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OPEN Response of Atlantic salmon *Salmo* salar to temperature and dissolved oxygen extremes established using animal-borne environmental sensors

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Understanding how aquatic species respond to extremes of DO and temperature is crucial for determining how they will be affected by climate change, which is predicted to increasingly expose them to levels beyond their optima. In this study we used novel animal-borne DO, temperature and depth sensors to determine the effect of extremes of DO and temperature on the vertical habitat use of Atlantic salmon Salmo salar in aquaculture cages. Salmon showed a preference for temperatures around 16.5 to 17.5 °C, however, selection of preferred temperatures was trumped by active avoidance of low DO (<35% saturation) at the bottom of the cage. In addition to low DO, salmon also avoided warm surface waters (>20.1 °C), which led to a considerable contraction in the available vertical habitat. Despite their avoidance behavior, fish spent a large amount of time in waters with suboptimal DO (<60% saturation). These results show that vertical habitat contraction could likely be a significant consequence of climate change if the reduction in DO outpaces the increase in hypoxia tolerance through local adaptation. They furthermore highlight that site-specific environmental conditions and stock-specific tolerance thresholds may need to be considered when determining stocking densities.

Oxygen is the final receptor in the electron transport chain underlying the aerobic respiration of most organisms¹. Hence, dissolved oxygen (DO) is a crucial and potentially limiting component of the metabolism of marine animals, due to the low availability and high cost of extraction of oxygen in aquatic environments^{2, 3}. Temperature is the most important environmental factor governing the metabolic rate in ectotherms⁴ and the effects of DO and temperature on marine fish are intrinsically linked as higher temperatures increase the metabolic rate and consequently oxygen demand but also lower the solubility of oxygen, reducing supply⁵. Together, these two environmental variables have a profound effect on the ecology of marine species, from the growth and health of individuals⁶ to the spatial distribution of populations and species⁵.

Understanding how marine fish respond to extremes of DO and temperature, both in physiology and behavior, is crucial for predicting how they will be affected by climate change, which is likely to increasingly expose them to levels beyond their optimal range, reducing the productivity of both wild and aquaculture populations. Salmonids are one of the families of species of particular concern, due to their commercial importance both as wild caught and aquaculture species⁶ and their suspected vulnerability to environmental change (see Jonsson and Jonsson⁹ for review), with increased mortality rates in some populations already attributed to warming temperatures¹⁰.

A plethora of laboratory studies have tested the effects of different temperatures and DO levels on the physiology of salmonids with the aim of establishing optimal ranges and tolerance to extremes (e.g. Anttila, et al. 11, Remen, et al.¹², Remen, et al.¹³, Kutty and Saunders¹⁴, Javaid and Anderson¹⁵, Barnes, et al.¹⁶). However, the use of those studies in predicting the behavioral response to extremes of DO and temperature in the wild is limited, as the range of responses available to fish and complex interactions of competing environmental influences cannot be replicated in the laboratory¹. Particularly the movement behavioral response of salmonids to environmental

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Fish ID	Cage	Tagging date	Date of last detection (fish alive and tag attached)	Data series duration (days)	Final outcome
1	1	14-12-2015	15-12-2015	<1	Mortality
2	1	14-12-2015	02-03-2016	78	Harvest
3	1	14-12-2015	04-01-2016	21	Tag loss
4	1	14-12-2015	25-12-2015	10	Tag failure
5	1	14-12-2015	15-02-2016	63	Mortality
6	1	14-12-2015	13-03-2016	89	Harvest
7	1	14-12-2015	13-03-2016	89	Harvest
8	1	14-12-2015	13-03-2016	89	Harvest
9	1	14-12-2015	13-03-2016	89	Harvest
10	1	14-12-2015	06-03-2016	83	Mortality
11	1	14-12-2015	21-01-2016	38	Tag failure
12	1	14-12-2015	23-02-2016	70	Harvest
13	1	14-12-2015	13-03-2016	89	Harvest
14	1	14-12-2015	13-03-2016	89	Harvest
15	2	14-12-2015	14-12-2015	<1	Mortality
16	2	14-12-2015	03-01-2016	19	Mortality
17	2	14-12-2015	09-03-2016	86	Harvest
18	2	14-12-2015	15-12-2015	1	Mortality
19	2	14-12-2015	13-03-2016	89	Harvest
20	2	14-12-2015	14-12-2015	<1	Mortality
21	2	14-12-2015	14-12-2015	<1	Mortality
22	2	14-12-2015	24-01-2016	41	Mortality
23	2	14-12-2015	17-02-2016	65	Harvest
24	2	14-12-2015	14-12-2015	<1	Mortality
25	2	14-12-2015	17-02-2016	65	Harvest
26	2	14-12-2015	10-01-2016	26	Mortality
27	2	14-12-2015	21-12-2015	7	Tag loss

Table 1. Overview of individual detection series duration. Fish removed from analysis are highlighted in grey. Final outcomes other than harvest were established based on the animals' depth record. Tags abruptly rising to the surface and remaining there were classed as lost tags, tags sinking to the bottom of the cage and remaining there were classed as mortalities, tags abruptly ending transmissions were classed as tag failures.

extremes is difficult to replicate in space-limited laboratory settings, however, understanding it is crucial, as any possible avoidance behavior is likely to alter other important ecological variables such as predation risk and feeding competition¹.

Studies of the response of salmonids to extreme DO and temperature values in the field have been carried out using echosounders to determine swimming depth of salmon in aquaculture cages in concurrence with high resolution environmental sensors^{17–19}. However, these studies can only measure group level rather than individual responses and only over relatively short timespans (days to weeks) and have not been able to resolve the question of whether salmon actively avoid extreme DO and temperature levels in the field²⁰. Biotelemetry methods on the other hand, which deploy environmental sensors on animals, allow the collection of individual level data over longer time spans (months). While more intrusive than echosounders, biotelemetry tags have been used successfully to measure DO²¹ and temperature values²² experienced by free-swimming salmonids.

The aim of this study was to use newly developed acoustic telemetry tags which record swimming depth and environmental DO and temperature experienced by Atlantic salmon *Salmo salar* in aquaculture cages in Macquarie Harbour, Tasmania, Australia to determine the behavioral response of salmon to seasonal extremes of DO and temperature during the hottest summer on record in Tasmania²³ and identify the trade-offs between the

	Cage 1	Cage 2
Number of fish	13	7
Mean swimming depth (m)	5.99 (2.30)	4.90 (1.94)
Mean DO experienced by fish (% sat)	60.80 (18.51)	67.12 (20.34)
Mean DO measured in cage (% sat)	50.58 (34.02)	49.98 (33.86)
Mean temperature experienced by fish (°C)	16.80 (1.51)	17.54 (1.43)
Mean temperature measured in cage (°C)	16.16 (2.38)	16.12 (2.41)

Table 2. Data overview. Standard deviations shown in brackets. To account for the fact that some stationary string tags malfunctioned during parts of the study period, cage water column data were grouped into 2.5 m depth bins and each data point weighted by the inverse of the total number of observations for its depth bin for the calculation of means and standard deviations.

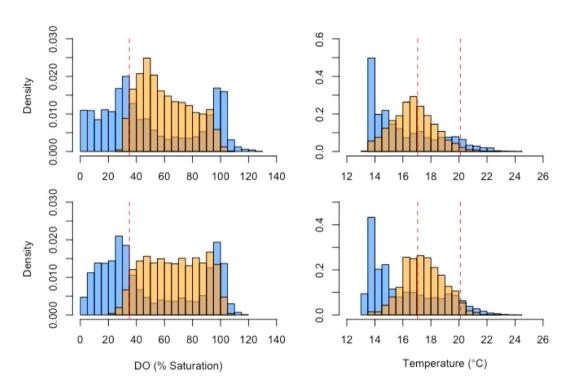


Figure 1. Distribution of DO and temperature values experienced by the fish and measured inside the cages. Values measured by fish are displayed in orange, those measured throughout the cage water column in blue. Data from cage 1 is shown in the top panel, data from cage 2 is shown in the bottom panel. Vertical dashed lines indicate the 35% DO saturation threshold in the DO distributions and the mean and 97.5 centile in the temperature distributions. To account for the fact that some stationary string tags malfunctioned during parts of the study period, water column data were grouped into 2.5 m depth bins and each data point weighted by the inverse of the total number of observations for its depth bin for the computation of density distributions.

opposing influences of the two environmental parameters. Macquarie Harbour presents an ideal site for this study as it is highly stratified with a high DO, high temperature top layer and low DO, low temperature bottom layer²⁴.

Results

Data overview. Initial processing resulted in the removal of 7 tagged fish, 6 mortalities and 1 fish that lost its tag, with the majority of mortalities occurring in cage 2. All 6 mortalities occurred within 48 hours of surgery, suggesting that handling stress rather than longer-term effects of the tag attachments were the cause (Table 1). A further 5 animals died during the study, however, detection series were deemed sufficiently long to be included in analyses. While the remaining 20 fish represent only a small fraction of the total stock in each pen, previous studies have shown that individual response to environmental factors was well correlated with average group behavior of farmed salmon both in the short and long term²⁰. Mean DO and temperature was similar in the two cages and in both cages mean DO and temperature values experienced by the fish were higher than those measured throughout the cage (Table 2). Fish in cage 1 swam on average approximately 1 m deeper than fish in cage 2 and consequently experienced cooler mean temperatures and lower DO values than fish in cage 2. In both cages fish spent a considerable amount of time in waters with DO values between 35 and 60% (Fig. 1) and particularly in cage 2 the distribution of DO values experienced by the fish resembled a uniform distribution with a lower

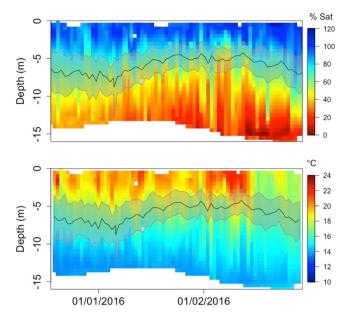


Figure 2. Profile of DO and temperature in experimental cage 1 during the study period. Black line indicates mean swimming depth averaged over all tagged fish in cage 1. Grey area shows \pm 1 SD.

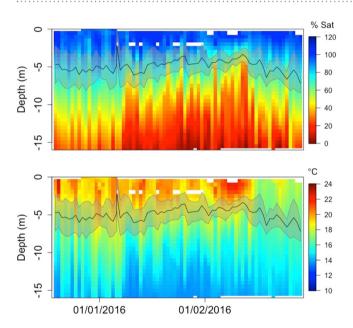


Figure 3. Profile of DO and temperature in experimental cage 2 during the study period. Black line indicates mean swimming depth averaged over all tagged fish in cage 2. Grey area shows \pm 1 SD.

bound of 30 to 40% and a higher bound of 95 to 100% whereas the temperatures experienced by the fish were approximately normally distributed, centered around a mean of 16.5 to 17.5 °C (Fig. 1). Maximum temperatures experienced by the fish $(23.0\,^{\circ}\text{C})$ were lower than those measured throughout the cages $(24.2\,^{\circ}\text{C})$.

Values measured by fish are displayed in orange, those measured throughout the cage water column in blue. Data from cage 1 is shown in the top panel, data from cage 2 is shown in the bottom panel. Vertical dashed lines indicate the 35% DO saturation threshold in the DO distributions and the mean and 97.5 centile in the temperature distributions. To account for the fact that some stationary string tags malfunctioned during parts of the study period, water column data were grouped into 2.5 m depth bins and each data point weighted by the inverse of the total number of observations for its depth bin for the computation of density distributions.

Seasonal patterns. Over the course of the study period, DO and temperature underwent strong temporal changes, with DO values in the 5–10 m depth range decreasing, surface water temperatures increasing and warm surface water temperatures penetrating deeper into the water column as summer progressed (Figs 2 and 3). Fish also showed a temporal pattern in their swimming depth, with both mean depth and standard deviation

	Parameter estimate	t-value	p		
Preference model					
Intercept	-5.672 (0.255)	-22.278	< 0.001		
Depth of preferred temperature layer	-0.033 (0.018)	-1.884	0.06		
Avoidance model					
Intercept	-2.659 (0.311)	-8.560	< 0.001		
Depth of low DO threshold	0.248 (0.014)	18.245	< 0.001		
Depth of high temperature threshold (1–2 m)	-0.101 (0.264)	-0.386	0.700 (0.649)		
Depth of high temperature threshold (2–3 m)	0.073 (0.266)	0.276	0.783 (0.649)		
Depth of high temperature threshold (>3 m)	-2.601 (0.566)	-4.593	< 0.001 (0.649)		
Depth of high temperature threshold (1–2 m) *depth of low DO threshold	-0.003 (0.025)	-0.119	0.905 (<0.001)		
Depth of high temperature threshold (2–3 m) *depth of low DO threshold	0.005 (0.025)	0.204	0.838 (<0.001)		
Depth of high temperature threshold (>3 m) *depth of low DO threshold	-0.241 (0.052)	-4.613	<0.001 (<0.001)		

Table 3. Model parameters for the preference and avoidance models. Standard error of parameter estimates given in brackets. Overall p-value of categorical variable given in brackets behind level specific p-values. For both models the inclusion of a temporal autocorrelation structure increased the AIC by more than 2, hence the final model was fitted without the correlation structure.

decreasing over the course of summer, reaching a minimum during the peak of summer (end of January/beginning of February) and increasing again towards the end of summer. Seasonal changes in swimming depth were more pronounced in cage 1 (Fig. 2) than cage 2 (Fig. 3), which showed an overall shallower mean swimming depth. However, the seasonal changes in standard deviation were consistent across the two cages.

Swimming behavior models. The linear mixed models of salmon vertical movement behaviour tested two opposing hypothesis: The preference hypothesis which stipulates that salmon position themselves at or near their optimum temperature, assumed to be 17 °C and the avoidance hypothesis which stipulates that salmon position themselves to avoid both harmfully low DO (<35%) and excessive surface temperatures (>20.1 °C). The preference model results showed that the observed changes in 6 hourly swimming depths could not be successfully explained by the preference hypothesis, with the depth of the 17 °C layer having no significant effect on fish swimming depth (Table 3). The avoidance model on the other hand, showed that depth of the 35% DO layer had a significant effect on fish swimming depth, which was mediated by the effect of the maximum depth of the 20.1 °C surface layer, as indicated by the significance of the interaction term (Table 3). This significance of the interaction term in the model was predominantly driven by a significantly different response to the 35% DO level when the warm surface layer extended below 3 m (Table 3). When the warm surface layer did not extend below 3 m depth, fish moved up in the water column when the depth of the 35% DO threshold decreased. When the warm surface layer extended below 3 m, however, the response of the fish was depressed and the change in swimming depth was minimal (Fig. 4). The variance of the random intercept, which accounts for individual variation between fish was 1.417.

Discussion

In this deployment of novel salmon-borne DO, temperature and depth sensors we showed for the first time that salmon in aquaculture cages adjusted their vertical positioning to actively avoid both low DO levels in the bottom layer and extreme surface temperatures. Tagged fish showed a gradual upward movement and reduction in depth range in response to rising low-DO waters which was suppressed by warm surface waters penetrating deeper into the water column at the peak of summer, resulting in the concentration of the salmon stock in the 4–6 m midlayer. Our results suggest that this avoidance behavior overrides the selection of optimal temperatures in the water column.

While the tagged fish only represent a small proportion of the total stock in the two experimental cages and an effect of the tags on salmon behavior and fitness cannot be discounted as the elevated mortality of tagged fish indicates, the fact that the observed patterns were consistent across cages and fish, both those that died during the study and those harvested at the end makes it likely that they are representative of the broader stock. Additionally, the observed concentration of fish in a small band in the midlayer of the cage is consistent with the results from a short term study on salmon swimming depth in Macquarie Harbour at the peak of summer using non-invasive echosounders¹⁹.

It is well established that like most aquatic animals, fish have the sensory capacity to detect and avoid waters with low oxygen levels²⁵ with a number of species exhibiting horizontal²⁶ or vertical avoidance behavior^{27, 28} even at the larval stage²⁹. Electronic tagging studies on wild salmon have provided circumstantial evidence that both Chinook *Oncorhynchus tshawytscha*³⁰ and Atlantic salmon²¹ may alter their migratory behavior to avoid waters with DO lower than 55% saturation. More recent studies on the vertical distribution of Atlantic salmon in aquaculture cages on the other hand have been inconclusive, with one study indicating avoidance of waters at or below 60%¹⁷ and two others finding no evidence of avoidance behavior in salmon with regards to low DO levels^{18, 31}. Johansson, *et al.*¹⁸ concluded that the salmons' preference for their optimum temperature overrides their response to potentially harmfully low DO levels, which was confirmed by a later experimental study, which showed an avoidance response of salmon to short-term drops in DO to intermediate levels (~60% saturation) but suggested that this was restricted to environments where more suitable DO levels are available within the animals 'preferred temperature band³². However, fish in neither of these studies experienced DO levels below 57–60%, which is higher than the lowest levels observed in our study (21–110% saturation) and the study on wild salmon²¹.

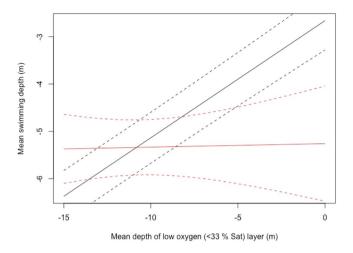


Figure 4. Model prediction of fish response to changes in the minimum depth of the low DO bottom layer mediated by the maximum depth of the high temperature surface layer. Black line indicates response to low DO layer depth when the high temperature surface layer was above 1 m depth, red line indicates response when the high temperature surface layer extended below 3 m. Dashed lines show 95% confidence interval.

Our study showed that avoidance of DO levels below 35% was a better predictor of salmon swimming depth than preference for optimal temperatures, although, the two models we tested are not necessarily mutually exclusive and the salmon in our study may switch to the temperature preference behavior observed by Johansson, *et al.*¹⁸ during times when DO levels are above 60% throughout the cage. Even though we did not quantify a specific DO threshold that triggered avoidance behavior in this study, it is clear from our results that the threshold falls below what is reported as the onset of harmful conditions in the literature. Salmon are generally considered to be a hypoxia sensitive species with high oxygen requirements to meet high routine metabolic demands and comparative studies have shown that Atlantic salmon have higher critical oxygen levels for swimming performance than rainbow trout *Salmo gairdneri*, goldfish *Carassius auratus* and *Tilapia* Moreover, growth of Atlantic salmon has been reported to be reduced below a threshold of 70%, hypoxia stress sets in below 60% as indicated by increases in plasma lactate levels in munune response is depressed at or below 52% and 39% has been suggested as the critical oxygen saturation and have to rely on anaerobic ATP production in the literature of salmon salmon are no longer able to maintain oxygen uptake rates through increased gill ventilation and have to rely on anaerobic ATP production in the levels in the laterature of salmon salmon are no longer able to maintain oxygen uptake rates through increased gill ventilation and have to rely on anaerobic ATP production in the levels in the laterature of salmon are no longer able to maintain oxygen uptake rates through increased gill ventilation and have to rely on anaerobic ATP production in the levels in the laterature of salmon are no longer able to maintain oxygen uptake rates through increased gill ventilation and have to rely on anaerobic ATP production in the levels in the laterature of salmon are no longer ab

While we were not able to assess growth in the tagged fish, the fact that they experienced and survived in DO values below these thresholds suggests that they may be able to cope with lower DO levels either due to acclimation or genetic adaptation. Remen, *et al.*¹² found no effect of acclimation on hypoxia tolerance. However, the periodic drops in DO simulated in their study were different to the chronic, gradual decrease in DO experienced by the salmon in Macquarie Harbor and a separate study showed that while the effect of hypoxia on the metabolic rate persisted, the initial stress response was down-regulated, suggesting that fish might become desensitized over time¹³. Similar to our study, Barnes, *et al.*¹⁶ found a higher degree of hypoxia tolerance in the Tasmanian aquaculture population in a laboratory study in freshwater and suggested as one possible explanation that a preconditioning event might have occurred since experimental fish had been exposed to low DO levels prior to procurement for the experiment. The authors also mention that the Tasmanian stock might have undergone genetic mutation³⁵ as another potential reason for its increased hypoxia tolerance. This hypothesis is supported by a study of the variation in temperature and hypoxia tolerance among families of Atlantic salmon, which indicated substantial genetic variation in these traits¹¹. This genetic variation could potentially allow for selective breeding in aquaculture or adaptation in the wild to respond to global warming¹¹.

Deutsch, et al.⁷ modeled the changes in metabolically viable habitat, where environmental oxygen supply exceeds the resting metabolic demand and predicted that marine species will have to adapt to a steep overall decline and vertical contraction in suitable habitat on a global scale. On a local scale, our study shows the effects temperature and DO extremes can have on salmon in aquaculture with low bottom DO and high surface temperatures severely limiting the suitable vertical habitat. During the peak of summer, a layer of approximately only 2 m of the 17 m cage depth was available to the fish. This means that schooling densities might have been up to 8.5 times higher than theoretical stocking densities calculated based on the full cage volume. This could potentially lead to overcrowding and increased stress, which can result in heightened susceptibility to disease and suppressed cytokine expression²⁰. Hence, based on a review of the environmental drivers of salmon behavior in aquaculture cages, Oppedal, et al.²⁰ suggested site-specific stocking densities guided by prevailing environmental conditions.

Conclusions

The results of the deployment of animal-borne, combined DO, temperature and depth sensors on Atlantic salmon in aquaculture showed that salmon actively avoided both low DO levels near the bottom of the aquaculture cage and warm surface water temperatures. This led to a considerable contraction in the vertical habitat available to them and potential overcrowding, with expected implications for fish welfare. Despite the avoidance behavior,

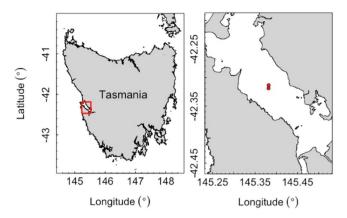


Figure 5. Map of experimental salmon cage locations in Macquarie Harbour on the west coast of Tasmania. Red rectangle in full-scale map of Tasmania on the left indicates extent of detailed map of Macquarie Harbour on the right. Red dots indicate experimental cage locations. Map was created in R version 3.3.1⁴⁴, using the 'PBSmapping' package version 2.69.76⁴⁸.

fish spent a large amount of time in waters with suboptimal DO levels, suggesting that the hypoxia tolerance in the Tasmanian population might be higher than that reported in the literature for other stocks. These results indicate that site-specific environmental conditions need to be taken into account when determining stocking densities and that vertical habitat contraction could likely play a significant part in the effects of climate change on salmon populations.

Methods

Study site and stock. Macquarie Harbour is a large (~280 km²) estuarine inlet on the west coast of Tasmania, Australia (Fig. 5). Depths range from 5–20 m in most parts of the harbour with the exception of a deeper basin at the center, which reaches depths of up to 50 m²4. Exchange of water with the Southern Ocean is limited to a shallow and narrow entrance at the North-western end of the harbour, freshwater input is from two main sources, the Gordon and King rivers which both carry high concentrations of dissolved organic matter²4. Due to limited vertical mixing, waters in the harbour are highly stratified with a persistent, brackish top layer, a middle layer with low DO and intermediate salinity and a bottom layer of seawater. While the limited water exchange and vertical mixing and high load of organic matter makes the harbour a DO depleted environment, with seasonal fluctuations causing challenging conditions for salmon growth, the brackish surface layer has made the harbour attractive for salmon farming as it limits the impact of amoebic gill disease on salmon stocks, which is a problem in other parts of the state³6,37.

Salmonid farming in the harbour started in the 1980s and has greatly increased since then with a major expansion in 2012 and farm leases now cover a total area of 9.26 km^{2.38}. Whilst both ocean rainbow trout *Oncorhynchus mykiss* and Atlantic salmon are farmed the majority of stock is Atlantic salmon. Tasmanian Atlantic salmon stocks originated from the Phillip River, Nova Scotia, Canada from where they were first imported to a hatchery on the Australian mainland in the 1960s³⁹. In the early 1980s the first fertilized eggs were then brought to Tasmania and the first fish transferred into ocean cages in 1985. Since the first import of ova to Australia in the 1960s new alleles have been observed in the Australian stock, suggesting the occurrence of genetic mutations³⁵ and possible adaptation to local environmental conditions¹⁶.

Study cages. The study was carried out between the 17th of December 2015 and the 29th of February 2016 in two cages owned by Huon Aquaculture, approximately 600 m apart, located in the central harbour (Fig. 5). Total depth at the two sites was between 40 and 45 m, cage depth was approximately 17 m, cage circumference was 160 m. Both cages were stocked with triploid Atlantic salmon hatched at Huon's Millybrook Hatchery and moved into the sea cages in Macquarie Harbour in July/August the previous year (2014) and scheduled for harvest at the end of the study period (March 2016). Total number of fish in cage 1 was 36512 with an average weight at time of tagging of 3625 g, total number of fish in cage 2 was 34270 with an average weight at time of tagging of 4175 g.

Tag deployment. To determine the behavioral response of salmon to changes in environmental conditions we deployed VEMCO acoustic telemetered sensors $(17.5 \times 106.5 \, \mathrm{mm}; 19.78 \, \mathrm{g})$ weight in seawater; ~160 db power output, 69 kHz frequency; VEMCO, Nova Scotia, Canada), which measure pressure $(0-102 \, \mathrm{m})$, temperature $(0-25 \, ^{\circ}\mathrm{C})$ and dissolved oxygen (0-140%) approximately simultaneously (within 250 milliseconds of one another) and transmit the resulting data together with the tags' ID coded as a series of acoustic pings (one ping series per environmental variable) which were recorded together with the date and time of transmission by an acoustic receiver deployed at the center, bottom of the cages, pointing up. To reduce the probability of signals from different tags colliding, transmission intervals were randomized within a set range. The DO sensor integrated in the tag was the optical RedEye oxygen sensor patch (Ocean Optics,, Dunedin, FL, USA) which has an accuracy range of 0.01 to 0.1% and a response time of less than 20 seconds. Since the tags did not measure salinity, DO was transmitted as % saturation rather than mg/L.

Tags were deployed both as strings of stationary sensors, attached at fixed depths to a rope at the center of the cages from the surface to the bottom and on fish. Stationary tags were deployed at 1, 3, 5, 7, 9, 12 and 15 m with one additional tag deployed below the cages at 20 m. These tags had a transmission delay of 800-1000 seconds, which means that measurements of the three variables were taken every 40-50 minutes.

Fish tags were deployed on 13 and 14 fish in the two cages and had a transmission delay of 180-260 seconds, which means measurements were taken every 9-13 minutes. This represents only a small fraction of the total stock in the cages, due to cost and receiver bandwidth constraints, however, Johansson, et al. 40 found that both short and long-term average individual behavior measured using data storage tags correlated well with overall group behavior. Tags were attached to the salmon using the method developed by Lacroix²², whereby the tag is secured to the fish via two monofilament tethers which form a lateral loop through the fish, directly underneath the pterygiophores. To ensure the tag was hovering just above the fish, it was made slightly positively buoyant by attaching a small cylindrical float to the end of it, replicating the hydrodynamics of a pop-up satellite archival tag, for which the method was designed²². Fish were brought to the surface using a crowd net and subsequently collected from the cage by farm staff using dip nets and immediately placed in an anesthetic bath containing Aqui-S (active ingredient Isoeugenol, Aqui-S, New Zealand). Once fish were sedate, they were moved to a padded cradle for tagging. During the tagging procedure, seawater containing a low dose of Aqui-S was continuously pumped over their gills to facilitate breathing and maintain anesthesia. Upon completion of the tagging procedure, fish were marked with Alcian Bue dye⁴¹ using a needleless inoculator (Dermo-jet Model G, Robbins Instruments, New Jersey, USA) for identification of tagged fish during harvest in the case of tag loss and placed in a recovery bath. Only tagged fish that fully regained balance and responded to manual agitation after recovery were released back into the cage. At the end of the study period, tags were recovered from fish during routine harvest operations and harvested tagged fish were deemed to be in good body condition (personal observation).

Animal ethics. Prior to the field deployment, tagging procedures and tag attachment method were trialed on 6 fish kept in a tank and monitored for 3 months after tagging. Fish were monitored and fed daily and physical condition assessed by the University veterinarian every two weeks. No difference in swimming or feeding behavior was observed between tagged fish and 6 control fish kept in the same tank. After a final physical and histological assessment by the veterinarian at the end of the trial, all methods were approved by the University of Tasmania animal ethics committee (Approval No. A0014977) and all procedures during both tagging trial and field deployment were performed in accordance with relevant guidelines and regulations.

Data processing. Prior to analyses, any data that might contain abnormal behavior was removed by excluding any fish that were detected and alive for less than 7 full days after tagging. Additionally, the first 24 hours after tagging were removed to exclude any potential effects of tagging and handling stress⁴² and for fish that were identified as mortalities, the last 24 hours before the time of death (the point where the depth measured by the tag remained constant at approximately the depth of the cage) were removed.

Following data cleanup, DO and temperature profiles were interpolated for the two cages from both fish and sensor string data on a 6 hours by 0.5 m grid using linear bivariate interpolation for irregularly distributed data, developed by Akima⁴³ and implemented in R⁴⁴ in the 'akima' package⁴⁵.

Data analysis. To determine the effect of DO and temperature on fish behavior we formulated two linear mixed models representing two contrasting hypotheses. The first model, termed the 'preference model' hypothesized that fish position themselves in the water column based on their preferred temperature¹⁸. The preferred temperature was hereby defined as the mean of all temperature values measured by the fish tags (17.05 °C; Fig. 1), which corresponds well with laboratory studies on the temperature preferences of the species^{15, 18}. Consequently, the model used the 6 hourly depth of the 17.05 °C temperature layer derived from the interpolated temperature profiles as the fixed effect, fish ID as the random effect and 6 hourly mean swimming depth of individual fish as the dependent variable. Cage ID was included as a fixed effect in the original model but was not significant (α =0.05) and therefore removed from the final version of the model. Mean swimming depth rather than raw values were used as the dependent variable to match the resolution of temperature and DO profiles and reduce temporal autocorrelation in the data.

The second model, termed the 'avoidance model' hypothesized that salmon position themselves vertically to avoid extremes of low DO and high temperatures. The low DO threshold was estimated to be at 35% saturation based on a previous study on the hypoxia tolerance of the Tasmanian salmon stock 16 , a value that corresponded well with the distribution of DO values experienced by the tagged salmon (Fig. 1). The high temperature threshold was estimated directly from the distribution of temperature values measured by the fish tags. Since measured temperatures were approximately normally distributed, the upper end of the 95% reference range (97.5 centile; $20.1\,^{\circ}\text{C}$) was used. Analogous to the 'preference model', depth of the DO threshold and temperature threshold layers as well as their interaction were used as fixed effects in the model, and fish ID used as the random effect. Since the depth of the $20.1\,^{\circ}\text{C}$ layer fell above the shallowest deployment depth of the tags during parts of the study period it could not be estimated for each 6 hour interval. Hence, it was converted from a continuous, numeric to a categorical variable, with depths assigned to the following bins: $<1\,\text{m}$, $1-2\,\text{m}$, $>3\,\text{m}$.

As with the 'preference model', Cage ID was initially included as a fixed effect but removed from the final model after it was found to be insignificant.

Both models were fitted in R using the 'nlme' package⁴⁶, significance of model parameters was estimated using the Wald Chi-squared test. Visual inspection of model residuals was used to ensure assumptions of homoscedasticity and normality of residuals were not violated. To test whether using 6 hourly averages of swimming depth was sufficient to eliminate temporal autocorrelation in the data, we fitted both models with and without a continuous, first order, autoregressive correlation structure and compared the resulting models using Akaike's Information Criterion (AIC) as suggested in Zuur, *et al.*⁴⁷.

References

- 1. Kramer, D. L. Dissolved oxygen and fish behavior. Environmental Biology of Fishes 18, 81-92, doi:10.1007/bf00002597 (1987).
- 2. Kramer, D. L. The evolutionary ecology of respiratory mode in fishes: an analysis based on the costs of breathing. *Environmental Biology of Fishes* 9, 145–158, doi:10.1007/bf00690859 (1983).
- 3. Brett, J. Ř. Metabolic demand for oxygen in fish, particularly salmonids, and a comparison with other vertebrates. *Respir. Physiol.* 14, 151–170, doi:10.1016/0034-5687(72)90025-4 (1972).
- 4. Brett, J. R. & Groves, T. D. D. In *Bio*energetics and *Growth Fish Physiology* (eds W.S. Hoar, D. J. Randall & J. R. Brett) 279–352 (Academic Press, 1979).
- 5. Pörtner, H. O. & Knust, R. Climate Change Affects Marine Fishes Through the Oxygen Limitation of Thermal Tolerance. *Science* 315, 95–97, doi:10.1126/science.1135471 (2007).
- Elliott, J. M. & Elliott, J. A. Temperature requirements of Atlantic salmon Salmo salar, brown trout Salmo trutta and Arctic charr Salvelinus alpinus: predicting the effects of climate change. *Journal of Fish Biology* 77, 1793-1817, doi:10.1111/j.1095-8649.2010.02762.x (2010).
- 7. Deutsch, C. et al. Climate change tightens a metabolic constraint on marine habitats. Science 348, 1132–1135, doi:10.1126/science.
- Brander, K. M. Global fish production and climate change. Proceedings of the National Academy of Sciences 104, 19709–19714, doi:10.1073/pnas.0702059104 (2007).
- 9. Jonsson, B. & Jonsson, N. A review of the likely effects of climate change on anadromous Atlantic salmon Salmo salar and brown trout Salmo trutta, with particular reference to water temperature and flow. *Journal of Fish Biology* 75, 2381–2447, doi:10.1111/j.1095-8649.2009.02380.x (2009).
- Eliason, E. J. et al. Differences in Thermal Tolerance Among Sockeye Salmon Populations. Science 332, 109–112, doi:10.1126/ science.1199158 (2011).
- 11. Anttila, K. et al. Variation in temperature tolerance among families of Atlantic salmon (Salmo salar) is associated with hypoxia tolerance, ventricle size and myoglobin level </div>. The Journal of Experimental Biology 216, 1183–1190, doi:10.1242/jeb.080556 (2013).
- 12. Remen, M., Oppedal, F., Imsland, A. K., Olsen, R. E. & Torgersen, T. Hypoxia tolerance thresholds for post-smolt Atlantic salmon: Dependency of temperature and hypoxia acclimation. *Aquaculture* 416–417, 41–47, doi:10.1016/j.aquaculture.2013.08.024 (2013).
- 13. Remen, M., Oppedal, F., Torgersen, T., Imsland, A. K. & Olsen, R. E. Effects of cyclic environmental hypoxia on physiology and feed intake of post-smolt Atlantic salmon: Initial responses and acclimation. *Aquaculture* 326–329, 148–155, doi:10.1016/j. aquaculture.2011.11.036 (2012).
- 14. Kutty, M. N. & Saunders, R. L. Swimming Performance of Young Atlantic Salmon (Salmo salar) as Affected by Reduced Ambient Oxygen Concentration. *Journal of the Fisheries Research Board of Canada* 30, 223–227, doi:10.1139/f73-039 (1973).
- 15. Javaid, M. Y. & Anderson, J. M. Thermal Acclimation and Temperature Selection in Atlantic Salmon, Salmo salar, and Rainbow Trout, S. gairdneri. *Journal of the Fisheries Research Board of Canada* 24, 1507–1513, doi:10.1139/f67-124 (1967).
- Barnes, R., King, H. & Carter, C. G. Hypoxia tolerance and oxygen regulation in Atlantic salmon, Salmo salar from a Tasmanian population. Aquaculture 318, 397–401, doi:10.1016/j.aquaculture.2011.06.003 (2011).
- 17. Johansson, D., Juell, J.-E., Oppedal, F., Stiansen, J.-E. & Ruohonen, K. The influence of the pycnocline and cage resistance on current flow, oxygen flux and swimming behaviour of Atlantic salmon (Salmo salar L.) in production cages. *Aquaculture* 265, 271–287, doi:10.1016/j.aquaculture.2006.12.047 (2007).
- 18. Johansson, D. *et al.* Effect of environmental factors on swimming depth preferences of Atlantic salmon (Salmo salar L.) and temporal and spatial variations in oxygen levels in sea cages at a fjord site. *Aquaculture* **254**, 594–605, doi:10.1016/j.aquaculture.2005.10.029
- 19. Dempster, T., Wright, D. & Oppedal, F. Identifying the nature, extent and duration of critical production periods for Atlantic salmon in Macquarie Harbour, Tasmania, during summer. FRDC Project N0 2016-229 (FRDC, Canberra, 2016).
- 20. Oppedal, F., Dempster, T. & Stien, L. H. Environmental drivers of Atlantic salmon behaviour in sea-cages: A review. *Aquaculture* 311, 1–18. doi:10.1016/j.aquaculture.2010.11.020 (2011).
- 21. Priede, I. G., Solbé, J. Fd. L. G., Nott, J. E., O'Grady, K. T. & Cragg-Hine, D. Behaviour of adult Atlantic salmon, Salmo salar L., in the estuary of the River Ribble in relation to variations in dissolved oxygen and tidal flow. *Journal of Fish Biology* 33, 133–139, doi:10.1111/j.1095-8649.1988.tb05567.x (1988).
- 22. Lacroix, G. L. Population-specific ranges of oceanic migration for adult Atlantic salmon (Salmo salar) documented using pop-up satellite archival tags. *Canadian Journal of Fisheries and Aquatic Sciences* 70, 1011–1030, doi:10.1139/cjfas-2013-0038 (2013).
- 23. BOM. Seasonal Climate Summary for Tasmania, http://www.bom.gov.au/climate/current/season/tas/archive/201602.summary.shtml (2016).
- 24. Teasdale, P. R., Apte, S. C., Ford, P. W., Batley, G. E. & Koehnken, L. Geochemical cycling and speciation of copper in waters and sediments of Macquarie Harbour, Western Tasmania. *Estuarine, Coastal and Shelf Science* 57, 475–487, doi:10.1016/S0272-7714(02)00381-5 (2003).
- 25. Wu, R. S. S. Hypoxia: from molecular responses to ecosystem responses. *Marine Pollution Bulletin* 45, 35–45, doi:10.1016/S0025-326X(02)00061-9 (2002).
- Suthers, I. M. & Gee, J. H. Role of Hypoxia in Limiting Diel Spring and Summer Distribution of Juvenile Yellow Perch (Perca flavescens) in a Prairie Marsh. Canadian Journal of Fisheries and Aquatic Sciences 43, 1562–1570, doi:10.1139/f86-194 (1986).
- 27. Hazen, E. L., Craig, J. K., Good, C. P. & Crowder, L. B. Vertical distribution of fish biomass in hypoxic waters on the Gulf of Mexico shelf. *Marine Ecology Progress Series* 375, 195–207 (2009).
- Rudstam, L. G. & Magnuson, J. J. Predicting the Vertical Distribution of Fish Populations: Analysis of Cisco, Coregonus artedii, and Yellow Perch, Perca flavescens. Canadian Journal of Fisheries and Aquatic Sciences 42, 1178–1188, doi:10.1139/f85-146 (1985).
- 29. Siefert, R. E. & Spoor, W. A. In T Early Life Hory of Fish: The Proceedings of an International Symposium Held at the Dunstaffnage Marine Research Laboratory of the Scottish Marine Biological Association at Oban, Scotland, from May 17–23, 1973 (ed John H. S. Blaxter) 487–495 (Springer Berlin Heidelberg, 1974).
- 30. Hallock, R. J., Elwell, R. F. & Fry, D. H. Jr. Migrations of adult king salmon Oncorhynchus tshawytscha in the San Joaquin Delta as demonstrated by the use of sonic tags. California Department of Fish Game Fisheries Bulletin 151 (1970)
- 31. Burt, K. et al. Environmental conditions and occurrence of hypoxia within production cages of Atlantic salmon on the south coast of Newfoundland. Aquaculture Research 43, 607–620, doi:10.1111/j.1365-2109.2011.02867.x (2012).
- 32. Oldham, T., Dempster, T., Fosse, J. O. & Oppedal, F. Oxygen gradients affect behaviour of caged Atlantic salmon Salmo salar. Aquaculture Environment Interactions 9, 145–153 (2017).
- 33. Kvamme, B. O. *et al.* Modulation of innate immune responses in Atlantic salmon by chronic hypoxia-induced stress. *Fish & Shellfish Immunology* **34**, 55–65, doi:10.1016/j.fsi.2012.10.006 (2013).
- 34. Stevens, E. D., Sutterlin, A. & Cook, T. Respiratory metabolism and swimming performance in growth hormone transgenic Atlantic salmon. *Canadian Journal of Fisheries and Aquatic Sciences* 55, 2028–2035, doi:10.1139/f98-078 (1998).
- 35. Innes, B. H. & Elliott, N. G. Genetic diversity in a Tasmanian hatchery population of Atlantic salmon (Salmo salar L.) compared with its Canadian progenitor population. *Aquaculture Research* 37, 563–569, doi:10.1111/j.1365-2109.2006.01463.x (2006).
- 36. Johnson-Mackinnon, J., Oldham, T. & Nowak, B. Amoebic gill disease: a growing threat. Microbiology Australia 37, 140–142 (2016).

- 37. Benedicenti, O., Secombes, C., McCarthy, U. & Collins, C. Atlantic salmon physiological and immune response to amoebic gill disease and insight into the biology of the amoeba. Fish & Shellfish Immunology 53, 83–84, doi:10.1016/j.fsi.2016.03.118 (2016).
- 38. Treloar, M. A., Barrett, N. S. & Edgar, G. J. Biology and ecology of Zearaja maugeana, an Endangered skate restricted to two southwestern Tasmanian estuaries. *Marine and Freshwater Research*. doi:10.1071/MF15478 (2016).
- 39. Reilly, A. et al. Genetic differentiation between Tasmanian cultured Atlantic salmon (Salmo salar L.) and their ancestral Canadian population: comparison of microsatellite DNA and allozyme and mitochondrial DNA variation. Aquaculture 173, 459–469, doi:10.1016/S0044-8486(98)00476-1 (1999).
- Johansson, D., Ruohonen, K., Juell, J.-E. & Oppedal, F. Swimming depth and thermal history of individual Atlantic salmon (Salmo salar L.) in production cages under different ambient temperature conditions. *Aquaculture* 290, 296–303, doi:10.1016/j. aquaculture.2009.02.022 (2009).
- 41. Hart, P. J. B. & Pitcher, T. J. Field Trials of Fish Marking using a Jet Inoculator. Journal of Fish Biology 1, 383-385, doi:10.1111/j.1095-8649.1969.tb03886.x (1969).
- 42. Bridger, C. J. & Booth, R. K. The Effects of Biotelemetry Transmitter Presence and Attachment Procedures on Fish Physiology and Behavior. *Reviews in Fisheries Science* 11, 13–34, doi:10.1080/16226510390856510 (2003).
- 43. Akima, H. A Method of Bivariate Interpolation and Smooth Surface Fitting for Irregularly Distributed Data Points. *ACM Trans. Math. Softw* 4, 148–159, doi:10.1145/355780.355786 (1978).
- 44. R: A Language and Environment for Statistical Computing v. 3.3.1 (R Foundation for Statistical Computing, Vienna, Austria, 2017).
- 45. Akima: Interpolation of Irregularly and Regularly Spaced Data v. 0.5–12 (2015).
- 46. Nlme: Linear and Nonlinear Mixed Effects Models v. 3.1-128 (2016).
- 47. Zuur, A., Ieno, E. N., Walker, N., Saveliev, A. A. & Smith, G. M. Mixed effects models and extensions in ecology with R (Springer Science & Business Media. 2009).
- 48. PBSmapping: Mapping Fisheries Data and Spatial Analysis Tools v. 2.69.76 (2015).

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Author Contributions

K.M.S., J.M.S. and J.D.M. devised and undertook the tagging study. K.M.S. carried out statistical analysis and figure preparation and drafted the manuscript and C.G.C., J.D.M., J.D.R. and J.M.S. contributed to its final form.

Additional Information

Competing Interests: The authors declare that they have no competing interests.

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