



Soil Bacterial Communities Exhibit Strong Biogeographic Patterns at Fine Taxonomic Resolution

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ABSTRACT Bacteria have been inferred to exhibit relatively weak biogeographic patterns. To what extent such findings reflect true biological phenomena or methodological artifacts remains unclear. Here, we addressed this question by analyzing the turnover of soil bacterial communities from three data sets. We applied three methodological innovations: (i) design of a hierarchical sampling scheme to disentangle environmental from spatial factors driving turnover; (ii) resolution of 16S rRNA gene amplicon sequence variants to enable higher-resolution community profiling; and (iii) application of the new metric zeta diversity to analyze multisite turnover and drivers. At fine taxonomic resolution, rapid compositional turnover was observed across multiple spatial scales. Turnover was overwhelmingly driven by deterministic processes and influenced by the rare biosphere. The communities also exhibited strong distance decay patterns and taxon-area relationships, with z values within the interquartile range reported for macroorganisms. These biogeographical patterns were weakened upon applying two standard approaches to process community sequencing data: clustering sequences at 97% identity threshold and/or filtering the rare biosphere (sequences lower than 0.05% relative abundance). Comparable findings were made across local, regional, and global data sets and when using shotgun metagenomic markers. Altogether, these findings suggest that bacteria exhibit strong biogeographic patterns, but these signals can be obscured by methodological limitations. We advocate various innovations, including using zeta diversity, to advance the study of microbial biogeography.

IMPORTANCE It is commonly thought that bacterial distributions show lower spatial variation than for multicellular organisms. In this article, we present evidence that these inferences are artifacts caused by methodological limitations. Through leveraging innovations in sampling design, sequence processing, and diversity analysis, we provide multifaceted evidence that bacterial communities in fact exhibit strong distribution patterns. This is driven by selection due to factors such as local soil characteristics. Altogether, these findings suggest that the processes underpinning diversity patterns are more unified across all domains of life than previously thought, which has broad implications for the understanding and management of soil biodiversity.

KEYWORDS biogeography, desert, soil bacteria, turnover, zeta diversity

central goal of microbial ecology is to link microbial distribution patterns to underlying ecological processes. Developing such links is important both for fundamental science and applied outcomes, for example to make accurate global biodiversity assessments and prioritize management goals in the face of both local and global change (1, 2). However, achieving this critically depends on our abilities to adequately characterize biodiversity at the first stage, with various methodological and

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🏏 Bay and colleagues show that soil bacteria exhibit much stronger spatial structure than previously thought, with scaling patterns similar to those of macroorganisms. They also introduce new methods to advance the field of bacterial biogeography.

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theoretical challenges limiting our understanding of microbial distribution patterns and their underlying ecological drivers.

Several principles have nevertheless become established in soil microbial ecology through cultivation-independent studies over the last two decades. First, it is appreciated that most soils harbor rich and abundant bacterial communities (3-6). In most soils, a small number of taxa are abundant and prevalent, while the remaining taxa have low abundance and frequency (the "rare" biosphere) (7, 8). In common with macroorganisms (9), four key ecological processes control microbial assembly across space and time: environmental selection, diversification, dispersal, and drift (10–12). While much work has emphasized the role of deterministic environmental selection in driving bacterial niche differentiation, especially edaphic factors such as pH (13-17), some studies have also inferred stochastic patterns of community structure, for example due to dispersal limitation or historical diversification (17-21). The relative strength of these factors can vary across time, for example with dispersal controlling recruitment and selection affecting retention during initial stages of primary succession (17, 22-24). As is also the case in the field of macroecology, the relative importance of deterministic and stochastic processes in shaping contemporary distributions of microorganisms continues to be debated and there is a large body of often divergent literature in this area. In this regard, a major methodological challenge is to perform sampling and analysis that sufficiently disentangles the autocorrelation between environmental and spatial factors in soil ecosystems (12, 25-27).

Also controversial is the extent to which microbial communities vary across space. Soil bacteria are generally thought to exhibit weaker biogeographic patterns than macroorganisms (15, 28). Most empirical studies have reported low exponents for taxa-area relationships (13, 15, 29–31) and low regression coefficients in distance decay curves (13, 19, 28, 32, 33), though exceptions have been reported (34-37). Several hypotheses have been put forward to explain these observations (38, 39). Primarily, bacteria are thought to be able to maintain wide geographic ranges in the face of environmental variation by entering dormant states (39, 40), leading to limited geographic turnover and shallow taxon-area curves (15, 28, 41). However, methodological artifacts may also account for some observations of weak spatial differences (28). Microbial biogeographic patterns are known to be sensitive to various factors, including spatial scale (42, 43), sampling effort (28, 44–46), and taxonomic resolution (15, 28, 44, 47-49). Communities are inherently prone to being undersampled, whether through insufficient sampling effort, low sequencing depth, or rarefying data (50, 51). In addition, the processing of 16S rRNA gene amplicon sequencing data typically used to profile communities can reduce data set resolution; reads are usually clustered into operational taxonomic units (OTUs) based on an arbitrary identity threshold (usually 97%), and the rare biosphere is regularly removed (52, 53). Compounding these issues, the pairwise analyses generally used to quantify community turnover inadequately partition variation from all community members: incidence-based measures are highly sensitive to the rare biosphere, and abundance-based measures focus on the common few (54, 55).

In this study, we employed three methodological innovations to address these common limitations of microbial biogeographic surveys and reassess patterns of bacterial community turnover. First, we adopted a hierarchical sampling scheme commonly used in macroecological surveys (56, 57); this enabled us to detect changes in community structure across multiple spatial scales and, in light of controversies in the literature, better distinguish the contributions of environmental and spatial drivers to community assembly processes (27). Second, we profiled community composition using high-resolution 16S rRNA gene amplicon sequence variants (ASVs), leveraging a new generation of sequence processing tools (58–60). We compared the effects of the commonly used approaches of filtering and clustering sequences reduces taxonomic resolution and thus may increase the overall similarity of the community, thereby weakening biogeographic patterns (35, 49). Finally, we used the multisite diversity





FIG 1 Occupancy frequency distribution of amplicon sequence variants (ASVs) at different taxonomic resolutions. The Kernelsmoothed density plot shows the number of sites that each taxon (ASVs) was detected in across the data set. (a) Effect of clustering taxa at either 100% or 97% identity threshold. (b) Effect of either including or removing taxa with lower than 0.05% relative abundance. The vertical dashed lines show distribution means.

metric zeta diversity to analyze spatial community turnover and predict the strength of underlying deterministic and stochastic drivers (55). Unlike the commonly used beta diversity that is calculated from pairwise comparisons, zeta diversity describes the number of taxa shared across multiple sites. As a result, this parameter can discriminate diversity patterns across the spectrum of common, intermediate, and rare taxa (55, 61–63) and infer deterministic and stochastic drivers of community assembly. On this basis, we provide evidence that at the level of exact sequences, bacterial biogeographic patterns are exceptionally stronger than previously reported.

RESULTS

Most community members have a low to moderate occupancy across soil transects. We analyzed 96 topsoil samples along two perpendicular transects (see Fig. S1a in the supplemental material): a 160-km latitudinal transect (north/south) spanning four climatic zones (subhumid, semiarid, arid, and hyperarid; 69 samples) and a 20-km longitudinal transect (east/west) in the arid zone (27 samples). Within each transect, samples were collected according to a hierarchical design (2 sites per zone \times 3 plots per site \times 3 samples per plot) (Fig. S1b). This sampling scheme was designed to enable the analysis of microbial community turnover at multiple spatial scales, capture a wide spectrum of distance classes (Fig. S1c), and discriminate underlying spatial and environmental drivers.

The bacterial and archaeal communities in each sample were profiled using both new and standard approaches for processing 16S rRNA gene amplicon sequencing data. Rarefaction curves (Fig. S2a to c) and richness estimators (Fig. S2d) confirmed that sequencing and sampling efforts sufficiently captured the diversity of taxa within and across samples. A high-resolution community profile was generated by processing reads using the deblur pipeline (59) to resolve 16S rRNA gene amplicon sequence variants (ASVs) at the single-nucleotide level (singletons removed) (see Data Set S1, tab 1, in the supplemental material). Most sequences were from the nine dominant soil phyla (7), especially *Actinobacteriota, Chloroflexota,* and *Proteobacteria,* as well as putatively ammonia-oxidizing archaea (Fig. S3). The occupancy frequency distribution (64) of the 11,335 taxa (ASVs) detected was positively skewed; ~67% of 7,602 taxa were detected in fewer than 10% of samples (Fig. 1a).

We then compared the effects of applying two standard approaches used to process sequencing data into OTUs: (i) clustering, i.e., combining sequences with an identity threshold of 97%, and (ii) filtering, i.e., removing sequences with lower than 0.05%





FIG 2 Patterns and drivers of beta diversity in the latitudinal transect. (a) Multidimensional scaling visualizing taxonomic and phylogenetic pairwise incidence and abundance dissimilarity of microbial communities. The axes show the explained variation of taxa (ASVs) between samples using four different dissimilarity metrics: Jaccard (taxonomic incidence-based), Bray-Curtis (taxonomic abundance-based), unweighted Unifrac (phylogenetic incidence-based), and weighted Unifrac (phylogenetic abundance-based) dissimilarity metrics. The PCoA ordination is compared at four different taxonomic resolutions (taxa clustered at 100% or 97% identity; taxa with <0.05% relative abundance included or removed). (b) Variation partitioning analysis delineating the relative contributions of environmental and spatial sources of variation on microbial community structure. The analysis shows the proportion of variation in microbial incidence between sample pairs as explained by environmental, spatial, overlapping, and unexplained sources of variation. These analyses were performed using data from each plot in the latitudinal transect. Results are compared at four different taxonomic resolutions (taxa clustered at 100% or 97% identity; taxa with <0.05% relative abundance included or removed). Data Set S2, tab 3, in the supplemental material summarizes the environmental variables that best explain the variation along the transect.

relative abundance (Data Set S1, tab 2). There was a sharp decrease in the number of taxa retained (2,943 clustered, 222 filtered, 403 clustered then filtered). Though clustering inevitably reduced richness, as well as the frequency of intermediate taxa, it did not affect the skew of the occupancy frequency distribution (Fig. 1a). However, when less abundant taxa were filtered from the data sets, occupancy frequency shifted from a positive skew to a modal distribution (Fig. 1b). These findings suggest that the prevalence of most community members is low to moderate; standard clustering and filtering approaches not only affect the "rare" biosphere, but a large percentage of community members with moderate range sizes. In turn, changing occupancy properties may underestimate ecological heterogeneity and markedly bias biogeographic interpretations.

Deterministic factors drive differences in community composition between soil samples. We subsequently used pairwise metrics (beta diversity) to analyze community composition between samples. We detected significant differences in microbial richness down to the site level (Data Set S1, tab 1) and community structure down to the plot level (Data Set S2, tab 2). The extent of compositional differences observed between sites depended on both the community property used (incidence versus abundance, taxonomic versus phylogenetic) and the taxonomic resolution of the data set. Principal-coordinate analysis (PCoA) ordinations showed prominent "V" patterns (Fig. 2a); this pattern, also known as the horseshoe effect, has been shown to indicate the presence of niche differentiation along environmental gradients (65). In line with the high environmental heterogeneity along the latitudinal transect, differentiation was more pronounced for the latitudinal transect than for the longitudinal transect (Fig. S4a).



A variation partitioning analysis was used to delineate the measured environmental and spatial predictor variables that best explain pairwise community structure. Across the latitudinal gradient, 45% of the community variation of the high-resolution data set was explained by measured edaphic factors (Fig. 2b; Data Set S1, tab 3) with pH, C/N ratio, aridity, and salinity explaining the greatest amount of variation (Data Set S2, tab 3). These results broadly reflect other studies in the Negev region and along aridity gradients globally (66–68). Less variation was explained for the more homogeneous longitudinal transect (35%) (Fig. S4; Data Set S2, tab 4). Altogether, these results suggest environmental effects predominate over distance effects in driving community composition.

In common with other biogeographic studies (13, 69, 70), a large proportion of variation was unexplained by the measured variables. A combination of factors could contribute to this unexplained variation, including deterministic processes driven by unmeasured abiotic and biotic factors, as well as neutral ecological drift and potentially sampling effects. In both the PCoA and variation partitioning analyses, less variation in community composition could be explained and partitioned for the high-resolution data set compared to filtered ones (Fig. 2 and Fig. S4). The rank importance and weight of environmental predictors also shifted depending on taxonomic resolution for both transects (Data Set S2, tabs 3 and 4). In support of recent findings (71), these results suggest that different environmental drivers structure common and rare microbial taxa.

Soil microbial communities exhibit rapid deterministically driven multisite turnover. We also analyzed spatial turnover in the community using the recently developed metric zeta diversity. As depicted in the infographic in Fig. S5, zeta diversity describes the number of taxa shared by multiple combinations of sites; whereas beta diversity (which it encompasses) is predisposed to detect turnover of rare taxa, zeta diversity discriminates patterns and drivers of turnover across the spectrum of common, intermediate, and rare taxa (55, 61).

For the high-resolution data set, zeta diversity rapidly declined toward zero within four orders in the latitudinal transect ($\zeta_4 = 0.0068$) (Fig. 3a). This means that the average number of taxa shared across any four plots was 0.68% of 10,826, indicating very rapid turnover. Similar patterns were observed across both transects and within each climatic zone; somewhat lower turnover was observed along the longitudinal transect ($\zeta_4 = 0.010$) (Fig. S6). Reducing taxonomic resolution markedly slowed compositional turnover (Fig. 3a); for the clustered and filtered data set, up to 30% of the community were shared across any four plots ($\zeta_4 = 0.18$ and 0.30 for the latitudinal and longitudinal transects, respectively). Such findings reflect that, given that common, intermediate, and rare community members show different distribution patterns, lowering taxonomic resolution distorts detection of microbial turnover and underlying drivers.

Derivations show that zeta decline most often follows either a power law or an exponential form, which are, respectively, associated with either deterministic or stochastic community assembly processes (55). Zeta decline much better fitted a power law form for both transects and within each climatic zone (Fig. 3c and Fig. S6). This therefore suggests that deterministic processes drive turnover and rejects the null hypothesis that microbial communities are randomly distributed. While power law support was overwhelming for the high-resolution data set, there was some support for exponential models in the low-resolution data sets; filtering microbial data sets, by obscuring biogeographic structure, may therefore cause false signals of stochastic assembly processes (Fig. S6).

Soil microbial communities exhibit strong distance decay and taxon-area relationships. We subsequently measured distance decay using a combination of pairwise (beta decay) and multisite (zeta decay) metrics. Based on pairwise comparisons, a strong decay of shared taxa was also detected across transects (P < 0.0001) (Fig. 4a and Fig. S4c; Data Set S2, tab 5). Lowering taxonomic resolution caused a large increase in community similarity, a steeper distance decay coefficient, and a lower rate of community turnover overall; across the 160-km latitudinal transect, there was a 82%





FIG 3 Multisite community turnover and assembly processes along the latitudinal transect at different taxonomic resolutions. (a) Zeta decline showing how the number of shared taxa (ASVs) decline with the addition of sites to the comparison (Order). (b) Zeta diversity ratio showing rate of taxon retention. This demonstrates the probability of retaining common over rare taxa at any particular order with the addition of an extra site. (c) Zeta decline followed a power law form in all cases, which is associated with deterministic processes driving community turnover. (d) Statistical support for the power law form, relative to an exponential form, varies depending on the taxonomic resolution.

reduction in community similarity for the high-resolution data set compared to 50% to 60% reductions for the clustered and/or filtered data sets. Given the concordant support for deterministic drivers, based on the zeta diversity (Fig. 3), variation partitioning analysis (Fig. 2b), and PCoA analysis (Fig. 2a), these decay patterns likely reflect environmental filtering rather than dispersal limitation.

To quantify how distance decay compares between rare, intermediate, and common taxa, distance decay was calculated for up to six zeta orders by using the mean distance between pairs for up to six plots. For both transects at high resolution, the gradient of the distance decay curve rapidly and significantly decreased with increasing zeta order (Fig. 4b and Fig. S4d). This provides additional evidence that these microbial communities are highly structured and that turnover is driven by loss of rare to intermediate members. In contrast, there were no significant changes in distance decay rates with zeta order for the less-resolved data sets, further demonstrating that clustering and/or filtering obscures biogeographic patterns.

Given these outcomes, we revisited the controversial taxa-area relationship for bacterial communities (15, 44) using these data sets. This universal relationship in ecology describes the increase in taxon richness with area sampled, i.e., $S = cA^z$ (where *S* is number of species, *A* is area sampled, and *c* and *z* are fitted constants), and its exponent *z* is a normalized measure of turnover rates that can be compared between organismal groups (15). A strong taxa-area relationship was also observed for both transects (P < 0.001) (Fig. 4c; Fig. S4e). The *z* exponents were 0.39 (latitudinal transect) and 0.40 (longitudinal transect) for the original high-resolution data sets, and the exponents decreased to 0.13 and 0.09 in the clustered then filtered data sets (Data Set





FIG 4 Distance decay in community similarity and the taxa-area relationship at different taxonomic resolutions. (a) Zeta distance decay relationship showing community turnover with increasing geographic distance based on pairwise comparisons (ζ_2) of sites along the latitudinal transect. (b) Differences in the slope (coefficient) of distance decay between pairwise and higher orders of zeta (>2) using the average distance between sites. (c) Taxa-area relationships of the increase in richness with area sampled along the latitudinal transect. This shows effects of taxonomic resolution for understanding how compositional heterogeneity scales with distance. (d) Violin plots showing the density distribution and interquartile range of the exponent *z* of the taxa-area slope reported here with those from other studies for bacteria and eukaryotes (Data Set S2, tab 6). Results are compared at four different taxonomic resolutions, whereby (i) taxa were clustered at either 100% or 97% identity threshold and (ii) taxa with lower than 0.05% relative abundance were either included or removed.

S2, tab 6). Such *z* exponents greatly exceed those reported for bacterial communities in most previous studies (median, 0.04), but are congruent with four studies (29, 34, 35, 37), two of which also performed hierarchical sampling. These exponents are of the same order of magnitude as those previously reported for animal and plant data sets (median, 0.27) (Fig. 4d; Data Set S2, tab 6), indicating that biogeographic patterns of bacteria and macroorganisms may not profoundly differ. However, more broad and detailed side-by-side sampling is required to compare scaling relationships between bacteria and macroorganisms.

Similar biogeographic patterns are observed using metagenomic sequences and global data sets. This study relies on 16S rRNA gene amplicon sequencing to profile the soil microbial communities. This approach remains standard practice for biogeographic studies, given that the alternative of metagenomic profiling requires much higher sequencing depths and yields either less information-rich short reads or more error-prone long reads (72). However, limitations of 16S rRNA gene sequencing include potential for amplification and sequencing errors, biases in the primer sets, and genome variability in 16S rRNA gene copy number (73). While it is possible that the data set includes some spurious sequences introduced through this approach, these are



unlikely to account for the surprising observations made here. First, a range of accuracy measures suggest deblur efficiently denoises sequencing data and that a 100% identity threshold resolved using the deblur denoising pipeline is optimal for community profiling with the V4 region (58, 59). Second, similar but weaker patterns of rapid deterministically driven community turnover were observed for the clustered (but not filtered) data sets, in which most spurious sequences should be removed (Fig. 2 and 3).

To test the reproducibility of our findings, we performed short-read metagenomic sequencing of a subset of 12 samples across the latitudinal transect and analyzed a single-copy ribosomal marker gene, *rplP*. Similar to the 16S rRNA gene amplicon data, samples showed a high estimated richness, comparable taxonomic composition, and rapid community turnover (Fig. S7). Zeta decline approached zero after three orders ($\zeta_3 = 0.06$) using a rarefied data set (Fig. S7c). In combination, this suggests that 16S rRNA gene ASVs are sufficient to estimate community turnover, whereas standard methods of clustering and filtering data obscure biogeographic patterns and inflate signals of taxon commonness.

Having detected these patterns at local and regional scales, we analyzed whether similar patterns were observable at the continental scale. To do so, we analyzed previously published 16S rRNA gene profiles of 237 soil samples collected from six continents (7). As with our original data set, we processed the 16S rRNA amplicon sequencing data into ASVs and analyzed the effects of clustering and/or filtering. The occupancy frequency distribution of the taxa showed a skew similar to that of the Israel data set (Fig. S8). Concordant with our previous observations, zeta diversity rapidly declined across the first few orders and followed a power law relationship with strong model support (Fig. S8). Clustering and filtering altered the occupancy frequency distribution, resulting in \sim 10% to 30% of taxa being retained at six zeta orders (Fig. S8). Thus, the key finding that soil bacterial communities exhibit strong biogeographic patterns is reproducible in data sets at local (longitudinal transect), regional (latitudinal transect), and global scales.

DISCUSSION

In this study, we analyze patterns and drivers of soil microbial composition across multiple scales. Steps were taken to overcome common limitations in microbial biogeographical studies by leveraging innovations in sampling design, amplicon processing, and diversity metrics. We found the following. (i) Soil bacterial communities exhibit strong biogeographic patterns. (ii) Spatial turnover is rapid, as most taxa have low to moderate levels of occupancy. (iii) Community structure is influenced more by niche differentiation due to environmental variation rather than dispersal limitation. Our findings agree with previous literature that reported the uneven distribution of bacteria across communities and the strong influence of deterministic drivers (3, 12). However, we observed much stronger spatial turnover than reported in most, though not all, previous literature (13, 15, 28). This is reflected by the compatible findings of four independent analyses using the original high-resolution data set. Occupancy frequency distributions revealed most taxa were shared across less than 10% of samples (Fig. 1). Through zeta decline analysis, we detected a logarithmic decrease in the number of taxa shared as the number of sites increased (Fig. 3). In addition, we observed strong distance decay (Fig. 4a) and taxon-area relationships (Fig. 4c), with z values one to two orders of magnitude higher than most previous observations (13, 29, 34, 35, 37).

Multiple factors may explain why we observed high environmentally driven turnover. These potentially include the choices of sampling site, sampling scheme, sequence processing, and downstream analyses. It is notable that our desert sampling sites contained loessial soils that facilitate dispersal and the regional transect contained high environmental heterogeneity, which is known to be associated with increased bacterial turnover (31, 66, 69); however, this is unlikely to primarily account for most discrepancies with previous literature, given that rapid turnover was also observed in the local transect where physicochemical variation was lower and similar findings were also made in the global analysis. A more significant factor may be that our study



adopted a hierarchical sampling design in order to quantify microbial variation across multiple spatial scales. In this regard, it is well-recognized that sampling design and sample size are critical determinants of taxa-area relationships (44, 74); this reflects that the detection of rare taxa largely determines species evenness and spatial structure, which in turn affects the exponent z (35). Methodological advances that improve the detection and inclusion of rare taxa are therefore predicted to align microbial z values more closely with those reported for animal and plant communities (44, 48). it is notable that other studies reporting high taxon-area exponents also used spatially explicit hierarchical designs (34, 35).

However, the biggest factor likely underlying these discrepancies is the treatment of sequencing data. A pervasive feature of 16S rRNA gene amplicon surveys is the clustering of similar sequences to remove potential "noise" and, less commonly, the filtering or undersampling of low-frequency sequences that constitute the rare biosphere. As summarized in Fig. 5, such processing greatly reduces and distorts the information in data sets, obscuring patterns in occupancy, turnover, and drivers. We avoided such downfalls by using a recently developed denoising algorithm to resolve sequence variants (59), while confirming through rarefaction curves that our sequencing efforts captured most rare taxa within and between samples. Through simulating sequence processing, we observed major differences in occupancy frequency, zeta diversity, distance decay, and taxon-area relationships upon filtering rare taxa and, to a lesser extent, clustering similar sequences (Fig. 5). It should be noted that these observations may appear to conflict with those of a recent study that reported clustering did not "change the rate of microbial taxonomic turnover" (28). However, this may be an issue of interpretation of distance decay curves. In common with this study (28), we also observed that the distance decay coefficient of bacteria and archaea remains similar between taxonomic resolutions, reflecting similar observations reported in fungal (75) and plant (76) communities. However, as the community similarity (y intercept) is lower at higher resolution, a higher proportion of taxa are lost overall in unclustered data sets compared to clustered data sets. Thus, it is reasonable to conclude that clustering masks microbial taxonomic turnover and broader biogeographic patterns.

This study also highlights the different patterns and drivers of community turnover between rare, common, and intermediate community members. As demonstrated by the occupancy frequency distribution, filtering sequences removes most rare species and retains most common ones. On the basis of the results of beta diversity analysis, we observed significant differences in the proportion of variation assigned to environmental, spatial, shared, or unexplained components at different taxonomic resolutions. This agrees with recent reports that environmental and spatial drivers differentially act on common and rare taxa (77, 78). Abundant generalists and rare specialists have been shown to differentially respond to environmental change, reflecting differences in niche breadth (11, 79, 80). Beyond these pairwise observations, we used zeta diversity to demonstrate that the turnover patterns reflect those typically observed in deterministically structured communities. Zeta decline consistently follows a power law, which indicates that communities are nonrandomly structured such as those with clear niche or range differentiation. However, upon lowering taxonomic resolution, these patterns degrade and increasingly resemble stochastic patterns such as seen in habitats with strong aeolian forces or aquatic flows (Fig. 5). These findings suggest that at lower taxonomic resolutions (49) or when rare taxa are removed (35), the community structure becomes more similar and thus predicted assembly processes switch from deterministic to stochastic. Through incorporating a multisite distance decay model, significant differences in the spatial structure of rare, intermediate, and common taxa could also be detected.

Looking forward, this work demonstrates how microbial biogeography can be advanced using readily implementable approaches. There is scope to use the methodological and theoretical innovations shown here to investigate these patterns across a broader range of environments and temporal scales. Detailed studies are needed to





FIG 5 Summary of biogeographic patterns of soil microbial communities at different taxonomic resolutions. (a) Principle of how zeta diversity encompasses turnover of rare, intermediate, and common community members. (b) Comparison of patterns of occupancy frequency, zeta decline, and distance decay for rare, intermediate, and common community members. In addition, the figure demonstrates how the common approaches of clustering and filtering can bias biogeographic interpretations of microbial communities.



better capture the biotic and abiotic subsets of drivers responsible for changes in community turnover across all occupancy classes; this has been achieved in plant ecology (61, 81) but is currently lacking in microbial ecology. Likewise, it is critical to compare the patterns and drivers of community turnover in parallel for microorganisms and macroorganisms. Indeed, a key observation of our study is that the *z* exponent for the taxon-area relationships microbial communities falls within the interquartile range of higher animal and plant communities, suggesting that microorganisms and macroorganisms exhibit similarly strong spatial structure. However, given that these exponents are highly sensitive to factors such as sampling design, sample size, and taxonomic resolution (44, 74), a rigorous comparison of turnover between domains requires side-by-side sampling. Finally, emerging advances in long-read and full-length 16S rRNA gene sequencing may enable resolution of biogeographic patterns of microorganisms at both the species and strain levels (82, 83).

MATERIALS AND METHODS

Soil survey design. Topsoil samples were collected along perpendicular latitudinal and longitudinal transects in the Judea Hills and Negev Desert regions, Israel. The latitudinal transect, which was designed to capture a high level of environmental heterogeneity, extended for 160 km in a north/south direction along a steep aridity gradient. This transect traversed four climatic zones that were differentiated by mean annual precipitation patterns: subhumid shrubland (300 to 400 mm/year), semiarid grassland (\sim 200 to 250 mm/year), arid desert (\sim 50 to 90 mm/year), and hyperarid desert (<20 mm/year). The longitudinal transect, sampled within the arid zone across a relative homogenous climate, extended perpendicular to the latitudinal transect for 20 km in an east/west direction.

A hierarchical sampling scheme was used to capture biogeographic patterns across multiple spatial scales and provide sufficient spatial resolution to cover the majority of distance classes between sites (see Fig. S1c in the supplemental material). Three spatial hierarchies were within each climatic zone: (i) site level (two representative sites of ~1,000 m²), (ii) plot level (three representative plots of ~100 m²), and (iii) sample level (random triplicates of ~100 cm²) (Fig. S1b). Site selection was based on four criteria: (i) soil type (wind-deposited loessic soils in the subhumid, semiarid, and arid zones and gypsic soils in the hyperarid zone), (ii) presence of soil crust to indicate no recent disturbance, (iii) vegetation-free soil to minimize a vegetation effect, and (iv) a buffer of 100 m from roads, slopes, and seasonal runoff water channels. No statistical methods were used to predetermine sample size.

Soil sampling and analysis. In total, 99 topsoil samples were collected across both transects over a 10-day period in May 2017. Prior to sampling, GPS coordinates and site metadata were recorded. Soil samples of approximately 50 g were collected in triplicate, using sterile techniques, by removing the soil crust (0- to 2-cm depth) and then sampling the underlying topsoil (2- to 10-cm depth). Samples were placed into individual 50-ml screw top falcon tubes and stored at 4°C until downstream analysis.

Within 24 h of sampling, all soils were homogenized by sieving (500 μ m) and soil water content (as a percentage) was measured gravimetrically in duplicate. All samples were then shipped to guarantine approved facilities at the School of Biological Sciences, Monash University. For soil chemistry analysis, samples were pooled to form one representative sample per plot and sent to the Environmental Analysis Laboratory, Southern Cross University. In total, 21 separate soil chemical parameters were selected for analysis, based on commonly reported drivers of soil microbial communities globally and those reported by previous studies of the Judea Hills and Negev Desert (66, 84). These parameters included the following: soil acidity (pH), electrical conductivity (EC), effective cation exchange capacity (ECEC), total organic carbon, total nitrogen, sodium (Na), sulfur (S), phosphate (P), potassium (K), nitrate (NO_3^{-}), and ammonium (NH_{*}^{+}) , as well as bioavailable minerals, including manganese (Mn), copper (Cu), zinc (Zn). boron (B), aluminum (AI), iron (Fe), and silicon (Si). Each chemical parameter was calculated following the methods of Rayment and Lyons (85). Aridity data for each site were obtained from a global geospatial data set (86) mapping the aridity index (MAP/PET, where MAP stands for mean annual precipitation and PET stands for potential evapotranspiration) at a resolution of 90 arcseconds (approximately 1 km at the equator) using a climatic time series from 1950 to 2000 (86). This data set includes samples from six continents, Africa, Europe, Asia, Australia, North America, and South America.

Community DNA extraction and sequencing. For all samples, total community DNA was extracted from 0.25 g of soil using the modified Griffith's protocol (87). We confirmed the DNA yield, purity, and integrity for each extraction using a Qubit fluorometer, Nanodrop 1000 spectrophotometer, and agarose gel electrophoresis. For each sample (88), the hypervariable V4 region of the 16S rRNA gene was amplified using the universal Earth Microbiome Project primer pairs F515 and R806 (89). The amplicons were subjected to Illumina paired-end sequencing at the Australian Centre for Ecogenomics, University of Queensland. Twelve samples were also subjected to shotgun metagenomics sequencing (SH.1.A3, SH.1.C2, SH.2.C3, SA.2.B3, SA.1.C3, SA.1.B1, AR.2.A3, AR.2.A1, AR.1.C2, HA.2.C2, HA.1.B1, and HA.1.C2). DNA was extracted from 0.25 g of soil using the MoBio PowerSoil isolation kit according to the manufacturer's instructions. Metagenomic shotgun libraries were prepared for the 12 samples using the Nextera XT DNA sample preparation kit (Illumina Inc., San Diego, CA, USA). Sequencing was performed on an Illumina NextSeq500 platform with 2 × 150 bp high-output run chemistry. For analysis of the previously published global data set (7), the raw 165 rRNA gene amplicon sequences were downloaded from Figshare (https://figshare.com/s/82a2d3f5d38ace925492).



Amplicon-based community profiling. Raw sequences from the Israel and global data sets were processed on the QIIME 2 platform (90) using the deblur pipeline (59) to resolve exact amplicon sequence variants (ASVs). In contrast to operational taxonomic unit (OTU)-based approaches that cluster sequences to a fixed identity threshold (usually 97%), deblur controls error rates (typically 0.1% per nucleotide) to resolve single-nucleotide differences over the sequenced gene region (59).

Paired-end raw reads were demultiplexed, and adapter sequences were trimmed, yielding 3,989,659 reads across all samples. Forward and reverse reads were joined using the q2-vsearch plugin (91). A quality filtering step was applied using a sliding window of four bases with an average base call accuracy of 99% (Phred score of 20). Low-quality reads were removed, and sequences were truncated at 250 bp before denoising using deblur (59). For downstream analysis, three samples with low read counts (<1,000 reads) were excluded (SH.1.B2, AR.1.B1, and AR.1.B2). An additional 414 ASV singletons were detected and removed. The final data set contained 96 samples and 11,335 ASVs (see Data Set S1, tab 1, in the supplemental material). In order to compare biogeographic patterns across different taxonomic resolutions, a second data set was created by clustering the ASVs at a 97% identity threshold using open reference OTU picking via q2-vsearch (91) (Data Set S1, tab 2). The third and fourth data sets were created by removing reads with lower than 0.05% relative abundance from the 100% and 97% identity threshold data sets using the Phyloseq filter_taxa function.

Taxonomic assignment was performed as per a previously described approach (https://osf.io/25djp/ wiki/home/). Briefly, all reference reads that matched the 515F/806R primer pair were extracted from the Genome Taxonomy Database (GTDB) (92) and used to train a naive Bayes classifier (93) by using the fit-classifier-naive-bayes function with default parameters. The classifier was then used to assign the taxonomy to the ASV feature table. Representative sequences were aligned using the multiple sequence alignment program MAFFT (94), and a phylogenetic tree was constructed using the fast-tree method in QIIME 2.

Metagenome-based community profiling. The 16S rRNA gene amplicon sequence is commonly used as a marker to profile microbial communities. However, a major limitation of this approach lies in the intragenomic and intergenomic variation in copy number of the 16S rRNA gene (73, 82). We conducted a comparative metagenomic analysis on a subset of 12 samples (biological triplicate from within each climatic zone) along the latitudinal transect using a single-copy ribosomal marker in order to test whether our observations of community turnover by 16S rRNA gene amplicon sequencing were affected by this variation. A total of 318,420,199 reads were obtained from metagenomic sequencing across the 12 samples. In contrast, the read counts for the negative controls were 6,547 (extraction control) and 1,360 (library preparation control). Raw sequence reads in each sample were stripped of adapter and barcode sequences, then contaminating PhiX sequences were identified and removed using the BBDuk function of BBTools v. 36.92 (https://sourceforge.net/projects/bbmap/) with a kmer size of 31 and hamming distance of 1. We then used SingleM (95), which uses hidden Markov model (HMM) searches of single-copy ribosomal markers, to generate de novo OTUs. In total, 28 HMM searches were performed against 14 single ribosomal single-copy marker genes. GraftM was used for taxonomic annotation of OTUs by searching sequences using hmmsearch (HMMER) (96). For downstream analysis, the single-copy marker gene rplP was used for comparison, encoding ribosomal protein L16 L10e. This marker was previously identified as a robust means of distinguishing between both closely and distantly related genomes (97). Sequences were then clustered *de novo* into OTUs using a sequence identity threshold of 97%. Taxonomic assignment was carried out using the GTDB taxonomy.

Due to large differences in sequencing depth, the amplicon and metagenomic sequences analyzed were both rarefied at 300, which was the minimum number of sequences observed for *rplP*. Rarefied data sets were used only in the supplemental analysis shown in Fig. S7, whereas the rest of the study used unrarefied data sets.

Richness analysis. Statistical analysis and visualizations were performed in R version 3.4.4 (2018-03-15) using the packages ggplot2 (98), phyloseq (99), vegan (100), and zetadiv (55). Occupancy frequency distributions (64) were used to visualize the distributions of the numbers of taxa occupying different numbers of areas and examine the distributional shifts at lower identity thresholds and after filtering rare taxa. Taxa accumulation curves were used to compare alpha diversity properties between sites and confirm adequate sampling of the microbial community. A sample-based rarefaction method was used to find the expected curve, namely, the Mao Tau estimate, and a moment-based standard deviation was estimated from the extrapolated number of ASVs surveyed (gamma diversity) using the "exact" method of the *specacum* function [Vegan | R] (100). Observed richness and estimated richness (Chao1 and abundance coverage estimate [ACE] methods) were calculated using the *estimate_richness* function [Phyloseq | R] (99). To test for significant differences in the mean observed and estimated richness at the site level, an analysis of variance (ANOVA) with a Shapiro-Wilk test to confirm normality was used.

Turnover analysis. The multisite diversity metric zeta diversity (ζ) was used [Zetadiv | R] (55) to examine incidence-based turnover in community composition (Fig. S5). Pairwise metrics of incidence-based turnover (e.g., Jaccard, Simpson index) are biased toward detecting turnover that is driven predominantly by the loss and addition of taxa from the rare biosphere, as by definition rare taxa are not shared by many sites. Zeta diversity overcomes this limitation by enabling discrimination between turnover of rare, intermediate, and common taxa. With increasing orders of zeta, the average number of taxa shared between sites declines and the contribution of increasingly more common taxa to the value of zeta diversity increases. Variation in the rate and form of zeta decline provides information on community structure and inference of the processes driving community assembly. If the zeta decline follows an exponential form (the ratio between ζ_i and ζ_{i-1} is constant), there is a similar probability of



finding a common or rare taxon with the addition of a site, suggesting that turnover is predominantly stochastic or dispersal limited. However, if zeta decline follows a power law form (the ratio between ζ_i and $\zeta_{l,1}$ increases at higher orders), then the chance of detecting a common taxon is greater than detecting a rare one with increasing orders, demonstrating structure in the community and suggesting that turnover is driven primarily by deterministic processes such as selection due to soil or climate (55).

Zeta decline using Monte Carlo sampling was calculated via the *zeta.decline.mc* function [Zetadiv | R] (55). Zeta diversity was calculated on nonweighted presence-absence data for ζ orders ζ_1 to ζ_{ci} ; this captured the extent at which the community was structured across each transect, as ζ values within each data set approached zero. To account for differences in richness between sites, all ζ_i values were normalized by using a Jaccard normalization with subsampling set to 1,000 permutations for each analysis. Power-law and exponential models were fitted to ζ_i decline curves and Akaike information criterion (AIC) were used to estimate the likelihood of either exponential or power law model describing the relationship between ζ diversity and order *i*.

Biogeographic analysis. We calculated the distance decay of similarity across both transects to quantify the number of shared ASVs over geographic distance and to explore turnover within the context of geographic distance. Pairwise distance decay was calculated using normalized ζ_2 , with subsampling set to 1,000 using the function *zeta.ddecay* [Zetadiv | R] (55). To quantify the contribution of rare and common ASVs to turnover, distance decay was calculated for orders ζ_1 to ζ_6 by using the mean distances between pairs of *n* sites via the *zeta.decays* function [Zetadiv | R] (55). Spatially explicit taxa-area relationships (74) were calculated by estimating richness as a function of the sample, plot, and site level spatial hierarchies using the *specnumber* function [Vegan | R] (100). The taxa-area curve was fitted using the Arrhenius model with the expression kA^2 , where *k* is the average number of taxa, *A* is the area (spatial hierarchy), and *z* is the steepness of the curve. For comparison, turnover rates from this study were compared against a total of 655 data sets, including bacteria and higher eukaryotes (101).

Community structure analysis. Principal-coordinate analyses (PCoA) were used on both weighted and unweighted distance matrices. Read counts were normalized to relative abundance, and a square root transformation was applied prior to calculating distances between samples using Bray-Curtis. For non-weighted analysis, read counts were transformed to incidence (presence-absence) and distances were calculated using the Jaccard index. A multivariate model-based framework was used to test for significant differences in community structure among spatial hierarchies and identify the subset of environmental drivers that best explain spatial patterns in community structure [MVAbund | R] (102). Microbial abundance and incidence data typically show a mean-variance relationship, which standard approaches such as permutational multivariate analysis of variance (PERMANOVA), analysis of similarity (ANOSIM), and redundancy analysis (RDA) fail to account for. Instead they rely on pairwise distance matrices which convert multivariate data sets to univariate ones, which has been shown to reduce statistical power. MVAbund solves this problem for nonnormal data by fitting a single generalized linear model (GLM) to each ASV separately and performing resampling of *P* values to determine significance of a shared predictor variable.

In this study, ASV incidence data were modeled using generalized linear models. Mean variance relationships of the data were confirmed by visually inspecting scatterplots showing mean variance as a function of ASV incidence. Model assumptions were validated by inspecting Dunn-Smyth residuals as a function of each predictor variable, and significance was established using a likelihood ratio test (LRT) with PIT-trap bootstrapping (103). To obtain the subset of predictor variables which best explain a multivariate response, significant predictor variables were passed through a forward selection in a multivariate linear model using the top 10 independent variables with the highest average R^2 . A variation partitioning analysis was performed to disentangle the autocorrelation between environmental and geographic distance and partition variation in community structure into its spatial and environmental components. Multisite generalized dissimilarity modeling (MS-GDM) was used to identify the importance of correlates of turnover by regressing ζ_2 against the subset of identified predictor variables at each taxonomic resolution using *zeta.msgdm* function [Zetadiv R] (55), which partitioned the variation into (i) variation explained by distance alone, (ii) variation explained by distance or environment, (iii) variation explained by environment alone, and (iv) unexplained variation.

Data availability. The amplicon and shotgun sequencing data sets generated for this study have been deposited in the NCBI Sequence Read Archive under the BioProject accession no. PRJNA642232.

SUPPLEMENTAL MATERIAL

Supplemental material is available online only. FIG S1, TIF file, 2.7 MB. FIG S2, TIF file, 1.4 MB. FIG S3, TIF file, 1.4 MB. FIG S4, TIF file, 1.6 MB. FIG S5, TIF file, 1 MB. FIG S6, TIF file, 1.5 MB. FIG S7, TIF file, 2 MB. FIG S8, TIF file, 1.5 MB. DATA SET S1, XLSX file, 4.5 MB. DATA SET S2, XLSX file, 0.02 MB.



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C.G., S.L.C., M.A.M., and O.G. conceived and supervised this study. S.K.B., C.G., M.A.M., S.L.C., and O.G. designed the study. S.K.B., O.G., and N.W. were responsible for field sampling. S.K.B. carried out all DNA extractions, sequence processing, and statistical analyses. S.K.B., C.G., M.A.M., S.L.C., D.J.B., and D.J.P. analyzed data. S.K.B. and C.G. wrote and edited the paper with input from all authors.

We declare that we have no conflicts of interest.

REFERENCES

- Cameron EK, Martins IS, Lavelle P, Mathieu J, Tedersoo L, Gottschall F, Guerra CA, Hines J, Patoine G, Siebert J, Winter M, Cesarz S, Delgado-Baquerizo M, Ferlian O, Fierer N, Kreft H, Lovejoy TE, Montanarella L, Orgiazzi A, Pereira HM, Phillips HRP, Settele J, Wall DH, Eisenhauer N. 2018. Global gaps in soil biodiversity data. Nat Ecol Evol 2:1042–1043. https://doi.org/10.1038/s41559-018-0573-8.
- Coyle DR, Nagendra UJ, Taylor MK, Campbell JH, Cunard CE, Joslin AH, Mundepi A, Phillips CA, Callaham MA. 2017. Soil fauna responses to natural disturbances, invasive species, and global climate change: current state of the science and a call to action. Soil Biol Biochem 110: 116–133. https://doi.org/10.1016/j.soilbio.2017.03.008.
- Fierer N. 2017. Embracing the unknown: disentangling the complexities of the soil microbiome. Nat Rev Microbiol 15:579–590. https://doi.org/ 10.1038/nrmicro.2017.87.
- Fierer N, Strickland MS, Liptzin D, Bradford MA, Cleveland CC. 2009. Global patterns in belowground communities. Ecol Lett 12:1238–1249. https://doi.org/10.1111/j.1461-0248.2009.01360.x.
- Terrat S, Horrigue W, Dequiedt S, Saby NPA, Lelièvre M, Nowak V, Tripied J, Régnier T, Jolivet C, Arrouays D, Wincker P, Cruaud C, Karimi B, Bispo A, Maron PA, Chemidlin Prévost-Bouré N, Ranjard L. 2017. Mapping and predictive variations of soil bacterial richness across France. PLoS One 12:e0186766. https://doi.org/10.1371/journal.pone .0186766.
- Serna-Chavez HM, Fierer N, Van Bodegom PM. 2013. Global drivers and patterns of microbial abundance in soil. Glob Ecol Biogeogr 22: 1162–1172. https://doi.org/10.1111/geb.12070.
- Delgado-Baquerizo M, Oliverio AM, Brewer TE, Benavent-González A, Eldridge DJ, Bardgett RD, Maestre FT, Singh BK, Fierer N. 2018. A global atlas of the dominant bacteria found in soil. Science 359:320–325. https://doi.org/10.1126/science.aap9516.
- Karimi B, Terrat S, Dequiedt S, Saby NPA, Horrigue W, Lelièvre M, Nowak V, Jolivet C, Arrouays D, Wincker P, Cruaud C, Bispo A, Maron P-A, Bouré NCP, Ranjard L. 2018. Biogeography of soil bacteria and archaea across France. Sci Adv 4:eaat1808. https://doi.org/10.1126/sciadv.aat1808.
- Shade A, Dunn RR, Blowes SA, Keil P, Bohannan BJM, Herrmann M, Küsel K, Lennon JT, Sanders NJ, Storch D, Chase J. 2018. Macroecology to unite all life, large and small. Trends Ecol Evol 33:731–744. https:// doi.org/10.1016/j.tree.2018.08.005.
- Nemergut DR, Schmidt SK, Fukami T, O'Neill SP, Bilinski TM, Stanish LF, Knelman JE, Darcy JL, Lynch RC, Wickey P, Ferrenberg S. 2013. Patterns and processes of microbial community assembly. Microbiol Mol Biol Rev 77:342–356. https://doi.org/10.1128/MMBR.00051-12.
- Zhou J, Ning D. 2017. Stochastic community assembly: does it matter in microbial ecology? Microbiol Mol Biol Rev 81:2–17. https://doi.org/ 10.1128/MMBR.00002-17.
- 12. Hanson CA, Fuhrman JA, Horner-Devine MC, Martiny J. 2012. Beyond

biogeographic patterns: processes shaping the microbial landscape. Nat Rev Microbiol 10:497–506. https://doi.org/10.1038/nrmicro2795.

- Fierer N, Jackson RB. 2006. The diversity and biogeography of soil bacterial communities. Proc Natl Acad Sci U S A 103:626–631. https:// doi.org/10.1073/pnas.0507535103.
- Rousk J, Bååth E, Brookes PC, Lauber CL, Lozupone C, Caporaso JG, Knight R, Fierer N. 2010. Soil bacterial and fungal communities across a pH gradient in an arable soil. ISME J 4:1340–1351. https://doi.org/10 .1038/ismej.2010.58.
- Horner-Devine MC, Lage M, Hughes JB, Bohannan B. 2004. A taxa–area relationship for bacteria. Nature 432:750–754. https://doi.org/10.1038/ nature03073.
- Wang J, Shen J, Wu Y, Tu C, Soininen J, Stegen JC, He J, Liu X, Zhang L, Zhang E. 2013. Phylogenetic beta diversity in bacterial assemblages across ecosystems: deterministic versus stochastic processes. ISME J 7:1310–1321. https://doi.org/10.1038/ismej.2013.30.
- Tripathi BM, Stegen JC, Kim M, Dong K, Adams JM, Lee YK. 2018. Soil pH mediates the balance between stochastic and deterministic assembly of bacteria. ISME J 12:1072–1083. https://doi.org/10.1038/s41396-018 -0082-4.
- Caruso T, Chan Y, Lacap DC, Lau MCY, McKay CP, Pointing SB. 2011. Stochastic and deterministic processes interact in the assembly of desert microbial communities on a global scale. ISME J 5:1406–1413. https://doi.org/10.1038/ismej.2011.21.
- Wang X-B, Lü X-T, Yao J, Wang Z-W, Deng Y, Cheng W-X, Zhou J-Z, Han X-G. 2017. Habitat-specific patterns and drivers of bacterial β-diversity in China's drylands. ISME J 11:1345–1358. https://doi.org/10.1038/ismej .2017.11.
- Evans S, Martiny JBH, Allison SD. 2017. Effects of dispersal and selection on stochastic assembly in microbial communities. ISME J 11:176–185. https://doi.org/10.1038/ismej.2016.96.
- Albright MBN, Martiny J. 2018. Dispersal alters bacterial diversity and composition in a natural community. ISME J 12:296–299. https://doi .org/10.1038/ismej.2017.161.
- Ferrenberg S, O'Neill SP, Knelman JE, Todd B, Duggan S, Bradley D, Robinson T, Schmidt SK, Townsend AR, Williams MW, Cleveland CC, Melbourne BA, Jiang L, Nemergut DR. 2013. Changes in assembly processes in soil bacterial communities following a wildfire disturbance. ISME J 7:1102–1111. https://doi.org/10.1038/ismej.2013.11.
- Dini-Andreote F, Stegen JC, van Elsas JD, Salles JF. 2015. Disentangling mechanisms that mediate the balance between stochastic and deterministic processes in microbial succession. Proc Natl Acad Sci U S A 112:E1326–E1332. https://doi.org/10.1073/pnas.1414261112.
- Delgado-Baquerizo M, Bardgett RD, Vitousek PM, Maestre FT, Williams MA, Eldridge DJ, Lambers H, Neuhauser S, Gallardo A, García-Velázquez L, Sala OE, Abades SR, Alfaro FD, Berhe AA, Bowker MA, Currier CM, Cutler NA, Hart SC, Hayes PE, Hseu Z-Y, Kirchmair M, Peña-Ramírez VM,

Pérez CA, Reed SC, Santos F, Siebe C, Sullivan BW, Weber-Grullon L, Fierer N. 2019. Changes in belowground biodiversity during ecosystem development. Proc Natl Acad Sci U S A 116:6891–6896. https://doi.org/ 10.1073/pnas.1818400116.

- 25. Legendre P. 1993. Spatial autocorrelation: trouble or new paradigm? Ecology 74:1659–1673. https://doi.org/10.2307/1939924.
- Baker KL, Langenheder S, Nicol GW, Ricketts D, Killham K, Campbell CD, Prosser JI. 2009. Environmental and spatial characterisation of bacterial community composition in soil to inform sampling strategies. Soil Biol Biochem 41:2292–2298. https://doi.org/10.1016/j.soilbio.2009.08.010.
- Diniz-Filho JAF, Bini LM, Hawkins BA. 2003. Spatial autocorrelation and red herrings in geographical ecology. Glob Ecol Biogeogr 12:53–64. https://doi.org/10.1046/j.1466-822X.2003.00322.x.
- Meyer KM, Memiaghe H, Korte L, Kenfack D, Alonso A, Bohannan B. 2018. Why do microbes exhibit weak biogeographic patterns? ISME J 12:1404–1413. https://doi.org/10.1038/s41396-018-0103-3.
- Bell T, Ager D, Song JI, Newman JA, Thompson IP, Lilley AK, Van Der Gast CJ. 2005. Larger islands house more bacterial taxa. Science 308: 1884. https://doi.org/10.1126/science.1111318.
- Zhou J, Kang S, Schadt CW, Garten CT. 2008. Spatial scaling of functional gene diversity across various microbial taxa. Proc Natl Acad Sci U S A 105:7768–7773. https://doi.org/10.1073/pnas.0709016105.
- Ranjard L, Dequiedt S, Chemidlin Prévost-Bouré N, Thioulouse J, Saby NPA, Lelievre M, Maron PA, Morin FE, Bispo A, Jolivet C, Arrouays D, Lemanceau P. 2013. Turnover of soil bacterial diversity driven by wide-scale environmental heterogeneity. Nat Commun 4:1434. https:// doi.org/10.1038/ncomms2431.
- Barreto DP, Conrad R, Klose M, Claus P, Enrich-Prast A. 2014. Distancedecay and taxa-area relationships for bacteria, archaea and methanogenic archaea in a tropical lake sediment. PLoS One 9:e110128. https:// doi.org/10.1371/journal.pone.0110128.
- Jiang Y, Liang Y, Li C, Wang F, Sui Y, Suvannang N, Zhou J, Sun B. 2016. Crop rotations alter bacterial and fungal diversity in paddy soils across East Asia. Soil Biol Biochem 95:250–261. https://doi.org/10.1016/j .soilbio.2016.01.007.
- Noguez AM, Arita HT, Escalante AE, Forney LJ, García-Oliva F, Souza V. 2005. Microbial macroecology: highly structured prokaryotic soil assemblages in a tropical deciduous forest. Glob Ecol Biogeogr 14: 241–248. https://doi.org/10.1111/j.1466-822X.2005.00156.x.
- Zinger L, Boetius A, Ramette A. 2014. Bacterial taxa-area and distance-decay relationships in marine environments. Mol Ecol 23: 954–964. https://doi.org/10.1111/mec.12640.
- Bell T. 2010. Experimental tests of the bacterial distance-decay relationship. ISME J 4:1357–1365. https://doi.org/10.1038/ismej.2010.77.
- Reche I, Pulido-Villena E, Morales-Baquero R, Casamayor EO. 2005. Does ecosystem size determine aquatic bacterial richness? Ecology 86: 1715–1722. https://doi.org/10.1890/04-1587.
- Jones SE, Lennon JT. 2010. Dormancy contributes to the maintenance of microbial diversity. Proc Natl Acad Sci U S A 107:5881–5886. https:// doi.org/10.1073/pnas.0912765107.
- 39. Ji M, Greening C, Vanwonterghem I, Carere CR, Bay SK, Steen JA, Montgomery K, Lines T, Beardall J, Van Dorst J, Snape I, Stott MB, Hugenholtz P, Ferrari BC. 2017. Atmospheric trace gases support primary production in Antarctic desert surface soil. Nature 552:400–403. https://doi.org/10.1038/nature25014.
- 40. Greening C, Grinter R, Chiri E. 2019. Uncovering the metabolic strategies of the dormant microbial majority: towards integrative approaches. mSystems 4:e00107-19. https://doi.org/10.1128/mSystems .00107-19.
- Locey K, Muscarella M, Larsen M, Bray S, Jones S, Lennon JT. 2020. Dormancy dampens the microbial distance-decay relationship. Philos Trans R Soc Lond B Biol Sci 375:20190243. https://doi.org/10.1098/rstb .2019.0243.
- Martiny JBH, Eisen JA, Penn K, Allison SD, Horner-Devine MC. 2011. Drivers of bacterial β-diversity depend on spatial scale. Proc Natl Acad Sci U S A 108:7850–7854. https://doi.org/10.1073/pnas.1016308108.
- O'Brien SL, Gibbons SM, Owens SM, Hampton-Marcell J, Johnston ER, Jastrow JD, Gilbert JA, Meyer F, Antonopoulos DA. 2016. Spatial scale drives patterns in soil bacterial diversity. Environ Microbiol 18: 2039–2051. https://doi.org/10.1111/1462-2920.13231.
- Woodcock S, Curtis TP, Head IM, Lunn M, Sloan WT. 2006. Taxa–area relationships for microbes: the unsampled and the unseen. Ecol Lett 9:805–812. https://doi.org/10.1111/j.1461-0248.2006.00929.x.
- 45. Hermans SM, Buckley HL, Lear G. 2019. Perspectives on the impact of



sampling design and intensity on soil microbial diversity estimates. Front Microbiol 10:1820. https://doi.org/10.3389/fmicb.2019.01820.

- Castle SC, Samac DA, Sadowsky MJ, Rosen CJ, Gutknecht JLM, Kinkel LL. 2019. Impacts of sampling design on estimates of microbial community diversity and composition in agricultural soils. Microb Ecol 78:753–763. https://doi.org/10.1007/s00248-019-01318-6.
- Yeh C, Soininen J, Teittinen A, Wang J. 2019. Elevational patterns and hierarchical determinants of biodiversity across microbial taxonomic scales. Mol Ecol 28:86–99. https://doi.org/10.1111/mec.14935.
- Terrat S, Dequiedt S, Horrigue W, Lelievre M, Cruaud C, Saby NPA, Jolivet C, Arrouays D, Maron P-A, Ranjard L, Chemidlin Prévost-Bouré N. 2015. Improving soil bacterial taxa-area relationships assessment using DNA meta-barcoding. Heredity (Edinb) 114:468–475. https://doi.org/ 10.1038/hdy.2014.91.
- Storch D, Šizling AL. 2008. The concept of taxon invariance in ecology: do diversity patterns vary with changes in taxonomic resolution? Folia Geobot 43:329–344. https://doi.org/10.1007/s12224-008-9015-8.
- Bent SJ, Forney LJ. 2008. The tragedy of the uncommon: understanding limitations in the analysis of microbial diversity. ISME J 2:689–695. https://doi.org/10.1038/ismej.2008.44.
- McMurdie PJ, Holmes S. 2014. Waste not, want not: why rarefying microbiome data is inadmissible. PLoS Comput Biol 10:e1003531. https://doi.org/10.1371/journal.pcbi.1003531.
- Nguyen N-P, Warnow T, Pop M, White B. 2016. A perspective on 16S rRNA operational taxonomic unit clustering using sequence similarity. NPJ Biofilms Microbiomes 2:16004. https://doi.org/10.1038/npjbiofilms .2016.4.
- Edgar RC. 2018. Updating the 97% identity threshold for 16S ribosomal RNA OTUs. Bioinformatics 34:2371–2375. https://doi.org/10 .1093/bioinformatics/bty113.
- Beck J, Holloway JD, Schwanghart W. 2013. Undersampling and the measurement of beta diversity. Methods Ecol Evol 4:370–382. https:// doi.org/10.1111/2041-210x.12023.
- Hui C, McGeoch MA. 2014. Zeta diversity as a concept and metric that unifies incidence-based biodiversity patterns. Am Nat 184:684–694. https://doi.org/10.1086/678125.
- Crist TO, Veech JA, Gering JC, Summerville KS. 2003. Partitioning species diversity across landscapes and regions: a hierarchical analysis of alpha, beta, and gamma diversity. Am Nat 162:734–743. https://doi .org/10.1086/378901.
- Wagner HH, Wildi O, Ewald KC. 2000. Additive partitioning of plant species diversity in an agricultural mosaic landscape. Landsc Ecol 15: 219–227. https://doi.org/10.1023/A:1008114117913.
- Nearing JT, Douglas GM, Comeau AM, Langille M. 2018. Denoising the denoisers: an independent evaluation of microbiome sequence errorcorrection approaches. PeerJ 6:e5364. https://doi.org/10.7717/peerj.5364.
- Amir A, McDaniel D, Navas-Molina JA, Kopylova E, Morton JT, Xu ZZ, Kightley EP, Thompson LR, Hyde ER, Gonzalez A, Knight R. 2017. Deblur rapidly resolves single-nucleotide community sequence patterns. mSystems 2:e00191-16. https://doi.org/10.1128/mSystems.00191-16.
- Callahan BJ, McMurdie PJ, Rosen MJ, Han AW, Johnson AJA, Holmes SP. 2016. DADA2: high-resolution sample inference from Illumina amplicon data. Nat Methods 13:581–583. https://doi.org/10.1038/nmeth.3869.
- Latombe G, Hui C, McGeoch MA. 2017. Multi-site generalised dissimilarity modelling: using zeta diversity to differentiate drivers of turnover in rare and widespread species. Methods Ecol Evol 8:431–442. https:// doi.org/10.1111/2041-210X.12756.
- McGeoch MA, Latombe G, Andrew NR, Nakagawa S, Nipperess DA, Roigé M, Marzinelli EM, Campbell AH, Vergés A, Thomas T, Steinberg PD, Selwood KE, Henriksen MV, Hui C. 2019. Measuring continuous compositional change using decline and decay in zeta diversity. Ecology 100:e02832. https://doi.org/10.1002/ecy.2832.
- Hui C, Vermeulen W, Durrheim G. 2018. Quantifying multiple-site compositional turnover in an Afrotemperate forest, using zeta diversity. For Ecosyst 5:15. https://doi.org/10.1186/s40663-018-0135-1.
- McGeoch MA, Gaston KJ. 2002. Occupancy frequency distributions: patterns, artefacts and mechanisms. Biol Rev Camb Philos Soc 77: 311–331. https://doi.org/10.1017/s1464793101005887.
- Morton JT, Toran L, Edlund A, Metcalf JL, Lauber C, Knight R. 2017. Uncovering the horseshoe effect in microbial analyses. mSystems 2:e00166-16. https://doi.org/10.1128/mSystems.00166-16.
- Angel R, Soares MIM, Ungar ED, Gillor O. 2010. Biogeography of soil archaea and bacteria along a steep precipitation gradient. ISME J 4:553–563. https://doi.org/10.1038/ismej.2009.136.



- Johnson RM, Ramond J, Gunnigle E, Seely M, Cowan DA. 2017. Namib Desert edaphic bacterial, fungal and archaeal communities assemble through deterministic processes but are influenced by different abiotic parameters. Extremophiles 21:381–392. https://doi.org/10.1007/s00792 -016-0911-1.
- Neilson JW, Califf K, Cardona C, Copeland A, van Treuren W, Josephson KL, Knight R, Gilbert JA, Quade J, Caporaso JG, Maier RM. 2017. Significant impacts of increasing aridity on the arid soil microbiome. mSystems 2:e00195-16. https://doi.org/10.1128/mSystems.00195-16.
- Ramette A, Tiedje JM. 2007. Multiscale responses of microbial life to spatial distance and environmental heterogeneity in a patchy ecosystem. Proc Natl Acad Sci U S A 104:2761–2766. https://doi.org/10.1073/ pnas.0610671104.
- Yergeau E, Bezemer TM, Hedlund K, Mortimer SR, Kowalchuk GA, Van Der Putten WH. 2010. Influences of space, soil, nematodes and plants on microbial community composition of chalk grassland soils. Environ Microbiol 12:2096–2106. https://doi.org/10.1111/j.1462-2920.2009.02053.x.
- Jiao S, Lu Y. 2020. Soil pH and temperature regulate assembly processes of abundant and rare bacterial communities in agricultural ecosystems. Environ Microbiol 22:1052–1065. https://doi.org/10.1111/ 1462-2920.14815.
- Bankevich A, Pevzner PA. 2018. Joint analysis of long and short reads enables accurate estimates of microbiome complexity. Cell Syst 7:192–200. https://doi.org/10.1016/j.cels.2018.06.009.
- Větrovský T, Baldrian P. 2013. The variability of the 16S rRNA gene in bacterial genomes and its consequences for bacterial community analyses. PLoS One 8:e57923. https://doi.org/10.1371/journal.pone.0057923.
- Scheiner SM, Chiarucci A, Fox GA, Helmus MR, McGlinn DJ, Willig MR. 2011. The underpinnings of the relationship of species richness with space and time. Ecol Monogr 81:195–213. https://doi.org/10.1890/10 -1426.1.
- Oono R, Rasmussen A, Lefèvre E. 2017. Distance decay relationships in foliar fungal endophytes are driven by rare taxa. Environ Microbiol 19:2794–2805. https://doi.org/10.1111/1462-2920.13799.
- Nekola JC, White PS. 1999. The distance decay of similarity in biogeography and ecology. J Biogeogr 26:867–878. https://doi.org/10.1046/ j.1365-2699.1999.00305.x.
- Liu L, Yang J, Yu Z, Wilkinson DM. 2015. The biogeography of abundant and rare bacterioplankton in the lakes and reservoirs of China. ISME J 9:2068–2077. https://doi.org/10.1038/ismej.2015.29.
- Logares R, Lindström ES, Langenheder S, Logue JB, Paterson H, Laybourn-Parry J, Rengefors K, Tranvik L, Bertilsson S. 2013. Biogeography of bacterial communities exposed to progressive long-term environmental change. ISME J 7:937–948. https://doi.org/10.1038/ismej.2012.168.
- Székely AJ, Langenheder S. 2014. The importance of species sorting differs between habitat generalists and specialists in bacterial communities. FEMS Microbiol Ecol 87:102–112. https://doi.org/10.1111/1574 -6941.12195.
- Chen Y-J, Leung PM, Bay SK, Hugenholtz P, Kessler AJ, Shelley G, Waite DW, Cook PLM, Greening C. 2020. Metabolic flexibility allows generalist bacteria to become dominant in a frequently disturbed ecosystem. bioRxiv. https://doi.org/10.1101/945220.
- Latombe G, Richardson DM, Pyšek P, Kučera T, Hui C. 2018. Drivers of species turnover vary with species commonness for native and alien plants with different residence times. Ecology 99:2763–2775. https:// doi.org/10.1002/ecy.2528.
- Johnson JS, Spakowicz DJ, Hong BY, Petersen LM, Demkowicz P, Chen L, Leopold SR, Hanson BM, Agresta HO, Gerstein M, Sodergren E, Weinstock GM. 2019. Evaluation of 16S rRNA gene sequencing for species and strain-level microbiome analysis. Nat Commun 10:5029. https://doi.org/10.1038/s41467-019-13036-1.
- Karst SM, Dueholm MS, Mcllroy SJ, Kirkegaard RH, Nielsen PH, Albertsen M. 2018. Retrieval of a million high-quality, full-length microbial 16S and 18S rRNA gene sequences without primer bias. Nat Biotechnol 36:190–195. https://doi.org/10.1038/nbt.4045.
- Šťovíček A, Kim M, Or D, Gillor O. 2017. Microbial community response to hydration-desiccation cycles in desert soil. Sci Rep 7:45735. https:// doi.org/10.1038/srep45735.

- Rayment GE, Lyons DJ. 2011. Soil chemical methods: Australasia. CSIRO Publishing, Collinwood, Australia.
- Trabucco A, Zomer RJ. 2009. Global aridity index (global-aridity) and global potential evapo-transpiration (global-PET) geospatial database. CGIAR CSI Consortium for Spatial Information.
- Paulin MM, Nicolaisen MH, Jacobsen CS, Gimsing AL, Sørensen J, Bælum J. 2013. Improving Griffith's protocol for co-extraction of microbial DNA and RNA in adsorptive soils. Soil Biol Biochem 63:37–49. https://doi.org/10.1016/j.soilbio.2013.02.007.
- Marotz C, Sharma A, Humphrey G, Gottel N, Daum C, Gilbert JA, Eloe-Fadrosh E, Knight R. 2019. Triplicate PCR reactions for 16S rRNA gene amplicon sequencing are unnecessary. Biotechniques 67:29–32. https://doi.org/10.2144/btn-2018-0192.
- Gilbert JA, Jansson JK, Knight R. 2014. The Earth Microbiome project: successes and aspirations. BMC Biol 12:69. https://doi.org/10.1186/ s12915-014-0069-1.
- 90. Bolyen E, Rideout JR, Dillon MR, Bokulich NA, Abnet CC, Al-Ghalith GA, Alexander H, Alm EJ, Arumugam M, Asnicar F, Bai Y, Bisanz JE, Bittinger K, Brejnrod A, Brislawn CJ, Brown CT, Callahan BJ, Caraballo-Rodríguez AM, Chase J, Cope EK, Da Silva R, Diener C, Dorrestein PC, Douglas GM, Durall DM, Duvallet C, Edwardson CF, Ernst M, Estaki M, Fouquier J, Gauglitz JM, Gibbons SM, Gibson DL, Gonzalez A, Gorlick K, Guo J, Hillmann B, Holmes S, Holste H, Huttenhower C, Huttley GA, Janssen S, Jarmusch AK, Jiang L, Kaehler BD, Kang KB, Keefe CR, Keim P, Kelley ST, Knights D, Koester I, et al. 2019. Reproducible, interactive, scalable and extensible microbiome data science using QIIME 2. Nat Biotechnol 37:852–857. https://doi.org/10.1038/s41587-019-0209-9.
- Rognes T, Flouri T, Nichols B, Quince C, Mahé F. 2016. VSEARCH: a versatile open source tool for metagenomics. PeerJ 4:e2584. https:// doi.org/10.7717/peerj.2584.
- Parks DH, Chuvochina M, Waite DW, Rinke C, Skarshewski A, Chaumeil P-A, Hugenholtz P. 2018. A standardized bacterial taxonomy based on genome phylogeny substantially revises the tree of life. Nat Biotechnol 36:996–1004. https://doi.org/10.1038/nbt.4229.
- Pedregosa F, Michel V, Grisel O, Blondel M, Prettenhofer P, Weiss R, Vanderplas J, Cournapeau D, Pedregosa F, Varoquaux G, Gramfort A, Thirion B, Grisel O, Dubourg V, Passos A, Brucher M, Perrot M, Duchesnay É. 2011. Scikit-learn: machine learning in Python. J Mach Learn Res 12:2825–2830.
- Katoh K, Standley DM. 2013. MAFFT multiple sequence alignment software version 7: improvements in performance and usability. Mol Biol Evol 30:772–780. https://doi.org/10.1093/molbev/mst010.
- Woodcroft BJ. 2019. SingleM. GitHub, Inc. https://github.com/wwood/ singlem.
- Boyd JA, Woodcroft BJ, Tyson GW. 2018. GraftM: a tool for scalable, phylogenetically informed classification of genes within metagenomes. Nucleic Acids Res 46:e59. https://doi.org/10.1093/nar/gky174.
- Lan Y, Rosen G, Hershberg R. 2016. Marker genes that are less conserved in their sequences are useful for predicting genome-wide similarity levels between closely related prokaryotic strains. Microbiome 4:18. https://doi.org/10.1186/s40168-016-0162-5.
- Wickham H. 2016. ggplot2: elegant graphics for data analysis. Springer-Verlag, New York, NY.
- McMurdie PJ, Holmes S. 2013. phyloseq: an R package for reproducible interactive analysis and graphics of microbiome census data. PLoS One 8:e61217. https://doi.org/10.1371/journal.pone.0061217.
- Oksanen J, Blanchet FG, Kindt R, Legendre P, Minchin PR, O'Hara RB, Simpson GL, Solymos P, Stevens MHH, Wagner H, Oksanen MJ. 2018. vegan: community ecology package. R package version 2.4-6.
- Drakare S, Lennon JJ, Hillebrand H. 2006. The imprint of the geographical, evolutionary and ecological context on species-area relationships. Ecol Lett 9:215–227. https://doi.org/10.1111/j.1461-0248.2005.00848.x.
- Wang Y, Naumann U, Wright ST, Warton DI. 2012. mvabund–an R package for model-based analysis of multivariate abundance data. Methods Ecol Evol 3:471–474. https://doi.org/10.1111/j.2041-210X.2012.00190.x.
- 103. Warton DI, Thibaut L, Wang YA. 2017. The PIT-trap—a "model-free" bootstrap procedure for inference about regression models with discrete, multivariate responses. PLoS One 12:e0181790. https://doi.org/ 10.1371/journal.pone.0181790.