

Screening of a New Kosakonia Species for Polyethylene Biodegradation

Jin-Hee Cho^{1,†}, Seung-Do Yun^{1,†}, Hyun-Woo Kim¹, Min-Ju Seo², Bong Hyun Sung^{3*}, and Soo-Jin Yeom^{1,2,4,5*}

¹School of Biological Sciences and Biotechnology, Graduate School, Chonnam National University, Gwangju 61186, Republic of Korea

²Institute of Synthetic Biology for Carbon Neutralization, Chonnam National University, Gwangju 61186, Republic of Korea

³Synthetic Biology Research Center, Korea Research Institute of Bioscience and Biotechnology, Daejeon 34141, Republic of Korea

⁴School of Biological Sciences and Technology, Chonnam National University, Gwangju 61186, Republic of Korea ⁵Institute of Systems Biology and Life Science Informatics, Chonnam National University, Gwangju 61186, Republic of Korea

Polyethylene (PE) is among the most widely used synthetic plastics globally, serving as an essential material in daily life and numerous industries, such as packaging for bottles and food, as well as in the production of toys and pipes. PE is used for various purposes owing to its high durability and low production costs, leading to a steadily increasing demand. However, PE waste is a significant contributor to environmental pollution, posing serious threats to marine and soil ecosystems. Therefore, the efficient decomposition of PE, a synthetic polymer known for its resistance to degradation, using bacteria offers a sustainable and effective method for reusing PE. In this study, we isolated a novel species of Kosakonia, designated Kosakonia cowanii JNU01, from a landfill site, capable of biodegrading PE. K. cowanii JNU01 exhibited the highest cell growth rate in media containing PE, indicating its effectiveness in decomposing PE for use as a sole carbon source in its metabolic pathway. Treatment of PE with K. cowanii JNU01 resulted in the emergence of new chemical functional groups, including hydroxyl, carboxyl, amide, and ether groups, within the inert hydrocarbon structure. Analysis of the PE film treated with K. cowanii JNU01 revealed considerable physical degradation on the film's surface. Additionally, various metabolites released from PE by K. cowanii JNU01 were identified. These findings suggest that K. cowanii JNU01 proves to be an effective candidate bacterium for PE degradation.

Keywords: Polyethylene, biodegradation, Kosakonia cowanii

Received: November 24, 2024 Revised: January 14, 2025 Accepted: January 14, 2025

*Corresponding authors BH Sung E-mail: bhsung@kribb.re.kr SJ Yeom E-mail: sooiin258@chonnam.ac.kr

[†]These authors contributed equally to this work.

pISSN 1017-7825 eISSN 1738-8872

Copyright © 2025 by the authors. Licensee KMB. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license.

Introduction

With the increasing use of plastics, production and waste generation have risen correspondingly [1]. This has led to the accumulation of plastic in landfills and natural environments, posing serious threats to ecosystems and causing harmful impacts on human health and numerous wildlife species. To address these issues, various preliminary studies and interventions have been undertaken, including recycling, thermal degradation, and biological degradation [2, 3]. Among these approaches, biological degradation is gaining attention as an economically viable and environmentally friendly solution for plastic waste reduction [4, 5]. Traditional plastic waste treatment methods such as incineration, release organic pollutants, particulate matter, and hazardous substances like dioxins and furans into the environment [6-9]. Therefore, the biodegradation of plastics using biocatalysts such as bacteria and their enzymes represents an ideal solution to mitigate adverse environmental effects [10].

The backbone of PE consists solely of C-C single bonds, lacking functional groups, and its structural stability grants it significant resistance to decomposition [11]. PE materials, such as low-density PE (LDPE) and high-density PE (HDPE), are the most widely used petroleum-based polymers for manufacturing disposable plastic items, accounting for up to 64% of disposable plastics. LDPE, in particular, is discarded within a short period after use, leading to its rapid accumulation in the environment [3, 12, 13]. Therefore, it is imperative to identify an eco-friendly and sustainable method to address this issue [3, 12]. Recent studies have isolated several PE-degrading bacteria from various environments: *Rhodococcus ruber* from agricultural waste in soil [14-16]; *Klebsiella pneumoniae* from soil [17]; *Desulfotomaculum nigrificans* and *Pseudomonas alcaligenes* from plastic waste contaminated soil [18]; *Stenotrophomonas* sp., *Comamonas* sp., and *Delftia* sp. from plastic debris in soil [19];

Enterobacter sp. D1 from the gut of Galleria mellonella [20]; Pseudomonas putida from garden soil [21]; as well as various PE-degrading bacteria (namely, Bacillus megaterium, B. thuringiensis, B. cereus, B. amylolyticus, B. subtilis, Pseudomonas putida, and P. fluorescens) from soil [22-24]. Enzymes reported to be involved in PE biodegradation include alkane-1-monooxygenase [25, 26], laccase-like multi-copper oxidase [17, 27], and cytochrome P450 [24], and PEase [28]. Although bacteria capable of degrading PE and certain enzymes predicted to be involved in PE degradation have been identified, the precise mechanism underlying PE degradation remains unclear. Therefore, further investigation to elucidate this mechanism is essential, and discovering biodegradable biocatalysts capable of breaking down PE is a critical step. Screening for promising bacteria with PE-degrading capabilities is necessary to identify additional biocatalysts involved in PE biodegradation.

In this study, we identified promising new microbial candidates for PE degradation. Specifically, we successfully isolated a PE-degrading strain, Kosakonia cowanii JNU01, from the Metropolitan City Sanitary Landfill in Gwangju, Korea. The strain demonstrated the ability to grow using PE as its sole carbon source. We confirmed the PE degradation ability of K. cowanii JNU01 by characterizing the chemical and physical changes in PE through Fourier transform infrared spectroscopy (FT-IR) and scanning electron microscopy (SEM) analyses. Additionally, we analyzed the metabolites released during incubation with PE and K. cowanii JNU01 using gas chromatography-mass spectrometry (GC-MS). The successful screening of K. cowanii JNU01 as a PE-degrading bacterium lays the groundwork for identifying additional PE-degrading enzymes and elucidating the mechanisms underlying PE biodegradation.

Material and Methods

Chemicals Used in the Study

This study utilized M9 minimal broth (6.0 g/l Na₂HPO₄, 3.0 g/l KH₂PO₄, 0.5 g/l NaCl, and 1.0 g/l NH₄Cl, 240.7 mg/l MgSO₄, 11.098 mg/l CaCl₂) along with a trace element solution (21.8 mg/l CoCl₂·6H₂O, 21.6 mg/l $NiCl_2 \cdot 6H_2O, 24.6 \ mg/l \ CuSO_4 \cdot 5H_2O, 1.62 \ g/l \ FeCl_3 \cdot 6H_2O, 0.78 \ g/l \ CaCl_2, and 14.7 \ mg/l \ MnCl_2 \cdot 4H_2O) \ were \ used.$ Luria-Bertani (LB) medium was procured from MB Cell (Republic of Korea), and methanol (≥99.8%) and heptane (≥98%) were purchased from Duksan (Republic of Korea). Chloroform (≥99.5%) and ethyl acetate $(\ge 99.5\%)$ were sourced from Daejung (Republic of Korea). Additionally, PE, with an average M_w (Weight-average molecular weight) of approximately 4,000 and an average M_n (Number-average molecular weight) of approximately 1,700, possessing typical physical properties of LDPE, was purchased from Sigma-Aldrich (USA). To improve the resolution of PE, the external surface area was increased through a reprecipitation process. The re-precipitation process for PE powder and film was the same as the method described in our previous study [24].

Identification of PE Biodegradable Bacterial Strains

Environmental samples, including soil and plastic waste, were collected from the Metropolitan City Sanitary Landfill located in Gwangju, Korea. The sample (50 g) was mixed with 450 ml of phosphate-buffered saline (PBS) and then filtered using Whatman paper with an 11 µm pore size. The filtered solutions were spread onto liquid M9 minimal broth agar plates supplemented with 1 g/l PE powder and incubated at 28°C for 7 days. Bacterial colonies that developed were re-streaked on M9 minimal broth agar plates containing 1 g/l PE. The cultured bacteria were subsequently incubated in liquid M9 minimal broth containing 1 g/l PE powder at 28°C with shaking at 150 rpm for 7 days. To evaluate bacterial growth, the optical density (OD_{600}) of 200 μ l of culture solution was measured on day 0 and day 7 using a VICTOR Nivo multimode microplate reader (VICTOR Nivo Multimode Microplate Reader, PerkinElmer, USA). The fold-change was calculated by dividing the OD_{600} reading on day 7 by the OD_{600} reading on day 0. Colony #2 was streaked onto LB agar plates and incubated overnight at 28°C. Subsequently, DNA extraction and 16S rRNA sequencing were conducted following the method established in our previous study [24]. In addition to the 16S rRNA sequences of the newly identified Kosakonia cowanii JNU01, sequences from nine other Kosakonia species, three Enterobacter species, and Vibrio aerogenes were obtained from the NCBI nucleotide database. Phylogenetic trees were constructed using the maximum-likelihood (ML) method with the Tamura-Nei model and 1,000 bootstrap replicates in MEGA11 (version 11.0.13) software [29].

Growth Curve and Dissolved Oxygen Analysis of K. cowanii JNU01 in PE Media

K. cowanii JNU01 was first cultured in LB media and incubated overnight at 28°C with shaking at 200 rpm. The culture was then centrifuged at 4,000 rpm for 20 min, and the supernatant was removed. The resulting bacterial pellet was washed with M9 minimal broth before being inoculated into M9 minimal broth containing PE powder $at concentrations of 10, 30, and 50 \,mg/l, respectively, supplemented with 0.1\% \,trace \,element \,solution. \,The \,cultures \,respectively and the concentration of the concentrat$ were incubated at 28°C with shaking at 200 rpm for 35 days. The initial OD_{600} was 0.2, and the experiment included two controls: one containing only PE powder in M9 media and the other containing bacteria without PE powder. Every five days, $500 \,\mu l$ of culture and controls were diluted with $500 \,\mu l$ of fresh M9 broth, and OD_{600} was measured using a UV-Vis spectrophotometer (UV-1900, Shimadzu, Japan).

Additionally, dissolved oxygen (DO) concentration in M9 minimal broth was monitored using a biochemical oxygen demand (BOD) probe and a DO meter (YSI 4100 BOD IDS PROBE & YSI 4010-1W DO meter, YSI Incorporated, USA). The DO measurements were conducted at an initial OD₆₀₀ of 0.2 in 60 ml BOD bottles containing PE at a concentration of 30 mg/ml, with incubation at 28°C and shaking at 140 rpm for 4 days.

GC-MS Analysis of PE Media after Treatment of K. cowanii JNU01

The cultured solution was filtered using Whatman filter paper with an 11 µm pore size, and any remaining PE particles were rinsed with methanol. The filtered culture solution was then combined with ethyl acetate in a 1:1 ratio, vortexed for 10 min, and centrifuged at 13,000 rpm for 20 min to isolate the supernatant. The extract was analyzed using a gas chromatograph-mass spectrometer (GC-MS) (QP2020 NX, Shimadzu) equipped with an electron impact ionization source and a DB-5MS capillary column (Agilent, USA). The column temperature was programmed to increase from 40° C to 280° C at a rate of 6° C min⁻¹, with an initial hold at 40° C for 2 min and a final hold at 280° C for 3 min. The chemical structures of the products were determined using the GC-MS data and the NIST/WILEY database.

FT-IR Analysis of K. cowanii JNU01-Treated PE Powder

Residual PE particles rinsed with methanol were subjected to Fourier transform infrared (FT-IR) analysis to characterize alterations in the chemical structure of the PE polymer chains. FT-IR measurements (Spectrum Two, PerkinElmer) were conducted to identify the chemical functional groups in the PE samples across a wavenumber range of $4,000-700 \, \mathrm{cm}^{-1}$, with a resolution of $4 \, \mathrm{cm}^{-1}$.

SEM Analysis of K. cowanii JNU01-Treated PE Film

Scanning electron microscope (SEM) was employed to examine the physical properties of PE films. The surface morphologies of pure and biologically degraded PE films were observed using a field-emission SEM (FE-SEM; Hitachi S-4800, Japan) operating at an accelerating voltage of 15 kV (Center for Scientific Instrument, Chosun University, Gwangju, Korea). *K. cowanii* JNU01 was inoculated into a flask containing 10 ml of liquid M9 minimal broth supplemented with PE film and 0.1% trace element solution. With an initial OD $_{600}$ of 2.0, *K. cowanii* JNU01 was cultured at 28°C and 80 rpm for 30 days. Following incubation, the biologically degraded and pure PE films were washed in a sodium dodecyl sulfate solution (SDS; 2%, w/v) for 6 h and subsequently rinsed twice with fresh methanol. This pretreatment procedure was employed to remove any bacterial cells adhering to the PE surface, allowing for accurate surface morphology assessment. To prepare the PE films for SEM imaging, a platinum (Pt) layer was sputtered at 20 mA for 60 s to generate a thin, conductive coating, thus facilitating clear top-view imaging of the non-conductive polymer samples.

Results and Discussion

Isolation and Selection of PE-Biodegrading Bacteria

Landfill soil samples were initially screened for PE-biodegrading bacteria, using PE powder as the sole carbon source (Fig. 1). Sixteen single colonies were obtained and cultured on M9 minimal broth agar plates supplemented with PE powder. These colonies were then grown in liquid M9 minimal broth containing only PE powder for 7 days. One bacterial species, identified as #2, exhibited a 4.1-fold increase in OD₆₀₀ from day 0 to day 7, indicating significant growth (Fig. 2A). A total of 14 different 16S rRNA sequences were analyzed and compared with those of closely related species (Fig. 2B). The bacterial species #2 formed a distinct monophyletic clade with *K. cowanii*, leading us to designate this strain as #2 *K. cowanii* JNU01 (Fig. 2B). Common microorganisms known to degrade PE include *Acinetobacter*, *Alcanivorax*, *Bacillus*, *Brevibacillus*, *Cobetia*, *Comamonas*, *Delftia*, *Enterobacter*, *Halomonas*, *Klebsiella*, *Paenibacillus*, *Pseudomonas*, and *Rhodococcus* species [13, 30-32]. In this study, *K. cowanii* JNU01 is identified as a novel PE-degrading bacterium with potential applications in PE degradation as well as in the decomposition of cotton textile waste [33].

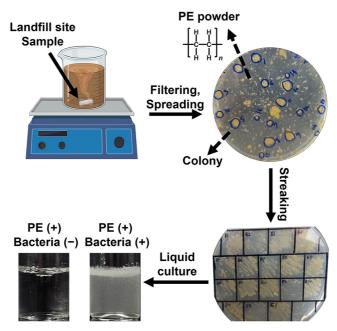


Fig. 1. Overview of the screening process for PE-biodegrading microorganisms. The landfill site sample was obtained from the metropolitan city sanitary landfill in Gwangju, Korea.

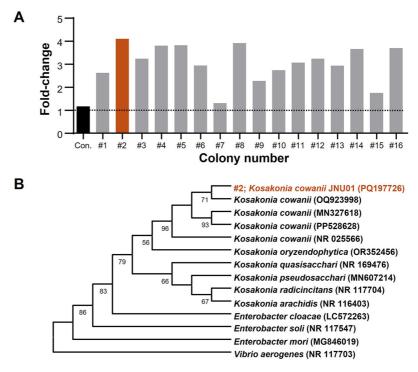


Fig. 2. Identification process of PE-degrading microorganisms. (A) Fold-change in OD_{600} (Day 7 OD_{600} /Day 0 OD_{600}) for 16 individual colonies. (B) Phylogenetic tree of *K. cowanii* JNU01 based on 16S rRNA gene sequences, with accession numbers in parentheses.

Growth of K. cowanii JNU01 in PE Media

The growth curve of *K. cowanii* JNU01 was determined by culturing the bacterium in M9 minimal broth medium using PE powder as the sole carbon source at concentrations of 10, 30, and 50 mg/ml for 35 days at 28°C. Significant increases in turbidity were observed at higher PE concentrations (Fig. 3A). At 10 mg/ml PE, the OD $_{600}$ value remained constant, indicating no significant growth in the absence of microbes. However, as the PE concentration increased to 30 and 50 mg/ml, the OD $_{600}$ values rose markedly. Notably, *K. cowanii* JNU01 achieved a very high OD $_{600}$ value of 0.44 within just 10 days at 50 mg/ml PE, demonstrating its exceptional capacity to grow

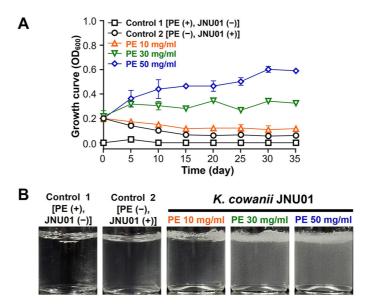


Fig. 3. Growth curve of *K. cowanii* **JNU01 in PE media.** (A) Growth curve of *K. cowanii* **JNU01** in PE media. The orange triangle, green inverted triangle, and blue diamond represent *K. cowanii* **JNU01** inoculated with 10, 30, and 50 mg/ml of PE powder, respectively. (B) Image of medias with *K. cowanii* **JNU01** and PE powder. Control 1 represents a sample containing only PE powder, and Control 2 contains only K. cowanii JNU01.

Table 1. Microorganisms capable of biodegrading polyethylene (PE).

Microorganism	РЕ Туре	Culture Conditions	Growth rate	FT-IR analysis	Metabolite analysis (MS)	Reference
Kosakonia cowanii JNU01	LDPE powder / film	28°C, Sole carbon source (PE), 35 days	OD ₆₀₀ = 0.62	Carboxylic acid, Hydroxyl group, Amide	Alkanes, ketones, and fatty acids	This study
Enterobacter sp. D1	LDPE film	37°C, Sole carbon source (PE), 31 days	$OD_{600} = 0.24$	Carbonyl bands, ether groups	Alkanes, alcohols, hydrocarbon, esters, and acids	[20]
Acinetobacter sp. NyZ450 & Bacillus sp. NyZ451	LDPE mulching film	28°C, Co-culture Sole carbon source (PE), 30 days	$OD_{600} = 0.065$	C=C bonds, O-H bonds (Hydroxyl group)	N/A	[34]
Klebsiella pneumoniae CH001	HDPE film	30 ± 2°C, Medium containing NB (Nutrient Broth), 2 months	N/A	Carboxylic group	carboxylic acids and alkanes	[42]
Bacillus velezensis C5	LDPE film	37°C, Sole carbon source (PE), 90 days	N/A	Alcohols, carbonyl bond, double bond, terminal double bond	n-alkanes $(C_{24}-C_{29})$	[43]
Pseudomonas aeruginosa SKN1	LDPE powder	29°C, Sole carbon source (PE), 60 days	$OD_{600} =$ approximat ely 0.24	C–H bond stretching, C–C bond stretching	N/A	[35]
Bacillus cereus, Bacillus megaterium, Bacillus subtilis	LDPE, LLDPE film	30 & 45°C, Co-culture Sole carbon source (PE), 90 days	N/A	Amide bond, hydroxyl bond,	Alkanes, alcohols	[23]
Bacillus pumilus, Bacillus halodenitrificans, Bacillus cereus	LDPE film	30°C, Sole carbon source (PE), 80 days	N/A	C=O stretch (ketone)	Ketones, alcohols, Alkanes	[44]

N/A: not applicable, LDPE: Low-Density Polyethylene, LLDPE: Linear Low-Density Polyethylene, HDPE: High-Density Polyethylene, MS: mass spectrometry

using PE as a carbon source over a 35-day incubation period (Fig. 3A). According to previous studies summarized in Table 1, *Enterobacter* sp. D1 was cultured at 37°C using PE film as the sole carbon source, achieving an OD_{600} value of 0.24 after 31 days [20]. Similarly, a consortium of *Acinetobacter* sp. NyZ450 and *Bacillus* sp. NyZ451 reached an OD_{600} value of 0.065 after 30 days at 28°C using PE particles as the carbon source [34]. Furthermore, *Pseudomonas aeruginosa* SKN1 achieved an OD_{600} value of approximately 0.24 after 60 days at 29°C using PE powder [35]. These results suggest that *K. cowanii* JNU01 can utilize PE as the sole carbon source independently, without forming a consortium, and achieve rapid and efficient growth under relatively moderate temperature conditions (28°C) within a short period. Additionally, the PE powder cultured with *K. cowanii* JNU01 exhibited a tendency to blend into the PE media, indicating that the PE surface becomes more hydrophilic as it is colonized by microorganisms (Fig. 3B).

This hydrophilicity likely facilitates further colonization and enhances microbial activity on the hydrophilized PE surface [36-40]. To evaluate the PE mineralization by *K. cowanii* JNU01, dissolved oxygen (DO) was measured. The initial DO was 8.93 ± 0.01 , decreasing to 2.79 ± 0.01 after 1 day, 0.54 ± 0.01 after 2 days, and 0.27 ± 0.10 on the 4th day (Table 2). Overall, these results provide additional evidence of PE mineralization [41], suggesting that *K. cowanii* JNU01 can effectively utilize PE as a carbon source.

Identification of Metabolites in Culture Media with PE

To identify metabolites produced when PE is used as a carbon source, the culture medium with K. cowanii JNU01 and added PE powder was analyzed using GC-MS. The results revealed the formation of alkanes, ketones and fatty acids (Fig. 4 and Table 3). Using the NIST/WILEY database, GC-MS analysis of K. cowanii JNU01 identified a variety of metabolites, including ketone compounds with carbon chain lengths ranging from C_{16} to C_{33} . Key compounds detected included 2-hexadecanone, 9-heptadecanone, 2-heptadecanone, 9-octadecanone,

Table 2. Dissolved oxygen (DO) levels in PE media inoculated with K. cowanii JNU01.

Bacterium	Initial DO (mg/l)	DO after 1 day (mg/l)	DO after 2 days (mg/l)	DO after 4 days (mg/l)
K. cowanii JNU01	8.93 ± 0.01	2.79 ± 0.09	0.54 ± 0.02	0.27 ± 0.10

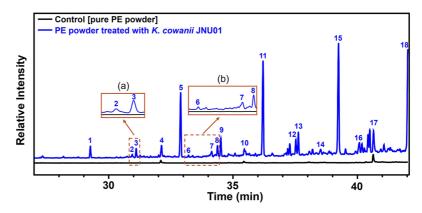


Fig. 4. GC-MS analysis of PE culture media inoculated with *K. cowanii* **JNU01.** Control is a sample containing only PE powder, and the experimental group includes PE powder treated with *K. cowanii* **JNU01.** (a) Magnified regions corresponding to peaks #2 and #3 are shown. (b) Magnified regions corresponding to peaks #6, #7, and #8, shown in the inset boxes.

2-nonadecanone, 10-nonadecanone, 2-nonacosanone, and 2-tritriacontanone. In addition, alkanes such as octadecane, eicosane, pentacosane, 2-methylhexacosane, and dotriacontane, along with several oxidized compounds, were also identified as metabolites (Table 3). Compared to the findings from previously reported studies summarized in Table 1, K. cowanii JNU01 exhibits distinct characteristics in its metabolic profile. For example, metabolites detected in Enterobacter sp. D1 primarily included alkanes, alcohols, hydrocarbons, esters, and acids with carbon chain lengths below C₂₀ [20]. Metabolites in Klebsiella pneumoniae CH001 were identified as carboxylic acids ranging from C2 to C18 and alkanes from C2 to C19 [42]. Bacillus velezensis C5 was reported to have n-alkanes identified within the range of C24 to C29 [43]. While metabolites, including alkanes and alcohols with chain lengths spanning from C₁₄ to C₃₆, were detected in a consortium of B. cereus, B. megaterium, and B. subtilis [23]. Additionally, the PE metabolites derived from B. pumilus, B. halodenitrificans, and B. cereus included ketone compounds with carbon chains below C₁₆, alkanols below C₂₀, and alkanes below C₃₀ [44]. These findings suggest that K. cowanii JNU01 metabolizes PE, resulting in the production of fatty acids, alkanes, and long-chain ketone compounds. The alkanes, ketones, and fatty acids produced during the PE degradation process by K. cowanii JNU01 are consistent with the previously proposed PE degradation pathway [10]. In this pathway, PE is initially hydroxylated at subterminal and in-chain carbons by hydroxylases such as cytochrome P450 (CYP) [24], unspecific peroxygenase (UPO), and alkane monooxygenase (AlkB). The resulting alcohols are further oxidized to ketones via alcohol dehydrogenase (Adh), producing metabolites such as 2-hexadecanone, 9heptadecanone, 2-heptadecanone, 9-octadecanone, 2-nonadecanone, 10-nonadecanone, 2-nonacosanone, and 2-tritriacontanone. In next step, Baeyer-Villiger monooxygenase (BVMO) catalyzes the insertion of an oxygen atom into the carbonyl group of the intermediate ketone, leading to the formation of an ester group which are

Table 3. Metabolites of polyethylene (PE) in PE media-inoculated with K. cowanii JNU01.

Number	Retention time (min)	Similarity (%)	Compound name	Molecular weight	Formula
1	29.263	96	2-Hexadecanone	240	C ₁₆ H ₃₂ O
2	30.940	80	9-Heptadecanone	254	$C_{17}H_{34}O$
3	31.100	94	Octadecane	254	$C_{18}H_{38}$
4	32.113	93	n-Hexadecanoic acid	256	$C_{16}H_{32}O_2$
5	32.883	90	2-Heptadecanone	254	$C_{17}H_{34}O$
6	33.203	90	Octadecanal	268	$C_{18}H_{36}O$
7	34.127	85	6,10,14-Trimethylpentadecan-2-one	268	$C_{18}H_{36}O$
8	34.363	82	9-Octadecanone	268	$C_{18}H_{36}O$
9	34.493	95	Eicosane	282	$C_{20}H_{42}$
10	35.440	89	Octadecanoic acid	284	$C_{18}H_{36}O_2$
11	36.190	84	2-Nonadecanone	282	$C_{19}H_{38}O$
12	37.503	82	10-Nonadecanone	282	$C_{19}H_{38}O$
13	37.607	95	Pentacosane	352	$C_{25}H_{52}$
14	38.497	90	2-Methylhexacosane	380	$C_{27}H_{56}$
15	39.217	82	2-Nonacosanone	422	$C_{29}H_{58}O$
16	40.053	85	Nonacosan-14-one	422	$C_{29}H_{58}O$
17	40.473	93	Dotriacontane	450	$C_{32}H_{66}$
18	41.997	91	2-Tritriacontanone	478	$C_{33}H_{66}O$

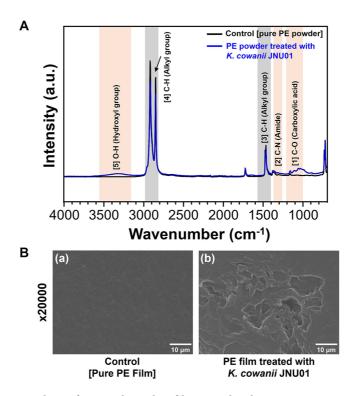


Fig. 5. FT-IR and **SEM** analyses of **PE** powder and **PE** film treated with *K. cowanii* JNU01. (A) FT-IR analysis of PE powder treated with *K. cowanii* JNU01. [1] C-O (carboxylic acid), [2] C-N (amide), [3] C-H (alkyl group), [4] C-H (alkyl group), [5] O-H (hydroxyl group). (B) SEM analysis of *K. cowanii* JNU01 treated PE film: (a) The control (pure PE film) sample, and (b) The PE film treated with *K. cowanii* JNU01.

hydrolyzed by esterase to form alkanoic acids (*e.g.*, octadecanoic acid) and alcohols. Additionally, alkanols further oxidized to ketones and carboxylic acids via Adh and aldehyde dehydrogenase (Aldh). The metabolites observed in this study show a notable correlation with the proposed PE degradation pathway. To further elucidate the enzymatic and metabolic networks involved in PE degradation, a whole genome analysis of *K. cowanii* JNU01 should be performed. This analysis could help identify key genes and enzymes responsible for mediating PE breakdown, providing valuable insights for optimizing microbial strategies to enhance PE degradation.

Functional Group Changes of PE Treated with K. cowanii JNU01

The newly formed chemical functional groups in PE powder treated with *K. cowanii* JNU01 were analyzed using FT-IR spectroscopy (Fig. 5A and Table 4). PE samples incubated with *K. cowanii* JNU01 exhibited increased characteristic peaks corresponding to hydroxyl groups (O–H; 3,200–3,400 cm⁻¹), carboxylic acids (C–O; 1,000–1,320 cm⁻¹), and amides (C–N; 1,200–1,350 cm⁻¹), compared to the control sample. These changes indicate the formation of oxygen- and nitrogen-rich functional groups through the biodegradation of PE powder [24, 34, 45, 46]. Notably, the decrease in characteristic peaks associated with alkyl groups (C–H; 2,850–3,000 and 1,445–1,485 cm⁻¹) can be attributed to the consumption of alkane chains [47]. This finding suggests the existence of multiple biodegradation pathways for PE, including the formation of oxygen-rich functional groups by *K. cowanii* JNU01. Hydroxylation of the C–C bonds in PE is a key step in the biodegradation mechanism previously proposed [10]. FT-IR analysis of *K. cowanii* JNU01-treated PE indicates that degradation occurred on the surface of the PE powder following treatment with the *Kosakonia* strain. In summary, our study demonstrates that *K. cowanii* JNU01-biodegraded PE exhibits diverse chemical functionalities, including hydroxyl groups, carboxylic acids, and amides, highlighting the various chemical transformations that occur during the biodegradation process.

Table 4. FT-IR analysis results of PE powder cultured with K. cowanii JNU01.

Number	Wavenumber (cm ⁻¹)	Bond	Functional group	Change
[1]	1,000-1,200	C-O	Carboxylic acid	Increase
[2]	1,200-1,350	C-N	Amide	Increase
[3]	1,445-1,485	C-H	Alkyl	Decrease
[4]	2,850-3,000	C-H	Alkyl	Decrease
[5]	3,200-3,400	C-OH	Hydroxyl group	Increase

Observation of Surface Changes in PE Film Treated with K. cowanii JNU01

The surface morphology of pure PE film and PE film biologically degraded by *K. cowanii* JNU01 after one month of incubation was examined using FE-SEM to assess the external morphological degradation of PE. The untreated PE film displayed a smooth surface, free of any defects such as holes or cracks (Fig. 5B (left, a)). In contrast, the *K. cowanii* JNU01-treated PE film exhibited numerous surface defects (Fig. 5B (right, b)), including micron-sized pores with a broad size distribution. These observations align with findings from previous studies on PE biodegradation by various bacterial strains, such as *Enterobacter* sp. D1, *Acinetobacter* sp. NyZ450, *Bacillus* sp. NyZ451, *Exiguobacterium* sp., *Ochrobactrum* sp., *Halomonas* sp., *B. amyloliquefaciens* BSM-1, *B. amyloliquefaciens* BSM-2, *Enterobacter asburiae* YT1, and *Bacillus* sp. YP1 [20, 34, 48-50]. Consequently, it was confirmed that *K. cowanii* JNU01 is potential PE decomposer. And then it can be used to generate new synthetic plastic-eating bacteria by introducing genome editing and providing new pathways for PE upcycling.

Conclusion

Environmental pollution caused by plastics poses a significant threat to ecosystems and human health, making it a critical issue to address. In this study, we isolated a novel Kosakonia strain from landfill soil capable of biodegrading PE. The strain, K. cowanii JNU01, demonstrated growth using PE as its sole carbon source, reaching an OD_{600} value of 0.62 at a PE concentration of 50 mg/ml. Observations of morphological changes on the PE surface, the detection of various new functional groups, and the release of degradation by-products strongly support the PE degradation capability of K. cowanii JNU01. Furthermore, various alkanones released following the PE biodegradation by K. cowanii JNU01 offer potential as renewable resources for producing high-value products. Thus, bacterial PE biodegradation represents a promising approach to plastic-refinery, which could transform plastic waste into valuable materials. This study highlights that the PE-degrading capability of the K. cowanii JNU01 strain could be pivotal for advancing methods to effectively biodegrade PE. In future research, we aim to conduct whole-genome sequencing and transcriptome analysis of K. cowanii JNU01 to investigate the mechanisms underlying PE biodegradation by characterizing the enzymes involved. This task presents challenges due to the potential existence of various unexpected pathways for PE biodegradation. Additionally, it is also essential to investigate the accessibility of PE into bacterial cells.

Acknowledgments

This work was supported by grants from the Enzyme Engineering for Next Generation Biorefinery (NRF-2022M3J5A1056169 and NRF-2022M3J5A1085239), the Bio & Medical Technology Development Program (RS-2024-00440681), the Synthetic Biology Core Technology Development Program (RS-2024-00398252), and the NRF grant (Project Number: 1711195195, RS-2023-00208002) of the National Research Foundation (NRF) funded by the Korean government (MSIT), and the KRIBB Research Initiative Program.

Conflict of Interest

The authors have no financial conflicts of interest to declare.

References

- 1. Pilapitiya PNT, Ratnayake AS. 2024. The world of plastic waste: a review. Clean. Mater. 100220.
- Datta J, Kopczyńska P. 2016. From polymer waste to potential main industrial products: actual state of recycling and recovering. Crit. Rev. Environ. Sci. Technol. 46: 905-946.
- 3. Ragaert K, Delva L, Van Geem K. 2017. Mechanical and chemical recycling of solid plastic waste. Waste Manage. 69: 24-58.
- 4. Lee A, Liew MS. 2021. Tertiary recycling of plastics waste: an analysis of feedstock, chemical and biological degradation methods. *J. Mater. Cycles Waste Manag.* 23: 32-43.
- 5. Ahmed T, Shahid M, Azeem F, Rasul I, Shah AA, Noman M, et al. 2018. Biodegradation of plastics: current scenario and future prospects for environmental safety. Environ. Sci. Pollut. Res. 25: 7287-7298.
- Jones H, Saffar F, Koutsos V, Ray D. 2021. Polyolefins and polyethylene terephthalate package wastes: recycling and use in composites. Energies 14: 7306.
- 7. Verma R, Vinoda K, Papireddy M, Gowda A. 2016. Toxic pollutants from plastic waste-a review. Procedia Environ. Sci. 35: 701-708.
- 8. Makarichi L, Jutidamrongphan W, Techato K-a. 2018. The evolution of waste-to-energy incineration: a review. Renew. Sust. Energ. Rev. 91: 812-821.
- 9. Ágnes N, Rajmund K. 2016. The environmental impact of plastic waste incineration. Acad. Appl. Res. Mil. Public Manag. Sci. 15: 231-237.
- 10. Yeom SJ, Le TK, Yun CH. 2021. P450-driven plastic-degrading synthetic bacteria. Trends Biotechnol. 40: 166-179.
- 11. Chamas A, Moon H, Zheng J, Qiu Y, Tabassum T, Jang JH, et al. 2020. Degradation rates of plastics in the environment. ACS Sustain. Chem. Eng. 8: 3494-3511.
- 12. Harshvardhan K, Jha B. 2013. Biodegradation of low-density polyethylene by marine bacteria from pelagic waters, Arabian Sea, India. *Mar. Pollut. Bull.* 77: 100-106.
- 13. Montazer Z, Habibi Najafi MB, Levin DB. 2020. Challenges with verifying microbial degradation of polyethylene. Polymers 12: 123.
- 14. Usha R, Sangeetha T, Palaniswamy M. 2011. Screening of polyethylene degrading microorganisms from garbage soil. *Libyan Agric. Res. Cent. J. Int.* 2: 200-204.
- Sivan A, Szanto M, Pavlov V. 2006. Biofilm development of the polyethylene-degrading bacterium Rhodococcus ruber. Appl. Microbiol. Biotechnol. 72: 346-352.
- 16. Santo M, Weitsman R, Sivan A. 2013. The role of the copper-binding enzyme laccase in the biodegradation of polyethylene by the actinomycete *Rhodococcus ruber*. *Int. Biodeterior. Biodegradation* 84: 204-210.
- 17. Zhang X, Feng X, Lin Y, Gou H, Zhang Y, Yang L. 2023. Degradation of polyethylene by *Klebsiella pneumoniae* Mk-1 isolated from soil. *Ecotox. Environ. Safe.* 258: 114965.
- 18. Begum MA, Varalakshmi B, Umamagheswari K. 2015. Biodegradation of polythene bag using bacteria isolated from soil. *Int. J. Curr. Microbiol. App. Sci.* 4: 674-680.

- 19. Peixoto J, Silva LP, Krüger RH. 2017. Brazilian Cerrado soil reveals an untapped microbial potential for unpretreated polyethylene biodegradation. *J. Hazard. Mater.* **324**: 634-644.
- 20. Ren L, Men L, Zhang Z, Guan F, Tian J, Wang B, et al. 2019. Biodegradation of polyethylene by Enterobacter sp. D1 from the guts of wax moth Galleria mellonella. Int. J. Environ. Res. Public Health 16: 1941.
- 21. Saminathan P, Sripriya A, Nalini K, Sivakumar T, Thangapandian V. 2014. Biodegradation of plastics by *Pseudomonas putida* isolated from garden soil samples. *J. Adv. Bot. Zool.* 1: 34-38.
- 22. Patil R. 2018. Screening and characterization of plastic degrading bacteria from garbage soil. Br. J. Environ. Sci. 6: 33-40.
- 23. Abrusci C, Pablos JL, Corrales T, López-Marín J, Marín I, Catalina F. 2011. Biodegradation of photo-degraded mulching films based on polyethylenes and stearates of calcium and iron as pro-oxidant additives. *Int. Biodeterior. Biodegrad.* 65: 451-459.
- 24. Yun S-D, Lee CO, Kim H-W, An SJ, Kim S, Seo M-J, et al. 2023. Exploring a new biocatalyst from *Bacillus thuringiensis* JNU01 for polyethylene biodegradation. *Environ. Sci. Technol. Lett.* **10:** 485-492.
- 25. Jeon HJ, Kim MN. 2016. Comparison of the functional characterization between alkane monooxygenases for low-molecular-weight polyethylene biodegradation. *Int. Biodeterior. Biodegrad.* 114: 202-208.
- Jeon HJ, Kim MN. 2015. Functional analysis of alkane hydroxylase system derived from *Pseudomonas aeruginosa* E7 for low molecular weight polyethylene biodegradation. *Int. Biodeterior. Biodegrad.* 103: 141-146.
- 27. Zampolli J, Mangiagalli M, Vezzini D, Lasagni M, Ami D, Natalello A, et al. 2023. Oxidative degradation of polyethylene by two novel laccase-like multicopper oxidases from Rhodococcus opacus R7. Environ. Technol. Innov. 32: 103273.
- 28. Sanluis-Verdes A, Colomer-Vidal P, Rodriguez-Ventura F, Bello-Villarino M, Spinola-Amilibia M, Ruiz-Lopez E, et al. 2022. Wax worm saliva and the enzymes therein are the key to polyethylene degradation by *Galleria mellonella*. Nat. Commun. 13: 5568.
- Kumar S, Stecher G, Li M, Knyaz C, Tamura K. 2018. MEGA X: molecular evolutionary genetics analysis across computing platforms. Mol. Biol. Evol. 35: 1547-1549.
- 30. Seo M-J, Yun S-D, Kim H-W, Yeom S-J. 2023. Polyethylene-biodegrading microbes and their future directions. *Biotechnol. Bioprocess Eng.* 28: 977-989.
- 31. Ghatge S, Yang Y, Ahn J-H, Hur H-G. 2020. Biodegradation of polyethylene: a brief review. Appl. Biol. Chem. 63: 1-14.
- 32. Ali SS, Elsamahy T, Al-Tohamy R, Zhu D, Mahmoud YA-G, Koutra E, et al. 2021. Plastic wastes biodegradation: mechanisms, challenges and future prospects. Sci. Total Environ. 780: 146590.
- 33. El-Sheshtawy H, Fahim I, Hosny M, El-Badry M. 2022. Optimization of lactic acid production from agro-industrial wastes produced by Kosakonia cowanii. Curr. Res. Green Sustain. 5: 100228.
- 34. Yin C-F, Xu Y, Zhou N-Y. 2020. Biodegradation of polyethylene mulching films by a co-culture of *Acinetobacter* sp. strain NyZ450 and *Bacillus* sp. strain NyZ451 isolated from *Tenebrio molitor* larvae. *Int. Biodeterior. Biodegradation.* **155:** 105089.
- 35. Nourollahi A, Sedighi-Khavidak S, Mokhtari M, Eslami G, Shiranian M. 2019. Isolation and identification of low-density polyethylene (LDPE) biodegrading bacteria from waste landfill in Yazd. *Int. J. Environ. Sci.* 76: 236-250.
- Restrepo-Flórez J-M, Bassi A, Thompson MR. 2014. Microbial degradation and deterioration of polyethylene A review. Int. Biodeterior. Biodegrad. 88: 83-90.
- Al-Makhlafi H, McGuire J, Daeschel M. 1994. Influence of preadsorbed milk proteins on adhesion of *Listeria monocytogenes* to hydrophobic and hydrophilic silica surfaces. *Appl. Environ. Microbiol.* 60: 3560-3565.
- 38. Cunliffe D, Smart C, Alexander C, Vulfson E. 1999. Bacterial adhesion at synthetic surfaces. Appl. Environ. Microbiol. 65: 4995-5002.
- 39. Donlan RM. 2002. Biofilms: microbial life on surfaces. Emerg. Infect. Dis. 8: 881.
- 40. Wang R, Neoh KG, Shi Z, Kang ET, Tambyah PA, Chiong E. 2012. Inhibition of *Escherichia coli* and *Proteus mirabilis* adhesion and biofilm formation on medical grade silicone surface. *Biotechnol. Bioeng.* 109: 336-345.
- Kim H-W, Jo JH, Kim Y-B, Le T-K, Cho C-W, Yun C-H, et al. 2021. Biodegradation of polystyrene by bacteria from the soil in common environments. J. Hazard. Mater. 416: 126239.
- 42. Awasthi S, Srivastava P, Singh P, Tiwary D, Mishra PK. 2017. Biodegradation of thermally treated high-density polyethylene (HDPE) by Klebsiella pneumoniae CH001. 3 Biotech. 7: 332.
- 43. Liu X, Zhang Y, Sun Q, Liu Z, Zhao Y, Fan A, et al. 2022. Rapid colonization and biodegradation of untreated commercial polyethylene wrap by a new strain of *Bacillus velezensis* C5. *J. Environ. Manage.* 301: 113848.
- 44. Roy PK, Titus S, Surékha P, Tulsi E, Deshmukh C, Rajagopal C. 2008. Degradation of abiotically aged LDPE films containing prooxidant by bacterial consortium. *Polym. Degrad. Stab.* 93: 1917-1922.
- 45. Coates J. 2000. Interpretation of infrared spectra, a practical approach. Encyclopedia Anal. Chem. 12: 10815-10837.
- 46. Wang P, Song T, Bu J, Zhang Y, Liu J, Zhao J, et al. 2022. Does bacterial community succession within the polyethylene mulching film plastisphere drive biodegradation? Sci. Total Environ. 824: 153884.
- 47. Hadiyanto H, Khoironi A, Dianratri I, Suherman S, Muhammad F, Vaidyanathan S. 2021. Interactions between polyethylene and polypropylene microplastics and *Spirulina* sp. microalgae in aquatic systems. *Heliyon* 7: e07676.
- Gao R, Sun C. 2021. A marine bacterial community capable of degrading poly(ethylene terephthalate) and polyethylene. J. Hazard. Mater. 416: 125928.
- 49. Das MP, Kumar S. 2015. An approach to low-density polyethylene biodegradation by Bacillus amyloliquefaciens. 3 Biotech. 5: 81-86.
- 50. Yang J, Yang Y, Wu W-M, Zhao J, Jiang L. 2014. Evidence of polyethylene biodegradation by bacterial strains from the guts of plastic-eating waxworms. *Environ. Sci. Technol.* 48: 13776-13784.