

39 **Abstract**

40

41 Coastal fish populations are threatened by multiple anthropogenic impacts, including the 42 accumulation of industrial contaminants and the increasing frequency of hypoxia. Some 43 populations of the Atlantic killifish (*Fundulus heteroclitus*), like those in New Bedford Harbor 44 (NBH), Massachusetts, have evolved a resistance to dioxin-like polychlorinated biphenyls 45 (PCBs) that may influence their ability to cope with secondary stressors. To address this 46 question, we compared hepatic gene expression and DNA methylation patterns in response to 47 mild or severe hypoxia in killifish from NBH and Scorton Creek (SC), a reference population 48 from a relatively pristine environment. We hypothesized that NBH fish would show altered 49 responses to hypoxia due to trade-offs linked to toxicant resistance. Our results revealed 50 substantial differences between populations. SC fish demonstrated a dose-dependent changes 51 in gene expression in response to hypoxia, while NBH fish exhibited a muted transcriptional 52 response to severe hypoxia. Interestingly, NBH fish showed significant DNA methylation 53 changes in response to hypoxia, while SC fish did not exhibit notable epigenetic alterations. 54 These findings suggest that toxicant-adapted killifish may face trade-offs in their molecular 55 response to environmental stress, potentially impacting their ability to survive severe hypoxia in 56 coastal habitats. Further research is needed to elucidate the functional implications of these 57 epigenetic modifications and their role in adaptive stress responses.

58 59

60 **Summary Statement** 61

62 This study reveals how evolved resistance to toxicants in killifish may compromise their ability to 63 respond to hypoxia, highlighting trade-offs that impact survival in stressed coastal environments.

2

79 **Introduction**

80

81 Adaptation to environmental stressors is critical for survival in rapidly changing ecosystems.

- 82 Understanding the physiological and molecular responses that underlie adaptive mechanisms is
- 83 essential for predicting organismal sensitivity. Fish populations, particularly those inhabiting
- 84 coastal waters, often face multiple environmental challenges simultaneously, which can
- 85 compound the stress on these organisms. One such common stressor is hypoxia—low oxygen
- 86 levels in the environment—which is increasingly documented in coastal regions due to
- 87 anthropogenic activities leading excess nutrient loading, harmful algal blooms and climate
- 88 change [1]. The presence and accumulation of anthropogenic chemicals in coastal ecosystems
- 89 poses an additional, co-occurring threat to the health of fish populations. The ability of fish to
- 90 cope with hypoxia is mediated through a range of physiological, transcriptional, and epigenetic
- 91 mechanisms. However, very little is known about how populations chronically exposed to
- 92 toxicants respond to secondary stressors such as hypoxia.
- 93

94 The Atlantic killifish (*Fundulus heteroclitus*) is one of the most ecologically important estuarine 95 fish distributed along the East coast of the United States. Their ability to tolerate wide changes 96 in environmental conditions, including temperature, salinity, oxygen and pH, have made them an 97 ideal model species to investigate the biochemical, physiological and evolutionary basis of 98 environmental adaptation [2-4]. Some populations of killifish are also valuable models for 99 understanding the mechanisms of evolved resistance to toxicants [5]. Populations of killifish 100 inhabiting contaminated coastal waters along the North Atlantic U.S. coast have evolved 101 resistance to some contaminants representing major categories of aryl hydrocarbon pollutants, 102 such as polynuclear aromatic hydrocarbons (PAHs), and halogenated aromatic hydrocarbons 103 such as polychlorinated biphenyls (PCBs), 2,3,7,8-tetrachlorodibenzo-p-dioxin (TCDD, 'dioxin') 104 and other dioxin-like compounds (DLCs) [6]. This evolved resistance involves alterations in 105 signaling through the aryl hydrocarbon receptor (AHR), a ligand-activated transcription factor 106 that forms a heterodimer with aryl hydrocarbon receptor translocator (ARNT), binding to dioxin-107 response elements to regulate the expression of target genes. While there is a great deal of 108 understanding about the physiological and biochemical basis of adaptation to a variety of 109 environmental conditions including toxicants in this species, very little is known about the impact 110 of resistance to toxicants on their ability to respond to subsequent stressors such as hypoxia [7- 111 9].

112

113 The hypoxia-inducible factor (HIF) signaling pathway is a critical cellular response mechanism 114 that enables organisms to respond to hypoxic conditions. Under normal oxygen levels 115 (normoxia), HIF-α subunits (mainly HIF-1α and HIF-2α) are hydroxylated by prolyl hydroxylase 116 enzymes, marking them for degradation via the von Hippel-Lindau (VHL) ubiquitin-proteasome 117 pathway [10, 11]. This prevents the accumulation of HIF-α under normoxia. Under hypoxic 118 conditions, the activity of prolyl hydroxylases is inhibited due to a lack of oxygen, leading to the 119 stabilization of HIF-α [12]. Once stabilized, HIF-α translocates into the nucleus, where it 120 dimerizes with ARNT, also known as HIF-1β. This HIF-α/ARNT complex binds to hypoxia 121 response elements (HREs) and regulates the transcription of target genes [13, 14] . Some of the 122 target genes regulate processes such as angiogenesis (e.g., VEGF), erythropoiesis, glucose 123 metabolism (e.g., GLUT1), and anaerobic metabolism (e.g., LDHA) [15]. Similar responses were 124 observed in killifish, suggesting conserved physiological and molecular mechanisms across 125 species [2-4, 16-20].

126

127 The AHR and HIF pathways exhibit crosstalk primarily through their shared use of the 128 heterodimerizing partner ARNT [21]. Since ARNT is a limiting factor, competition between AHR 129 and HIF for ARNT can influence the balance of responses to environmental toxins and hypoxic 130 stress. This crosstalk may result in altered cellular outcomes, particularly in situations where 131 both pathways are activated simultaneously, such as under environmental stress. One study 132 tested this hypothesis in Atlantic killifish by exposing them to a dioxin-like PCB for three days, 133 followed by a hypoxia challenge [22]. Prior PCB exposure disrupted the classical hypoxia 134 response by increasing hepatic glycolytic enzyme activity, suggesting that dioxin-induced AHR 135 activation could limit ARNT availability for the hypoxia response.

136

137 The objective of this study was to investigate the impact of evolved resistance to toxicants on 138 response to acute hypoxia. We characterized the hepatic gene expression, and DNA 139 methylation patterns in response to two levels of hypoxia in two distinct populations of Atlantic 140 killifish (*Fundulus heteroclitus*). One population is from Scorton Creek, Sandwich, MA (SC), a 141 relatively pristine environment, and is considered sensitive to environmental toxicants [23]. The 142 other population originates from New Bedford Harbor, MA (NBH), a Superfund site heavily 143 contaminated with dioxin-like PCBs, where the killifish have evolved resistance to toxicants. The 144 NBH population represents a unique case study in how toxicant-adapted organisms might 145 exhibit trade-offs in their ability to cope with additional stressors, such as hypoxia. We 146 hypothesized that fish from NBH would exhibit altered responses to acute hypoxia compared to

- 147 the sensitive Scorton Creek (SC) population. Specifically, we predicted that the NBH fish would
- 148 show differential hepatic gene expression and DNA methylation patterns in response to hypoxia,
- 149 potentially compromising their ability to mount an optimal response to low oxygen conditions
- 150 compared to the SC population. Our results demonstrate substantial differences between the
- 151 two populations in their transcriptional and epigenetic responses to hypoxia.
- 152

153 **Results**

154

155 *Effect of hypoxia on loss of equilibrium*

- 156 Neither mild nor severe hypoxia exposure had any effect on the loss of equilibrium during the 6-
- 157 hour exposure period in fish from both populations. Upon initial transfer into the hypoxia
- 158 chamber, fish from both populations exhibited a rapid swimming response for the first 10-15
- 159 minutes. This was followed by a noticeable reduction in swimming activity accompanied by rapid
- 160 ventilation (opercular movements). By the end of the exposure period, the fish were consistently
- 161 found at the bottom of the container, exhibiting slow opercular movements but there was no loss
- 162 of equilibrium as evidenced by their ability to maintain position and coordinated movement.
- 163

164 *Gene expression changes in response to hypoxia* 165

166 Strand-specific RNA sequencing of NBH and SC samples yielded an average of 17.2 million

- 167 reads per sample. Of these, 83% of the reads were uniquely mapped to the genome. The
- 168 summary of mapping statistics and read counts for annotated genes is provided in
- 169 Supplementary Material (RNAseq_supplementaryInformation.xlsx). Principal component
- 170 coordinate analysis revealed one NBH hypoxia sample to be an outlier, which was omitted from
- 171 statistical analysis (Supplementary Figure S1).
- 172

173 *Scorton Creek*

174 We observed a dose-dependent effect of hypoxia on differential gene expression in SC fish.

175 Exposure of SC fish to mild hypoxia revealed 2,241 differentially expressed genes (DEGs), with

- 176 1,170 upregulated and 1,071 downregulated. In response to severe hypoxia, 4191 DEGs were
- 177 observed, with 2,221 upregulated and 1,970 downregulated. A total of 1794 DEGs were shared
- 178 between the two hypoxia groups, with 980 upregulated and 814 downregulated genes (**Figure**
- 179 **2A**). Gene Ontology (GO) analysis of DEGs from mild and severe hypoxia treatment groups
- 180 revealed overrepresentation of GO molecular function (MF) terms related to ATPase activity,
- 181 RNA binding, proteosome and extracellular matrix functions. The list of top 10 overrepresented

182 GO:MF terms among up- and downregulated DEG in SC mild and severe hypoxia groups are 183 shown in **Figure 3**.

184

185 *New Bedford Harbor*

186 In NBH fish, exposure to mild and severe hypoxia elicited differential expression of 3,328 and 187 2,423 genes, respectively (**Table 1**). Among the 3328 genes differentially expressed in response 188 to mild hypoxia, 1717 were upregulated and 1611 genes were downregulated. Whereas in 189 response to severe hypoxia, 1,272 of the 2,423 DEGs were upregulated, and 1,151 genes were 190 downregulated. Comparison of the upregulated DEGs from the mild and severe hypoxia groups 191 revealed 847 genes shared between the two groups (**Figure 2B**), while a similar comparison of 192 downregulated DEGs revealed 796 shared DEGs. 193

194 GO analysis of mild and severe hypoxia upregulated DEGs showed overrepresentation of

195 GO:MF terms related to mRNA splicing, translation, and proteasomal degradation. The terms

196 enriched among downregulated DEGs include specific pathways such as cell adhesion and

- 197 extracellular matrix functions. The top 10 overrepresented terms among up and downregulated
- 198 DEGs are shown in **Figure 4**.
- 199

200 *Population Differences*

201 Comparison of NBH and SC control groups revealed differential expression of 307 genes.

202 Among them 159 and 148 are up- and downregulated in NBH, respectively, in comparison to

203 SC. The GO terms enriched among these genes are shown in Supplementary Figure S2.

204 Comparison of mean expression (log counts per million) of all the DEGs demonstrates that NBH

205 fish have a more muted gene expression response to severe hypoxia in comparison to SC fish

206 (**Figure 5**).

207

208 We compared the two hypoxia treatment groups between NBH and SC to identify unique genes 209 across treatments. A total of 453 upregulated and 349 downregulated genes were shared

210 between the two populations and two hypoxia treatment groups (**Figure 6**). Heatmap

211 representation of these genes shows magnitude of change between the two hypoxia groups in

212 both populations (Supplementary Figure S3). Overrepresented GO terms of common up- and

213 downregulated genes are shown in Supplementary Figure S4. These terms are similar to those

- 214 observed in SC and NBH in response to hypoxia.
- 215

216 *DNA methylation changes in response to hypoxia*

- 217 Reduced representation bisulfite sequencing yielded 547.1 million total paired reads (14.4 to 218 33.3 million paired reads per sample). Of these reads, 544.5 million (99.5%) remained after 219 quality-trimming. From the trimmed reads, 88.6% to 94.5% total reads were aligned to the *F.* 220 *heteroclitus* genome. The detailed list of mapping statistics, raw and trimmed reads per sample 221 is provided in the supplementary information (RRBS Supplementary information.xlsx). A total 222 of 439,469 CpGs (5.4% of the 8,094,243 CpGs in the *F. heteroclitus* genome) had between 10x 223 and 500x coverage in at least one sample after CpG filtering. Among them, 148,752 CpG sites 224 (35.2% methylated and 64.8% unmethylated) were present in all samples and were used in 225 downstream analysis.
- 226

227 There was no statistically significant difference in global methylation level between NBH (28.2%) 228 and SC liver samples (28.8%). **Figure 7** shows the mean CpG methylation density plots and 229 total number of CpG sites in all the treatment groups. Only one DMR was identified between 230 NBH and SC fish exposed to control conditions. This DMR is in chromosome 10 in an intron of 231 SHISA6, a gene that encodes AMPA glutamate receptor subunit (Klassen et al., 2016). This 232 DMR also overlaps with an annotated CpG island (identified as CpG island 77 in the genome)

- 233 and is hypomethylated in NBH in comparison to SC fish.
- 234

235 Hypoxia exposure did not substantially alter global DNA methylation levels in both populations. 236 In SC fish, global CpG DNA methylation levels were 28.8% in the controls and mild hypoxia, and 237 29.2% in severe hypoxia group. No DMR were observed in response to mild or severe hypoxia 238 in this population. In NBH fish, global DNA methylation level was 27.9% in the control group, 239 27.2% in mild hypoxia, and 29.7% in severe hypoxia. Comparison of control and mild hypoxia 240 groups revealed 10 DMR, all of which were hypermethylated. Similar comparison between 241 control and severe hypoxia revealed 59 DMRs. Among them five were hypomethylated and 54 242 were hypermethylated (**Figure 8**). The majority of the NBH DMR were found in CpG islands 243 annotated in the genome. Only two DMR were shared between the mild and severe hypoxia 244 groups, and they were hypermethylated. The genomic coordinates of these DMR are provided 245 in supplementary information. We did not observe any significant correlation between DNA 246 methylation in DMRs and the expression level of the associated genes (Supplemental Figures 247 S5 and S6.)

- 248
- 249 **Discussion**

250

251 The findings from this study demonstrate a shift in physiological responses in fish adapted to 252 toxicants and this could compromise their capacity to respond to secondary stressors such as 253 hypoxia. Our results show that both toxicant-sensitive (SC) and resistant killifish (NBH) respond 254 to hypoxia but there are considerable differences in their responses as measured by gene 255 expression and DNA methylation patterns. Atlantic killifish from SC show dose dependent 256 changes in gene expression patterns in response to hypoxia and no changes in DNA 257 methylation patterns, whereas the fish from NBH showed a muted gene expression response to 258 severe hypoxia suggesting compromised ability to mount a stress response. However, NBH fish 259 showed a modest but significant DNA methylation changes in response to hypoxia. Together, 260 these results reveal different response mechanisms and strategies to cope with hypoxia 261 between sensitive and resistant fish populations.

262

263 *HIF signaling in toxicant-sensitive and toxicant-resistant populations*

264

265 Transcriptional responses to hypoxia are well-documented in fish species [24-28], including 266 Atlantic killifish [29-31]. Our findings demonstrate that both SC and NBH killifish exhibit 267 responses to hypoxia, though their reactions differ significantly. The hypoxia-inducible 268 transcription factors (HIF-1α, HIF-2α, HIF-3α) play a major role in the transcriptional activation of 269 the hypoxic response [32]. The expression of these genes under hypoxic conditions depends on 270 the intensity and duration of hypoxia exposure. Interestingly, we did not observe any changes in 271 the expression of HIF1α and HIF2α genes in response to mild or severe hypoxia in either of the 272 populations. However, we observed an increased expression of HIF3 α only in NBH fish (mild 273 hypoxia – log FC 1.75, FDR 7.45E-04; severe hypoxia – log FC 1.54, FDR 4.07E-03). A 274 previous study in a closely related species, *Fundulus grandis*, showed no significant differences 275 in mRNA expression of any of these HIF genes in response to hypoxia (1 mg oxygen/L) after 6 276 and 24 hours post-exposure [33]. However, studies in mammalian cell culture systems have 277 shown differences in temporal profiles of HIF genes in response to hypoxia. In human 278 endothelial cells, HIF-1α expression is maximal after 4 hours of hypoxia and then is dramatically 279 reduced by 8 hours [34]. In contrast, HIF-2α is maximal at 8 hours and remains elevated up to 280 24 hours [34-36]. Unlike HIF-1α and -2α, moderate hypoxia exposure induced HIF3α mRNA 281 expression within 2 hours [35]. However, very little is known about the functional role of HIF3α in 282 hypoxia responses. The upregulation of HIF3α in response to hypoxia only in NBH fish suggests 283 that adaptation to contaminated environment may have altered its regulation, with potential

284 implications for the hypoxia response. Further studies are needed to understand the roles of all 285 three HIFs under different hypoxic conditions and at different life stages.

286

287 Another set of HIF pathway genes that were differentially expressed are the prolyl hydroxylase 288 domain-containing proteins (PHD) [10]. There are three 2-oxoglutarate-dependent PHD proteins 289 (PHD1, PHD2, and PHD3, encoded by EGLN2, EGLN1, and EGLN3, respectively) that are 290 involved in HIFα ubiquitination and proteasomal degradation under normoxic conditions [37]. 291 Hypoxia has been shown to induce the expression of EGLN1 and -3 mRNAs, but not EGLN2, in 292 several cell types [12]. In most cases, the induction of EGLN3 mRNA is much more prominent 293 than that of EGLN1. We observed a several-fold increase in the *egln3* expression (SC: log FC 294 8.77, FDR 1.82E-06; NBH: log FC 3.51, FDR 0.028) and modest but significant *elgn1a* 295 upregulation in response to severe hypoxia in both populations. We also observed upregulation 296 of *egln2* (logFC 1.33, FDR 9.59E-04) in response to severe hypoxia in SC fish, while it was not 297 significant in NBH fish. Although very little is known about the role of *egln2* in oxygen sensing 298 and hypoxia tolerance, the differences in expression between the two populations are intriguing. 299 It has been suggested that increase in EGLN activity during hypoxia acts as a regulatory 300 feedback loop for fast elimination of HIF after reoxygenation [38].

301

302 In addition to the direct regulation of HIF transcription by hypoxia, multiple signaling pathways 303 have been shown to play a role in the regulation of HIF gene expression in a variety of model 304 systems. These include the PI3K-mTOR, interleukin-6 (IL-6), ERK, and MAPK signaling 305 pathways [15]. We observed differential expression of several genes associated with these 306 signaling pathways only in NBH fish, suggesting a role in adaptation to toxicants. Despite the 307 lack of an increase in HIF-1α expression, the differential expression of these signaling pathways 308 suggests that NBH fish utilize cross-talk between multiple signaling pathways to cope with 309 hypoxia in a way that appears to be distinct from SC fish.

310

311 *Transcriptional responses to hypoxia*

312

313 While responses to hypoxia are well-documented across various organisms, encompassing 314 metabolic changes, angiogenesis, and vascularization [3, 4, 15, 16], our results, based on a 6- 315 hour hypoxia exposure, revealed gene expression changes that were more focused on 316 transcription, translation, and cell cycle-related genes, rather than the classical hypoxia 317 responses. This contrasts with studies in *in vitro* mammalian systems, where hypoxia target

318 genes are differentially expressed in shorter time scales [39]. The lack of similar responses in 319 our study could be due to differences in the timescales of transcription and translation between 320 fish maintained at 20°C and mammalian cells held at 37°C [39, 40]. SC fish exhibited dose-321 dependent changes in gene expression, indicating a gradual and proportional response to 322 varying levels of hypoxia. In contrast, NBH fish demonstrated drastic gene expression changes 323 in response to mild hypoxia, but a muted response to severe hypoxia. This pattern suggests that 324 NBH fish have a lower tolerance threshold for hypoxic conditions and are unable to induce a 325 robust transcriptional response to severe hypoxia. This could be potentially due to higher 326 energetic costs for survival in contaminated environment leading to compromised ability to 327 respond to additional stressors [7, 8].

328 In SC fish, mild hypoxia exposure caused an overrepresentation of genes associated with efflux 329 transporters, transcription factor activity, mRNA and nucleic acid binding, and proteasomal 330 functions. All these functions have been previously shown to be altered by hypoxia [11, 41-44]. 331 For instance, efflux pumps belonging to the ATP-binding cassette (ABC) superfamily of 332 membrane transporters are known to play a significant role in cellular protection against 333 oxidative stress [45]. Several genes within the ABC family of efflux transporters are expressed 334 in the liver and are involved in the transport of glutathione and glucuronide conjugates [46-48]. 335 These findings suggest that mild hypoxia increases reactive oxygen species (ROS), with 336 glutathione playing a crucial role in neutralizing free radicals and protecting the liver. The 337 glutathione conjugates are subsequently eliminated from hepatocytes by the efflux pumps, 338 contributing to cellular detoxification [48]. Additionally, mild hypoxia caused the upregulation of 339 genes associated with RNA polymerase II (RNAPII) general transcription factor activity and the 340 downregulation of cis-regulatory sequence-specific DNA binding RNAPII activity. This is not 341 surprising, given the evidence that hypoxia affects the transcription of hundreds of genes [13]. 342 Under normoxic conditions, almost all HIF target genes display an open chromatin structure and 343 harbor transcriptionally active but paused RNA polymerase II [49]. Changes in the expression of 344 RNAPII activity genes in response to hypoxia suggests the release of paused Pol II into 345 productive RNA synthesis by recruiting various coactivators, repressors, and chromatin 346 remodelers, resulting in either the activation or inhibition of transcription of target genes [49].

347 In response to severe hypoxia, SC fish showed an overrepresentation of genes associated with 348 RNA helicases. They play an important role in cellular RNA metabolism, including transcription, 349 pre-mRNA splicing, RNA export, storage, decay, and translation. Recently, RNA helicases have 350 been shown to be involved in many biological processes, including DNA damage repair, cellular

351 stress response, hypoxia and antiviral defense [50, 51]. Another major overrepresented group of 352 genes in response to severe hypoxia in SC fish are the RNA binding proteins (RBPs) – 353 important players in mRNA turnover (decay and stabilization) and translation [52]. RNA 354 helicases and RBPs have been shown to be essential for the adaptive cellular response to 355 hypoxia [50, 53]. While there is very little understanding on the role of RBPs in environmental 356 model species such as killifish, the upregulation of these genes in response to hypoxia suggests 357 highly conserved cellular mechanisms.

358 Interestingly, the genes and pathways that were upregulated in SC fish under severe hypoxia 359 were similarly upregulated in NBH fish exposed to mild hypoxia. This suggests that the toxicant-360 adapted NBH fish are more sensitive to mild hypoxia, undergoing more drastic transcriptional 361 changes. The upregulation of RNA helicases and RBPs indicates post-transcriptional regulation 362 of pre-existing RNAs [53], suggesting that NBH fish may reduce transcription under stress. 363 Indeed, this was observed under severe hypoxia in NBH fish, where the number of DEGs was 364 lower compared to SC fish, indicating a drastic reduction in transcription.

365

366 Hypoxia exposure downregulated genes related to calcium signaling, oxidoreductase activity, 367 and extracellular matrix (ECM) modeling proteins. These pathways were altered in both hypoxia 368 treatments and across both populations, suggesting they are vital for cellular response to 369 hypoxia. It is well-established that hypoxia impairs mitochondrial respiration and ATP synthesis, 370 leading to an increased production of reactive oxygen species (ROS) and calcium release from 371 the ER into the cytosol [15]. The resulting elevated cytosolic calcium levels cause increased 372 calcium uptake into mitochondria and mitochondrial calcium overload, which in turn leads to 373 mitochondrial depolarization and the initiation of cell death [54]. The decreased expression of 374 genes associated with calcium signaling suggests an adaptation toward hypoxia tolerance. 375 Similarly, hypoxia involves a switch from oxidative phosphorylation to glycolysis, resulting in 376 increased production of NADH and an imbalance in the NAD+ and NADH ratio, causing altered 377 redox potential [14, 55]. These conditions favor the overexpression of many redox enzymes, 378 such as cytochrome P450 reductase and nitroreductases. Even though these genes were not 379 altered, genes that are dependent on NAD+ or NADH and play critical roles in metabolism were 380 downregulated. They include *agmo* (alkylglycerol monooxygenase), *sc5d* (sterol-C5- 381 desaturase), *hsd11b2* (11-β-hydroxysteroid dehydrogenase 2), *bdh1* (d-beta-hydroxybutyrate 382 dehydrogenase), and *foxred1* (FAD-dependent oxidoreductase domain-containing 1) [14]. It

383 remains to be determined whether the downregulation of these genes under hypoxia is adaptive 384 or maladaptive.

385

386 Another significant group of genes that are commonly downregulated in response to hypoxia are 387 ECM components genes. The ECM is a complex network of proteins and other molecules that 388 provide structural and biochemical support to surrounding cells [56]. Its composition and 389 function are crucial for tissue integrity, cell behavior, and overall organism health [57]. Hypoxia 390 has been shown to cause significant effects on the ECM in various aquatic species [58-61]. In 391 addition, there is growing evidence suggesting the role of hypoxia and HIFs in reprogramming 392 cancer cells by regulating extracellular matrix (ECM) deposition, remodeling and degradation, 393 thereby promoting cancer metastasis [62]. We observed downregulation genes associated with 394 collagen synthesis, which are essential for maintaining the structural integrity of tissues. This 395 could lead to weakened tissue architecture or cause tissue remodeling such as changes in 396 vascularization, which can be adaptive to survive under hypoxic conditions. The molecular 397 mechanisms associated with these changes could be either by direct regulation of ECM 398 pathway genes by HIF proteins or indirectly by the induction of oxidative stress [57, 63]. Overall, 399 the effects of hypoxia on the ECM in environmental species such as killifish could be critical, as 400 they frequently encounter hypoxic conditions under diel cycle and ECM-related adaptations are 401 critical for survival.

402

403 *Epigenetic changes in responses to hypoxia*

404 Epigenetic effects, particularly DNA methylation changes in response to hypoxia, are well 405 documented across a variety of fish species [64-68]. Hypoxia has been shown to cause both 406 global and gene-specific alterations in DNA methylation patterns, which can influence gene 407 expression, development, and stress response pathways. For instance, in species like zebrafish 408 (*Danio rerio*) and Atlantic salmon (*Salmo salar*), hypoxic conditions have been associated with 409 both hypermethylation and hypomethylation of key regulatory genes involved in metabolic 410 adaptation and oxidative stress responses [64, 68]. These epigenetic changes are hypothesized 411 to help fish cope with hypoxia by modulating critical pathways for survival, growth, and 412 development.

413 To our knowledge, this is the first study where genome-wide DNA methylation patterns are 414 profiled in Atlantic killifish**.** In this study, we hypothesized that resistant and sensitive populations 415 of Atlantic killifish would exhibit distinct hepatic DNA methylation patterns, possibly reflecting

416 their differential tolerance to dioxin-like PCBs. However, we did not observe distinct population 417 differences in DNA methylation (NBH vs SC control groups), suggesting that parental exposure 418 to dioxin-like PCBs did not have a detectable effect on DNA methylation in the offspring held 419 under normoxic conditions.

420 Hypoxia exposure induced DNA methylation changes that were markedly different in the two 421 populations**.** In the sensitive SC population, no differentially methylated regions (DMRs) were 422 identified in response to either mild or severe hypoxia. In contrast, hypoxia resulted in significant 423 changes in DNA methylation in the toxicant-resistant NBH population. Most of these DMRs are 424 located in intronic regions and were enriched within CpG islands, which are known to play a 425 critical role in regulating gene expression [69]. These findings suggest that the NBH population 426 may have a more plastic epigenetic response to hypoxia compared to the SC population, which 427 could reflect underlying differences in their adaptive capacities. While further research is needed 428 to elucidate the functional consequences of these methylation changes, this study provides new 429 insights into the epigenetic mechanisms by which fish populations respond to environmental 430 stressors like hypoxia.

431 The majority of hypoxia induced DMRs are hypermethylated supporting previous observations 432 that hypoxia cause DNA hypermethylation [70-72]. Associating DMRs with genes revealed that 433 some of the DMRs are related to genes involved in sialylation, vascularization and development. 434 Four DMRs were associated with the gene *St6galnac3*, which is involved in the sialylation 435 pathway. Sialylation refers to the addition of sialic acid units to oligosaccharides and 436 glycoproteins [73]. Sialic acids moieties act as bridging molecules, facilitating communication 437 between cells and the extracellular matrix. Additionally, one DMR was associated with *beta-*438 *chimaerin*, a protein linked to vascularization [74]. Hypoxia has been shown to influence both 439 sialylation and vascularization processes, particularly in cancer models [74, 75], suggesting that 440 similar mechanisms may be involved in the hypoxic responses observed in this study. However, 441 we did not observe any significant correlation between these DNA methylation changes and 442 expression of the associated genes. None of the 58 genes associated with DMRs in NBH fish 443 were differentially expressed suggesting a temporal lag in DNA methylation changes and gene 444 expression. In addition, several studies have shown that majority of the DNA methylation 445 changes are not correlated with gene expression [76-78]. This suggests that epigenetic 446 regulation of gene expression is multilayered with many levels of control, involving DNA 447 methylation, histone modifications and chromatin organization.

448 While differential methylation did not correlate with gene expression changes in NBH, we 449 observed differential expression of several chromatin modifier genes, particularly histone lysine 450 demethylases (KDMs), in response to hypoxia in both fish populations. This is expected, as 451 KDMs belong to the family of 2-oxoglutarate-dependent dioxygenases, which function as 452 oxygen sensors [13]. The differentially expressed KDM genes include *kdm1aa*, *kdm2aa*, 453 *kdm2ab*, *kdm2ba*, *kdm3b*, *kdm4b*, *kdm5ba*, *kdm5bb*, *kdm5c*, *kdm6a*, *kdm6ba*, and *kdm7aa*. 454 These genes have been shown to be directly regulated by hypoxia-inducible factors (HIFs), 455 linking their expression to the cellular response to hypoxic stress [79, 80]. Given the differential 456 expression of histone lysine demethylases in response to hypoxia, it would be intriguing to 457 investigate the role of chromatin modifiers in the adaptation to hypoxia in environmental 458 species, as they may be key regulators of gene expression in low-oxygen environments.

459 **Conclusions**

460 This study highlights the complex and distinct physiological and epigenetic responses of Atlantic 461 killifish populations adapted to toxicants when exposed to hypoxia. Our findings suggest that the 462 capacity to respond to secondary stressors such as hypoxia may be altered in populations 463 adapted to environmental contaminants. As expected, the toxicant-sensitive SC fish displayed a 464 dose-dependent response to hypoxia exposure. However, the toxicant-resistant NBH fish, 465 exhibited muted transcriptional responses but a more pronounced DNA methylation response to 466 severe hypoxia, suggesting different molecular mechanisms in this population. Importantly, the 467 differential DNA methylation patterns in response to hypoxia between the two populations 468 indicate differences in epigenetic plasticity, which needs further investigation. Overall, this 469 research provides valuable insights into the diverse molecular mechanisms by which fish 470 populations, with different environmental histories, respond to hypoxia and highlights the need 471 for further exploration of epigenetic and chromatin-level responses in the context of 472 environmental adaptation.

473

477

474 **Materials and Methods**

475
476 **Experimental fish**

478 The animal husbandry and experimental procedures used in this study were approved by the 479 Animal Care and Use Committee of the Woods Hole Oceanographic Institution. Mature adult 480 male and female killifish from Scorton Creek (SC; Sandwich, MA) and New Bedford Harbor

481 (NBH; New Bedford, MA) were collected using minnow traps, as described previously [81]. Fish 482 were maintained in the Redfield Laboratory (WHOI) with continuous flow-through seawater 483 (SW) at 18–20°C, saturated dissolved oxygen (21% oxygen saturation or 7.21 mg O₂ L⁻¹), and 484 14h:10h light/dark photoperiod conditions.

485

486 F1 generation of embryos from SC and NBH were obtained by *In vitro* fertilization following 487 established protocols (Karchner et al., 1999). Briefly, 4-5 female fish from SC or NBH (15.5 ± 488 1.2 g mean wet mass) were lightly anesthetized with Tricaine (MS222; buffered with sodium 489 bicarbonate, Sigma-Aldrich, St. Louis, MO, USA) and oocytes were obtained for *in vitro* 490 fertilization by gently squeezing the abdomen. Oocytes were collected in glass petri dishes with 491 filtered SW (30 parts per thousand; ppt). Milt was obtained by euthanizing 2-3 mature males 492 $(13.5 \pm 1.1 \text{ g}$ mean wet mass) from the same population in MS222, dissecting out the gonads, 493 and chopping them with a scalpel blade in seawater. A few drops of milt were added to the 494 oocytes for fertilization. Approximately 20 minutes after the addition of milt, embryos were rinsed 495 with filtered SW to remove any excess sperm. Fertilized embryos were reared at 23°C under 496 14h:10h light/dark photoperiod conditions until hatching. Larvae were raised in 2-gallon 497 aquarium tanks in aerated seawater for six months. During larval rearing, fish were fed brine 498 shrimp daily and water was exchanged every 2-3 weeks. Oxygen concentration was measured 499 in the tanks once every 2-3 days and oxygen saturation was above 20% throughout the rearing 500 period.

501

503

502 *Hypoxia exposure*

504 Six-month-old killifish juveniles from SC (149 \pm 43 mg mean wet mass) and NBH (145 \pm 39 mg 505 mean wet mass) were exposed to either mild (10% oxygen saturation, 3.46 mg O₂ L⁻¹; n = 5 per 506 population) or severe hypoxia (5% oxygen saturation, 1.72 mg O_2 L⁻¹; n = 5 per population) for 6 507 hours. These two hypoxia levels were chosen based on preliminary experiments with the same 508 cohort of fish, where the loss of equilibrium (LOE) was assessed under 1% and 5% oxygen 509 saturation in both NBH and SC fish. Six hours of exposure to 5% oxygen saturation did not 510 cause LOE in fish from either population ($n = 5$ individual fish per population), whereas 1% 511 oxygen saturation caused LOE within 6 hours in fish from both populations. 512

513 Hypoxia exposure set up includes pyrex glass dishes (270 mL volume) equipped with oxygen 514 sensor spots (PreSens Precision Sensing GmbH, Germany) placed inside hypoxia chambers

515 (STEMCELL Technologies Inc.) with pre-mixed air set to 5 or 10% oxygen pumped into the

516 chambers continuously. A control group (normoxia, 20.9% oxygen saturation; n = 5 per

- 517 population) was maintained on the benchtop (**Figure 1**). Prior to introducing the fish to hypoxia,
- 518 250 mL filtered seawater was added to the pyrex glass dishes and was allowed to equilibrate
- 519 overnight to ensure that the water has reached the respective treatment conditions. Oxygen
- 520 levels in individual beakers were checked prior to introducing the fish using a FireString oxygen
- 521 sensor (PyroScience, Germany) and were found to be at the treatment conditions in each of the
- 522 individual dishes. At the start of the experiment, individual fish were quickly introduced into the
- 523 pyrex dishes and chambers closed quickly. Fish were maintained at treatment conditions for 6
- 524 hours. At the end of the exposure period, oxygen levels were measured.
- 525

527

526 *Isolation of total RNA and genomic DNA from liver samples*

- 528 Simultaneous isolation of genomic DNA and total RNA from liver tissues was performed using
- 529 the ZR-Duet DNA/RNA Mini Prep kit (Zymo Research, California). RNA was treated with DNase
- 530 during the isolation process. DNA and RNA were quantified using the Nanodrop
- 531 Spectrophotometer. The quality of DNA and RNA was checked using the Agilent 4200 and 2200
- 532 Tape Station systems, respectively. The DNA and RNA integrity numbers of all samples were
- 533 between 9 and 10.
- 534

535 *RNA sequencing*

- 536 Libraries were constructed using Illumina stranded library preparation kit following
- 537 manufacturer's protocol. Single end 50bp reads were sequenced using Illumina HiSeq2500
- 538 platform. RNA sequencing library construction and sequencing were done at the Tufts university 539 core facility.
- 540

541 *Reduced Representation Bisulfite Sequencing (RRBS)*

542 Library preparation was performed using the Premium RRBS kit (Diagenode). In brief, 100 ng 543 DNA from each sample were enzymatically digested by the restriction enzyme MspI at 37°C for 544 12 hours. Following ends preparation, a different set of adaptors was added to each sample and 545 adaptor ligation was performed by the addition of ligase. Size selection of adaptor-ligated DNA 546 fragments was performed by Agencourt AMPure XP beads (Beckman Coulter) and the DNA was 547 eluted in Resuspension buffer. Part of the eluted sample was subjected to qPCR using 2X KAPA 548 HiFi HotStart ReadyMix (Kapa Biosystems) for quantification and subsequent pooling per 9 549 samples. The pooling was performed according to two parameters: the Ct value and the adaptor

550 ID of each sample. The pooling was followed by a cleanup with AMPure XP beads to reduce the

551 volumes. Bisulfite treatment was performed, and bisulfite-converted DNA was eluted twice in BS 552 Elution buffer. Part of the bisulfite converted library was used in qPCR for the determination of 553 the optimal cycle number for the enrichment PCR. 2X MethylTaq Plus Master Mix was used for 554 the amplification PCR and a last cleanup with AMPure XP beads followed. PCR product was run 555 on an 2% agarose gel to remove adaptor dimers. The quality of the final libraries was checked 556 on an Agilent 2100 High Sensitivity DNA chip. The concentration was determined by performing 557 qPCR on the samples using a dilution of PhiX index3 as standard. Paired end 50bp reads were 558 sequenced on an Illumina HiSeq4000 platform by a commercial facility (NXT-Dx, Ghent, 559 Belgium).

560

561 *Genome Information and Feature Tracks*

562 The *F. heteroclitus* genome (https://www.ncbi.nlm.nih.gov/datasets/genome/GCF_011125445.2/)

563 was used for all analyses. Genome feature information was pulled directly from the genome to 564 generate gene, coding sequence (CDS), exon, and lncRNA genome feature tracks using 565 Gnomom, RefSeq, cmsearch, and tRNAscan-SE annotations. Chromosome name and length 566 information was also extracted from the genome to generate additional feature tracks using 567 bedtools v2.31.1 [82]. A non-coding sequence track was created by using the complement of the 568 CDS track (complementBed). Similarly, the intergenic genome feature track was created using 569 the complement of the gene track. The intersection (intersectBed) between the non-coding 570 sequence and gene tracks were used to create an intron track. All genome feature tracks are 571 available in an Open Science Framework repository (doi.org/10.17605/OSF.IO/NZRA8).

572

573 *RNA sequencing and analysis*

574 Raw data files were assessed for quality using FastQC Version 0.11.9 [83] prior to

575 preprocessing. Preprocessing was done by trimming the adaptor sequences using Trimmomatic

576 (Version 0.25) and removing any reads with low sequence quality (Phred score < 20) [84].

577 Trimmed sequence reads were mapped to the *F. heteroclitus* genome using the STAR aligner

- 578 v.2.6.1d [85]. The number of reads mapped to annotated regions of the genome was obtained
- 579 using HTSeq-count v.0.11.1 [86]. Statistical analysis was conducted using edgeR v.3.40.2, a
- 580 Bioconductor package [87]. Transcripts from all samples were compiled into a DGEList, and
- 581 lowly expressed transcripts were filtered out using the filterByExpr function. Sample ordination
- 582 was visualized using multidimensional scaling analysis with the ape v5.8 package [88],
- 583 revealing an outlying sample that was removed from subsequent analysis (Fig. S1). We used
- 584 the quasi-likelihood model in edgeR (glmQLFTest) to perform differential gene expression

585 analysis. Only genes with false discovery rate (FDR) of <5% were considered to be differentially 586 expressed. Raw data has been deposited in gene expression omnibus (Accession number 587 GSE278569).

588

589 Functional annotation of DEGs was done using gene ontology (GO) biological process (GO:BP) 590 and molecular function (GO:MF) terms. Identification of overrepresented GO terms (*p* value < 591 0.05) among sets of DEGs was done using the enricher function in ClusterProfiler v.4.6.2 [89]. 592 The background gene list included all expressed transcripts in the filtered DGEList. Similar 593 GO:BP or GO:MF terms were clustered based on the frequency of shared genes using the R 594 package *rrvgo* 1.16.0 [90]. Representative parent terms from each cluster were chosen based 595 on the lowest *p value*.

596

597 *DNA methylation profiling by Reduced Representation Bisulfite Sequencing (RRBS)*

598 The Bisulfite Analysis Toolkit (BAT) [91] was used for RRBS analysis. Prior to analysis, raw data

599 was quality trimmed with TrimGalore! v.0.6.6 [92]. Trimming was performed on non-directional (- 600 -non_directional) paired-end reads (--paired). An additional 2 bp were trimmed from the 3' end of

601 the first read and 5' end of the second read (--rrbs). Sequence quality was assessed with

602 FastQC v0.11.9 [83] and MultiQC v1.11 [93] after trimming.

603

604 Trimmed paired reads were aligned to the genome using BAT_mapping module specifying non-605 directional input (-F 2). Mapping statistics and methylation calling was done using 606 BAT_mapping_stat and BAT_calling modules, respectively. CpG methylation data was filtered to 607 retain only a minimum 10 and maximum 500 reads per sample (--MD_min 10, –MD_max 500 -- 608 CG). The data were sorted using bedGraphs (sortBed v2.29.1; [82]) and merged into treatment-609 specific groups (BAT_summarize). Within each group, one sample was allowed to have missing 610 data for a CpG locus (--mis1 1, --mis2 1). If data were missing for more than one sample at a 611 particular CpG, it was not included in the downstream analysis. Chromosome lengths were 612 specified (--cs) for merging methylation information accurately. BAT_overview was used to 613 obtain average methylation rate per sample in each group, hierarchical clustering of sample 614 methylation rates, distribution of CpG methylation, comparison of methylation rate between 615 groups for common loci, and differences in mean methylation rate between groups. Differentially 616 methylated regions (DMR) — defined as regions with at least 10 CpGs, a minimum methylation 617 rate difference of 0.1, and q-value < 0.05 — were identified for each comparison using

- 618 BAT_DMRcalling module. The closest genome feature to each DMR was characterized using
- 619 closestBed.

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949 950 **Figure 6.** Comparison of all treatment groups. Venn diagram of upregulated (A) and 951 downregulated (B) genes in all treatment groups revealed a core set of genes (802 DEGs) 952 altered by hypoxia exposure irrespective of the level of hypoxia and population. GO analysis 953 and heatmap representation of these core genes are provided in the supplementary information. 954 955 **Figure 7.** DNA methylation landscape in *F. heteroclitus*. **(A)** CpG DNA methylation density plots 956 showing proportion of CpG methylation in different population and treatment groups. Inset 957 shows the density plots of CpG sites with methylation levels below 25%. **(B)** Percent of 958 methylated (>0% methylation; dark grey) and unmethylated (0% methylated; light grey) CpGs in 959 various population and treatment groups. The average number of methylated and unmethylated 960 CpGs are shown. 961 962 **Figure 8.** Volcano plot showing differentially methylated regions (DMRs) in response to **(A)** 10% 963 (mild) and **(B)** 5% (severe) hypoxia exposure in New Bedford Harbor fish. Mean methylation 964 difference (x-axis) between severe hypoxia and control group is plotted against q-value (y-axis). 965 Each green and red spot represents a statistically significant hypo- and hypermethylated region, 966 respectively. 967 968

Figure 1.

A

Figure 2.

A. Scorton Creek

B. New Bedford Harbor

A Figure 4. Mild hypoxia (NBH)

Figure 5.

Figure 6.

A. Upregulated DEGs B. Downregulated DEGs

Figure 8.

