


 Cite this: *RSC Adv.*, 2020, 10, 228

# The combined effect of light irradiation and chloride on the physicochemical properties of silver nanoparticles†

 Bojie Yuan, Minghao Sui, \* Hongtao Lu, Jingyu Wang  and Jie Qin

Understanding the effect of various environmental factors on the transformation of silver nanoparticles (Ag NPs) is crucial to determine their toxicity and fate in the environment. Ag NPs are inevitably exposed to sunlight and will be in contact with chloride ( $\text{Cl}^-$ ), a ubiquitous ligand in natural water, once released into the environment. In this study, the combined effect of  $\text{Cl}^-$  and light on various physicochemical properties (optical property, dissolution, morphology, surface charge) of different sizes Ag NPs was studied. The results showed that light irradiation, in the presence of  $\text{Cl}^-$ , led to a great decrease in the concentration of dissolved Ag and a remarkable increase in the zeta potential of Ag NPs, as well as the generation of some tiny Ag NPs and fusion aggregates. AgCl was suggested to rapidly coat onto Ag NPs after exposure to  $\text{Cl}^-$ . And the AgCl layer was obviously destroyed by photoreduction under light irradiation. Meanwhile, the Ag NP size exhibited a great impact on the destruction of the AgCl layer. It was further observed that the AgCl layer gradually re-formed when the light was removed, which suggested that Ag NPs might present different states during the daytime and at night in aquatic environments.

 Received 8th November 2019  
 Accepted 10th December 2019

DOI: 10.1039/c9ra09261g

[rsc.li/rsc-advances](http://rsc.li/rsc-advances)

## Introduction

Nanomaterials have been widely employed in medical, commercial and industrial products in the past several decades due to their unique chemical and physical properties.<sup>1–5</sup> Among these nanomaterials, silver nanoparticles (Ag NPs) with outstanding optical, catalytic, electrical, and antibacterial properties are currently one of the most widely used engineered nanoparticles. As a result, Ag NPs will be inevitably released from the related products of Ag NPs during production, use and washing and eventually introduced into the environment.<sup>6</sup> Ag NPs have been reported to be highly toxic to a variety of model organisms by many studies.<sup>7–12</sup> Hence, there is a growing concern about the potential risk of released Ag NPs to human and ecological health. Therefore, it is crucial to understand the transformation and fate of Ag NPs in the environment.

After releasing into the environment, Ag NPs are inevitably exposed to sunlight. It was reported that light irradiation had

a strong effect on various properties of Ag NPs. Under light irradiation, the dissolution of Ag NPs was obviously accelerated.<sup>13</sup> Furthermore, the dissolution rate was found to be affected by irradiation conditions (UV-365 > xenon lamp > UV-254) and surface coating (bare-Ag NPs > PVP-Ag NPs > citrate-Ag NPs).<sup>14</sup> In addition, exposing Ag NPs to light irradiation also induced the destabilization of Ag NPs and promoted formation of aggregates.<sup>15,16</sup>

The behavior and fate of Ag NPs in aquatic environment are also sensitive to environment conditions. In the presence of natural organic materials (NOMs), the stability of Ag NPs was enhanced due to the electrostatic and steric repulsion caused by the adsorption of NOMs on the surface of Ag NPs.<sup>17–20</sup> Owing to the photoreduction of NOMs,  $\text{Ag}^+$  would be reduced to  $\text{Ag}^0$ , which also greatly suppressed the dissolution of Ag NPs.<sup>17–19,21</sup>  $\text{Cl}^-$ , a ligand with strong affinity for oxidized silver, ubiquitously presents in natural water and Ag NPs are inevitably in contact with  $\text{Cl}^-$  in aquatic environment. It was reported that  $\text{Cl}^-$  could significantly affect the dissolution of Ag NPs, and the kinetics of dissolution were closely related to the ratio of  $\text{Cl}/\text{Ag}$ .<sup>22</sup> At low  $\text{Cl}/\text{Ag}$  ratio,  $\text{Cl}^-$  reacted with dissolved silver and AgCl coating was generated on Ag NPs.<sup>22,23</sup> The formation of AgCl coating, a passivating layer, would remarkably suppress the dissolution rate of Ag NPs.<sup>24</sup> However, the dissolution rate of Ag NPs was obviously enhanced at high  $\text{Cl}/\text{Ag}$  ratio owing to the generation of soluble  $\text{AgCl}_x^{(x-1)-}$  species. The increase in the amount of dissolved silver also led to increased toxicity.<sup>22</sup>

Shanghai Institute of Pollution Control and Ecological Security, State Key Laboratory of Pollution Control and Resource Reuse, School of Environmental Science and Engineering, Tongji University, 1239 Siping Road, Shanghai 200092, People's Republic of China. E-mail: minghaosui@tongji.edu.cn

† Electronic supplementary information (ESI) available: UV-vis spectra of Ag NPs during dark treatment in 0.5 mM NaCl, UV-vis spectra of Ag NPs during light treatment in ultrapure water and 0.5 mM  $\text{NaNO}_3$ , XRD patterns of Ag NPs, TEM images of Ag NPs after treating with 0.5 mM NaCl for 9 h in dark conditions and the residual rate of Ag NPs absorbance with time after introducing  $\text{H}_2\text{O}_2$  (dark control). See DOI: 10.1039/c9ra09261g



Additionally, the intrinsic properties of Ag NPs, like particle size, also have significant influence on its behavior. Particles size was found to have an inverse effect on Ag NPs dissolution: small Ag NPs exhibited higher solubility than large ones.<sup>25,26</sup> Small Ag NPs with higher surface-to-volume ratio have higher surface energy and reactivity than large particles. Ivanova *et al.* have demonstrated that the standard redox potential of metal nanoparticles decreased with decreasing size.<sup>27</sup>

Many studies have focused on the fate and behavior of Ag NPs during light irradiation,<sup>17–19,28</sup> however, little is known about the combined effect of light irradiation and  $\text{Cl}^-$ , as well as particle size, on various physicochemical properties of Ag NPs. Here, polyvinyl pyrrolidone (PVP) coated Ag NPs with different size (20, 40 and 57 nm) were synthesized. And then, we investigated the evolution of physicochemical properties (optical property, dissolution, zeta potential and morphology) of three sizes Ag NPs after exposing to  $\text{Cl}^-$  solutions under light irradiation, and this study was expected to provide more information on the behavior and fate of Ag NPs in natural water.

## Materials and methods

### Materials

Tannic acid (TA), poly(vinylpyrrolidone) (PVP, K29-32,  $M_w = 58\ 000$ ), trisodium citrate ( $\text{Na}_3\text{C}_6\text{H}_5\text{O}_7 \cdot 5\text{H}_2\text{O}$ ), hydrogen peroxide ( $\text{H}_2\text{O}_2$ ), sodium chloride (NaCl) and sodium nitrate ( $\text{NaNO}_3$ ) were purchased from Aladdin Chemistry Co. Ltd. Silver nitrate ( $\text{AgNO}_3$ ) was purchased from Sigma-Aldrich. Ultrapure water (Aqualore 2S, USA,  $18.2\ \text{M}\Omega\ \text{cm}$ ) was used to prepare all solutions.

### Synthesis and characterization of Ag NPs

According to the method proposed by Bastus *et al.*,<sup>29</sup> PVP coated Ag NPs with different sizes were synthesized. And then, Ag NPs were dispersed in ultrapure water and stored in dark conditions at  $4\ ^\circ\text{C}$  for later use. The total concentration of Ag NPs stock suspension was measured by inductively coupled plasma mass spectrometry (ICP-MS, Agilent 7700). Samples were digested with  $\text{HNO}_3$  and  $\text{H}_2\text{O}_2$  before ICP-MS analysis. Transmission electron microscopy (TEM, JEM2011, Jeol) was used to characterize the morphology of synthesized Ag NPs. TEM size distributions were obtained from more than 200 particles of each Ag NPs type. The UV-vis spectra from 300–700 nm were collected by using a UV-3600 spectrophotometer. Hydrodynamic diameter and the polydispersity index (PDI) of Ag NPs, as well as zeta potential, were determined by a zeta potential analyzer (Nano Z, Malvern).

### The light irradiation experiment

The experiments were carried out in a solar simulator equipped with a 500 W Xe lamp to simulate the sunlight. The reaction vessels were 50 mL quartz tubes and rotated around the light source to ensure each sample obtain similar levels of light exposure. Meanwhile, the incident light would pass through the cooling water layer to remove infrared rays before reaching the reaction vessels. The reaction vessels filled with 30 mL Ag NPs

suspensions ( $6.4\ \text{mg}\ \text{L}^{-1}$ ,  $59.3\ \mu\text{M}$ ) and accompanied with magnetic stirring. And then, 0.5 mM of NaCl was introduced to the reaction solutions. The dark control was conducted in the same simulator without light irradiation. At selected time point, samples were taken from the reaction tubes to determine the UV-vis spectra, concentration of dissolved Ag ( $\text{Ag}_{(\text{dis})}$ ) and zeta potential of particles. In addition, the  $\text{H}_2\text{O}_2$  mediated oxidation experiments were conducted after the characterization of UV-vis spectra.  $\text{H}_2\text{O}_2$  ( $200\ \text{mg}\ \text{L}^{-1}$ ) was introduced into the sample, and then the change of absorbance over time at the maximum absorption wavelength was monitored immediately.  $\text{Ag}_{(\text{dis})}$  was separated from Ag NPs by Amicon centrifugal ultrafilters (Amicon Ultra-15 3K, Millipore, Billerica, MA). Samples were centrifuged at 15 000 rpm for 10 min in centrifugal ultrafilters. All Ag in collected solutions was defined as  $\text{Ag}_{(\text{dis})}$  in this study and the concentration of  $\text{Ag}_{(\text{dis})}$  was measured by ICP-MS after acidifying with  $\text{HNO}_3$ . It was worth mentioning that some tiny Ag NPs with a size smaller than the pore size of the filter membrane might be generated during light irradiation. Hence,  $\text{Ag}_{(\text{dis})}$  in this study included mainly free  $\text{Ag}^+$  and some tiny Ag NPs that could pass the filter membrane. After 9 h of light and dark treatment, samples were also characterized by TEM to monitor the change of morphology of Ag NPs. X-ray diffraction (XRD) of Ag NPs before and after being incubated with NaCl in dark and light conditions was collected by an X-ray diffractometer (D-8 Advance, Bruker).

## Results and discussion

### Characterization of Ag NPs

TEM images of prepared three sizes Ag NPs were showed in Fig. 1. All of the prepared Ag NPs were in quasi-spherical shape and presented monodispersed. TEM size distributions of synthesized Ag NPs were statistically obtained, which ranged in diameters from 13 to 25 nm, 34 to 45 nm, and 51 to 61 nm, respectively. And the average diameters were  $20 \pm 6\ \text{nm}$  (nAg-20),  $40 \pm 6\ \text{nm}$  (nAg-40) and  $57 \pm 5\ \text{nm}$  (nAg-57). *In situ* DLS showed hydrodynamic diameters of 43.6 nm, 61.2 nm and 82.2 nm with PDI of 0.184, 0.135 and 0.088, respectively, suggesting that the three sizes of Ag NPs were well monodispersed without significant aggregation. Compared with TEM size, the larger hydrodynamic diameter of Ag NPs is probably due to the presence of some aggregates and PVP coating which decrease the diffusivity of particles.<sup>25</sup> The UV-vis measurements revealed that the surface plasmon resonance (SPR) peaks of synthesized Ag NPs were single peak with sharp shape, which also demonstrated that the products were well size-distributed and highly monodispersed.<sup>30</sup> In addition, the absorbance peaks of Ag NPs were 401, 416, and 435 nm for nAg-20, nAg-40, and nAg-57, respectively. The position of extinction peaks presented red-shifting as particle size increased, which was in consistence with previous studies.<sup>26</sup> The zeta potential of prepared Ag NPs was  $-7.05\ \text{mV}$ ,  $-8.37\ \text{mV}$  and  $-10.4\ \text{mV}$  for nAg-20, nAg-40, and nAg-57, respectively. Similar zeta potential values have been also reported in many studies on PVP-coated Ag NPs.<sup>22,31–33</sup> The negative charge of PVP-coated Ag NPs might be attributed to the adsorption of some side products on the surface of Ag NPs

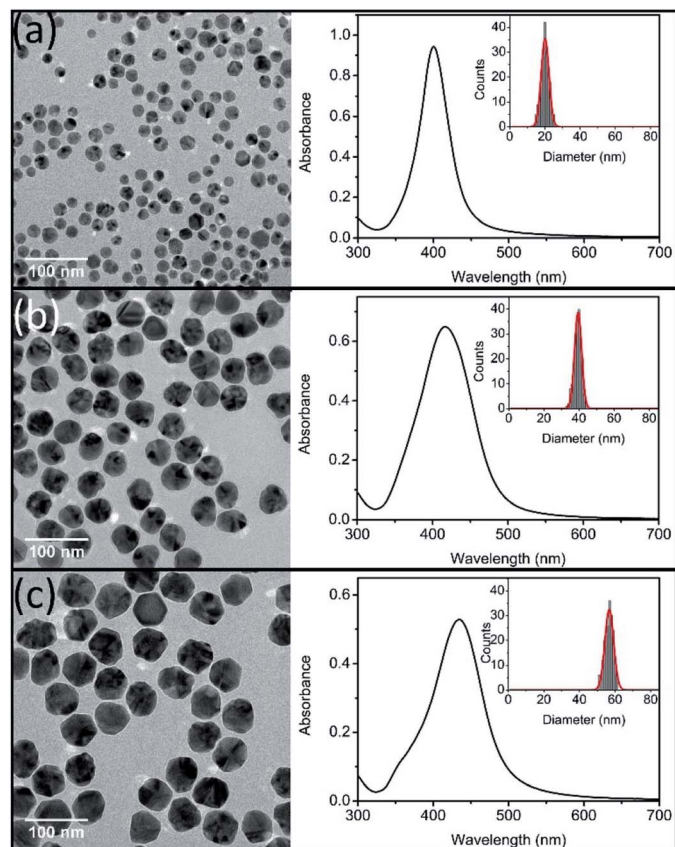


Fig. 1 TEM images (left), UV-vis spectra (right) and size distributions (inset) of synthesized Ag NPs (nAg-20 (a), nAg-40 (b), and nAg-57 (c)).

during the synthesis of Ag NPs.<sup>34</sup> The surface area of single Ag NPs increases with the increase of particle size. Thus, more side products might adsorb on the surface of large Ag NPs, resulting in the zeta potential of large Ag NPs become more negative than small one.

### Effect on UV-vis spectra of Ag NPs

The evolution of UV-vis spectra of Ag NPs was monitored during 9 h of treatment in 0.5 mM NaCl. As illustrated in Fig. S1† in dark conditions, a slight decline was observed in the absorption

peak intensity of the three sizes of Ag NPs. It has been reported that the addition of electrolytes (such as NaCl) could promote the dissolution of Ag NPs.<sup>23,35</sup> In the presence of electrolytes, Ag<sup>+</sup> absorbed on Ag NPs would be redistributed and/or replaced by the electrolyte ions and the dissolution of surface oxide layer was promoted.<sup>23</sup> This process also exposed the internal metallic Ag, leading to the enhancement of further dissolution of Ag NPs,<sup>35</sup> which should be responsible for the slight decrease of absorption peak (Fig. S1†). However, under light irradiation, it was interesting to observe that the absorption peak of Ag NPs didn't decline in the presence of 0.5 mM NaCl for nAg-40 and nAg-57, and even the absorption peak of nAg-20 presented a slight increase (Fig. 2). To further investigate the reason for the evolution of absorption peak during light treatment, the experiments were conducted in ultrapure water (to check the role of light) and 0.5 mM NaNO<sub>3</sub> (to check the role of Na<sup>+</sup>). The results showed that the absorption peak of Ag NPs presented obvious decline after exposing to light, especially in the presence of NaNO<sub>3</sub> (Fig. S2†). Therefore, the abnormal evolution of Ag NPs absorption peak during light irradiation in 0.5 mM NaCl should be attributed to the presence of Cl<sup>-</sup>.

### Effect on Ag<sub>(dis)</sub> concentration

Fig. 3 showed the evolution of Ag<sub>(dis)</sub> concentration for Ag NPs suspensions in 0.5 mM Cl<sup>-</sup> solution during 9 h of dark and light treatment. It should first notice that more dissolved Ag presented in pristine Ag NPs suspensions with the decreasing of Ag NPs size. At the beginning, 315.2 μg L<sup>-1</sup> of Ag<sub>(dis)</sub> was detected for nAg-20, while the concentration of Ag<sub>(dis)</sub> was 191.1 μg L<sup>-1</sup> and 132.8 μg L<sup>-1</sup> for nAg-40 and nAg-57. From Fig. 3, obviously, the concentration of Ag<sub>(dis)</sub> exhibited different changing trend in dark and light conditions. During dark treatment, obvious decrease in the concentration of Ag<sub>(dis)</sub> was observed. Many studies have suggested that the introduction of Cl<sup>-</sup> to Ag NPs suspensions would lead to the generation of AgCl coating on Ag NPs.<sup>22,32</sup> In our study, XRD presented in Fig. S3† confirmed the generation of AgCl as evidenced by the appearing of diagnostic AgCl diffraction peaks after incubating with NaCl. Therefore, the decrease of Ag<sub>(dis)</sub> concentration should be resulted from the generation of AgCl. The introduction of NaCl was also believed to enhance the dissolution of Ag NPs. However, no

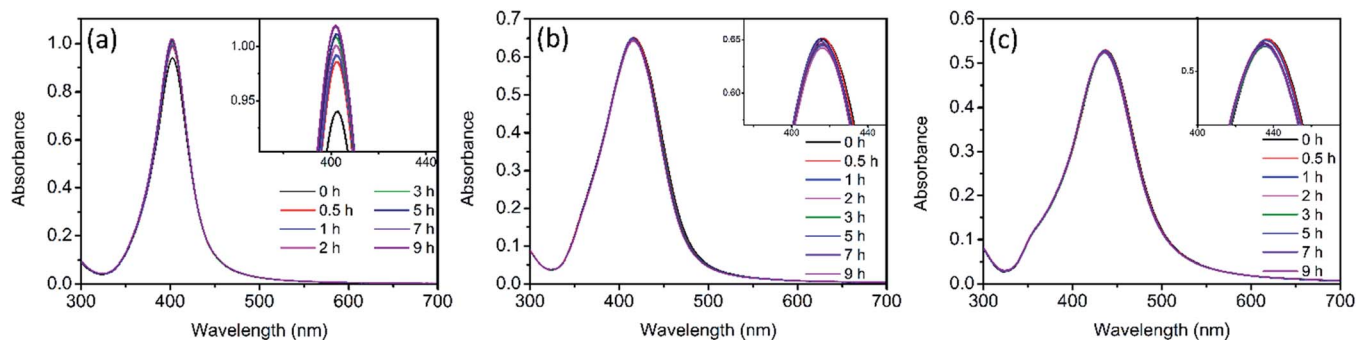


Fig. 2 UV-vis spectra of Ag NPs (nAg-20 (a), nAg-40 (b), nAg-57 (c)) over time during light irradiation in 0.5 mM Cl<sup>-</sup>.

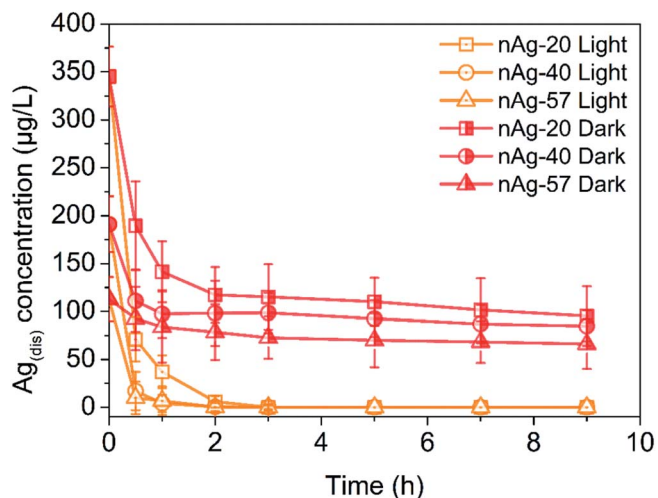


Fig. 3  $\text{Ag}_{(\text{dis})}$  concentration of different size Ag NPs over time during light and dark treatment in  $0.5 \text{ mM Cl}^-$ .

increase in  $\text{Ag}_{(\text{dis})}$  concentration was observed, which might be ascribed to the strong affinity between  $\text{Cl}^-$  and  $\text{Ag}^+$ . The amount of  $\text{Ag}^+$  consumed by the formation of  $\text{AgCl}$  was much greater than that released from Ag NPs. Therefore, the concentration of  $\text{Ag}_{(\text{dis})}$  decreased after the addition of  $\text{NaCl}$ . In addition, the generation of a passivating layer ( $\text{AgCl}$  coating) on Ag NPs could remarkably inhibit the dissolution of Ag NPs.<sup>24</sup> Hence, the decline in absorption peak of Ag NPs slowed down as the time of contact with  $\text{NaCl}$  increased (Fig. S1†).

Under light irradiation, the evolution of  $\text{Ag}_{(\text{dis})}$  concentration was different from that in dark conditions. During light treatment, the concentration of  $\text{Ag}_{(\text{dis})}$  remarkably decreased and almost no  $\text{Ag}_{(\text{dis})}$  could be detected after only 3 h light treatment. Yin *et al.* have reported that the co-occurrence of  $\text{Cl}^-$  could enhance the photo-conversion of  $\text{Ag}^+$  into Ag NPs owing to the formation of  $\text{AgCl}$ .<sup>36</sup>  $\text{AgCl}$ , as a common photocatalyst, can be excited under UV irradiation and the photogenerated

electrons can reduce the  $\text{Ag}^+$  of the  $\text{AgCl}$  lattice to  $\text{Ag}^0$ . This process has been demonstrated in many studies.<sup>37–41</sup> Therefore, we speculated that the great decreasing of  $\text{Ag}_{(\text{dis})}$  concentration after exposing to simulated sunlight irradiation might be attributed to the  $\text{Cl}^-$ -assisted photoreduction of free  $\text{Ag}^+$ . After 9 h of light irradiation, much weaker  $\text{AgCl}$  XRD patterns than that under dark were observed (Fig. S3†), which was consistent with our speculation. Plenty of studies have reported that dissolved Ag (commonly regarded as the source of Ag NPs toxicity) exhibited highly toxicity to many aquatic organisms.<sup>11,22,42</sup> Hence, in aquatic environment, the toxicity of Ag NPs is expected to be diminished to a certain degree.

#### Effect on the morphology of Ag NPs

TEM images of Ag NPs in  $0.5 \text{ mM Cl}^-$  electrolyte solutions were obtained after 9 h of treatment in dark and light irradiation. In dark control, no obvious change on Ag NPs was observed (Fig. S4†). However, after exposing to light irradiation, many tiny nanoparticles were found in TEM images (Fig. 4). On the basis of discussion above, these tiny nanoparticles were believed to be newly produced Ag NPs from the  $\text{Cl}^-$ -assisted photoreduction of  $\text{Ag}^+$ . Since much  $\text{Ag}^+$  presented in pristine nAg-20 suspensions (Fig. 3), plenty of tiny Ag NPs were formed in nAg-20 after light treatment, which should also be responsible for the slight increase of absorption peak of nAg-20 (Fig. 2a). Some tiny Ag NPs were also generated in nAg-40 and nAg-57, but no increase in the absorption peak was observed (Fig. 2b and c). We speculated that this might be attributed to the large difference between the size of re-formed tiny Ag NPs and pristine Ag NPs in nAg-40 and nAg-57. Hence, the absorbance of these tiny Ag NPs had little contribution to the absorption peak of nAg-40 and nAg-57. From TEM images, we also found that some Ag NPs adhered to each other and formed irregularly shaped fusion aggregates. The photoreduction of  $\text{Ag}^+$  to  $\text{Ag}^0$  under light irradiation might generate nanobridges to connect nearby particles together and led to the formation of fusion aggregates. Much more fusion aggregates were observed in nAg-20, which should be resulted from

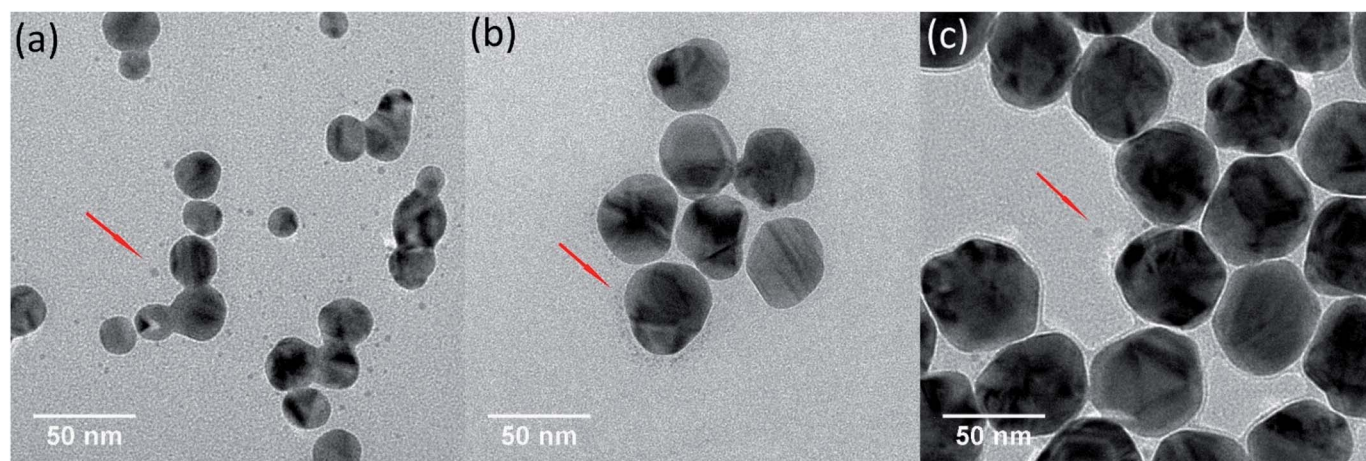


Fig. 4 TEM images of Ag NPs (nAg-20 (a), nAg-40 (b), and nAg-57 (c)) after 9 h of light irradiation in  $0.5 \text{ mM Cl}^-$ .

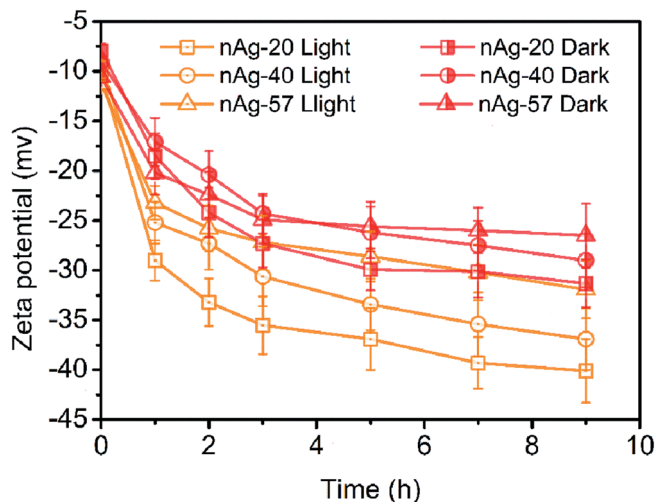


Fig. 5 Zeta potential of Ag NPs over time during light and dark treatment in 0.5 mM  $\text{Cl}^-$ .

the presence of much higher  $\text{Ag}_{(\text{dis})}$  concentration and particles density in nAg-20.

#### Effect on the surface charge of Ag NPs

Zeta potential measurements for Ag NPs in  $\text{Cl}^-$  electrolyte solutions during 9 h of dark and light treatment were measured. As displayed in Fig. 5, zeta potential of Ag NPs presented increasing (more negative) in the dark conditions. Since the concentration of  $\text{Cl}^-$  (0.5 mM) was much higher than the concentration of  $\text{Ag}^+$  (the concentration of total Ag in this study was 59.3  $\mu\text{M}$ ), AgCl was most likely terminated by  $\text{Cl}^-$ , resulting in negative charge of AgCl. With the formation of AgCl coating on the surface of Ag NPs, zeta potential of Ag NPs increased. Similar results have also been reported in many studies.<sup>22,32,43</sup> Compare to dark conditions, more increase in the zeta potential of Ag NPs was observed under light irradiation (Fig. 5). Two possible reasons might be responsible for the remarkable increase of zeta potential during light treatment: (1) light irradiation promoted the formation of AgCl on the surface of Ag NPs. The dissolution of Ag NPs has been reported to be obviously accelerated under light.<sup>13</sup>  $\text{Ag}^+$  was captured by  $\text{Cl}^-$  once released from Ag NPs, promoting the formation of AgCl coating on Ag NPs. However, as discussed above, AgCl might also be photoreduced to  $\text{Ag}^0$  under sunlight and the weak AgCl diffraction peak was observed in XRD patterns after light irradiation (Fig. S3†). It suggested that just a little AgCl existed. Hence, promoting the formation of AgCl layer under light seemed not to be the reason for the remarkable increasing of zeta potential. (2) Light irradiation promoted the adsorption of  $\text{Cl}^-$  on Ag NPs.  $\text{Cl}^-$ , as a nucleophilic reagent, could donate a pair of electrons to the unoccupied orbital of Ag and adsorb on the surface of Ag NPs. The adsorption of  $\text{Cl}^-$  on Ag NPs might be promoted under light irradiation, thereby leading to greatly increasing in zeta potential.

The increase in zeta potential of Ag NPs could enhance the stability of Ag NPs due to the extra electrostatic repulsion.

Although some fusion aggregates were observed after light irradiation in our study, the possibility of the formation of fusion aggregates in real aquatic environment was significantly reduced due to much lower concentration of Ag NPs presented in real aquatic environment than that used in this study. Thus, it is expected that Ag NPs may be dispersed as a single particle and suspended in natural water for a long time, displaying a persistent potential risk to aquatic vertebrates and invertebrates.

#### Study of AgCl on the surface of Ag NPs

In the presence of  $\text{Cl}^-$ , light irradiation was found to exhibit a remarkable influence on various physicochemical properties of Ag NPs. In dark conditions, AgCl layer was believed to generate on the surface of Ag NPs after the addition of  $\text{Cl}^-$ . However, the change on the surface of Ag NPs during light treatment was still unclear. The presence of AgCl on the surface of Ag NPs required further investigation. As a passivating layer, the AgCl coating could efficiently block the contact of inner Ag NPs with oxidants, thereby preventing Ag NPs from further oxidation.<sup>24</sup> Here, we used  $\text{H}_2\text{O}_2$  mediated oxidation experiments to verify the presence of AgCl layer. After introducing  $\text{H}_2\text{O}_2$ , a strong oxidant, the change of absorbance at the maximum absorption wavelength was recorded, and the result was shown in Fig. 6. The blank control (pristine Ag NPs) was conducted in 0.5 mM  $\text{NaNO}_3$  to ensure the same ionic strength, and obvious decrease in the absorbance was observed. It should be noted that oxidation of a thin surface layer of Ag NPs by  $\text{H}_2\text{O}_2$  might result in the deterioration of the free electron density on the surface, thereby strong dampening of SPR. The major part of Ag NPs remains unoxidized but no absorbance can be observed. Thus, the decrease of absorbance might be attributed to the formation of “white” Ag NPs rather than the dissolution of Ag NPs. To verify whether Ag NPs was dissolved by  $\text{H}_2\text{O}_2$  in our system, the size and morphology of nAg-57 after being treated with  $\text{H}_2\text{O}_2$  was investigated by TEM (Fig. S5†). The size of residual particles significantly decreased and the morphology of particles was greatly damaged, indicating that the actual oxidation of Ag NPs occurred. And the concentration of dissolved Ag increased from 153.6  $\mu\text{g L}^{-1}$  to 2694.5  $\mu\text{g L}^{-1}$  after reacting with  $\text{H}_2\text{O}_2$  for 180 s. Hence, the decreasing of absorbance should be resulted from actual dissolution of Ag NPs. Therefore, the dissolution of Ag NPs could be qualitatively reflected by the decreasing of absorbance, and He *et al.* also used this method to characterize the decay of Ag NPs after the introduction of  $\text{H}_2\text{O}_2$ .<sup>44</sup> In dark control, as shown in Fig. S6,† three sizes of Ag NPs presented remarkable resistance to the oxidation of  $\text{H}_2\text{O}_2$  after the addition of  $\text{Cl}^-$  and just a little decreasing in absorbance was observed, which further confirmed the formation of AgCl layer on the surface of Ag NPs. Moreover, during 9 h incubation of Ag NPs with  $\text{Cl}^-$  in dark, samples, taken at various time points, exhibited almost the same decreasing trend in absorbance after introducing  $\text{H}_2\text{O}_2$ . The sample at 0 h was a sample taken immediately after the introduction of  $\text{Cl}^-$  to Ag NPs suspensions. Therefore, the AgCl layer was supposed to be formed on the surface of Ag NPs as

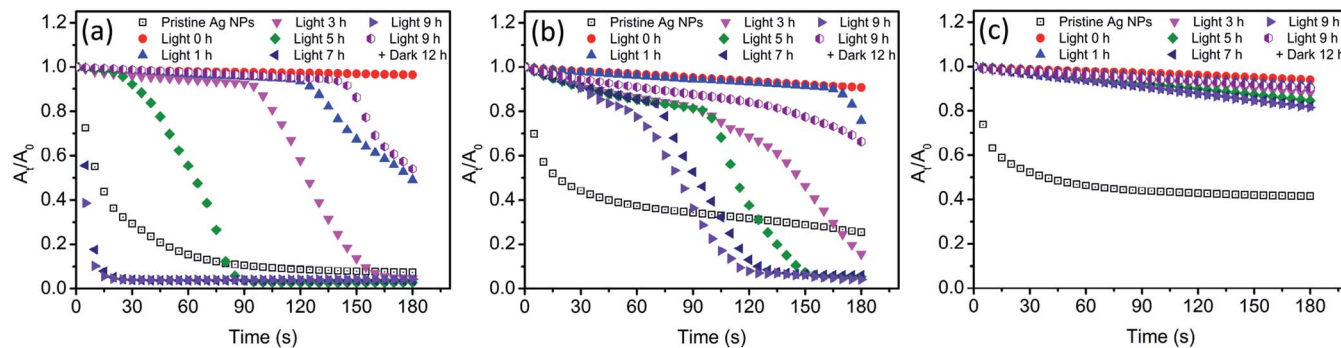


Fig. 6 The residual rate of Ag NPs (nAg-20 (a), nAg-40 (b), nAg-57 (c)) absorbance ( $A_t/A_0$ ) with time after introducing  $H_2O_2$  ( $200 \text{ mg L}^{-1}$ ), where  $A_0$  and  $A_t$  were the absorbance of Ag NPs at the maximum absorption wavelength before and after oxidation period of time, respectively. Before the introduction of  $H_2O_2$ , Ag NPs were incubated with  $0.5 \text{ mM Cl}^-$  under light irradiation.

soon as Ag NPs contacted with  $Cl^-$ . The strong affinity between  $Cl^-$  and  $Ag^+$  should be responsible for the rapid formation of AgCl layer.

After light irradiation, the antioxidant ability of Ag NPs was remarkably reduced and the absorbance of Ag NPs obviously decreased (Fig. 6). It seemed that the AgCl layer on Ag NPs was destroyed during light treatment. The photoreduction of AgCl as proposed above was believed to be the reason for the destruction of AgCl layer. Hence, the inner Ag NPs was exposed to outside, which allowed  $H_2O_2$  to erode along the exposed part to the interior Ag NPs. And then the core-shell structure of Ag-AgCl was further destroyed. With the loss of AgCl protection, Ag NPs was significantly oxidized by  $H_2O_2$ , leading to a sharp decline in the absorbance of Ag NPs (Fig. 6a and b). Meanwhile, we also noticed that this sharp decline was much more remarkable when compared with the decreasing of absorbance in blank control. Moreover, after 7 h of irradiation, the decreasing rate of nAg-20 absorbance was found to be obviously higher than blank control. The results suggested that Ag NPs was easier to be oxidized after lighting treatment in the presence of  $Cl^-$ . Previous study has demonstrated that the oxidation of Ag NPs would be facilitated in the presence of nucleophiles owing to the adsorption of nucleophiles on Ag NPs, and the oxidation rate was dependent on the strength of Ag-nucleophile.<sup>45</sup> Hence, in our study, the enhancement of  $H_2O_2$ -mediated dissolution of Ag NPs might be resulted from the adsorption of  $Cl^-$ . It should also be the reason for the remarkable increase in the value of Ag NPs zeta potential during light irradiation (Fig. 5). Due to the strong interaction between  $Cl^-$  and Ag NPs, the adsorption of  $Cl^-$  might also replace the group of PVP absorbed on Ag NPs, leading to the detachment of PVP.<sup>46</sup> As the prolonging of irradiation time, the time point when Ag NPs began to be significantly dissolved by  $H_2O_2$  was constantly advanced (Fig. 6a and b). After 7 h of irradiation, nAg-20 was remarkably dissolved as soon as the introducing of  $H_2O_2$ , which indicated that almost no AgCl presented on Ag NPs. It illustrated that the degree of the damage of the AgCl layer during light irradiation was positively related to the irradiation time. In addition, we also found that the size of Ag NPs had a great effect on the destruction of AgCl layer during light

irradiation. As Ag NPs size increased, the effect of light irradiation on  $H_2O_2$ -mediated decrease in Ag NPs absorbance was weakened. After 9 h of irradiation, just a slight enhancement was observed in the decreasing of nAg-57 absorbance, which implied that only slight photoreduction was occurred on AgCl layer. It might also be one of the reasons that less fusion aggregates were observed in the TEM images of large Ag NP than small one (Fig. 4). Under light, Ag NPs absorb visible photons and produce photogenerated electrons and holes, which could be separated by the surface plasmon resonance (SPR)-induced local electromagnetic field.<sup>38,41</sup> The photogenerated electrons might transfer from Ag NPs to AgCl layer, promoting the decomposition of AgCl. Small Ag NPs absorb shorter wavelength light, with higher energy, than large Ag NPs, inducing more intense SPR. We speculated that photogenerated electrons were easier to transfer from small Ag NPs to AgCl layer, and then the AgCl layer on small Ag NPs was destroyed more seriously during light treatment. After 9 h of irradiation, the light was removed and samples were kept in dark for 12 h. Then, we found that the antioxidant ability of Ag NPs was enhanced and the  $H_2O_2$ -mediated decreasing of Ag NPs absorbance was obviously inhibited compare with the sample after 9 h of irradiation, which implied that AgCl was re-formed on the surface of Ag NPs in dark conditions. During the incubation in dark, surface Ag atoms were expected to be oxidized by dissolved  $O_2$  and then reacted with  $Cl^-$  to form AgCl, which should be the reason of the re-formation of AgCl layer.

### Schematic diagram

Based on the experimental results and discussion above, a possible schematic diagram was proposed to describe the change on Ag NPs after exposing to  $Cl^-$  containing water under light and dark conditions (Fig. 7). In the presence of  $Cl^-$ ,  $Ag^+$  was captured by  $Cl^-$  and a AgCl layer rapidly coated on the surface of Ag NPs. This process also resulted in a decrease in dissolved Ag concentration and an increase in zeta potential of Ag NPs. The formation of the AgCl layer, a passivating coating, could prevent Ag NPs from further oxidation, even in the presence of a strong oxidant,  $H_2O_2$ . However, under sunlight, the AgCl layer was destroyed due to photoreduction. AgCl could be

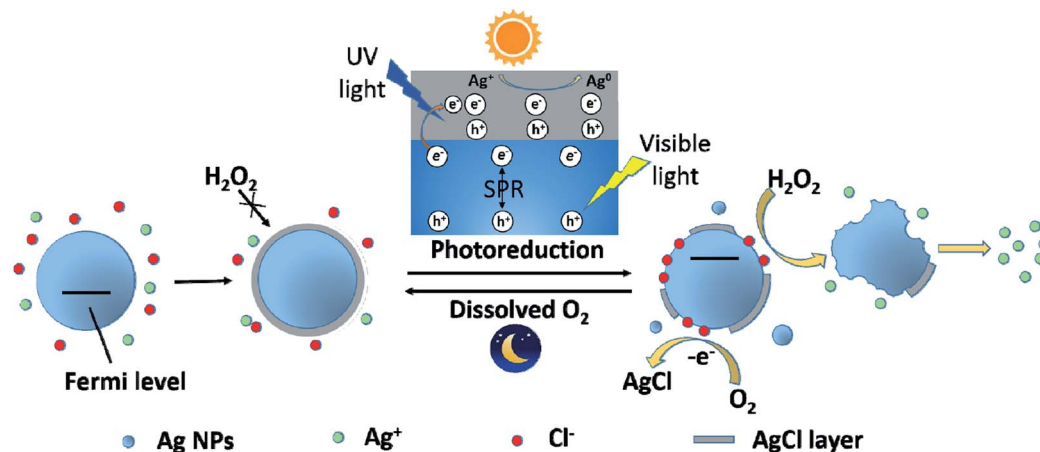


Fig. 7 Schematic diagram illustrating the photoreduction and re-formation of AgCl layer on the surface of Ag NPs under light and dark conditions.

excited by the UV light in sunlight and produced photoelectrons. The photogenerated electrons reduced the  $\text{Ag}^+$  of the AgCl lattice to form  $\text{Ag}^0$ . In addition, Ag NPs absorbed the visible light in sunlight and produced photoelectrons by SPR. These photoelectrons might transfer from Ag NPs to AgCl layer, promoting the photoreduction of AgCl. The photoreduction of AgCl led to a great decrease in dissolved Ag concentration and the formation of some tiny Ag NPs, as well as the destruction of AgCl layer on Ag NPs. The destruction of AgCl layer was greatly affected by irradiation time and Ag NPs size. With the prolonging of irradiation time, the AgCl layer was more seriously destroyed. And the AgCl layer on small Ag NPs was more easily photoreduced than large one. With the loss of AgCl protection, inner Ag NPs was exposed to the outside and could be oxidized by external oxidants (like  $\text{H}_2\text{O}_2$  and dissolved  $\text{O}_2$ ). Under sunlight, the adsorption of  $\text{Cl}^-$  on Ag NPs was also promoted, which led to a great increase in the zeta potential of Ag NPs. Moreover,  $\text{Cl}^-$ , adsorbed on Ag NPs, would donate a pair of electrons to the unoccupied orbital of Ag, shifting the Fermi level of Ag NPs upward.<sup>45</sup> Ag NPs with higher Fermi level was more easily to loss  $e^-$ . Therefore, the oxidation of Ag NPs (nAg-20 and nAg-40) induced by  $\text{H}_2\text{O}_2$  was obviously enhanced after light treatment. Unlike the violent reaction between  $\text{H}_2\text{O}_2$  and Ag NPs, the reaction of dissolved  $\text{O}_2$  with Ag NPs was relatively mild, which allowed surface silver atoms to interact with adsorbed  $\text{Cl}^-$  after being oxidized. Therefore, after removing light, the AgCl layer gradually re-formed on Ag NPs owing to the oxidation of dissolved  $\text{O}_2$ .

## Conclusion

In aquatic environment, Ag NPs are inevitably affected by various environmental factors, such as sunlight and  $\text{Cl}^-$ . Our study demonstrated that light irradiation had a great impact on various physicochemical properties of Ag NPs in the presence of  $\text{Cl}^-$ . After exposing to  $\text{Cl}^-$ , AgCl layer was rapidly formed on the surface of Ag NPs, leading to a decrease in the concentration of dissolved Ag and an increase in zeta potential of Ag NPs. Under

light irradiation, the decrease in dissolved Ag concentration was obviously enhanced owing to the photoreduction of AgCl. The AgCl layer on Ag NPs was also destroyed, and the adsorption of  $\text{Cl}^-$  on Ag NPs was promoted, resulting in the great increase of zeta potential. In addition, we also found that the destruction of AgCl layer during light treatment was significantly affected by Ag NPs size. AgCl layer on large Ag NPs was much more difficult to be photoreduced than that on small one. Due to the protection of AgCl layer, large Ag NPs are expected to be more stable and exist longer time than small one in aquatic environment. After removing light, surface silver atoms of Ag NPs were gradually oxidized by dissolved  $\text{O}_2$  and AgCl layer re-formed on the surface of Ag NPs. It was suggested that the destruction of AgCl layer during daytime and re-formation at night was a circle process and Ag NPs might present different states in aquatic environment during daytime and night.

## Conflicts of interest

There are no conflicts to declare.

## Acknowledgements

This study was supported by the National Key Research and Development Project (Grant 2019YFC0408801), the Major National Water Pollution Control and Management Project (Grant 2014ZX07405-003, 2017ZX07201001).

## References

- 1 M. N. Nadagouda, T. F. Speth and R. S. Varma, *Acc. Chem. Res.*, 2011, **44**, 469–478.
- 2 T. Benn, B. Cavanagh, K. Hristovski, J. D. Posner and P. Westerhoff, *J. Environ. Qual.*, 2010, **39**, 1875–1882.
- 3 Q. J. Huang and Y. Zhu, *Adv. Mater. Technol.*, 2019, **4**(5), 1800546.
- 4 F. Lu and D. Astruc, *Coord. Chem. Rev.*, 2018, **356**, 147–164.

- 5 H. Y. Jin, C. X. Guo, X. Liu, J. L. Liu, A. Vasileff, Y. Jiao, Y. Zheng and S. Z. Qiao, *Chem. Rev.*, 2018, **118**, 6337–6408.
- 6 N. C. Mueller and B. Nowack, *Environ. Sci. Technol.*, 2008, **42**, 4447–4453.
- 7 T. D. Huang, M. H. Sui, X. Yan, X. Zhang and Z. Yuan, *Colloids Surf., A*, 2016, **509**, 492–503.
- 8 E. J. Gubbins, L. C. Batty and J. R. Lead, *Environ. Pollut.*, 2011, **159**, 1551–1559.
- 9 L. Y. Yin, Y. W. Cheng, B. Espinasse, B. P. Colman, M. Auffan, M. Wiesner, J. Rose, J. Liu and E. S. Bernhardt, *Environ. Sci. Technol.*, 2011, **45**, 2360–2367.
- 10 K. J. Lee, P. D. Nallathamby, L. M. Browning, C. J. Osgood and X. H. N. Xu, *ACS Nano*, 2007, **1**, 133–143.
- 11 X. Y. Yang, A. P. Gondikas, S. M. Marinakos, M. Auffan, J. Liu, H. Hsu-Kim and J. N. Meyer, *Environ. Sci. Technol.*, 2012, **46**, 1119–1127.
- 12 P. V. AshaRani, G. L. K. Mun, M. P. Hande and S. Valiyaveetil, *ACS Nano*, 2009, **3**, 279–290.
- 13 N. Grillet, D. Manchon, E. Cottancin, F. Bertorelle, C. Bonnet, M. Broyer, J. Lerme and M. Pellarin, *J. Phys. Chem. C*, 2013, **117**, 2274–2282.
- 14 Y. Li, W. Zhang, J. F. Niu and Y. S. Chen, *Environ. Sci. Technol.*, 2013, **47**, 10293–10301.
- 15 Y. W. Cheng, L. Y. Yin, S. H. Lin, M. Wiesner, E. Bernhardt and J. Liu, *J. Phys. Chem. C*, 2011, **115**, 4425–4432.
- 16 J. P. Shi, B. Xu, X. Sun, C. Y. Ma, C. P. Yu and H. W. Zhang, *Aquat. Toxicol.*, 2013, **132**, 53–60.
- 17 W. C. Hou, B. Stuart, R. Howes and R. G. Zepp, *Environ. Sci. Technol.*, 2013, **47**, 7713–7721.
- 18 W. Zhou, Y. L. Liu, A. M. Stallworth, C. Ye and J. J. Lenhart, *Environ. Sci. Technol.*, 2016, **50**, 12214–12224.
- 19 S. J. Yu, Y. G. Yin, J. B. Chao, M. H. Shen and J. F. Liu, *Environ. Sci. Technol.*, 2014, **48**, 403–411.
- 20 X. Y. Yang, S. H. Lin and M. R. Wiesner, *J. Hazard. Mater.*, 2014, **264**, 161–168.
- 21 B. M. Angel, G. E. Batley, C. V. Jarolimek and N. J. Rogers, *Chemosphere*, 2013, **93**, 359–365.
- 22 C. Levard, S. Mitra, T. Yang, A. D. Jew, A. R. Badireddy, G. V. Lowry and G. E. Brown, *Environ. Sci. Technol.*, 2013, **47**, 5738–5745.
- 23 X. Li, J. J. Lenhart and H. W. Walker, *Langmuir*, 2010, **26**, 16690–16698.
- 24 C. M. Ho, S. K. W. Yau, C. N. Lok, M. H. So and C. M. Che, *Chem.-Asian J.*, 2010, **5**, 285–293.
- 25 R. Ma, C. Levard, S. M. Marinakos, Y. W. Cheng, J. Liu, F. M. Michel, G. E. Brown and G. V. Lowry, *Environ. Sci. Technol.*, 2012, **46**, 752–759.
- 26 T. S. Peretyazhko, Q. Zhang and V. L. Colvin, *Environ. Sci. Technol.*, 2014, **48**, 11954–11961.
- 27 O. S. Ivanova and F. P. Zamborini, *J. Am. Chem. Soc.*, 2010, **132**, 70–72.
- 28 A. R. Poda, A. J. Kennedy, M. F. Cuddy and A. J. Bednar, *J. Nanopart. Res.*, 2013, **15**(5), 1673.
- 29 N. G. Bastus, F. Merkoci, J. Piella and V. Puntes, *Chem. Mater.*, 2014, **26**, 2836–2846.
- 30 L. Li, J. Sun, X. R. Li, Y. Zhang, Z. X. Wang, C. R. Wang, J. W. Dai and Q. B. Wang, *Biomaterials*, 2012, **33**, 1714–1721.
- 31 C. Levard, B. C. Reinsch, F. M. Michel, C. Oumahi, G. V. Lowry and G. E. Brown, *Environ. Sci. Technol.*, 2011, **45**, 5260–5266.
- 32 A. M. E. Badawy, T. P. Luxton, R. G. Silva, K. G. Scheckel, M. T. Suidan and T. M. Tolaymat, *Environ. Sci. Technol.*, 2010, **44**, 1260–1266.
- 33 J. Farkas, P. Christian, J. A. Gallego-Urrea, N. Roos, M. Hasselov, K. E. Tollefsen and K. V. Thomas, *Aquat. Toxicol.*, 2011, **101**, 117–125.
- 34 K. A. Huynh and K. L. Chen, *Environ. Sci. Technol.*, 2011, **45**, 5564–5571.
- 35 X. Li, J. J. Lenhart and H. W. Walker, *Langmuir*, 2012, **28**, 1095–1104.
- 36 Y. G. Yin, W. Xu, Z. Q. Tan, Y. B. Li, W. D. Wang, X. R. Guo, S. J. Yu, J. F. Liu and G. B. Jiang, *Environ. Pollut.*, 2017, **220**, 955–962.
- 37 L. Han, P. Wang, C. Z. Zhu, Y. M. Zhai and S. J. Dong, *Nanoscale*, 2011, **3**, 2931–2935.
- 38 R. F. Dong, B. Z. Tian, C. Y. Zeng, T. Y. Li, T. T. Wang and J. L. Zhang, *J. Phys. Chem. C*, 2013, **117**, 213–220.
- 39 P. Wang, B. B. Huang, Y. Dai and M. H. Whangbo, *Phys. Chem. Chem. Phys.*, 2012, **14**, 9813–9825.
- 40 L. H. Dong, D. D. Liang and R. C. Gong, *Eur. J. Inorg. Chem.*, 2012, 3200–3208, DOI: 10.1002/ejic.201200172.
- 41 B. Z. Tian, R. F. Dong, J. M. Zhang, S. Y. Bao, F. Yang and J. L. Zhang, *Appl. Catal., B*, 2014, **158**, 76–84.
- 42 Z.-m. Xiu, Q.-b. Zhang, H. L. Puppala, V. L. Colvin and P. J. J. Alvarez, *Nano Lett.*, 2012, **12**, 4271–4275.
- 43 D. H. Lin, S. Ma, K. J. Zhou, F. C. Wu and K. Yang, *J. Colloid Interface Sci.*, 2015, **450**, 272–278.
- 44 D. He, S. Garg and T. D. Waite, *Langmuir*, 2012, **28**, 10266–10275.
- 45 P. Mulvaney, T. Linnert and A. Henglein, *J. Phys. Chem.*, 1991, **95**, 7843–7846.
- 46 Y. Tang, W. He, G. Y. Zhou, S. X. Wang, X. J. Yang, Z. H. Tao and J. C. Zhou, *Nanotechnology*, 2012, **23**(35), 355304.