cationic surfactants showed different cmcs in these regions because the counterions of the surfactants were electrostatically affected by the negatively charged surface. On the basis of the findings of this study, sonolysis of surfactants has the potential for higher efficiency and optimization by controlling the cmc.

CRediT authorship contribution statement

Ben Nanzai: Conceptualization, Methodology, Writing - original draft, Visualization. **Seiya Suzuki:** Formal analysis, Investigation, Data curation. **Kenji Okitsu:** Writing - review & editing.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi.org/10.1016/j.ultsonch.2020.105354.

References

- [1] K. Okitsu, T. Suzuki, N. Takenaka, H. Bandow, R. Nishimura, Y. Maeda, Acoustic multibubble cavitation in water: A new aspect of the effect of a rare gas atmosphere on bubble temperature and its relevance to sonochemistry, J. Phys. Chem. B 110 (2006) 20081–20084.
- [2] W.B. McNamara III, Y.T. Didenko, K.S. Suslick, Pressure during sonoluminescence, J. Phys. Chem. B 107 (2003) 7303–7306.
- [3] A.E. Alegria, Y. Lion, T. Kondo, P. Riesz, Sonolysis of aqueous surfactant solutions: probing the interfacial region of cavitation bubbles by spin trapping, J. Phys. Chem. 93 (12) (1989) 4908–4913.
- [4] C.M. Krishna, T. Kondo, P. Riesz, Sonochemistry of alcohol-water mixtures: spintrapping evidence for thermal decomposition and isotope-exchange reactions, J. Phys. Chem. 93 (1989) 5166–5172.
- [5] B. Nanzai, K. Okitsu, N. Takenaka, H. Bandow, Y. Maeda, Sonochemical degradation of various monocyclic aromatic compounds: Relation between hydrophobicities of organic compounds and the decomposition rates, Ultrason. Sonochem. 15 (4) (2008) 478–483.
- [6] K. Okitsu, B. Nanzai, K. Kawasaki, N. Takenaka, H. Bandow, Sonochemical decomposition of organic acids in aqueous solution: Understanding of molecular behavior during cavitation by the analysis of a heterogeneous reaction kinetics model, Ultrason. Sonochem. 16 (1) (2009) 155–162.

- [7] K. Hirai, Y. Nagata, Y. Maeda, Decomposition of chlorofluorocarbons and hydrofluorocarbons in water by ultrasonic irradiation, Ultrason. Sonochem. 3 (1996) S205–S207, https://doi.org/10.1016/S1350-4177(96)00027-2.
- [8] K.S. Suslick, D.A. Hammerton, The site of sonochemical reactions, IEEE Trans. Ultrason. Ferroelectr. Freq. Control 33 (1986) 143–147.
- [9] M. Ashokkumar, R. Hall, P. Mulvaney, F. Grieser, Sonoluminescence from aqueous alcohol and surfactant solutions, J. Phys. Chem. B 101 (50) (1997) 10845–10850.
 [10] F. Grieser, M. Ashokkumar, The effect of surface active solutes on bubbles exposed
- to ultrasound, Adv. Colloid Interface Sci. 89-90 (2001) 423–438.
- [11] K. Vinodgopal, M. Ashokkumar, F. Grieser, Sonochemical degradation of a polydisperse nonylphenol ethoxylate in aqueous solution, J. Phys. Chem. B 105 (16) (2001) 3338–3342.
- [12] R. Tronson, M. Ashokkumar, F. Grieser, Multibubble sonoluminescence from aqueous solutions containing mixtures of surface active solutes, J. Phys. Chem. B 107 (30) (2003) 7307–7311.
- [13] J. Lee, S.E. Kentish, M. Ashokkumar, The Effect of Surface-Active Solutes on Bubble Coalescence in the Presence of Ultrasound, J. Phys. Chem. B 109 (11) (2005) 5095–5099.
- [14] J. Lee, S. Kentish, M. Ashokkumar, Effect of surfactants on the rate of growth of an air bubble by rectified diffusion, J. Phys. Chem. B 109 (2005) 16860–16865.
- [15] D. Sunartio, K. Yasui, T. Tuziuti, T. Kozuka, Y. Iida, M. Ashokkumar, F. Grieser, Correlation between Na* emission and "Chemically Active" acoustic cavitation bubbles, ChemPhysChem 8 (16) (2007) 2331–2335.
- [16] R. Singla, F. Grieser, M. Ashokkumar, The mechanism of sonochemical degradation of a cationic surfactant in aqueous solution, Ultrason. Sonochem. 18 (2) (2011) 484–488.
- [17] J.Z. Sostaric, M. Ashokkumar, F. Grieser, Sodium atom emission from aqueous surfactant solutions exposed to ultrasound, Langmuir 32 (2016) 12387–12393.
- [18] J.Z. Sostaric, P. Riesz, Sonochemistry of surfactants in aqueous solutions: An EPR spin-trapping study, J. Am. Chem. Soc. 123 (2001) 11010–11019.
- [19] J.Z. Sostaric, P. Riesz, Adsorption of surfactants at the gas/solution interface of cavitation bubbles: an ultrasound intensity-independent frequency effect in sonochemistry, J. Phys. Chem. B 106 (2002) 12537–12548.
- [20] H. Destaillats, H.-M. Hung, M.R. Hoffmann, Degradation of alkylphenol ethoxylate surfactants in water with ultrasonic irradiation, Environ. Sci. Technol. 34 (2) (2000) 311–317.
- [21] G.-Y. Pee, J.F. Rathman, L.K. Weavers, Effects of surface active properties on the cavitational degradation of surfactant contaminants, Ind. Eng. Chem. Res. 43 (17) (2004) 5049–5056.
- [22] L. Yang, J.F. Rathman, L.K. Weavers, Degradation of alkylbenzene sulfonate surfactants by pulsed ultrasound, J. Phys. Chem. B 109 (33) (2005) 16203–16209.
- [23] L. Yang, J.F. Rathman, L.K. Weavers, Sonochemical degradation of alkylbenzene sulfonate surfactants in aqueous mixtures, J. Phys. Chem. B 110 (37) (2006) 18385–18391.
- [24] L. Yang, J.Z. Sostaric, J.F. Rathman, L.K. Weavers, Effect of ultrasound frequency on pulsed sonolytic degradation of octylbenzene sulfonic acid, J. Phys. Chem. B 112 (3) (2008) 852–858.
- [25] B. Nanzai, K. Okitsu, N. Takenaka, H. Bandow, Effects of initial concentration of LASs on the rates of sonochemical degradation and cavitation efficiency, Res. Chem. Intermed. 35 (8-9) (2009) 841–849.
- [26] B. Nanzai, K. Okitsu, N. Takenaka, H. Bandow, Sonochemical degradation of alkylbenzene sulfonate and kinetics analysis with a langmuir type mechanism, J. Phys. Chem. C 113 (2009) 3735–3739, https://doi.org/10.1021/jp809509g.
- [27] M. Takahashi, ζ Potential of microbubbles in aqueous solutions: electrical properties of the gas-water interface, J. Phys. Chem. B 109 (2005) 21858–21864, https://doi.org/10.1021/jp0445270.
- [28] B. Nanzai, K. Okitsu, N. Takenaka, H. Bandow, N. Tajima, Y. Maeda, Effect of reaction vessel diameter on sonochemical efficiency and cavitation dynamics, Ultrason. Sonochem. 16 (1) (2009) 163–168.
- [29] K. Okitsu, B. Nanzai, K. Kawasaki, N. Takenaka, H. Bandow, Sonochemical decomposition of organic acids in aqueous solution: understanding of molecular behavior during cavitation by the analysis of a heterogeneous kinetics model, Ultrason. Sonochem. 16 (2009) 155–162, https://doi.org/10.1016/j. ultsonch.2008.06.006.
- [30] A. Tauber, G. Mark, H.-P. Schuchmann, C. von Sonntag, Sonolysis of tert-butyl alcohol in aqueous solution, J. Chem. Soc., Perkin Trans. 2 (6) (1999) 1129–1136, https://pubs.rsc.org/-/content/articlepdf/1999/p2/a901085h.
- [31] Y.-C. Lee, K.J. Stebe, H.-S. Liu, S.-Y. Lin, Adsorption and desorption kinetics of C_mE₈ on impulsively expanded or compressed air–water interfaces, Colloids Surf. A 220 (1-3) (2003) 139–150.
- [32] M.J. Rosen, J.T. Kunjappu, Surfactants and Interfacial Phenomena, 4th ed., John Wiley & Sons Inc, New Jersey, 2012 https://onlinelibrary.wiley.com/doi/book/ 10.1002/9781118228920.
- [33] H. Okuno, B. Yim, Y. Mizukoshi, Y. Nagata, Y. Maeda, Sonolytic degradation of hazardous organic compounds in aqueous solution, Ultrason. Sonochem. 7 (4) (2000) 261–264.
- [34] Y. Nagata, K. Hirai, H. Bandow, Y. Maeda, Decomposition of hydroxybenzoic and humic acids in water by ultrasonic irradiation, Environ. Sci. Technol. 30 (4) (1996) 1133–1138.
- [35] G.D. Christian, P.K. Dasgupta, K.A. Schug, Analytical Chemistry, 7th ed., John Wiley & Sons Inc, New Jersey, 2013 https://www.wiley-vch.de/en/areas-interest/ natural-sciences/analytical-chemistry-978-0-470-88757-8.
- [36] T.W. Healy, C.J. Drummond, F. Grieser, B.S. Murray, Electrostatic surface potential and critical micelle concentration relationship for ionic micelles, Langmuir 6 (2) (1990) 506–508.

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- [37] E.J. Hart, A. Henglein, Free radical and free atom reactions in the sonolysis of
- aqueous iodide and formate solutions, J. Phys. Chem. 89 (20) (1985) 4342–4347.
 [38] J. Rae, M. Ashokkumar, O. Eulaerts, C. von Sonntag, J. Reisse, F. Grieser, Estimation of ultrasound induced cavitation bubble temperatures in aqueous solutions, Ultrason. Sonochem. 12 (5) (2005) 325-329.
- [39] E. Ciawi, J. Rae, M. Ashokkumar, F. Grieser, Determination of temperatures within acoustically generated bubbles in aqueous solutions at different ultrasound frequencies, J. Phys. Chem. B 110 (27) (2006) 13656–13660.