# Environment dependence of the expression of mutational load and species' range limits

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## Abstract

Theoretical and empirical research on the causes of species' range limits suggest the contribution of several intrinsic and extrinsic factors, with potentially complex interactions among them. An intrinsic factor proposed by recent theory is mutational load increasing towards range edges because of genetic drift. Furthermore, environmental quality may decline towards range edges and enhance the expression of load. Here, we tested whether the expression of mutational load associated with range limits in the North American plant Arabidopsis lyrata was enhanced under stressful environmental conditions by comparing the performance of within- versus between-population crosses at common garden sites across the species' distribution and beyond. Heterosis, reflecting the expression of load, increased with heightened estimates of genomic load and with environmental stress caused by warming, but the interaction was not significant. We conclude that range-edge populations suffer from a twofold genetic Allee effect caused by increased mutational load and stressdependent load linked to general heterozygote deficiency, but there is no synergistic effect between them.

#### KEYWORDS

Arabidopsis lyrata, environmental stress, genetic drift, heterosis, mutational load, range limit, temperature gradient

# 1 | INTRODUCTION

The geographic distribution of species can be constrained by several factors, and they may interact in complex manners (Connallon & Sgrò, 2018; Gaston, 2009; Roy et al., 2009; Sexton et al., 2009; Willi & Van Buskirk, 2019). An important factor put forward by relatively recent theory is the accumulation of mutational load in rangeedge populations due to a history of enhanced genetic drift (Henry et al., 2015; Peischl et al., 2013; Willi, 2019). In parallel, environmental conditions may degrade towards the range limits of a species (Brown, 1984; Cahill et al., 2014; Hargreaves et al., 2014; Lee-Yaw

et al., 2016). In addition to their respective negative effect on population performance, mutational load and environmental stress may interact synergistically in their negative effect on performance, as was suggested and found for inbreeding depression (Roff, 1997, pages 285-338; Reed et al., 2012), both in animals (e.g. Fox & Reed, 2011) and in plants (e.g. Cheptou et al., 2000). The contribution of mutational load to range limits could, therefore, be much more substantial than expected based on previous theoretical and empirical work, and relevant for shaping the distribution of many species.

Two types of demographic scenarios at range edges are predicted to lead to increased mutational load. First, theoretical models

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predict that range expansion through serial demographic bottlenecks enhances genetic drift and lowers the efficacy of purifying selection, which leads to the accumulation of deleterious mutations (Peischl et al., 2013, 2015; Peischl & Excoffier, 2015). High load at the expansion front is predicted to slow the spread of a species or halt it if genome-wide recombination is low (Peischl et al., 2015). The so-called expansion load is predicted to involve mainly recessive deleterious mutations (Peischl & Excoffier, 2015), to establish even in the absence of environmental gradients and to persist in populations for thousands of generations (Peischl et al., 2013). Second, if populations at range edges are small and isolated, genetic drift is also enhanced and should increase mutational load. Simulations showed that when the carrying capacity of suitable habitat declines along a line of habitat patches, the occupied range becomes shorter than the distribution of habitat would suggest because of mutation accumulation (Henry et al., 2015). The effect of load is larger if dispersal and population growth rates are low (Henry et al., 2015), and this scenario of mutational load limiting ranges may apply best to rear edges of species' distributions (Hampe & Petit, 2005). Genomic evidence for the accumulation of mutational load under range expansion was found in humans (Lohmueller et al., 2008; Peischl et al., 2018) and under expansion and rear-edge isolation in plants (González-Martínez et al., 2017; Koski et al., 2019; Willi et al., 2018; Zhang et al., 2016).

At range limits, the environment may also be more stressful. Empirical studies document a general decline in habitat suitability towards range limits in many taxa (Pironon et al., 2017; Sexton et al., 2009). A consistent and pronounced effect of increasing habitat marginality was found in species that were experimentally transplanted beyond their range edges as their lifetime performance was significantly lower beyond the edge compared with sites within the range (Hargreaves et al., 2014). The decline in performance was found to coincide with a decline in predicted habitat suitability revealed by niche modelling (Lee-Yaw et al., 2016). The two metaanalyses suggest that the decay in habitat suitability is general. Such unfavourable conditions may act in isolation and lower population performance, or they may enhance the expression of mutational load at range edges, further reducing population performance.

Potential mechanisms for a synergistic interaction between mutational load and environmental stress have rarely been explored *per se*, but have been mainly discussed and tested in the context of inbreeding depression (Armbruster & Reed, 2005; Fox & Reed, 2011; Willi et al., 2007a). Three mechanisms have been proposed (Reed et al., 2012). The first is that stress induces the expression of deleterious mutations that are silent under benign conditions (Elena & de Visser, 2003; Kondrashov & Houle, 1994). For example, in *Drosophila*, recessive alleles were linked to increased mortality in inbred lines under temperature stress (Vermeulen & Bijlsma, 2004). Inbreeding and extreme stress impede the function of heat shock proteins such as Hsp90, which are essential in buffering the expression of deleterious mutations (Bergman & Siegal, 2003; Queitsch et al., 2002; Rutherford & Lindquist, 1998). A second mechanism is that mutational load may lower stress resistance or tolerance, by destabilizing cellular homeostasis under stress, or affecting tissue and genome repair after stress exposure (Agrawal & Whitlock, 2010; Reed et al., 2012). For example, hybrids of *Arabidopsis thaliana* expressed more metabolites from central pathways linked to higher freezing tolerance or disease resistance compared with inbred lines (Korn et al., 2010; Yang et al., 2015), and the regulation of stress response pathways led to higher recovery after stress (Miller et al., 2015). The third mechanism is that stress increases phenotypic variance and with it the opportunity for relative fitness measures to differ more between inbred and outbred individuals (Waller et al., 2008). Reed et al. (2012) concluded that variation in inbreeding depression was mainly predicted by stress itself and only weakly by increased phenotypic variance.

In this study, we tested for stress dependence of the expression of mutational load at the range limits of North American Arabidopsis lyrata subsp. lyrata (L.). The species has been strongly impacted by the last Pleistocene glaciation cycle, with more than half of its contemporary distribution resulting from rapid post-glacial range expansion from two separate refugia (Griffin & Willi, 2014; Willi et al., 2018). Both range expansion and rear-edge dynamics were found to be associated with an increase in presumably deleterious mutations assessed on the level of coding DNA sequences, leading to a current pattern of heightened deleterious mutations in range-edge populations (Willi et al., 2018). Another study showed that this heightened genomic load was correlated with expressed load, estimated by heterosis-the fitness increase in between-population compared with within-population crosses (Perrier et al., 2020). Here, we tested whether the expression of mutational load, again estimated by heterosis, additionally depended on the interaction between genomic load and environmental stressfulness in the direction of a synergistic effect. Offspring of within-population crosses (WPC) and betweenpopulation crosses (BPC) were raised in a latitudinal transplant experiment with five common garden sites within and beyond the distribution range (plant data as in Perrier et al. (2020)). Environmental stressfulness was estimated by (a) the differences in climatic conditions experienced by populations in each common garden relative to their site of origin, assuming local adaptation to climate (Hoffmann & Hercus, 2000), and by (b) the relative performance of each population at a garden site compared with the best garden. Finally, we tested whether heterosis in range-edge populations-expressing the highest mutational load compared with populations from elsewhere of the range-was enhanced when offspring of crosses were transplanted to environments beyond range limits.

## 2 | MATERIALS AND METHODS

#### 2.1 | Experimental design

### 2.1.1 | Populations

Twenty populations of A. *lyrata* subsp. *lyrata* from across the species' distribution were selected for producing WPC and BPC (Figure 1;

Table S1). The populations represented the two main ancestral clusters (Griffin & Willi, 2014; Willi et al., 2018), the latitudinal range and variation in the genomic signatures of mutational load (Willi et al., 2018). Our sampling also represented the two mating systems: predominant outcrossing and predominant selfing (MO2, ON1 and ON11). In the study system, selfing populations are mainly restricted to the edges of distribution of the two ancestral clusters (Griffin & Willi, 2014). They have an especially high genomic load, on the one hand because of their expansion history, and on the other hand because of the mating system shift (Perrier et al., 2020; Willi et al., 2018). Eighteen populations (nine per cluster) were considered target populations in the assessment of expressed mutational load, and two were partner populations for BPC (NY1 for populations of the eastern cluster, and IA1 for populations of the western cluster). Partner populations were selected because of their core location in each genetic cluster, high genomic diversity and a history of little genetic drift. Between-population crosses with these partner populations were, therefore, expected to overcome mainly the load of target populations. Seeds of individual plants had been collected in each population in the field over an area of about  $450 \text{ m}^2$ .



FIGURE 1 Distribution of Arabidopsis lyrata in eastern North America, with information on habitat suitability, the location of the 20 populations studied, and the 5 common garden sites. The range of A. lyrata is represented by the dotted line, and habitat suitability by shades of blue, with darker blue indicating higher suitability (Lee-Yaw et al., 2018). Populations are shown by circles, with abbreviations for state (USA) or province (Canada) and a number (Willi et al., 2018). Blue and red circles represent northern- and southern-edge populations, respectivley, in our analysis. Green triangles indicate the five common garden (CG) sites; numbers added to labels are in sequence of north to south. State outlines for the USA are shown, and the split between eastern and western genetic cluster is represented by the dashed line. Of the 20 populations, two were used as partner populations for betweenpopulation crosses, NY1 for crosses with eastern populations, and IA1 for crosses with western populations

# 2.1.2 | Crossing

One plant of each of 26 randomly selected and presumably unrelated seed families per population was raised under controlled, indoor conditions. Plants were randomly appointed to 'mothers'/ pollen recipients (12), 'fathers'/pollen donors (12) and backups (2). Each 'mother' was crossed using pollen from one 'father' of the same population (WPC). For target populations, each 'mother' was additionally crossed using pollen from one 'father' of the partner population (BPC). Crosses were non-reciprocal. Pollination was performed at the bud stage to prevent unintended cross- or self-pollination. We repeated each cross-combination to obtain enough healthy-looking seeds to perform the transplant experiment (min. 60). In the case of systematic cross failure (overall fractions of failed crosses were similar between WPC and BPC, about 8%), we replaced the father plant with a backup plant. In total, 401 family-cross type combinations were used for the transplant experiment (Table S2).

#### 2.1.3 | Transplant experiment

Five common garden sites (CG) were selected along a ca. 1400 km latitudinal gradient crossing the distribution of A. lyrata in the eastern USA (Figure 1; Table S3 with WorldClim data v 2.0 (Fick & Hijmans, 2017)). The five sites represented the centre of the species' range (CG3, Harrisonburg, VA), the southern and northern range edges (CG4, Winston-Salem, NC, and CG2, Williamstown, MA, respectively), and areas beyond the edges (CG5, Athens, GA, and CG1, in the Adirondacks, NY). In 2017, in each common garden, two seeds (one if few seeds were available) of each cross were sown in each of three pots (5.7 cm wide and 5.9 cm deep, with holes at the bottom) randomly distributed across three spatial blocks, resulting in three to six seeds per cross in each common garden (more details in Perrier et al. (2020)). A total of 12,933 seeds were sown across the five common gardens. Five data loggers (iButton<sup>®</sup>, Maxim Integrated Products, Inc) monitored air temperature every hour (1.5 m above ground, in the shadow) in each site for the duration of the experiment.

### 2.1.4 | Performance

Germination was recorded three days a week until peak germination was over (after four to five weeks) and afterwards once a week until thinning, starting 11 weeks after sowing. Seedlings were randomly thinned to one per pot. Reproductive output was assessed several weeks after peak flowering in 2018 and 2019 and calculated for each individual as the sum of fruits, pedicels (flowers that did not develop into fruits), flowers and flower buds. Flowers that did not develop into fruits were included because they contributed to the performance capacity of plants, and their pollen may have fertilized ovules on other plants. However, seeds per fruit were not estimated as pollen transfer was not controlled for in the experiment. In an

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additional experiment, we assessed seed survival over winter, with seeds from the same crosses as used in the transplant experiment, to estimate how seeds that had not germinated the first fall could have contributed to population performance. For each population-cross type combination, we pooled 100 seeds from five to twelve seed families, split them in 10 bags (made of nonwoven polypropylene-felt, 40 g/m<sup>2</sup> to allow air and moisture exchange) and placed 2 bags per garden on bare ground next to pots in October 2018. Bags were retrieved in late spring 2019 and visually screened for germinated seedlings and seeds. Remaining seeds were tested for germination in the laboratory. Seed survival over the winter was calculated for each bag as the fraction of germinated seedlings outdoors and in the laboratory and used for estimating population growth rate (for each combination of population, cross type and garden).

# 2.2 | Statistical analysis

Initial analyses were based on multiplicative performance (MP) calculated on the level of the pot as the product of the fraction of germinated seeds and the sum of reproductive output until year 3 (year 2 for CG 3 that had to be given up early). We tested for a general effect of cross type (BPC compared with WPC, which were coded as 1 and 0, respectively), genomic load, environmental stress and all two- and three-way interactions between them. Genomic load was estimated based on nuclear DNA sequences. Population pools of equimolar DNA of 25 plants per population were wholegenome sequenced, nuclear single-nucleotide polymorphisms were retrieved, and those in coding regions were split into nonsynonymous and synonymous. For each population, the ratio of non-synonymous polymorphic sites (P<sub>n</sub>) to synonymous polymorphic sites (P<sub>2</sub>) adjusted for their mean derived allele frequency (relative to A. thaliana,  $f_{\rm p}$  and  $f_{\rm s}$ , respectively),  $P_{\rm p}f_{\rm p}/P_{\rm s}f_{\rm s}$  was calculated (Willi et al., 2018). This estimate of mutational load should depict the excess number and frequency of derived and presumably deleterious mutations due to genetic drift. The estimate of mutational load correlated well with population genomic diversity in intergenic DNA, which itself is affected little by interpopulation gene flow in the species. Between-population crosses were assigned the genomic load of the mother population as all BPC shared the same pollen donor population within a genetic cluster. Environmental stress was depicted by the difference in climate between garden site and site of origin of a population (Hoffmann & Hercus, 2000; Rutter & Fenster, 2007). Distribution modelling had shown that average minimum temperature in March and April described best the niche and latitudinal range limits of A. lyrata (Lee-Yaw et al., 2018). Therefore, environmental stress was calculated as the difference in average minimum temperature in early spring (March and April) between common garden sites and site of origin of populations:  $\Delta T_{\min} = T_{\min CG} - T_{\min origin}$ .  $T_{\min CG}$  was calculated based on records of temperature loggers (Table S4), and  $T_{\min \text{ origin}}$  based on the WorldClim 2.0 database (spatial resolutions of 30 seconds, Fick & Hijmans, 2017; Table S1) and averaged over target and partner

population for BPC. Positive and negative values of  $\Delta T_{min}$  indicated transplanting towards warmer and colder temperatures, respectively. We assumed that both positive and negative deviations could be perceived as stressful, resulting in a bell-shaped relationship between  $\Delta T_{min}$  and performance. Therefore, we included the square term of  $\Delta T_{min}$  in the model. Analyses were performed in a Bayesian framework on 10 parallel chains (no. of iterations: 50,000, burnin: 5,000, thinning: 100, leading to 450 posterior estimates; code in Method S1), with the package MCMCgImm (Hadfield, 2010, 2019) in R (R Core Team, 2021). The model accounted for the zero-inflation of MP by considering a logistic process (probability of 1 assigned to MP > 0) and a Gaussian process (MP > 0, hereafter referred to as the log-normal process). Random effects were mother plant nested within maternal population, and maternal population, for which variances in intercepts and slopes on cross type were estimated, and block nested within common garden, and common garden. Covariates were mean-centred, and square terms were calculated based on mean-centred values.

Further analyses were-for ease of interpretation-performed on heterosis estimated on the level of the population. Heterosis was calculated as the increase in performance of BPC relative to WPC:  $(W_{BPC} - W_{WPC})/W_{WPC}$ , where  $W_{WPC}$  and  $W_{BPC}$  were population averages based on seed family averages in a common garden.  $W_{WPC}$  was further averaged across target and partner populations. Heterosis was also assessed for the finite rate of increase per year,  $\lambda$ , from stage-classified matrices (Caswell, 2001) based on the mean performance of plants of each population-cross type-garden combination, over the three years (detailed in Perrier et al. (2020)). These matrices comprised the stages: (1) healthy seeds in year 1, (2) reproducing individuals of year 2 and (3) reproducing individuals of year 3, with a one-year projection interval between each stage. Heterosis estimates were log<sub>10</sub>-transformed and analysed by hierarchical mixedeffects models based on restricted maximum likelihood using the packages Ime4 (Bates et al., 2015) and LmerTest (Kuznetsova et al., 2017) in R. Fixed effects were the genomic estimate of mutational load of the mother population and environmental stress. Crossed random effects were population and common garden (detailed in Method S2A). Here,  $\Delta T_{min}$  was calculated with  $T_{min \text{ origin}}$  again averaged over target and partner population.

To explore the effect of stress, we depicted it by either signed or absolute  $\Delta T_{min}$ , included an orthogonal quadratic term or not, and included interaction terms or not. We performed model selection based on the corrected Akaike information criterion, AICc (Sugiura, 1978). We selected the model with the lowest AICc value, with the additional criterion that its AICc had to be at least 2 units below the next better model. For the best model, we tested the significance of each fixed effect with a likelihood-ratio  $\chi^2$  test using the *Anova* function (type III) of the package *car* (Fox & Weisberg, 2019). We also used the same model to test for effects on  $W_{WPC}$  to estimate the likely effect genomic load may have under stress in the natural populations. A confounding role of genetic incompatibility estimated by neutral genetic divergence and of divergent adaptation between target and partner populations estimated by environmental divergence could be excluded. Details on these analyses are presented in Appendix S1.

We repeated the analysis on heterosis on a second estimate of stress, the relative decline in population-level performance at a site relative to the best site  $(1 - W_{WPC}/W_{max})$  (Reed et al., 2012). Stress varied between 0 (conditions leading to maximal performance) and 1 (conditions with  $W_{WPC} = 0$ ). Relative performance was then averaged over both target and partner populations.

Finally, we tested whether heterosis was higher when rangeedge populations (with highest load) were transplanted beyond the range limits compared with sites at the range limits. We assessed this effect in two distinct data sets: one considering the six northern edge populations (ON8, ON11, ON12, NY4, NY5 and NY6; in blue in Figure 1) raised at the northern edge site (CG2) and beyond the northern edge (CG1), and one considering the six southern-edge populations (MO1, MO2, IN1, NC2, NC4 and VA1; in red in Figure 1) raised at the southern-edge site (CG4) and beyond the southern edge (CG5). For both data sets,  $log_{10}$ transformed heterosis was tested in hierarchical mixed-effect models based on restricted maximum likelihood. Transplant was a fixed effect (edge = 0), and population was a random effect (detailed in Method S2B).

## 3 | RESULTS

Analyses on the level of the pot revealed a significant positive effect of cross type on performance in the logistic process (Table 1). Performance was higher in BPC compared with WPC, supporting a general heterosis effect. Performance was also significantly negatively affected by genomic load and by a positive cross typeby-genomic load interaction in the log-normal process (MP > 0). Performance generally declined with increasing genomic load but less in BPC than WPC. Performance was further marginally negatively affected by the linear term of stress ( $\Delta T_{\min}$ ) and significantly negatively affected by the quadratic term of stress ( $\Delta T_{min}^2$ ), in both the logistic and log-normal process. Performance generally declined when conditions in the common garden diverged from conditions at the site of origin ( $\Delta T_{\min} \neq 0$ ). Furthermore, the interaction cross type-by- $\Delta T_{\min}$  was significantly positive for the logistic process, indicating that the decline due to stress was stronger for WPC compared with BPC, especially under warmer conditions ( $\Delta T_{min} > 0$ ). The logistic process determining performance was also significantly affected by a negative interaction between genomic load and  $\Delta T_{min}$ . Performance of populations with higher load declined more when exposed to warmer temperatures. Finally, three-way interactions between fixed effects were not significant.

Results on population-level heterosis based on *MP* or  $\lambda$  revealed significant positive associations between genomic load and heterosis and between the linear term of stress and heterosis (Table 1; Figure 2a), in line with the significant cross type-by-genomic load and cross type-by- $\Delta T_{min}$  interactions in the individual-level analysis. Heterosis was highest in populations with more mutational load, and in populations raised under warmer conditions than at their site of origin ( $\Delta T_{min} > 0$ ). Neither the quadratic term of stress nor the interactions between stress and genomic load were significant. Heterosis based on *MP* and  $\lambda$  in the five common gardens ranged from -0.96 to 23.50 (mean: 1.88) and -0.53 to 7.29 (mean: 0.73), respectively (Table S5).

Model selection pointed to this full model on heterosis being the best supported (Table S6). Among the models compared, the one with genomic load, the signed term of stress ( $\Delta T_{min}$ ), its square term and interactions between genomic load and stress had by far the lowest AICc (next better model with +6.7 AICc units). Similar to the analysis of heterosis, its component of  $W_{WPC}$  revealed an effect of genomic load,  $\Delta T_{min}$ , and  $\Delta T_{min}^2$  (Table S7), but no effect of the genomic load-by- $\Delta T_{min}$  interaction. The performance of WPC declined with increasing genomic load and, in an accelerating manner, with warmer conditions compared with the site of origin of populations (Figure S1). The model-predicted maximal decline in WPC performance for a population with average genomic load was 92% when exposed to warmer temperatures (baseline of  $\Delta T_{min} = 0$ ; Table S8).

Analyses on the effect of stress expressed by performance decline on heterosis revealed similar results as those based on  $\Delta T_{min}$ . The best model predicting heterosis included the linear and quadratic terms for stress and interactions between stress and genomic load (next better model with +7.3 AICc units; Table S9). Heterosis was significantly positively affected by genomic load (marginal for  $\lambda$ ) and both the linear (except *MP*) and the quadratic term of stress (Table S10) but was again unaffected by interactions. Heterosis based on *MP* increased almost symmetrically towards low ( $W_{WPC} = W_{max}$ ) and high stressfulness ( $W_{WPC} = 0$ ; Figure S2), while heterosis based on  $\lambda$ increased more strongly towards high stressfulness (Figure 2b).

Finally, heterosis of range-edge populations did not significantly differ between the garden sites at the range-edge and the garden sites beyond the range, neither for the southern nor for the northern range edge (Table S11; Figure S3).

## 4 | DISCUSSION

Recent evolutionary theory proposes that species' range limits may result from high mutational load in range-edge populations due to prolonged exposure to genetic drift caused by past range expansion or rear-edge demographics (Willi, 2019). Here, we tested whether the expression of such load was enhanced by environmental stress in a large-scale transplant experiment with *Arabidopsis lyrata*. We estimated expressed mutational load by heterosis based on the multiplicative performance of between- compared with within-population crosses, assuming that load was mainly due to recessive deleterious mutations. To estimate stress, we used divergent thermal conditions during a critical part of the year between gardens and sites of origin of populations or relative performance decline of each population. Expressed mutational load was stronger in populations with higher deleterious counts inferred from coding DNA sequences, as previously found (Perrier et al., 2020; Willi et al., 2018). However, stress

(S; $\Delta T_{min} = T_{min CG}$ - common garden site	T <sub>min origin</sub> ) and tl	heir inter	ractions or	n multipli	cative perfo	rmance (N	1P), or on l	heterosis ba	ised on MP o	r the finite ra	te of increa	ise (1) (wit	hout testi	ng for cross type	, at the five
	Cross type	Genor	nic load	Stress		<b>S</b> <sup>2</sup>		CT * GL	CT * S	CT * S <sup>2</sup>	GL * S		GL * S <sup>2</sup>	CT * GL * S	$CT * GL * S^2$
Performance	Mean	Mean		Mean		Mean		Mean	Mean	Mean	Mean		Mean	Mean	Mean
Logistic MP	0.46***	-2.69		-56.76(*	(*	-51.61**	*	2.29	54.34***	-14.77	-338.34*	*	69.24	-95.19	87.68‡
Log-normal MP	0.11(*)	-1.31*		-16.14(*	(*	-7.27**		1.36*	1.73	-6.52(*)	-56.74		-7.69	17.14	18.69†, ‡
Heterosis		β	$\chi^{2}$	β	$\chi^{2}$	β	$\chi^{2}$				β	×2	B X <sup>2</sup>		
МР		1.92	6.63**	1.03	5.77*	-0.73	2.98(*)				-3.21	0.48	3.72 0.7	7†, ‡	
У		0.84	5.36*	0.79	14.78***	-0.07	0.15				0.22	0.01	1.68 0.7	7†	
<i>Note: MP</i> was assume process (binary varia For this model, estim scale.	ed to follow a G ble depicting ge ates of coefficie	iaussian d erminatioı ents are n	listribution n combinec nodes of M	(log <sub>10</sub> -tra 1 with sur ICMC sam	insformed if vival and th nples from t	>0) with ( e capacity he posteri	D-inflation. to initiate or distribut	Therefore, 1 flowering) a tion of paran	the upper mo nd the Gaussi neters ( <i>mean</i> ).	del assessed a an process (to . The logistic p	ll fixed and the number of the m	random ef of flowers ıodel pred	fects for th during one icts non-ze	neir importance in e or two reproduct eros in the distribu	both the logistic ive seasons). :ion on the logit
Estimates of populat to improve converge respectively, and the	ion heterosis we nce. Test statist conditional R <sup>2</sup> c	ere log <sub>10</sub> - tics incluc of the mo	-transforme de regressic odel was of	ed prior tc on coeffic 0.37 and	o analysis. Si ient $(\beta)$ and 0.46, respectively	the $\chi^2$ -valuctively.	<sub>in</sub> , was an a Je of each	iverage over fixed effect.	target and p <i>e</i> For heterosis	irtner populati based on MP	ion. Both m and on ג, th	odels on h	eterosis w I R <sup>2</sup> of the	ere run with the <i>b</i> c model was of 0.18	<i>byqa</i> optimizer and 0.23,

TABLE 1 Results of models testing for the effect of cross type (CT; between- compared with within-population crosses), genomic estimate of mutational load (GL), environmental stress S ü Genomic load and stress were standardized prior to analyses (mean = 0). Model fits with significant (positive) intercept are indicated by  $\pm$ . Estimates and regression coefficients with *P*-values <0.05 are written in bold; significance is indicated: (\*) p < 0.05, \*\* p < 0.05, \*\* p < 0.01. Results for random effects are not shown. For one of the five common gardens (CG3), the experiment stopped early and variables consider performance to year 2 only (indicated by  $\ddagger$ ).



FIGURE 2 Stress dependence of the expression of mutational load estimated by heterosis in *Arabidopsis lyrata*. Population heterosis was estimated based on the finite rate of increase,  $\lambda$ , of within-population crosses and between-population crosses with a partner population. Outcrossing populations are indicated by dots, selfing populations by squares. Stress was depicted as either the difference in minimum temperature in early spring between common garden and the site of origin of a population ( $\Delta T_{min} = T_{min CG} - T_{min origin}$ ; positive values to the right of the vertical dashed line indicate a warmer environment) (a) or the relative decline in population-level performance *W* in a garden relative to the garden of highest performance for that population ( $1 - W_{WPC}/W_{max}$ ) (b). For both stress estimates, means of target and partner populations were considered. The horizontal dashed lines indicate when heterosis drops below 0, and outbreeding depression dominates. The black lines represent model-predicted relationships between heterosis and environmental stress (test statistics in Table 1, Table S10; \* *p* < 0.05, \*\*\* *p* < 0.001). Genomic load, the ratio of genome-wide non-synonymous to synonymous polymorphic sites adjusted by their mean derived frequency,  $P_nf_n/P_sf_s$  (Willi et al., 2018), is represented in shades of yellow (low) to red (high)

did not clearly synergistically interact with genomic load to enhance heterosis. Nonetheless, we found a general effect of stress on heterosis linked to a stronger performance decline of within- compared with between-population crosses under warmer conditions than at sites of origin of populations. Our study, therefore, suggests that a general heterozygote deficit is more detrimental under temperature stress and that interpopulation outcrossing can alleviate this effect.

A main result of this study is that expressed mutational load estimated by heterosis was not significantly enhanced under thermal stress. Indeed, while analyses on pot-level performance were supportive of an effect of the cross type, genomic load and environmental stress on performance, these analyses did not support a three-way interaction between effects. Analyses on populationlevel heterosis were not clearly supportive of an interaction either. Even though the model including interactions between stress and genomic load was better supported than the model without interactions, the interaction was not significant in the  $\chi^2$  test. This suggests that some variation in heterosis may be the result of an interaction, but other variation may easily blur the effect. Such environment dependence has often been evoked due to the similarities in the underlying genetics between heterosis (if caused by dominance) and inbreeding depression (Fenster & Galloway, 2000; Oakley et al., 2015). Indeed, a previous meta-analysis on the stress dependence of heterosis found that the benefit of outcrossing was considerably higher under stress (Frankham, 2015); however, the study only focussed on small and inbred populations and did not include larger populations for comparison. Some of them may also show a heterotic effect under outbreeding that is enhanced under stress, as found in our study.

We think that our result for A. lyrata is robust despite caveats that were previously mentioned for the study of environment dependence of load. It was argued that environment dependence of the expression of load depends on the strength of stress (Fox & Reed, 2011). In our study, exposure to warmer conditions strongly reduced WPC population performance, by up to 92%. Type and novelty of stress were also suggested to affect stress dependence of the expression of load (Sandner & Matthies, 2016; Willi et al., 2007a). If A. lyrata had regularly experienced conditions we considered stressful here across its evolutionary history, deleterious alleles expressed under these conditions may have already been purged (Bijlsma et al., 1999; Enders & Nunney, 2016; Hedrick, 1994). However, the expression of mutational load was also not enhanced by stress based on relative performance, a broader measure of stress likely depicting also novel stressors. The interaction between genomic load and stress could have been masked by outbreeding depression, previously reported to counteract heterosis under environmental stress such as drought (Prill et al., 2014). Outbreeding depression is expected to be stronger in crosses involving populations of highest divergence (Fenster & Galloway, 2000; Lynch, 1991; Oakley et al., 2015), although more recent studies find this relationship to be weak or absent in large data sets (Clo et al., 2021; Vasseur et al., 2019). In our study, partner populations came from the centres of geographic distribution and were unlikely highly divergent, given the shared post-glacial history and moderate climatic differences between populations. Furthermore, outbreeding depression was only found in few populations, mostly of the core of distribution, and it seemed unaffected by stress. The high performance gain, especially in populations with high load, also speaks in favour of an overwhelming positive effect of heterosis and

little action of outbreeding depression. For these reasons, we conclude that the absence of significant environment dependence of the expression of mutational load seems real.

Heterosis was nonetheless affected by stress in our study, especially warm stress. Heterosis increased under warmer conditions than at the sites of origin of populations (Figure 2). Perrier et al. (2020) previously linked patterns of heightened heterosis with increasing genomic load to a performance decline in WPC, with BPC being fairly constant across that gradient. Here, we observed that WPC and BPC performance generally declined with increasing warm stress, but less in BPC, as suggested by the pot-level analysis on performance. Lower sensitivity to stress in outbred plants could result from recessive deleterious mutations being frequently heterozygous and their better buffering under stress by, for example, heat shock proteins (Bergman & Siegal, 2003; Queitsch et al., 2002; Rutherford & Lindquist, 1998). Higher performance of heterotic hybrids under stress in A. thaliana has been linked to disruptions of stress response pathways in inbred parental lines (Korn et al., 2010; Miller et al., 2015; Yang et al., 2015). Previous research in the context of agriculture or genetic rescue has also provided many examples of heterosis being increased under stress (Blum, 2013; Frankham, 2015; Fujimoto et al., 2018). Here, the overall increase in heterosis under warm stress independent of genomic load suggests that most populations of A. lyrata suffer from a general heterozygote deficit whose expression is heat-sensitive.

Our results offer some important insights to the understanding of heterosis and its stress dependence. First, model selection showed that the effect of increasing stressfulness was not linear but better described by a positive accelerating relationship. This was the case both when stress was expressed as the difference in temperature regime between common garden and site of origin of populations during a critical part of the year or as the performance decline in a garden relative to the best garden for a population. Second, when stress was depicted by temperature difference, it was the unsigned temperature difference that depicted stress better. While cooler conditions led to relatively high plant performance, it was the warmer conditions that were stressful and that led to higher heterosis. Third, even though interactions were not significant between genomic load and stressfulness on heterosis, our study suggests that they may still be of some relevance. One indication is that model selection pointed to the inclusion of interactions being a better model predicting heterosis. Another indication is that pot-level performance revealed significant negative interactions between genomic load and the linear term of stress, suggesting that populations with high genomic load performed worse under increasingly warmer conditions than at the site of origin, independent of cross type.

Our study has two main implications beyond the study of heterosis. The first is the role of expressed mutational load for range limits. Populations expanding beyond range limits may suffer from a twofold genetic Allee effect (Luque et al., 2016). A first is linked to the process of range expansion itself: in the absence of strong environmental gradients, range expansion may lead to the accumulation

of high mutational load in populations at range limits (Gilbert et al., 2017; Peischl et al., 2013, 2015). Further range expansion originating from these range-edge populations may, therefore, lead to an even higher accumulation of mutational load beyond tolerable levels to maintain population demographic rates (Perrier et al., 2020). A second genetic Allee effect is due to a general heterozygote deficit expressed under stressful, particularly warm-stress conditions. However, this problem was not confirmed for range-edge populations at range limits. Northern and southern range-edge populations did not significantly differ in heterosis when transplanted at their respective range edge or beyond. Nonetheless, the stress-dependent genetic Allee effect could be triggered by extreme environmental events that may be more frequent beyond range limits and by longrange dispersal into less suitable habitats than those immediately beyond range limits. Future simulation studies on range dynamics should, therefore, consider the consequences of stress-dependent genetic Allee effects on range expansion.

The second implication of our study is that interpopulation outbreeding can have a generally positive effect in populations across the distribution of a species. A great majority of populations were found to benefit from outbreeding, especially under warmer conditions than what the populations had experienced in the past. Such heterotic effects can often be beneficial beyond the first generation (Frankham, 2016; Willi et al., 2007b). Our results, therefore, motivate the study of assisted gene flow in the context of climate warming to preserve species across the entire range. Such management, potentially combined with a climate-adjusted provenancing strategy (Prober et al., 2015), may be promising particularly for weak dispersers. Populations at the southern range limits, with high genomic load, should be supplemented with material from similar climatic conditions but with different geographic histories to maximize heterozygote advantage and ensure the local persistence of species.

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## CONFLICT OF INTEREST

The authors declare that there is no conflict of interest.

## AUTHOR CONTRIBUTIONS

All authors contributed to the study design. YW collected the seeds in the field, and AP and DS raised and crossed plants. AP analyzed the data and wrote the manuscript, with support by YW.

#### PEER REVIEW

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#### DATA AVAILABILITY STATEMENT

All data is stored in Dryad (https://doi.org/10.5061/dryad.cc2fq z642).

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#### REFERENCES

- Agrawal, A. F., & Whitlock, M. C. (2010). Environmental duress and epistasis: How does stress affect the strength of selection on new mutations. *Trends in Ecology and Evolution*, 25, 450–458. https://doi. org/10.1016/j.tree.2010.05.003
- Armbruster, P., & Reed, D. H. (2005). Inbreeding depression in benign and stressful environments. *Heredity*, 95, 235–242. https://doi. org/10.1038/sj.hdy.6800721
- Bates, D., Mächler, M., Bolker, B., & Walker, S. (2015). Fitting linear mixed-effects models using lme4. *Journal of Statistical Software*, 67, 1–48. https://doi.org/10.18637/jss.v067.i01
- Bergman, A., & Siegal, M. L. (2003). Evolutionary capacitance as a general feature of complex gene networks. *Nature*, 424, 549–552. https:// doi.org/10.1038/nature01765
- Bijlsma, R., Bundgaard, J., & Van Putten, W. F. (1999). Environmental dependence of inbreeding depression and purging in *Drosophila melanogaster. Journal of Evolutionary Biology*, 12, 1125–1137. https://doi. org/10.1046/j.1420-9101.1999.00113.x
- Blum, A. (2013). Heterosis, stress, and the environment: a possible road map towards the general improvement of crop yield. *Journal of Experimental Botany*, 64, 4829–4837. https://doi.org/10.1093/jxb/ ert289
- Brown, J. H. (1984). On the relationship between abundance and distribution of species. The American Naturalist, 124, 255–279. https://doi.org/10.1086/284267
- Cahill, A. E., Aiello-Lammens, M. E., Fisher-Reid, M. C., Hua, X., Karanewsky, C. J., Ryu, H. Y., Sbeglia, C. G., Spagnolo, F., Waldron, J. B., & Wien, J. J. (2014). Causes of warm-edge range limits: systematic review, proximate factors and implications for climate change. *Journal of Biogeography*, 41, 429–442. https://doi.org/10.1111/ jbi.12231

Caswell, H. (2001). Matrix population models, 2nd ed. Sinauer Associates. Cheptou, P. O., Berger, A., Blanchard, A., Collin, C., & Escarre,

J. (2000). The effect of drought stress on inbreeding

depression in four populations of the Mediterranean outcrossing plant *Crepis sancta* (Asteraceae). *Heredity*, *85*, 294–302. https://doi.org/10.1046/j.1365-2540.2000.00759.x

- Clo, J., Ronfort, J., & Gay, L. (2021). Fitness consequences of hybridization in a predominantly selfing species: insights into the role of dominance and epistatic incompatibilities. *Heredity*, 127, 393–400. https://doi.org/10.1038/s41437-021-00465-2
- Connallon, T., & Sgrò, C. M. (2018). In search of a general theory of species' range evolution. PLoS Biology, 16, e2006735. https://doi. org/10.1371/journal.pbio.2006735
- Elena, S. F., & de Visser, J. A. G. M. (2003). Environmental stress and the effects of mutation. *Journal of Biology*, *2*, 12. https://doi. org/10.1186/1475-4924-2-12
- Enders, L. S., & Nunney, L. (2016). Reduction in the cumulative effect of stress-induced inbreeding depression due to intragenerational purging in *Drosophila melanogaster*. *Heredity*, 116, 304–313. https:// doi.org/10.1038/hdy.2015.103
- Fenster, C. B., & Galloway, L. F. (2000). Population differentiation in an annual legume: Genetic architecture. *Evolution*, 54, 1157–1172. https://doi.org/10.1111/j.0014-3820.2000.tb00551.x
- Fick, S. E., & Hijmans, R. J. (2017). WorldClim 2: New 1-km spatial resolution climate surfaces for global land areas. *International Journal* of Climatology, 37, 4302–4315. https://doi.org/10.1002/joc.5086
- Fox, C. W., & Reed, D. H. (2011). Inbreeding depression increases with environmental stress: An experimental study and meta-analysis. *Evolution*, 65, 246–258. https://doi. org/10.1111/j.1558-5646.2010.01108.x
- Fox, J., & Weisberg, S. (2019). An R Companion to Applied Regression, 3rd ed. Sage.
- Frankham, R. (2015). Genetic rescue of small inbred populations: meta-analysis reveals large and consistent benefits of gene flow. *Molecular Ecology*, 24, 2610–2618. https://doi.org/10.1111/ mec.13139
- Frankham, R. (2016). Genetic rescue benefits persist to at least the F3 generation, based on a meta-analysis. *Biological Conservation*, 195, 33–36. https://doi.org/10.1016/j.biocon.2015.12.038
- Fujimoto, R., Uezono, K., Ishikura, S., Osabe, K., Peacock, W. J., & Dennis, E. S. (2018). Recent research on the mechanism of heterosis is important for crop and vegetable breeding systems. *Breeding Science*, 68, 145-158. https://doi.org/10.1270/ jsbbs.17155
- Gaston, K. (2009). Geographic range limits of species. *Proceedings of the Royal Society B: Biological Sciences*, 276, 1391–1393. https://doi. org/10.1098/rspb.2009.0100
- Gilbert, K. J., Sharp, N. P., Angert, A. L., Conte, G. L., Draghi, J. A., Guillaime, F., Hargreaves, A. L., Matthey Doret, R., & Whitlock, M. C. (2017). Local adaptation interacts with expansion load during range expansion: maladaptation reduces expansion load. *The American Naturalist*, 189, 368–380. https://doi. org/10.1086/690673
- González-Martínez, S. C., Ridout, K., & Pannell, J. R. (2017). Range expansion compromises adaptive evolution in an outcrossing plant. *Current Biology*, 27, 2544–2551. https://doi.org/10.1016/j. cub.2017.07.007
- Griffin, P. C., & Willi, Y. (2014). Evolutionary shifts to self-fertilisation restricted to geographic range margins in North American Arabidopsis lyrata. Ecology Letters, 17, 484–490. https://doi.org/10.1111/ ele.12248
- Hadfield, J. D. (2010). MCMC methods for multi-response generalized linear mixed models: the MCMCgImm R package. *Journal* of Statistical Softwares, 33, 1-22. https://doi.org/10.18637/jss. v033.i02
- Hadfield, J. D. (2019). MCMCgImm course notes. [WWW document] URL https://cran.r-project.org/web/packages/MCMCgImm/vigne ttes/CourseNotes.pdf [accessed 24 June 2020]

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- Hampe, A., & Petit, R. J. (2005). Conserving biodiversity under climate change: the rear edge matters. *Ecology Letters*, 8, 461–467. https:// doi.org/10.1111/j.1461-0248.2005.00739.x
- Hargreaves, A. L., Samis, K. E., & Eckert, C. G. (2014). Are species' range limits simply niche limits writ large? A review of transplant experiments beyond the range. *The American Naturalist*, 183, 157–173. https://doi.org/10.1086/674525
- Hedrick, P. W. (1994). Purging inbreeding depression and the probability of extinction: full-sib mating. *Heredity*, 73, 363–372. https://doi. org/10.1038/hdy.1994.183
- Henry, R. C., Barton, K. A., & Travis, J. M. J. (2015). Mutation accumulation and the formation of range limits. *Biology Letters*, 11, 20140871. https://doi.org/10.1098/rsbl.2014.0871
- Hoffmann, A. A., & Hercus, M. J. (2000). Environmental stress as an evolutionary force. *BioScience*, 50, 217-226. https://doi. org/10.1641/0006-3568(2000)050[0217:ESAAEF]2.3.CO;2
- Kondrashov, A. S., & Houle, D. (1994). Genotype-environment interactions and the estimation of the genomic mutation rate in *Drosophila melanogaster*. *Proceedings of the Royal Society B: Biological Sciences*, 258, 221–227. https://doi.org/10.1098/rspb.1994.0166
- Korn, M., Gärtner, T., Erban, A., Kopka, J., Selbig, J., & Hincha, D. K. (2010). Predicting *Arabidopsis* freezing tolerance and heterosis in freezing tolerance from metabolite composition. *Molecular Plant*, 3, 224–235. https://doi.org/10.1093/mp/ssp105
- Koski, M. H., Layman, N. C., Prior, C. J., Busch, J. W., & Galloway, L. F. (2019). Selfing ability and drift load evolve with range expansion. *Evolution Letters*, 3, 500–512. https://doi.org/10.1002/evl3.136
- Kuznetsova, A., Brockhoff, P. B., & Christensen, R. H. B. (2017). ImerTest Package: tests in linear mixed-effects models. *Journal of Statistical Software*, 82, 1–26. https://doi.org/10.18637/jss.v082.i13
- Lee-Yaw, J. A., Fracassetti, M., & Willi, Y. (2018). Environmental marginality and geographic range limits: a case study with Arabidopsis lyrata ssp. lyrata. Ecography, 41, 622–634. https://doi.org/10.1111/ ecog.02869
- Lee-Yaw, J. A., Kharouba, H. M., Bontrager, M., Mahony, C., Csergő, A. M., Noreen, A. M., Li, Q., Schuster, R., & Angert, A. L. (2016). A synthesis of transplant experiments and ecological niche models suggests that range limits are often niche limits. *Ecology Letters*, 19, 710–722. https://doi.org/10.1111/ele.12604
- Lohmueller, K. E., Indap, A. R., Schmidt, S., Boyko, A. R., Hernandez, R. D., Hubisz, M. J., Sninsky, J. J., White, T. J., Sunyaev, S. R., Nielsen, R., Clark, A. G., & Bustamante, C. D. (2008). Proportionally more deleterious genetic variation in European than in African populations. *Nature*, 451, 994–998. https://doi.org/10.1038/nature06611
- Luque, G. M., Vayssade, C., Facon, B., Guillemaud, T., Courchamp, F., & Fauvergue, X. (2016). The genetic Allee effect: a unified framework for the genetics and demography of small populations. *Ecosphere*, 7, e01413. https://doi.org/10.1002/ecs2.1413
- Lynch, M. (1991). The genetic interpretation of inbreeding depression and outbreeding depression. *Evolution*, 45, 622–629. https://doi. org/10.1111/j.1558-5646.1991.tb04333.x
- Miller, M., Song, Q., Shi, X., Juenger, T. E., & Chen, Z. J. (2015). Natural variation in timing of stress-responsive gene expression predicts heterosis in intraspecific hybrids of Arabidopsis. Nature Communications, 6, 7453. https://doi.org/10.1038/ncomms8453
- Oakley, C. G., Ågren, J., & Schemske, D. W. (2015). Heterosis and outbreeding depression in crosses between natural populations of Arabidopsis thaliana. Heredity, 115, 73–82. https://doi.org/10.1038/ hdy.2015.18
- Peischl, S., Dupanloup, I., Foucal, A., Jomphe, M., Bruat, V., Grenier, J. C., Gouy, A., Gilbert, K. J., Gbeha, E., Bosshard, L., Hip-Ki, E., Agbessi, M., Hodgkinson, A., Vézina, H., Awadalla, P., & Excoffier, L. (2018). Relaxed selection during a recent human expansion. *Genetics*, 208, 763–777. https://doi.org/10.1534/genetics.117.300551
- Peischl, S., Dupanloup, I., Kirkpatrick, M., & Excoffier, L. (2013). On the accumulation of deleterious mutations during range expansions.

Molecular Ecology, 22, 5972-5982. https://doi.org/10.1111/ mec.12524

- Peischl, S., & Excoffier, L. (2015). Expansion load: recessive mutations and the role of standing genetic variation. *Molecular Ecology*, 24, 2084–2094. https://doi.org/10.1111/mec.13154
- Peischl, S., Kirkpatrick, M., & Excoffier, L. (2015). Expansion load and the evolutionary dynamics of a species range. *The American Naturalist*, 185, E81–E93. https://doi.org/10.1086/680220
- Perrier, A., Sánchez-Castro, D., & Willi, Y. (2020). Expressed mutational load increases toward the edge of a species' geographic range. *Evolution*, 74, 1711–1723. https://doi.org/10.1111/evo.14042
- Pironon, S., Papuga, G., Villellas, J., Angert, A. L., García, M. B., & Thompson, J. D. (2017). Geographic variation in genetic and demographic performance: new insights from an old biogeographical paradigm. *Biological Reviews*, 92, 1877–1909. https://doi.org/10.1111/ brv.12313
- Prill, N., Bullock, J. M., van Dam, N. M., & Leimu, R. (2014). Loss of heterosis and family-dependent inbreeding depression in plant performance and resistance against multiple herbivores under drought stress. *Journal of Ecology*, 102, 1497–1505. https://doi. org/10.1111/1365-2745.12327
- Prober, S. M., Byrne, M., McLean, E. H., Steane, D. A., Potts, B. M., Vaillancourt, R. E., & Stock, W. D. (2015). Climate-adjusted provenancing: a strategy for climate-resilient ecological restoration. *Frontiers in Ecology and Evolution*, *3*, 65. https://doi.org/10.3389/ fevo.2015.00065
- Queitsch, C., Sangster, T. A., & Lindquist, S. (2002). Hsp90 as a capacitor of phenotypic variation. *Nature*, 417, 618–624. https://doi. org/10.1038/nature749
- R Core Team. (2021). R: a language and environment for statistical computing. R Foundation for Statistical Computing.
- Reed, D. H., Fox, C. W., Enders, L. S., & Kristensen, T. N. (2012). Inbreeding-stress interactions: evolutionary and conservation consequences. Annals of the New York Academy of Sciences, 1256, 33-48. https://doi.org/10.1111/j.1749-6632.2012.06548.x
- Roff, D. A. (1997). Evolutionary Quantitative Genetics. Chapman and Hall.
- Roy, K., Hunt, G., Jablonski, D., Krug, A. Z., & Valentine, J. W. (2009). A macroevolutionary perspective on species range limits. *Proceedings* of the Royal Society B: Biological Sciences, 276, 1485–1493. https:// doi.org/10.1098/rspb.2008.1232
- Rutherford, S. L., & Lindquist, S. (1998). Hsp90 as a capacitor for morphological evolution. *Nature*, 396, 336–342. https://doi. org/10.1038/24550
- Rutter, M. T., & Fenster, C. B. (2007). Testing for adaptation to climate in Arabidopsis thaliana: A calibrated common garden approach. Annals of Botany, 99, 529–536. https://doi.org/10.1093/aob/mcl282
- Sandner, T. M., & Matthies, D. (2016). The effects of stress intensity and stress type on inbreeding depression in *Silene vulgaris*. *Evolution*, 70, 1225–1238. https://doi.org/10.1111/evo.12929
- Sexton, J. P., McIntyre, P. J., Angert, A. L., & Rice, K. J. (2009). Evolution and ecology of species range limits. Annual Review of Ecology, Evolution, and Systematics, 40, 415–436. https://doi.org/10.1146/ annurev.ecolsys.110308.120317
- Sugiura, N. (1978). Further analysts of the data by Akaike's information criterion and the finite corrections. *Communications in Statistics*Theory and Methods, 7, 13–26. https://doi.org/10.1080/03610 927808827599
- Vasseur, F., Fouqueau, L., De Vienne, D., Nidelet, T., Violle, C., & Weigel, D. (2019). Nonlinear phenotypic variation uncovers the emergence of heterosis in Arabidopsis thaliana. PLoS Biology, 17, e3000214. https://doi.org/10.1371/journal.pbio.3000214
- Vermeulen, C. J., & Bijlsma, R. (2004). Characterization of conditionally expressed mutants affecting age-specific survival in inbred lines of *Drosophila melanogaster*: lethal conditions and temperature-sensitive periods. *Genetics*, 167, 1241–1248. https:// doi.org/10.1534/genetics.103.023721

- Waller, D. M., Dole, J., & Bersch, A. J. (2008). Effects of stress and phenotypic variation on inbreeding depression in *Brassica rapa*. Evolution, 62, 917-931. https://doi.org/10.1111/j.1558-5646.2008.00325.x
- Willi, Y. (2019). The relevance of mutation load for species range limits. American Journal of Botany, 106, 757–759. https://doi.org/10.1002/ ajb2.1296
- Willi, Y., Dietrich, S., van Kleunen, M., & Fischer, M. (2007a). Inter-specific competitive stress does not affect the magnitude of inbreeding depression. Evolutionary Ecology Research, 9, 959–974.
- Willi, Y., Fracassetti, M., Zoller, S., & Van Buskirk, J. (2018). Accumulation of mutational load at the edges of a species range. *Molecular Biology and Evolution*, 35, 781–791. https://doi.org/10.1093/molbe v/msy003
- Willi, Y., & Van Buskirk, J. (2019). A practical guide to the study of distribution limits. The American Naturalist, 193, 773–785. https://doi. org/10.1086/703172
- Willi, Y., van Kleunen, M., Dietrich, S., & Fischer, M. (2007b). Genetic rescue persists beyond first-generation outbreeding in small populations of a rare plant. Proceedings of the Royal Society B: Biological Sciences, 274, 2357–2364. https://doi.org/10.1098/ rspb.2007.0768

Yang, L., Li, B., Zheng, X. Y., Li, J., Yang, M., Dong, X., He, G., An, C., & Deng, X. W. (2015). Salicylic acid biosynthesis is enhanced JOURNAL OF Evolutionary Biology 。へく 🕮

and contributes to increased biotrophic pathogen resistance in *Arabidopsis* hybrids. *Nature Communications*, *6*, 7309. https://doi.org/10.1038/ncomms8309

Zhang, M., Zhou, L., Bawa, R., Suren, H., & Holliday, J. A. (2016). Recombination rate variation, hitchhiking, and demographic history shape deleterious load in poplar. *Molecular Biology and Evolution*, *33*, 2899–2910. https://doi.org/10.1093/molbev/msw169

## SUPPORTING INFORMATION

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