










Common Presence of Phototrophic *Gemmatimonadota* in Temperate Freshwater Lakes

 Izabela Mujakić,^{a,b}
 Adrian-Ştefan Andrei,^c
 Tanja Shabarova,^c
 Lívia Kolesár Fecskeová,^a
 Michaela M. Salcher,^c
 Kasia Piwosz,^{a*}
 Rohit Ghai,^c
 Michal Koblížek^{a,b}

^aLaboratory of Anoxygenic Phototrophs, Centre Algatech, Institute of Microbiology of the Czech Academy of Sciences, Třeboň, Czechia

^bDepartment of Ecosystem Biology, Faculty of Science, University of South Bohemia, České Budějovice, Czechia

^cDepartment of Aquatic Microbial Ecology, Institute of Hydrobiology, Biology Centre of the Czech Academy of Sciences, České Budějovice, Czechia

ABSTRACT Members of the bacterial phylum *Gemmatimonadota* are ubiquitous in most natural environments and represent one of the top 10 most abundant bacterial phyla in soil. Sequences affiliated with *Gemmatimonadota* were also reported from diverse aquatic habitats; however, it remains unknown whether they are native organisms or represent bacteria passively transported from sediment or soil. To address this question, we analyzed metagenomes constructed from five freshwater lakes in central Europe. Based on the 16S rRNA gene frequency, *Gemmatimonadota* represented from 0.02 to 0.6% of all bacteria in the epilimnion and between 0.1 and 1% in the hypolimnion. These proportions were independently confirmed using catalyzed reporter deposition-fluorescence *in situ* hybridization (CARD-FISH). Some cells in the epilimnion were attached to diatoms (*Fragilaria* sp.) or cyanobacteria (*Microcystis* sp.), which suggests a close association with phytoplankton. In addition, we reconstructed 45 metagenome-assembled genomes (MAGs) related to *Gemmatimonadota*. They represent several novel lineages, which persist in the studied lakes during the seasons. Three lineages contained photosynthesis gene clusters. One of these lineages was related to *Gemmatimonas phototrophica* and represented the majority of *Gemmatimonadota* retrieved from the lakes' epilimnion. The other two lineages came from hypolimnion and probably represented novel photoheterotrophic genera. None of these phototrophic MAGs contained genes for carbon fixation. Since most of the identified MAGs were present during the whole year and cells associated with phytoplankton were observed, we conclude that they represent truly limnic *Gemmatimonadota* distinct from the previously described species isolated from soils or sediments.

IMPORTANCE Photoheterotrophic bacterial phyla such as *Gemmatimonadota* are key components of many natural environments. Its first photoheterotrophic cultured member, *Gemmatimonas phototrophica*, was isolated in 2014 from a shallow lake in the Gobi Desert. It contains a unique type of photosynthetic complex encoded by a set of genes which were likely received via horizontal transfer from *Proteobacteria*. We were intrigued to discover how widespread this group is in the natural environment. In the presented study, we analyzed 45 metagenome-assembled genomes (MAGs) that were obtained from five freshwater lakes in Switzerland and Czechia. Interestingly, it was found that phototrophic *Gemmatimonadota* are relatively common in euphotic zones of the studied lakes, whereas heterotrophic *Gemmatimonadota* prevail in deeper waters. Moreover, our analysis of the MAGs documented that these freshwater species contain almost the same set of photosynthesis genes identified before in *Gemmatimonas phototrophica* originating from the Gobi Desert.

KEYWORDS *Gemmatimonadota*, *Gemmatimonadetes*, anoxygenic phototrophs, aquatic bacteria, MAGs, metagenome, photosynthesis gene cluster, freshwater ecology, CARD-FISH


Citation Mujakić I, Andrei A-Ş, Shabarova T, Fecskeová LK, Salcher MM, Piwosz K, Ghai R, Koblížek M. 2021. Common presence of phototrophic *Gemmatimonadota* in temperate freshwater lakes. *mSystems* 6:e01241-20. <https://doi.org/10.1128/mSystems.01241-20>.

Editor Steven J. Hallam, University of British Columbia

Copyright © 2021 Mujakić et al. This is an open-access article distributed under the terms of the [Creative Commons Attribution 4.0 International license](https://creativecommons.org/licenses/by/4.0/).

Address correspondence to Michal Koblížek, koblizek@alga.cz.

* Present address: Kasia Piwosz, Department of Fisheries Oceanography and Marine Ecology, National Marine Fisheries Research Institute, Gdynia, Poland.

 Analysis of metagenome assembled genomes documented a widespread presence of phototrophic *Gemmatimonadota* in euphotic zones of freshwater lakes in central Europe.

Received 8 January 2021

Accepted 21 February 2021

Published 16 March 2021

Microorganisms conduct key biogeochemical processes involved in the main fluxes of matter and energy on Earth. Most microbial diversity remains uncultured, and only analyses of environmental DNA samples have made it possible to unravel existing microbial diversity and to identify the main species involved (1). Indeed, out of 112 known bacterial phyla in the Genome Taxonomy Database (GTDB) (2), more than one half are still recognized only from environmental sequences (3).

One of the phyla that was first identified using molecular phylogenetic methods was *Gemmatimonadota* (also called *Gemmatimonadetes* [4]), which was originally established as the so-called BD group based on five 16S rRNA gene sequences, which originated from deep-sea sediments, soils, and reactor sludge (5, 6). Independently, the group was also proposed as a candidate division KS-B based on three 16S rRNA gene sequences retrieved from coastal sediment samples from French Guiana (7).

The first cultured strain T27 belonging to the BD/KS-B group was isolated from a wastewater treatment plant in Japan. The isolate was named *Gemmatimonas aurantiaca* and established the new phylum *Gemmatimonadota*, along with its first class *Gemmatimonadetes* and genus *Gemmatimonas* (8). Subsequently, three more *Gemmatimonadota* genera (*Gemmatirosa*, *Longimicrobium*, and *Roseisolibacter*) with type strains were described from various soil environments (9–11). Apart from class *Gemmatimonadetes*, phylum *Gemmatimonadota* consists of four more class-level groups which include class *Longimicrobia*, two terrestrial groups (BD2-11 and S0134), and one marine benthic group (PAUC43f) (10, 12). The four cultured representatives from the phyla were all chemo-organoheterotrophs that require oxygen and grow under fully aerobic or semiaerobic conditions (8–11). An interesting metabolic potential and ecological role was reported for *G. aurantiaca*, as this species has the ability to reduce the greenhouse gas N₂O (13). However, with the discovery of *Gemmatimonas phototrophica*, which contains photosynthetic reaction centers (14, 15), *Gemmatimonadota* were added to several bacterial phyla containing anoxygenic phototrophic species alongside *Proteobacteria*, *Chlorobi* (now included as a class-level lineage in *Bacteroidota* [2]), *Chloroflexota*, *Firmicutes* (*Bacillota*), *Acidobacteriota*, and the newly discovered phylum “*Candidatus* Eremiobacterota” (WPS-2) (14, 16, 17). Anoxygenic phototrophs, such as *G. phototrophica*, are able to support their metabolism by harvesting light using bacteriochlorophylls; however, they require a supply of organic substrate for growth (18). Another characteristic found in *G. phototrophica* is the organization of its photosynthesis genes into a cluster called the photosynthesis gene cluster (PGC). Interestingly, the gene arrangement in the PGC of *G. phototrophica* is very similar to the one found in *Proteobacteria*, so it has been suggested that phototrophy in *Gemmatimonadota* originates from an ancient horizontal gene transfer of the entire PGC from *Proteobacteria* (14). As yet, this is the only known case of horizontal transfer of an entire set of photosynthesis genes between distant bacterial phyla (14, 19).

Information about the prevalence of *Gemmatimonadota* in different habitats is continuously growing, although information about their ecology is scarce. Members of this phylum were found in many natural environments (12, 20–22) and represent the eighth most abundant phylum in soils, accounting for about 1 to 2% of bacteria in soils worldwide (23). Their highest contributions are typically found in fertile agricultural and forest soils (20) but are also present in more unique soil environments, such as arid Antarctic Dry Valley soils (24, 25). It has been suggested that *Gemmatimonadota* may be relatively more abundant in dry soils (26). On the other hand, from the available data, it is known that they are also present in aquatic environments, such as freshwater lakes (27), sediments (22, 28–30), and estuaries (31, 32). In addition, *G. phototrophica* was isolated from a freshwater lake in Inner Mongolia (18). However, this organism does not grow in liquid culture and requires microaerophilic conditions, which are more typical for sediment-dwelling species. Thus, the data showing that *Gemmatimonadota* prefer dry environments does not seem to be universal. There is probably a large ecological and functional diversity

TABLE 1 Basic characteristics of the lakes studied

Characteristic	Římov Reservoir	Jiřická pond	Lake Zurich	Lake Constance	Lake Thun
Country ^a	CZ	CZ	CH	CH	CH
Latitude	48.50°N	48°36'N	47°18'N	47°32'N	46°41'N
Longitude	14.29°E	14°40'E	8°34'E	9°31'E	7°43'E
Altitude (m)	470	892	406	395	558
Area (km ²)	2.10	0.035	88.66	536	48.3
Volume (m ³)	34.5 × 10 ⁶	6.59 × 10 ³	3.3 × 10 ⁹	48 × 10 ⁹	6.5 × 10 ⁹
Avg depth (m)	16.5	1.8	49	90	136
Maximum depth (m)	43	4.5	136	251	217
Mean hydraulic residence time	77 days	9 days	1.4 yrs	5 yrs	1.8 yrs
Trophic status	Mesoeutrophic	Dystrophic	Oligomesotrophic	Oligomesotrophic	Ultraoligotrophic
Mixing type	Dimictic	Polymictic	Monomictic	Monomictic	Monomictic
Sampling	2015–2017	2016–2017	2010–2019	2018	2018
No. of samples ^b	E = 10, H = 8	E = 5	E = 6, H = 4	E = 2, H = 2	E = 1, H = 1

^aCZ, Czechia; CH, Switzerland.

^bE refers to epilimnion, and H refers to hypolimnion. The exact depths for each of the lakes are provided in Materials and Methods.

among the members of *Gemmatimonadota*. The question remains whether *Gemmatimonadota*-related sequences identified in lakes originate from strictly limnic species, or perhaps they are just a passive component that enters the lakes along with runoff waters from surrounding soil.

Current progress in sequencing technologies and bioinformatics has circumvented the necessity for cultivated representatives and allowed biological and ecological inferences to be drawn by using genomic data recovered directly from microbial communities. Over the past few years, the usage of metagenome-assembled genomes (MAGs) has allowed the description of many novel bacterial divisions and unearthed large radiations in the prokaryotic tree of life (33). This approach has already led to the discovery of new phototrophic organisms belonging to the yet uncultured candidate phylum “*Ca. Eremiobacterota*” (16, 34). Also, two MAGs belonging to *Gemmatimonadota* were recovered from Lake Baikal. One MAG was more similar to *Gemmatirosa kalamazonensis* found in soils and seemed more abundant at a depth of 5 m. The other MAG-encoded rhodopsin gene and was closely related to the phototrophic species *G. phototrophica* and showed a higher abundance at 20 m (35).

Therefore, in order to address the question whether there are any truly limnic *Gemmatimonadota* and to investigate their diversity, we analyzed metagenome data from five freshwater lakes in Czechia and Switzerland. The lakes were chosen based on their trophic status and included a representative mesoeutrophic Římov Reservoir, a dystrophic Jiřická pond, oligomesotrophic Lake Zurich and Lake Constance, and an ultraoligotrophic Lake Thun (Table 1). The metagenome sequences were collected over several years and seasons. The reconstructed MAGs were analyzed with the aim to identify the most common freshwater and photoheterotrophic *Gemmatimonadota* and to analyze their spatiotemporal variability. Using catalyzed reporter deposition-fluorescence *in situ* hybridization (CARD-FISH), cells of *Gemmatimonadota* were visualized for the first time in their natural environment, and their association with other organisms was observed.

RESULTS

Abundance of *Gemmatimonadota*. The presence of *Gemmatimonadota* in freshwater lakes was first assessed by using relative 16S rRNA abundances extracted from our metagenomic data sets. In all studied lakes, *Gemmatimonadota* formed only a small part of the bacterial community with relative abundances typically below 1%. The highest relative abundance was found in Římov Reservoir, where *Gemmatimonadota* sequences were present over the entire sampling period (2015 to 2017). Interestingly, their relative abundance was higher in the hypolimnion ($0.58\% \pm 0.23\%$; $n=8$) than in the epilimnion ($0.24\% \pm 0.21\%$; $n=10$), with the highest numbers (1.03%) in August 2016. A similar pattern of higher contribution in the hypolimnion ($0.44\% \pm 0.27\%$; $n=4$) than in the

epilimnion ($0.28\% \pm 0.25\%$; $n = 6$) was also observed in Lake Zurich, with the highest relative abundance occurring in spring (13 May 2013, 0.77%). The general pattern of higher abundance in the hypolimnion was observed also in both Lake Constance and Lake Thun and was recorded in summer with 0.81% and 0.58% relative abundance, respectively. Finally, the lowest contribution of *Gemmatimonadota* was found in Jiřícká pond, where they represented less than 0.1% of the prokaryotic community, with a maximum of 0.09% recorded in summer, in August 2017.

Statistical analysis using distance-based linear models (DistLM) showed that the only environmental factor driving the *Gemmatimonadota* community in Římov Reservoir was water temperature, which explained 47% of the variability in the data set ($P = 0.0002$, pseudo- $F = 13.303$). A separate analysis of epilimnion and hypolimnion samples showed that temperature was also important in epilimnion ($P = 0.0079$, pseudo- $F = 5.7106$, and 41.6% of the explained variability in data set) but not in hypolimnion.

Metagenome-assembled genomes of *Gemmatimonadetes* and their distribution.

To explore the diversity of aquatic *Gemmatimonadota*, we performed genome-resolved metagenomics analyses. The freshwater MAGs ($n = 45$) obtained in our study were analyzed together with those publicly available ($n = 226$) in March 2019 (see Table S1 in the supplemental material for a complete list). Most of the MAGs assembled by us were obtained from Římov Reservoir ($n = 27$), followed by Lake Zurich ($n = 10$), Lake Constance ($n = 5$), Lake Thun ($n = 2$), and Jiřícká pond ($n = 1$). Publicly available MAGs originated from various environments: mostly from marine habitats ($n = 36$) and soils ($n = 37$), followed by sediments of soda lakes ($n = 38$), permafrost ($n = 23$), and other ($n = 11$). Only three MAGs from freshwater environments were available in the database at the time of the study.

Phylogenomically, based on Genome Taxonomy Database (GTDB) taxonomy (39), all MAGs recovered in this study clustered within class *Gemmatimonadetes* (Fig. 1). Although they were all recovered from a freshwater environment, six of them were more closely related to publicly available MAGs from environments like soils and sediments. MAGs CH-Jul18-bin44, ZH-3nov15-207, and TH-Jun18-bin75 clustered with MAGs from soils, and CH-Jul18-bin76, CH-Jul18-bin112, and ZH-3nov15-212 clustered with MAGs from sediments. Similarly, in 16S rRNA phylogeny (see Fig. S1 in the supplemental material), all freshwater MAGs with recovered 16S rRNA genes clustered within class *Gemmatimonadetes*, with the exception of two previously mentioned MAGs (CH-Jul18-bin112 and ZH-3nov15-212) that clustered within the BD2-11 terrestrial group based on SILVA SSU v138 database taxonomy (12).

16S rRNA gene diversity. The 16S rRNA gene fragments from metagenomes were classified taxonomically. As most of the *Gemmatimonadota* phylum consists of environmental sequences that remain uncultured, many of the fragments could not be classified. Therefore, in order to increase taxonomic resolution of unclassified *Gemmatimonadota*, we have used 16S rRNA gene sequences from our MAGs (Fig. S1) to define six new clusters (five from class *Gemmatimonadetes* and one from BD2-11 terrestrial group). In Římov Reservoir, the *Gemmatimonadota* community was solely composed of the class *Gemmatimonadetes* (Fig. 2). The genus *Gemmatimonas* was present in all seasons and both depths, but it was more abundant in the epilimnion, especially in summer (15-Aug-16). Epilimnion was dominated by the genus *Gemmatimonas* and unclassified *Gemmatimonadaceae*, with the exception of three samples (two in November and one in April) where other clusters were also present. In contrast, hypolimnion samples contained several different clusters, including phototrophic cluster PG1 that was one of the dominant clusters in every season. Similarly, clusters GRI1 and GSR7 were predominantly present in the hypolimnion. The presence of these three clusters in the epilimnion occurred only in November and April, when the reservoir was mixed. In Jiřícká pond, the *Gemmatimonadota* community was composed of three clusters from the class *Gemmatimonadetes*, i.e., genus *Gemmatimonas*, cluster GSR7, and cluster PG1, the latter being mostly dominant in summer. Additionally, the BD2-11 terrestrial group was also present in low

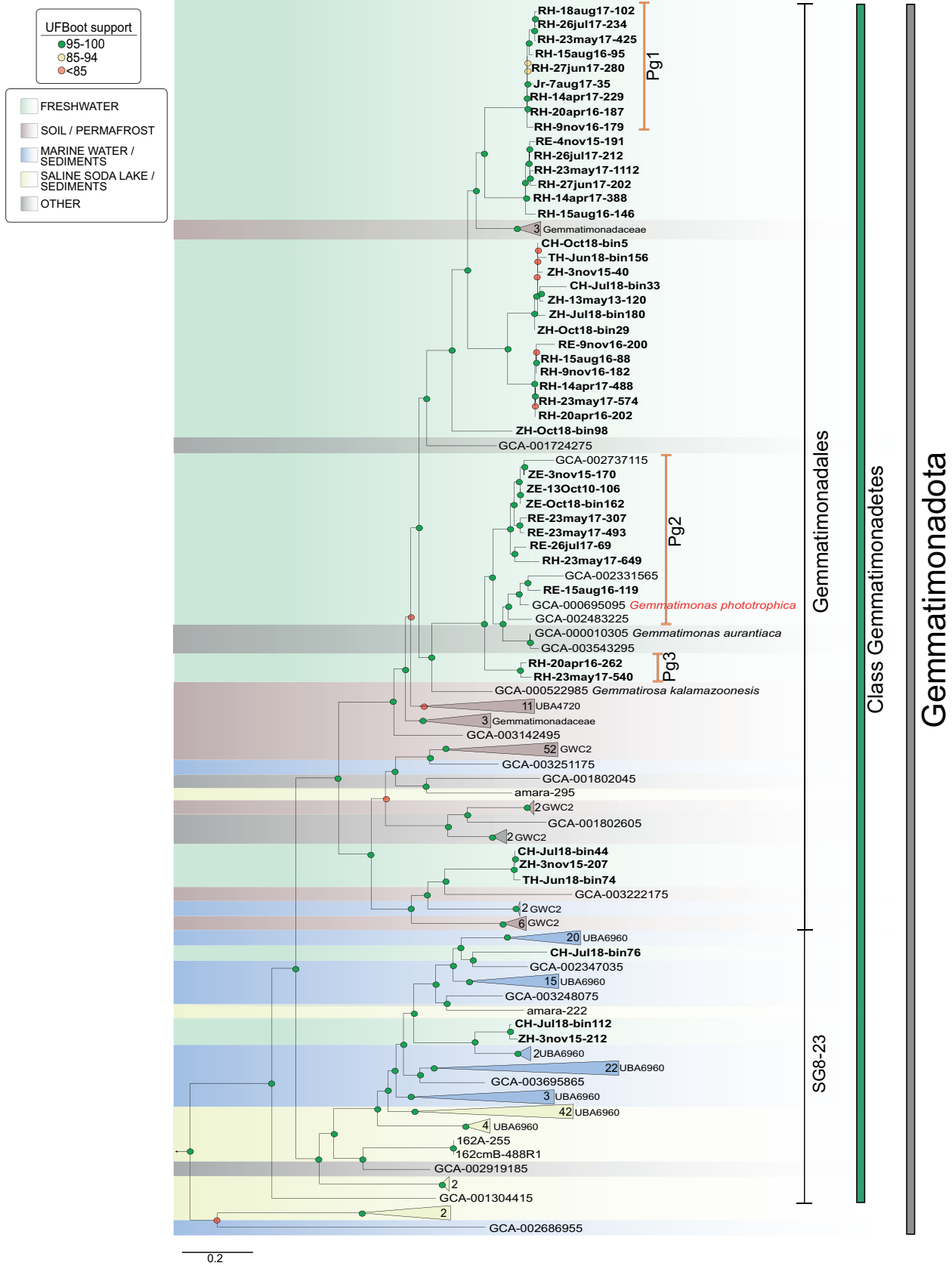


FIG 1 Maximum-likelihood tree (1,000 ultrafast bootstrap replicates, 121 conserved concatenated marker proteins) showing 45 MAGs of freshwater *Gemmatimonadota* as well as 226 MAGs and genomes of cultures collected from GTDB originating from different environments. The (Continued on next page)

proportions in summer. The highest taxonomic diversity within *Gemmatimonadota* was observed in Swiss lakes, where three other class-level groups in addition to *Gemmatimonadetes* were recovered, i.e., representatives of classes *Longimicrobia*, BD2-11 and S0134, groups that were previously known only from terrestrial 16S rRNA gene sequences (12). Here, genus *Gemmatimonas* was also predominantly present in the epilimnion, whereas cluster GSR7 dominated the hypolimnion. In Lake Zurich, a large contribution of class *Longimicrobia* was recorded in a spring sample (13-May-13) in both the epilimnion and hypolimnion.

Photosynthesis genes in *Gemmatimonadota*. Further analyses of the assembled genomes of *Gemmatimonadota* revealed that 19 out of 45 freshwater MAGs contained phototrophic genes. In the phylogenetic tree, we could differentiate them into three distinct phototrophic groups (Fig. 1). Phototrophic group 1 contained eight MAGs from the hypolimnion of Římov Reservoir and one from Jiřická pond. Phototrophic group 2 was formed of nine MAGs (three from the epilimnion of Lake Zurich, five from the epilimnion of Římov Reservoir, and one from the hypolimnion of Římov Reservoir). Phototrophic group 2 also included the only cultured photoheterotrophic strain of the phylum, *G. phototrophica*, as well as the only three MAGs from freshwater environments that were publicly available. Phototrophic group 3 contained two MAGs from the hypolimnion of Římov Reservoir. All these MAGs were recovered from different time points (several years and different seasons).

Using representatives from each of these groups, we reconstructed the photosynthetic gene cluster and compared its structure and organization with the PGC found in *G. phototrophica* (14). For phototrophic groups 1 and 2, we reconstructed the entire PGC. In the case of phototrophic group 3, the PGC was missing several genes due to the incompleteness of the genomes (MAGs had completeness of 71.07. and 83.16%). From our findings, it appears that limnic *Gemmatimonadetes* have the same or very similar organization of phototrophic genes as *G. phototrophica* AP64 (Fig. 3 and Table S4). A distinct trait of *G. phototrophica* is a fragmented PGC divided by a set of hypothetical genes. The same split PGC was observed in three out of nine MAGs from phototrophic group 2, but it was absent in phototrophic groups 1 and 3. Furthermore, we observed the presence of the gene *frhB* (coenzyme F420-reducing hydrogenase, beta subunit) in the operon *bchP2G*, in the PGC of all members of phototrophic group 1, at the position where all other phototrophic *Gemmatimonadetes* have *bch2*. A full list of phototrophic genes for each MAG, as well as their distribution on contigs, is provided as Table S4.

As the presence of ribulose 1,5-bisphosphate carboxylase enzyme (RuBisCO) genes was previously reported in some sediment *Gemmatimonadota* MAGs, we also searched for them in our freshwater MAGs. We identified type IV RubisCO-like genes in eight phototrophic freshwater MAGs from Římov Reservoir (Fig. S3). Seven of these MAGs belonged to phototrophic group 1 recovered from hypolimnion, and one to phototrophic group 2 recovered from the epilimnion of Římov Reservoir.

Average amino acid identities and 16S rRNA and photosynthetic gene similarities of photoheterotrophic *Gemmatimonadetes*. We calculated average amino acid identities (AAI) between all the freshwater MAGs and the three genomes of cultured reference strains (*Gemmatimonas phototrophica*, *Gemmatimonas aurantiaca*, and *Gemmatirosa kalamazoonensis*) (Fig. S4). All MAGs of the phototrophic group 1 represent the same species, with AAI values of 99 and 100%, that was recovered in both Římov Reservoir and Jiřická pond, multiple times through different years and seasons. When phototrophic

FIG 1 Legend (Continued)

different environments are shown by different colors (left legend). The numbers shown at collapsed branches (i.e., 11 and 52) indicate the numbers of genomes (not shown) comprising the respective taxonomic categories. Genomes belonging to *Fibrobacterota* were used as an outgroup to root the tree. The annotations Pg1 to Pg3 define groups of photoheterotrophic *Gemmatimonadota*. Order, class-level, and phylum taxonomic labels are indicated through vertical delimiters (right part of the figure). The strength of support for internal nodes (assessed by ultrafast bootstrapping) is shown through colored circles (left legend). Details on all genomes can be found in Table S1 in the supplemental material.

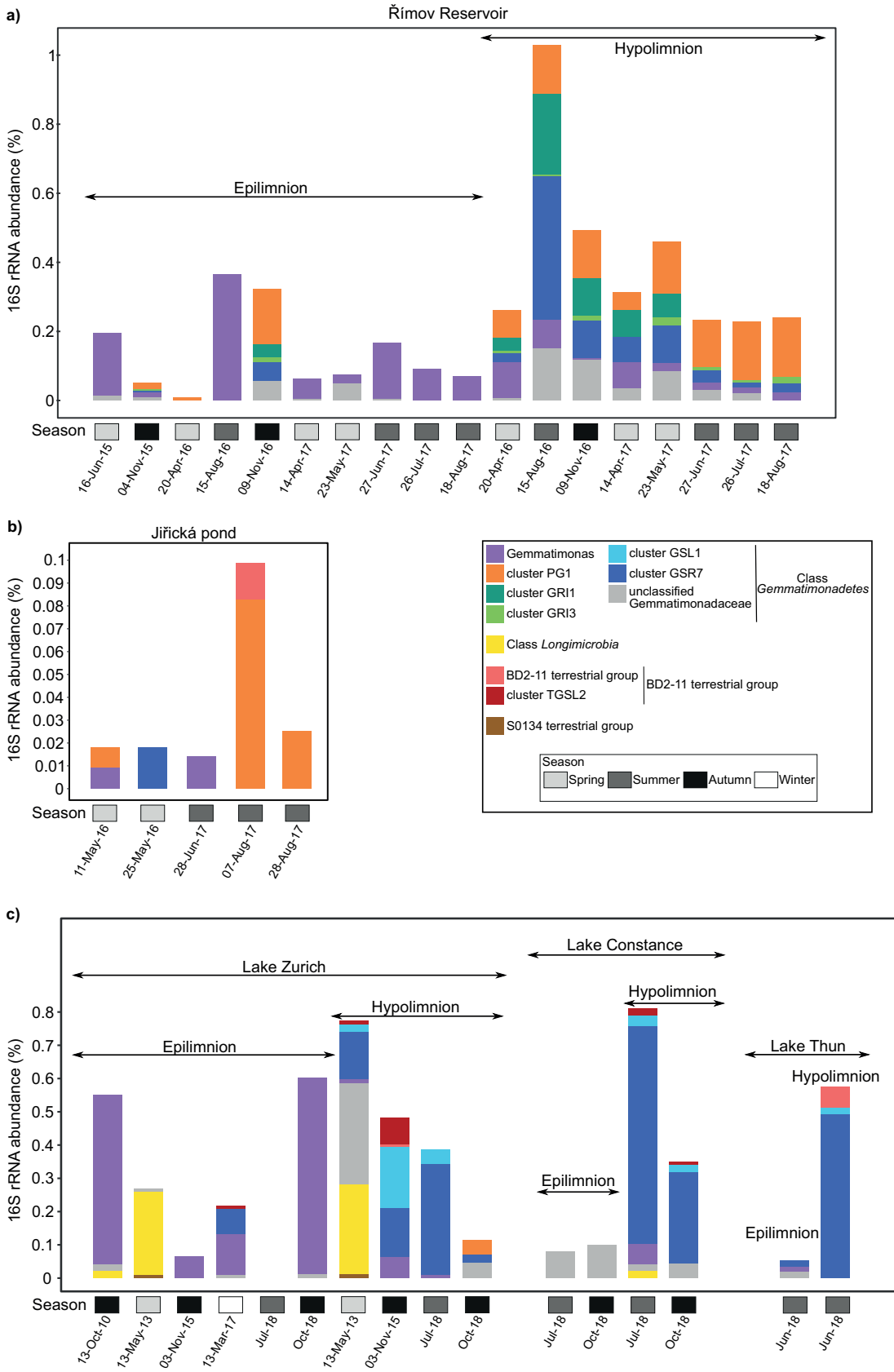


FIG 2 Relative abundance of *Gemmatimonadota* based on 16S rRNA gene fragments from metagenomes of five contrasting freshwater environments. These five freshwater environments are as follows: Římov Reservoir (a), Jiřická pond (b), and Lake Zurich (c), Lake Constance (c), and Lake Thun (c). (Continued on next page)

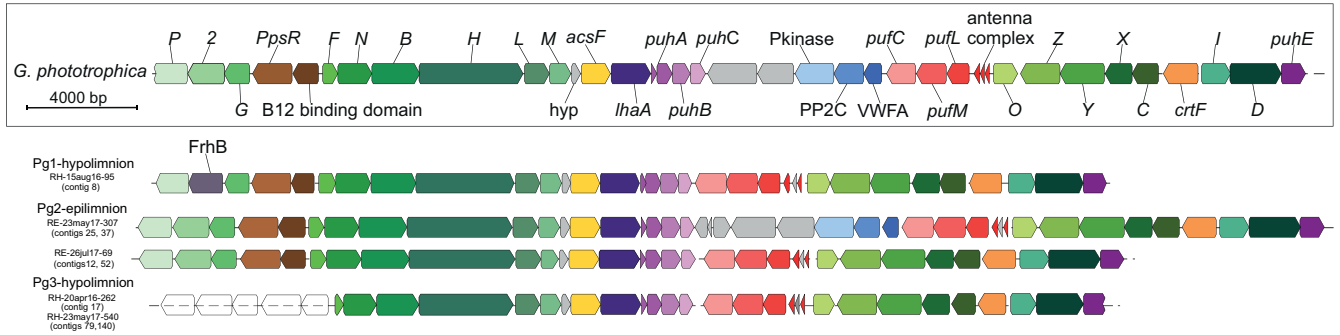


FIG 3 Structure and composition of photosynthetic gene clusters of *Gemmatimonas phototrophica* (top) and phototrophic MAGs of *Gemmatimonadota* that clustered into three photoheterotrophic groups. The names of the MAGs used to depict each photosynthetic group (photosynthetic group 1 [Pg1] to Pg3) are shown at the right part of the figure. *bch* genes involved in bacteriochlorophyll biosynthesis (green), *puf* operon encoding the reaction center (red), *puh* operon (pink/purple), carotenoid biosynthesis genes (orange), *ppsR* gene and B12 binding domain (brown), *acsF* (yellow), *lhaA* (dark purple), genes not involved in photosynthesis (blue), and hypothetical genes (gray) are indicated. A list of all genes is provided in Table S4.

group 1 was compared with phototrophic groups 2 and 3, AAI values were below 65%, suggesting that these groups represent different genera. Similarly, phototrophic group 1 showed AAI values below 65% (specifically 55 and 56%) compared with cultured representatives. In phototrophic group 2, the three MAGs from Lake Zurich represent the same species (AAI, 95 to 100%) recovered in different years, whereas other MAGs from Řimov Reservoir in this group seem to belong to the same genus but are all different species (AAI, 77 to 82%). Phototrophic group 3 consists of two MAGs that represent the same species (AAI, 96 and 100%). Both phototrophic groups 2 and 3 (AAI, 66 to 68%) belong to the same genus as the cultured representatives *G. phototrophica* and *G. aurantiaca* (AAI 66%) but are different from *Gemmatirosa kalamazonensis* (AAI, 56%).

From the phototrophic MAGs (phototrophic group 1), one MAG from the hypolimnion of Řimov Reservoir (RH-18aug17-102) contained a 16S rRNA gene, which showed 92.8% identity with *G. phototrophica* and 91.5% and 89.6% identity with *G. aurantiaca* and *Gemmatirosa kalamazonensis*, respectively (Fig. S1), and one MAG from Jiřická pond (Jr-7aug17-35) showed 92.1% 16S rRNA gene sequence identity with *G. phototrophica*.

To further analyze the differences among the identified phototrophic groups, we compared the *pufM* and *acsF* gene sequence identities between them and type species *G. phototrophica*. Both of these genes are markers for anoxygenic photoheterotrophic bacteria (22, 40). For the *pufM* gene (encoding the M subunit of the bacterial photosynthetic reaction center), phototrophic group 2 showed only 80 to 85% identity with *G. phototrophica*. Much lower identities were found for hypolimnion groups 3 and 1 (77 to 80% and 72%, respectively). The *acsF* gene (aerobic oxidative cyclase gene) gave similar results with only 71% identity between phototrophic group 1 and *G. phototrophica*. MAGs from phototrophic groups 2 and 3 had 79 to 89% and 78% identity of the *acsF* gene of *G. phototrophica*, respectively.

Relative abundance of MAGs of phototrophic and nonphototrophic *Gemmatimonadota*. In order to compare the relative abundance of phototrophic and nonphototrophic *Gemmatimonadota*, we used 28 metagenomes from Řimov Reservoir from which 27 MAGs were recovered for fragment recruitment (Fig. 4). Nonphotosynthetic MAGs showed the highest contribution in the hypolimnion, especially in summer (15 August 2016); however, their abundance was not constant and changed over time. Similarly, photosynthetic group 1 was also present mostly in the hypolimnion, and its

FIG 2 Legend (Continued)

Zurich, Lake Constance, and Lake Thun (c). The figure depicts the classification of 16S rRNA gene fragments (as unassembled shotgun reads) retrieved from five freshwater data sets. The x axis shows the sampling dates, while the y axis indicates the percentage of *Gemmatimonadota* within the prokaryotic communities. The sample collection time, following a four-season breakdown, is indicated by the gray-scale-colored boxes arranged along the x axis.

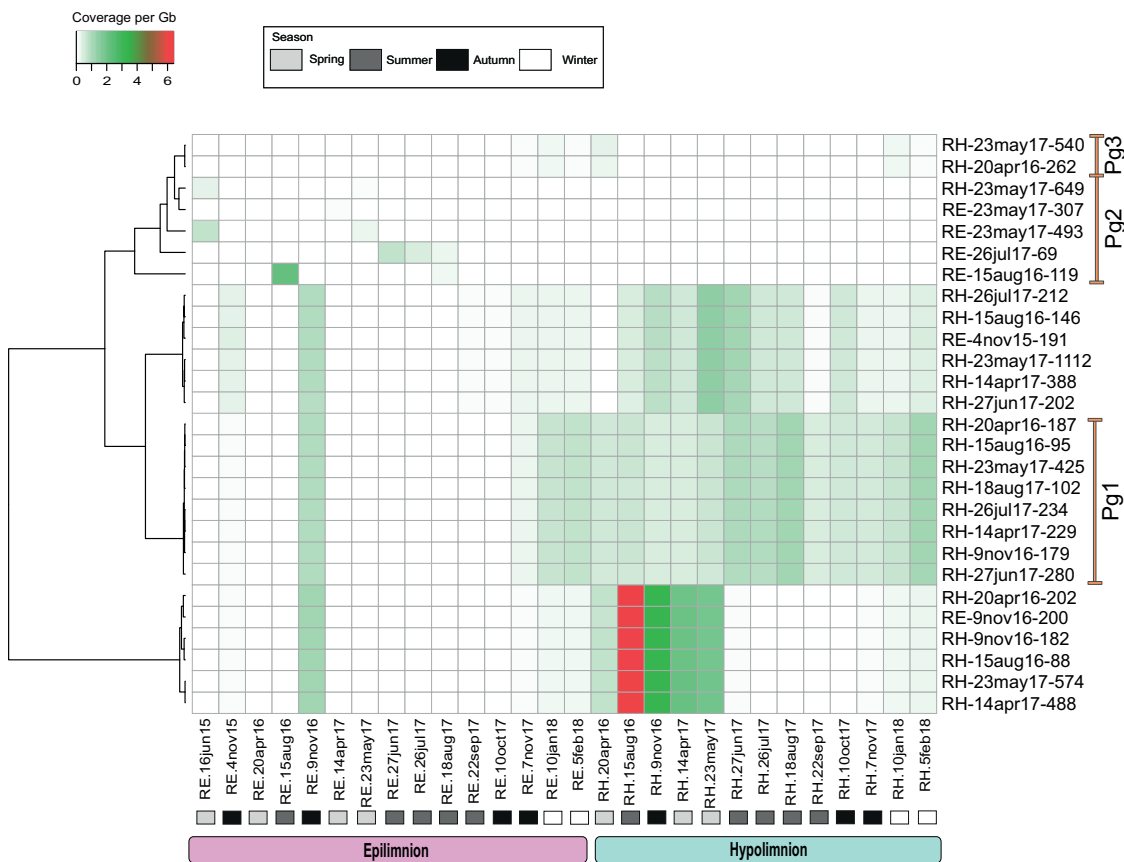


FIG 4 Heatmap of the relative abundance of 27 *Gemmatimonadota* MAGs gained from Řimov Reservoir in 28 metagenomes from Řimov Reservoir (Table S3). MAGs are clustered using complete linkage and Euclidian correlation. Columns are labeled with the sampling date, the depth of the sample (RE = epilimnion [0.5 m] and RH = hypolimnion [30 m]) and season. Three groups of MAGs (Pg1 to Pg3) containing photosynthetic gene clusters representing photoheterotrophic *Gemmatimonadota* are labeled with orange lines. Color key of abundance (coverage per gigabase) is shown in the top left corner, with red indicating the highest abundance and white indicating that a MAG is not present in a metagenome.

relative abundance varied during the season. The only occurrence of this group in the epilimnion was in late autumn (9 November 2016), at times of deep mixing of the reservoir (41). In contrast, photosynthetic groups 2 and 3 were present mostly in the epilimnion but remained at very low relative abundances in all seasons and at all depths.

CARD-FISH analyses. In order to visualize members of the class *Gemmatimonadetes* and access more information on their distribution and potential associations within freshwater environments, we designed the CARD-FISH probe Gemma_801 and applied it to a set of samples obtained from Řimov Reservoir, comprising eight longitudinal transects collected during summer 2015. The hybridized cells were also counted; however, since the probe *in silico* matches only 35% of all *Gemmatimonadetes*, the presented numbers, while corresponding with 16S rRNA gene abundance from metagenomes, reflect only the detected, and not total *Gemmatimonadetes*, implying that the observed numbers can underestimate the absolute abundance of *Gemmatimonadetes* in the environments studied. The lowest relative (mean = 0.09%, minimum [min] = 0.00, maximum [max] = 0.19%) and absolute (mean = 3.25×10^3 cells ml⁻¹, min = 0.00, max = 7.9×10^3 cells ml⁻¹) abundances were observed at the river station of the reservoir (Fig. S2). However, at stations 2 and 3 (stations following the “river station” in the longitudinal transect), the contribution of hybridized planktonic cells had already reached up to 1.0% (4.3×10^4 cells ml⁻¹) during a bloom of *Fragilaria* sp. at the end of June. Besides free-living *Gemmatimonadetes*, we observed hybridized cells that seemed to be attached to diatoms (Fig. 5a). On some *Fragilaria* sp. colonies, hybridized *Gemmatimonadetes* contributed up to 22% of all bacterial cells detected on the diatom

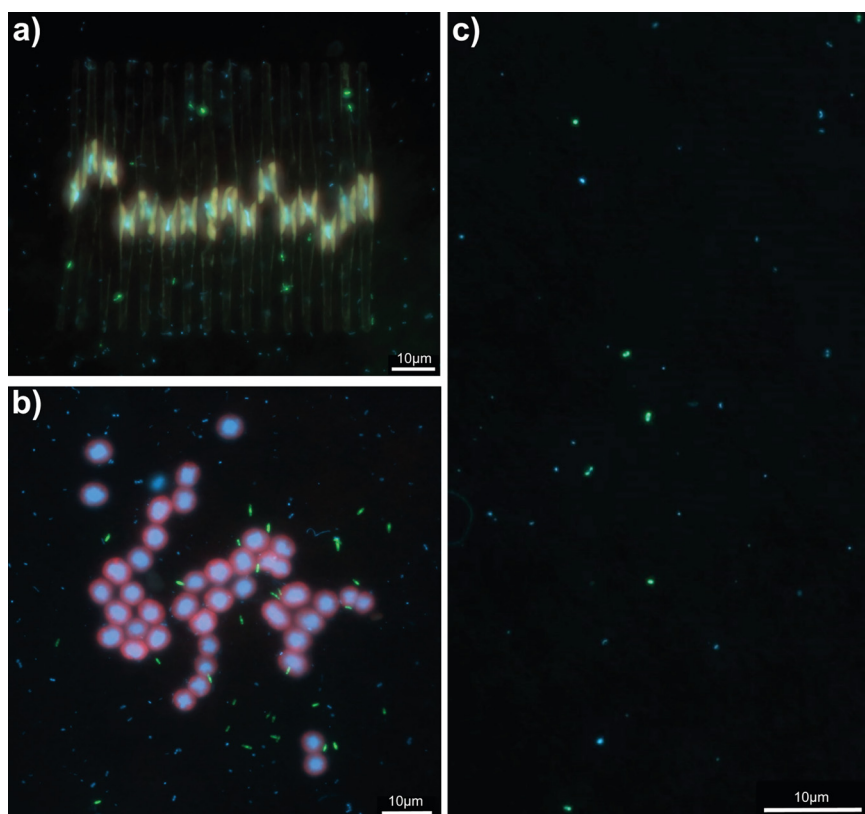


FIG 5 Overlay images of *Gemmatimonadetes* cells in Āřimov Reservoir visualized by CARD-FISH with probe Gemma_801. (a) *Gemmatimonadetes* associated with a *Microcystis* sp. colony. (b) *Gemmatimonadetes* observed on a colony of *Fragilaria* sp. (c) Free-living *Gemmatimonadetes* observed in the hypolimnion of Āřimov Reservoir. The probe signal is displayed in green, DAPI staining is displayed in blue, and autofluorescence of *Cyanobacteria* is displayed in red.

surface. In addition, *Gemmatimonadetes* seemed to be associated with cyanobacteria, as higher densities were detected with agglomerations of *Microcystis* sp. colonies (Fig. 5b and Fig. S5a). However, the highest numbers of hybridized cells were observed in the hypolimnion (≥ 10 m) of Āřimov Reservoir (up to 4.43% and 2.75×10^4 cells ml^{-1} , respectively; Fig. 5c, Fig. 6, and Fig. S5b). These cells displayed different shapes and were smaller than those detected in the epilimnion.

DISCUSSION

In the presented work, we clearly demonstrated that *Gemmatimonadota* were permanently present in all of the studied lakes showing the ubiquitous nature of this phylum in contrasting freshwater lakes ranging from ultraoligotrophic to mesoeutrophic. The freshwater *Gemmatimonadota* community followed seasonal patterns with water temperature as the main driving variable, especially in the epilimnion. This observation further supports the idea that the studied *Gemmatimonadota* are truly limnic, since abundances of passively transferred microorganisms correlate with water mass movements rather than with temperature (42). The stratification of the lake also seemed to have a significant effect on the *Gemmatimonadota* community, as clusters inhabiting the epilimnion differed from those present in the hypolimnion, suggesting that they can be metabolically diverse and be specialized in different ecological roles. While *Gemmatimonadota* were relatively stable in the hypolimnion, their numbers in the epilimnion varied with higher numbers in late summer and autumn and lower numbers in spring and early summer. The stratification effect was previously studied in Grand Lake (OK, USA) where the *Gemmatimonadota* community was present in late summer in the free-living fraction within the thermocline and hypolimnion. The highest abundance in

often increase carbon assimilation in the presence of light, which allows them to have larger cells (46, 47). Subsequently, the distribution of *Gemmatimonadota* in the epilimnion could be dependent upon the distribution and occurrence of diatoms like *Fragilaria* sp. and cyanobacteria like *Microcystis* sp., as their blooms can influence and alter bacterioplankton communities (45, 48). Likewise, this dependence is a plausible explanation for the reason why photoheterotrophic members of this phylum are proving to be difficult to culture.

The presented 45 metagenome-assembled genomes of *Gemmatimonadota* in this study represent the largest collection of freshwater metagenome-assembled genomes (MAGs) of this phylum so far available. The obtained MAGs further confirmed the limnic nature of *Gemmatimonadota*. With recovery of several MAGs that represent the same species based on AAI (with 99 or 100% similarity), we show that *Gemmatimonadota* MAGs are reassembled from metagenomes and recovered periodically from different years as well as different seasons. This confirms the strong resilience of their microdiversity in freshwater and cannot be taken as a random occurrence. Furthermore, 39 of the obtained MAGs are different from soil species (Fig. 1; also see Fig. S1 and S4 in the supplemental material) and form distinct phylogenetic groups (with AAI between 45 and 65%) with both photoheterotrophic and heterotrophic representatives (38). Six MAGs, gained from the hypolimnion of deep lakes, three from Lake Constance (CH-Jul18-bin44, CH-Jul18-bin76, and CH-Jul18-bin112), two from Lake Zurich (ZH-3nov15-207 and ZH-3nov15-212), and one from Lake Thun (TH-Jun18-bin75), (Fig. 1 and Fig. S1) are more closely related to genomes from soil and sediments. In the 16S rRNA phylogeny (Fig. S1), two previously mentioned MAGs (CH-Jul18-bin112 and ZH-3nov15-212) clustered within the BD2-11 terrestrial group based on SILVA SSU v138 database taxonomy (12). This is consistent with 16S rRNA gene abundance in Swiss lakes where we could episodically observe the occurrence of 16S rRNA sequences related with terrestrial groups, such as *Longimicrobia*, BD2-11, and S0134 that has, as yet, been connected only with soil environments (10). Environmental 16S rRNA gene sequences from freshwater have until now been associated only with the class *Gemmatimonadetes* (10–12), and all other groups were formed with soil and sediment representatives. While the six MAGs could represent new freshwater members of the soil-connected groups, due to the close phylogenomic similarity with soil representatives and not with other freshwater MAGs, it is more probable that they represent genomes recovered due to soil runoff. Nevertheless, the MAGs present in these different phylogenetic groups were assembled from metagenomes gained from different sampling times which shows *Gemmatimonadota* were recovered repeatedly from freshwater environments. All this evidence shows that at least 39 of the identified MAGs represent truly limnic and planktonic species that do not come from soils. Furthermore, with the notable genomic diversity of limnic *Gemmatimonadota*, we demonstrate the ecological relevance of this group, as different members are clearly able to persist in the water column, occupying different ecological niches as they occur both in the hypolimnion and in the epilimnion. Moreover, the distribution of *Gemmatimonadota* in contrasting freshwater lakes showed they are adapted to different types of aquatic environments. The highest diversity in terms of different genus level groups or species of *Gemmatimonadota*, especially photoheterotrophic representatives, was observed in Římov Reservoir. Despite this, we cannot associate the higher diversity to mesoeutrophic lakes, as the data set obtained from Římov Reservoir was substantially larger, allowing for an increased chance of recovering higher diversity. A higher trophic status of any lake is connected with higher phytoplankton productivity; therefore, lakes and reservoirs often show variations of microbial communities connected with phytoplankton productivity (49, 50). Specifically, the connection of bacterial communities with changes in phytoplankton has already been recorded in Římov Reservoir (51). Furthermore, anoxygenic phototrophic bacteria are often found in close association with algae (45, 52, 53), and since they seem to follow seasonal phytoplankton blooms in freshwater lakes (45, 48), it is suggested that they represent an important functional group in freshwater environments (54).

Phototrophic gene, phylogenomic, and AAI analyses have shown that phototrophy in the *Gemmatimonadota* spans multiple genera. The identified phototrophic *Gemmatimonadota* represent three different groups but share the same set of phototrophic genes with *G. phototrophica*. All identified PGCs share a very similar gene inventory (see Table S4 in the supplemental material) and organization (Fig. 3). It seems that the convergent orientation ($\rightarrow \leftarrow$) of superoperons *bchFNBHLM* and *crtFbchXYZ-puf* is conserved among all the phototrophic *Gemmatimonadota*, in contrast to phototrophic *Proteobacteria*, where orientation of these superoperons may be divergent, convergent, or colinear (55). Moreover, the split of the PGC with inserted genes in the type strain *G. phototrophica* AP64 (14) is also present in several MAGs from the epilimnion. Previously, the PGC of *G. phototrophica* was compared with two PGCs reconstructed from the Odense wastewater metagenome (OdenseWW) and the Aalborg wastewater metagenome 2 (AalborgWW-2) which did not contain the insert of several hypothetical genes between two operons. The explanation proposed for the difference in having a split PGC or not was that it could represent different evolutionary history of phototrophic *Gemmatimonadota* originating from different environments (21). Apart from the split PGC present in some of the MAGs, we observed that members of phototrophic group 1 have coenzyme F420-reducing hydrogenase (*frhB* gene) instead of bacteriochlorophyll synthase 4.5-kDa chain (*bch2*), as part of the *bchP2G* operon. Coenzyme F420-reducing hydrogenase enzyme seems to have an important role in energy conservation and methanogenesis from CO₂ (56).

In support of the true photoheterotrophic nature of *Gemmatimonadota*, a previous study conducted in Římov Reservoir, found active expression of their *pufM* genes (a common molecular marker gene for aerobic anoxygenic phototrophs) (27). Expression of the photosynthetic apparatus of *Gemmatimonadota* showed that they are an active part of bacterial community and do not just passively contain the phototrophic genes (27).

Recently, several *Gemmatimonadota* MAGs that originate from sediments of a soda lake were reported to contain both phototrophic genes and genes related to the large subunit of ribulose 1,5-bisphosphate carboxylase/oxygenase enzyme (57), suggesting that these soda lake MAGs represent the first photoautotrophic *Gemmatimonadota* (58). In contrast, some of our freshwater *Gemmatimonadota* contain phototrophic genes as well as type IV RuBisCO (Fig. S3), which is considered only a homologue of RuBisCO, since it does not have any carboxylation activity (59, 60). Type IV RuBisCO genes are present in many microorganisms, including both phototrophic and heterotrophic bacteria and *Archaea* and are thought to participate in some other metabolic pathways different from the Calvin-Benson cycle (61, 62). Therefore, these freshwater *Gemmatimonadota* MAGs are not photoautotrophs, rather have a photoheterotrophic metabolism, typical for aerobic anoxygenic photoheterotrophic bacteria. These bacteria do not fix inorganic carbon and need to rely on organic carbon produced by other organisms, so the ability to harvest light is used to supply energy for their mostly organoheterotrophic metabolism (40, 63, 64).

In conclusion, with MAGs from these contrasting freshwater lakes, we not only reveal the existence of several new phototrophic species that differ phylogenetically from the already cultured and characterized *G. phototrophica* but also show the considerable diversity of both photoheterotrophic and heterotrophic *Gemmatimonadota* in freshwater.

MATERIALS AND METHODS

Sampling for metagenomics analysis. Samples from five European freshwater lakes representing a large range of size, depth, and/or trophic status (Table 1) were used to obtain genomic information from *Gemmatimonadota*. Samples were taken from the surface mixed-sun exposed layer known as the epilimnion and the deeper, colder layer where typically there is reduced turbulence and a smaller amount of light is present. This deeper layer is called the hypolimnion. Římov Reservoir is a mesoeutrophic, canyon-shaped dimictic water body that was built during 1970s by damming a 13.5-km-long section of the River Malše (65). The sampling was performed between June 2015 and August 2017, above the deepest point of the reservoir by using a Friedinger sampler. A volume of 20 liters of water was collected from

both the epilimnion (0.5 m; $n = 10$) and hypolimnion (30 m; $n = 8$) and subjected to sequential peristaltic filtration through a series of 20-, 5-, and 0.2- μm -pore-size polycarbonate membrane filters (142 mm diameter) (Sterilitech Corporation, USA). Characteristics of the water column, depth, temperature, oxygen (GRYF XBQ4; Havlíčkův Brod, Czechia) and chlorophyll *a* (FluoroProbe TS-16-12; bbe Moldaenke, Kiel, Germany) were also measured. The sample collection and filtration steps were identical for the rest of the lakes unless otherwise stated. Jířická pond is a dystrophic humic water body from which five epilimnion (0.5-m depth) water samples were collected between May 2016 and August 2017. Lake Zurich is an oligomesotrophic, perialpine monomictic water body, from which nine water samples were collected in a period between 2010 and 2018 from the epilimnion (5-m depth; $n = 5$) and hypolimnion (80/120-m depth; $n = 4$) of the lake and processed as described above. Lake Thun is an ultraoligotrophic, alpine water body. Two water samples were collected in June 2018 from 5-m and 180-m depth. Lake Constance is a large oligomesotrophic, perialpine lake. Four samples, which were collected in July and October 2018 from the epilimnion (5 m) and hypolimnion (200 m) were used for this study.

DNA extraction, sequencing, and assembly. DNA was extracted from the 0.22- μm filters (0.2- to 5- μm fraction) using the ZR Soil Microbe DNA MiniPrep kit (Zymo Research, Irvine, CA, USA) in accordance with the manufacturer's instructions. The total quantity of DNA was estimated using the Qubit dsDNA BR assay kit (Life Technologies, Foster City, CA, USA) on a Qubit 2.0 fluorometer (Life Technologies). DNA integrity was assessed by agarose gel (2%) electrophoresis and SYBR green I staining. Shotgun sequencing was performed using the Novaseq 6000 sequencing platform (2×150 bp) (Novogene, Hong Kong, China). Raw Illumina metagenomic reads were quality filtered in order to remove low-quality bases/reads and adapter sequences using bbmap package (66, 67). Briefly, the paired-end (PE) reads were interleaved by reformat.sh and quality trimmed by bbduk.sh ($\text{qtri} = \text{rltrimq} = 18$) (68). Subsequently, bbduk.sh (68) was used for adapter trimming and identification/removal of possible PhiX and p-Fosil2 contamination ($k = 21$ ref = vector file ordered cardinality). Additional checks (i.e., *de novo* adapter identification with bbmerge.sh [67]) were performed in order to ensure that the data sets met the quality threshold necessary for assembly. The obtained quality-filtered data sets were then assembled independently with MEGAHIT (v1.1.5) (69) using the *k*-mer sizes: 39, 49, 69, 89, 109, 129, 149, and default parameters.

16S rRNA abundance-based taxonomic classification. The obtained quality-filtered data sets were converted to FASTA format and randomly subsampled to 20 million reads by using reformat.sh (68). These subsets (containing 20 million sequences each) were queried against the SILVA SSU database, release 132 (70), in order to identify RNA-like sequences by using MMSeqs2 (71) and an E-value cutoff of $1e-3$. The bona fide 16S rRNA gene sequences (as identified by SSU-ALIGN [72]) were further compared by blastn, in nucleotide space (using as cutoff the E value $1e-5$), against the SILVA SSU database amended with 16S rRNA genes recovered from *Gemmatimonadota* MAGs (see below), and classified if the sequence identity was $\geq 80\%$ and the alignment length was ≥ 90 bp (sequences failing these thresholds were not used for downstream analyses). The taxonomic affiliation of each identified 16S rRNA read was inferred based on its best blastn hit. The relative abundances of *Gemmatimonadota* taxonomic categories were calculated as a percentage of total 16S rRNA reads.

The statistical relationships between environmental data (oxygen, temperature, and chlorophyll *a*) (41) and *Gemmatimonadota* abundance in Řimov Reservoir was analyzed by distance-based linear models and nonmetric multidimensional scaling (nMDS) in the PERMANOVA+ add-on package of the PRIMER7 software (Primer Ltd., Luton, UK). Abundance data were square root transformed, and analysis was done using a stepwise selection procedure. The best model was selected based on statistical significance (9,999 permutations), and the value of the Akaike's information criterion (AICc) (73, 74). The same analysis was also done separately for the epilimnion and the hypolimnion.

Recovery of bacterial genomes. Quality-filtered metagenomics data sets were mapped using bbwrap.sh (75) ($k\text{filter} = 31$ $\text{subfilter} = 15$ $\text{maxindel} = 80$) against the assembled contigs (longer than 3 kb) in a lake-dependent fashion. The resulting BAM files were used to generate contig abundance files with jgi_summarize_bam_contig_depths ($-\text{percentidentity} 97$) (76). The contigs and their abundance files were used for binning with MetaBAT2 (default settings) (76). Bin completeness, contamination, and strain heterogeneity were estimated using CheckM (with default parameters) (77). Bins with estimated completeness above 40% and contamination below 5% were denominated as metagenome-assembled genomes (MAGs). MAGs were taxonomically classified with GTDB-Tk (2) using default settings.

Phylogenomics. MAGs belonging to *Gemmatimonadota* ($n = 45$), together with reference genomes ($n = 226$) recovered from public repositories (see Table S1 in the supplemental material) were annotated using the TIGRFAMs database (78). A total of 121 conserved marker proteins (Table S2) were extracted from the annotated *Gemmatimonadota* genomes. MAGs that had more than 40% markers present, together with reference genomes were used for phylogenetic reconstruction. Briefly, homologous proteins were independently aligned with PRANK (default settings) (79), trimmed with BMGE ($-\text{t AA} -\text{g} 0.5 -\text{b} 3 -\text{m BLOSUM30}$) (80), and concatenated. A maximum-likelihood phylogeny was constructed using IQ-TREE (81) with the LG+*F*+*R10* substitution model (chosen as the best-fitting model by ModelFinder [82]) and 1000 ultrafast bootstrap replicates. Genomes appertaining to *Fibrobacterota* were used as an outgroup to root the tree.

The average amino acid identity (AAI) within coherent phylogenomic groups was determined by performing whole-genome pairwise coding DNA sequence (CDS) comparison, using BLAST, as previously described (83). Taxonomic categories for the MAGs were defined using the suggested standards (38).

The photosynthetic gene clusters within obtained phylogenomic groups were analyzed in Geneious Prime 2019.2.3. For each group, MAGs with the most complete photosynthetic gene cluster were chosen

as representatives. The alignment and identity matrix for nucleotide *pufM* and *acsF* gene sequences from photoheterotrophic MAGs and *G. phototrophica* was done with ClustalW 2.0.10 (84).

Phylogenetics. The 16S rRNA sequences present in the recovered MAGs were identified by SSU-ALIGN (72). The ones with a length longer than 200 nucleotides (nt) were merged with a data set comprising *Gemmatimonadota* sequences recovered from SILVA SSU v138 database (reference NR 99; length, 1200 bp) (70). The sequences were aligned with PASTA v1.8.3 (85) and used to construct a maximum-likelihood phylogeny with IQ-TREE v1.6.10 (-m GTR+F+R10; chosen as the best fitting model by ModelFinder) (81, 82). Several sequences pertaining to *Cyanobacteria* and *Fibrobacterota* were used as outgroup to root the tree.

Eighteen proteins belonging to the large subunit of the ribulose 1,5-bisphosphate carboxylase enzyme (RuBisCO) were identified in the assembled MAGs ($n = 106$). These proteins, together with a data set comprised of 146 RuBisCO (types I to III) and RuBisCO-like (type IV) proteins (86) were treated with PREQUAL v1.02 (87) prior to alignment with PASTA v1.8.3 (85). The obtained alignment (854 aligned positions) was used to construct a maximum-likelihood phylogeny with IQ-TREE v1.6.10 and the LG+F+R6 substitution model (chosen as the best-fitting model by ModelFinder) (81, 82).

Fragment recruitment. The obtained MAGs and the metagenomic data sets were used to compute genome coverage per gigabase using RazerS 3 (using cutoffs of >95% identity and alignment lengths of ≥ 50 bp) (88). All rRNA sequences (5S, 16S, and 23S) present in the MAGs were identified using *rrna_hmm* (89) and were masked prior to comparisons with quality-filtered metagenomic sequences. The quality-filtered data sets were used to compute abundance profiles for the *Gemmatimonadota* MAGs. Raw data showing coverage per gigabase is shown in Table S3, and a heatmap (Fig. 4) was created using <http://heatmapper.ca>.

Design of oligonucleotide probes. We designed a new oligonucleotide probe that targets a part of the class *Gemmatimonadetes* (Fig. S1). Originally, the probe was designed as a *Gemmatimonadetes*-specific broad-range (degenerate) PCR primer. *Gemmatimonadetes* and non-*Gemmatimonadetes* 16S rRNA gene sequences were retrieved from SSURef SILVA database (release 132 [70]), and aligned with the MUSCLE60 algorithm in MEGA7 software (90). The primers were then designed manually to target conserved regions specific for *Gemmatimonadetes*. The specificity of the primers was first tested *in silico* with the SILVA database (70) and then with PCR using DNA from *Gemmatimonas phototrophica* which was used as a positive control. Optimized PCR conditions were as follows: initial denaturation at 98°C for 3 min; 26 cycles with 1 cycle consisting of 98°C for 10 s, 60°C for 20 s, and 72°C for 20 s; and final elongation for 3 min. Primers were additionally tested with several other bacterial strains, which served as negative controls and showed no amplification. Finally, the primers were used on an environmental sample where their specificity was confirmed with Illumina amplicon sequencing. Briefly, three different primer sets were applied to an environmental sample collected at station DAM of Římov Reservoir. After Illumina sequencing, read quality was evaluated using FastQC v0.11.7 (Babraham Bioinformatics, Cambridge, UK). Primer sequences were trimmed using cutadapt v1.16 (91) and subsequently analyzed in the R/Bioconductor environment using the dada2 v1.6 package (37). Taxonomic assignment was performed using the SILVA 132 database (70). On the basis of these results, we opted to use the reverse primer (5' TCG CTC CCC CAR RSA CCT AGT 3') as a CARD-FISH-probe Gemma_801. As the probe-binding site was located within a hairpin loop and the accessibility of the binding site was low (92), we designed two 18-bp-long helpers to open the loop and facilitate hybridization (93): Gemma_801 H1 (5'-GCG CCG GCA YYC GAG GGG-3') and Gemma_801 H2 (5'-GGG GDA CTT AAT GCG TTA-3'). The specificity of the probe was tested *in silico* using the ARB Probe_Match function and the SILVA online ProbeCheck tool against the ENA database (EMBL-EBI). The probe matches only six untargeted sequences and six more if one weighted mismatch is allowed (<0.0021% of all bacterial sequences), and all of them originate from soil or the rhizosphere. The coverage of the probe for class *Gemmatimonadetes* is 35%, so the usage of this probe was primarily justified with the intent to visualize the *Gemmatimonadetes* cells in the natural environment, and not to use it for absolute quantification. Hybridization conditions were optimized at 35°C and 46°C by a stepwise increase of formamide concentrations in the hybridization buffer (20% to 70% by 5%) using pure cultures of *G. phototrophica* AP64 and photoheterotrophic strain TET16 (36) as positive controls, and *Limnohabitans planktonicus* (*Gammaproteobacteria/Betaproteobacteriales*) and *Sphingomonas* sp. strain AAP5 (*Alphaproteobacteria*) as negative controls. The final protocol is described below.

Sampling for catalyzed reporter deposition-fluorescence *in situ* hybridization (CARD-FISH) analysis. We applied the newly designed probe on a sample set collected between 14 May and 24 August 2015 at Římov Reservoir. Six permanent stations (river, stations 2, 3, 4, and 7, and dam) along the longitudinal transect of the reservoir (94) (Fig. S2) were sampled with a Friedinger sampler at 3-week intervals. Samples were taken from 0.5-m depth at all stations and at additional depths at lacustrine stations: 5 m and 10 m at stations 4 and 7; and 5 m, 10 m, 20 m, 30 m, and 40 m at the dam site. Water samples were fixed on-site with formaldehyde (2% final concentration [vol/vol]) and transported to the laboratory within 2 h. For the enumeration of bacterial densities, 1 to 2 ml was stained with 4',6'-diamidino-2-phenylindole (DAPI) and filtered onto black 0.2- μ m-pore-size polycarbonate membranes (25 mm diameter; SPI Supplies, USA), and bacterial abundance was determined via epifluorescence microscopy (95).

CARD-FISH analysis. Fixed formaldehyde water samples containing 1×10^6 to 2×10^6 bacteria were filtered onto polycarbonate membranes with a 0.2- μ m pore size (47-mm diameter; Millipore, USA), and filters were stored at -20°C. The catalyzed reporter deposition-fluorescence *in situ* hybridization (CARD-FISH) analysis was conducted as in the previously described protocol (96) with slight modification in digestion procedure, optimized for samples from Římov Reservoir. In brief, the microorganisms were immobilized on the membranes with low-melting-point agarose and permeabilized with lysozyme (10 μ g

ml⁻¹, 1 h) and achromopeptidase (60 U ml⁻¹, 20 to 25 min). Digestion was followed by neutralization of endogenous peroxidases and hybridization with the horseradish peroxidase-labeled probe Gemma_801 (biomers.net GmbH) and helpers in a hybridization buffer with 35% formamide concentration (vol/vol) at 35°C for 3 h. After the subsequent washing steps, amplification of the hybridization signal was performed with fluorescein-labeled tyramides (Invitrogen, Carlsbad, CA, USA) at 37°C for 30 min. Washed and dried filters were counterstained with DAPI (1 μg ml⁻¹) and analyzed with epifluorescence microscopy (Olympus BX-53F) using UNWU, U-WB, and U-WG optical filter sets. Proportions of CARD-FISH stained bacteria were determined by inspecting more than 1000 DAPI-stained cells per sample.

Data availability. Sequence data for all metagenomes generated in this work are archived at EBI European Nucleotide Archive and can be accessed under the BioProject accession number [PRJEB35770](https://doi.org/10.6084/m9.figshare.12662132.v2). All metagenome-assembled genomes are also available at figshare (<https://doi.org/10.6084/m9.figshare.12662132.v2>).

SUPPLEMENTAL MATERIAL

Supplemental material is available online only.

FIG S1, PDF file, 0.3 MB.

FIG S2, PDF file, 0.1 MB.

FIG S3, PDF file, 0.04 MB.

FIG S4, PDF file, 0.1 MB.

FIG S5, PDF file, 0.1 MB.

TABLE S1, XLSX file, 0.04 MB.

TABLE S2, XLSX file, 0.01 MB.

TABLE S3, XLSX file, 0.01 MB.

TABLE S4, XLSX file, 0.02 MB.

ACKNOWLEDGMENTS

This research was supported by the *PhotoGemm+* project (grant 19-28778X) from the Czech Science Foundation. R.G. was supported by the research grant 20-12496X (Grant Agency of the Czech Republic). A.-Ş.A. was supported by the research grants 17-04828S (Grant Agency of the Czech Republic) and MSM200961801 (Academy of Sciences of the Czech Republic). T.S. was supported by the research grant 20-23718Y (Grant Agency of the Czech Republic). M.M.S. was supported by the research grant 19-23469S (Grant Agency of the Czech Republic).

We are thankful to Yonghui Zeng and Thanassis Zervas (Aarhus University) for help in initial stages of this study. Petr Znachor and Pavel Rychtecký are acknowledged for providing samples from Římov Reservoir, and Petr Porcal for providing samples from Jiřícká pond. We are also thankful to Thomas Posch and Eugen Loher (University of Zurich) for help with sampling of Lake Zurich, Laboratory for Water and Soil Protection of the Canton of Bern (GBL), Switzerland, for help with sampling of Lake Thun and the crew of the research vessel “Kormoran” (LUBW-ISF Langenargen) for help with sampling of Lake Constance. We are thankful to Alastair T. Gardiner for language corrections.

REFERENCES

- Giovannoni SJ, Britschgi TB, Moyer CL, Field KG. 1990. Genetic diversity in Sargasso Sea bacterioplankton. *Nature* 345:60–63. <https://doi.org/10.1038/345060a0>.
- Parks DH, Chuvochina M, Waite DW, Rinke C, Skarshewski A, Chaumeil PA, Hugenholtz P. 2018. A standardized bacterial taxonomy based on genome phylogeny substantially revises the tree of life. *Nat Biotechnol* 36:996–1004. <https://doi.org/10.1038/nbt.4229>.
- Hug LA, Baker BJ, Anantharaman K, Brown CT, Probst AJ, Castelle CJ, Butterfield CN, Hemsdorf AW, Amano Y, Ise K, Suzuki Y, Dudek N, Relman DA, Finstad KM, Amundson R, Thomas BC, Banfield JF. 2016. A new view of the tree of life. *Nat Microbiol* 1:16048. <https://doi.org/10.1038/nmicrobiol.2016.48>.
- Whitman WB, Oren A, Chuvochina M, da Costa MS, Garrity GM, Rainey FA, Rossello-Mora R, Schink B, Sutcliffe I, Trujillo ME, Ventura S. 2018. Proposal of the suffix *-ota* to denote phyla. Addendum to ‘Proposal to include the rank of phylum in the International Code of Nomenclature of Prokaryotes.’ *Int J Syst Evol Microbiol* 68:967–969. <https://doi.org/10.1099/ijsem.0.002593>.
- Hugenholtz P, Tyson GW, Webb RI, Wagner AM, Blackall LL. 2001. Investigation of candidate division TM7, a recently recognized major lineage of the domain Bacteria, with no known pure-culture representatives. *Appl Environ Microbiol* 67:411–419. <https://doi.org/10.1128/AEM.67.1.411-419.2001>.
- Mummey DL, Stahl PD. 2003. Candidate division BD: phylogeny, distribution and abundance in soil ecosystems. *Syst Appl Microbiol* 26:228–235. <https://doi.org/10.1078/072320203322346074>.
- Madrid VM, Aller JY, Aller RC, Chistoserdov AY. 2001. High prokaryote diversity and analysis of community structure in mobile mud deposits off French Guiana: identification of two new bacterial candidate divisions. *FEMS Microbiol Ecol* 37:197–209. <https://doi.org/10.1111/j.1574-6941.2001.tb00867.x>.
- Zhang H, Sekiguchi Y, Hanada S, Hugenholtz P, Kim H, Kamagata Y, Nakamura K. 2003. *Gemmatimonas aurantiaca* gen. nov., sp. nov., a Gram-negative, aerobic, polyphosphate-accumulating micro-organism, the first cultured representative of the new bacterial phylum Gemmatimonadetes phyl. nov. *Int J Syst Evol Microbiol* 53:1155–1163. <https://doi.org/10.1099/ijso.0.02520-0>.

9. DeBruyn JM, Fawaz MN, Peacock AD, Dunlap JR, Nixon LT, Cooper KE, Radosevich M. 2013. *Gemmatirosa kalamazoonensis* gen. nov., sp. nov., a member of the rarely-cultivated bacterial phylum Gemmatimonadetes. *J Gen Appl Microbiol* 59:305–312. <https://doi.org/10.2323/jgam.59.305>.
10. Pascual J, García-López M, Bills GF, Genilloud O. 2016. Longimicrobium terrae gen. nov., sp. nov., an oligotrophic bacterium of the under-represented phylum Gemmatimonadetes isolated through a system of miniaturized diffusion chambers. *Int J Syst Evol Microbiol* 66:1976–1985. <https://doi.org/10.1099/ijsem.0.000974>.
11. Pascual J, Foesel BU, Geppert A, Huber KJ, Boedeker C, Luckner M, Wanner G, Overmann J. 2018. *Roseisolibacter agri* gen. nov., sp. nov., a novel slow-growing member of the under-represented phylum Gemmatimonadetes. *Int J Syst Evol Microbiol* 68:1028–1036. <https://doi.org/10.1099/ijsem.0.002619>.
12. Hanada S, Sekiguchi Y. 2014. The phylum Gemmatimonadetes, p 677–681. In Rosenburg E, DeLong EF, Lory S, Stackebrandt E, Thompson F (ed), *The prokaryotes*, 4th ed, vol 11. Springer, Berlin, Germany.
13. Park D, Kim H, Yoon S. 2017. Nitrous oxide reduction by an obligate aerobic bacterium, *Gemmatimonas aurantiaca* strain T-27. *Appl Environ Microbiol* 83:e00502-17. <https://doi.org/10.1128/AEM.00502-17>.
14. Zeng Y, Feng F, Medová H, Dean J, Koblížek M. 2014. Functional type 2 photosynthetic reaction centers found in the rare bacterial phylum Gemmatimonadetes. *Proc Natl Acad Sci U S A* 111:7795–7800. <https://doi.org/10.1073/pnas.1400295111>.
15. Dachev M, Bina D, Sobotka R, Moravcová L, Gardian Z, Kaftan D, Šlouf V, Fuciman M, Polívka T, Koblížek M. 2017. Unique double concentric ring organization of light harvesting complexes in *Gemmatimonas phototrophica*. *PLoS Biol* 15:e2003943. <https://doi.org/10.1371/journal.pbio.2003943>.
16. Ward LM, Cardona T, Holland-Moritz H. 2019. Evolutionary implications of anoxygenic phototrophy in the bacterial phylum Candidatus Eremiobacterota (WPS-2). *Front Microbiol* 10:1658. <https://doi.org/10.3389/fmicb.2019.01658>.
17. Cardona T. 2019. Thinking twice about the evolution of photosynthesis. *Open Biol* 9:180246. <https://doi.org/10.1098/rsob.180246>.
18. Zeng Y, Selyanin V, Lukeš M, Dean J, Kaftan D, Feng F, Koblížek M. 2015. Characterization of the microaerophilic, bacteriochlorophyll a-containing bacterium *Gemmatimonas phototrophica* sp. nov., and emended descriptions of the genus *Gemmatimonas* and *Gemmatimonas aurantiaca*. *Int J Syst Evol Microbiol* 65:2410–2419. <https://doi.org/10.1099/ijms.0.000272>.
19. Cardona T. 2016. Origin of bacteriochlorophyll a and the early diversification of photosynthesis. *PLoS One* 11:e0151250. <https://doi.org/10.1371/journal.pone.0151250>.
20. DeBruyn JM, Nixon LT, Fawaz MN, Johnson AM, Radosevich M. 2011. Global biogeography and quantitative seasonal dynamics of Gemmatimonadetes in soil. *Appl Environ Microbiol* 77:6295–6300. <https://doi.org/10.1128/AEM.05005-11>.
21. Zeng Y, Baumbach J, Barbosa EGV, Azevedo V, Zhang C, Koblížek M. 2016. Metagenomic evidence for the presence of phototrophic Gemmatimonadetes bacteria in diverse environments. *Environ Microbiol Rep* 8:139–149. <https://doi.org/10.1111/1758-2229.12363>.
22. Huang Y, Zeng Y, Lu H, Feng H, Zeng Y, Koblížek M. 2016. Novel acsF gene primers revealed a diverse phototrophic bacterial population, including Gemmatimonadetes, in Lake Taihu (China). *Appl Environ Microbiol* 82:5587–5594. <https://doi.org/10.1128/AEM.01063-16>.
23. Delgado-Baquerizo M, Oliverio AM, Brewer TE, Benavent-González A, Eldridge DJ, Bardgett RD, Maestre FT, Singh BK, Fierer N. 2018. A global atlas of the dominant bacteria found in soil. *Science* 359:320–325. <https://doi.org/10.1126/science.aap9516>.
24. Cary SC, McDonald IR, Barrett JE, Cowan DA. 2010. On the rocks: the microbiology of Antarctic Dry Valley soils. *Nat Rev Microbiol* 8:129–138. <https://doi.org/10.1038/nrmicro2281>.
25. Bakermans C, Skidmore ML, Douglas S, McKay CP. 2014. Molecular characterization of bacteria from permafrost of the Taylor Valley, Antarctica. *FEMS Microbiol Ecol* 89:331–346. <https://doi.org/10.1111/1574-6941.12310>.
26. Ren C, Chen J, Lu X, Doughty R, Zhao F, Zhong Z, Han X, Yang G, Feng Y, Ren G. 2018. Responses of soil total microbial biomass and community compositions to rainfall reductions. *Soil Biol Biochem* 116:4–10. <https://doi.org/10.1016/j.soilbio.2017.09.028>.
27. Fecskéová LK, Piwosz K, Hanusová M, Nedoma J, Znachor P, Koblížek M. 2019. Diel changes and diversity of pufM expression in freshwater communities of anoxygenic phototrophic bacteria. *Sci Rep* 9:18766. <https://doi.org/10.1038/s41598-019-55210-x>.
28. Song H, Li Z, Du B, Wang G, Ding Y. 2012. Bacterial communities in sediments of the shallow Lake Dongping in China. *J Appl Microbiol* 112:79–89. <https://doi.org/10.1111/j.1365-2672.2011.05187.x>.
29. Zhang J, Yang Y, Zhao L, Li Y, Xie S, Liu Y. 2015. Distribution of sediment bacterial and archaeal communities in plateau freshwater lakes. *Appl Microbiol Biotechnol* 99:3291–3302. <https://doi.org/10.1007/s00253-014-6262-x>.
30. Gugliandolo C, Michaud L, Lo Giudice A, Lentini V, Rochera C, Camacho A, Maugeri TL. 2016. Prokaryotic community in lacustrine sediments of Byers Peninsula (Livingston Island, Maritime Antarctica). *Microb Ecol* 71:387–400. <https://doi.org/10.1007/s00248-015-0666-8>.
31. Gołębiewski M, Calkiewicz J, Creer S, Piwosz K. 2017. Tideless estuaries in brackish seas as possible freshwater-marine transition zones for bacteria: the case study of the Vistula river estuary. *Environ Microbiol Rep* 9:129–143. <https://doi.org/10.1111/1758-2229.12509>.
32. Traving SJ, Rowe O, Jakobsen NM, Sørensen H, Dinascuet J, Stedmon CA, Andersson A, Riemann L. 2017. The effect of increased loads of dissolved organic matter on estuarine microbial community composition and function. *Front Microbiol* 8:351. <https://doi.org/10.3389/fmicb.2017.00351>.
33. Parks DH, Rinke C, Chuvochina M, Chaumeil PA, Woodcroft BJ, Evans PN, Hugenholtz P, Tyson GW. 2017. Recovery of nearly 8,000 metagenome-assembled genomes substantially expands the tree of life. *Nat Microbiol* 2:1533–1542. <https://doi.org/10.1038/s41564-017-0012-7>.
34. Holland-Moritz H, Stuart J, Lewis LR, Miller S, Mack MC, McDaniel SF, Fierer N. 2018. Novel bacterial lineages associated with boreal moss species. *Environ Microbiol* 20:2625–2638. <https://doi.org/10.1111/1462-2920.14288>.
35. Cabello-Yeves PJ, Zemskaya TI, Rosselli R, Coutinho FH, Zakharenko AS, Blinov VV, Rodriguez-Valera F. 2017. Genomes of novel microbial lineages assembled from the sub-ice waters of Lake Baikal. *Appl Environ Microbiol* 84:e02132-17. <https://doi.org/10.1128/AEM.02132-17>.
36. Zeng Y, Nupur, Wu N, Madsen AM, Chen X, Gardiner AT, Koblížek M. 2021. *Gemmatimonas groenlandica* sp. nov. is an aerobic anoxygenic phototroph in the phylum Gemmatimonadetes. *Front Microbiol* 11:606612. <https://doi.org/10.3389/fmicb.2020.606612>.
37. Callahan BJ, McMurdie PJ, Rosen MJ, Han AW, Johnson AJA, Holmes SP. 2016. DADA2: high-resolution sample inference from Illumina amplicon data. *Nat Methods* 13:581–583. <https://doi.org/10.1038/nmeth.3869>.
38. Konstantinidis KT, Rosselló-Móra R, Amann R. 2017. Uncultivated microbes in need of their own taxonomy. *ISME J* 11:2399–2406. <https://doi.org/10.1038/ismej.2017.113>.
39. Chaumeil P-A, Mussig AJ, Hugenholtz P, Parks DH. 2019. GTDB-Tk: a toolkit to classify genomes with the Genome Taxonomy Database. *Bioinformatics* 36:1925–1927. <https://doi.org/10.1093/bioinformatics/btz848>.
40. Koblížek M. 2015. Ecology of aerobic anoxygenic phototrophs in aquatic environments. *FEMS Microbiol Rev* 39:854–870. <https://doi.org/10.1093/femsre/fuv032>.
41. Kavagutti VS, Andrei AŞ, Mehrshad M, Salcher MM, Ghai R. 2019. Phage-centric ecological interactions in aquatic ecosystems revealed through ultra-deep metagenomics. *Microbiome* 7:135. <https://doi.org/10.1186/s40168-019-0752-0>.
42. Ameryk A, Kownacka J, Zalewski M, Piwosz K. 2021. Typical freshwater and marine bacterial lineages dynamics at salinity between 0 and 4 in the Vistula Lagoon. *Estuar Coast Shelf Sci* 250:107100. <https://doi.org/10.1016/j.ecss.2020.107100>.
43. Morrison JM, Baker KD, Zamor RM, Nikolai S, Elshahed MS, Youssef NH. 2017. Spatiotemporal analysis of microbial community dynamics during seasonal stratification events in a freshwater lake (Grand Lake, OK, USA). *PLoS One* 12:e0177488. <https://doi.org/10.1371/journal.pone.0177488>.
44. Cai H, Jiang H, Krumholz LR, Yang Z. 2014. Bacterial community composition of size-fractionated aggregates within the phycosphere of cyanobacterial blooms in a eutrophic freshwater lake. *PLoS One* 9:e102879. <https://doi.org/10.1371/journal.pone.0102879>.
45. Shia L, Cai Y, Wang X, Li P, Yu Y, Kong F. 2010. Community structure of bacteria associated with *Microcystis* colonies from cyanobacterial blooms. *J Freshw Ecol* 25:193–203. <https://doi.org/10.1080/02705060.2010.9665068>.
46. Fauteux L, Cottrell MT, Kirchman DL, Borrego CM, Garcia-Chaves MC, Del Giorgio PA. 2015. Patterns in abundance, cell size and pigment content of aerobic anoxygenic phototrophic bacteria along environmental gradients in northern lakes. *PLoS One* 10:e0124035. <https://doi.org/10.1371/journal.pone.0124035>.
47. Hauruseu D, Koblížek M. 2012. Influence of light on carbon utilization in aerobic anoxygenic phototrophs. *Appl Environ Microbiol* 78:7414–7419. <https://doi.org/10.1128/AEM.01747-12>.

48. Xing P, Kong F, Cao H, Zhang M, Tan X. 2007. Variations of bacterioplankton community composition during *Microcystis* spp. blooms in a shallow eutrophic lake. *J Freshw Ecol* 22:61–67. <https://doi.org/10.1080/02705060.2007.9664146>.
49. Šimek K, Hornák K, Jezbera J, Mašín M, Nedoma J, Gasol JM, Schauer M. 2005. Influence of top-down and bottom-up manipulations on the R-BT065 subcluster of β -proteobacteria, an abundant group in bacterioplankton of a freshwater reservoir. *Appl Environ Microbiol* 71:2381–2390. <https://doi.org/10.1128/AEM.71.5.2381-2390.2005>.
50. Salcher MM, Pernthaler J, Posch T. 2011. Seasonal bloom dynamics and ecophysiology of the freshwater sister clade of SAR11 bacteria that rule the waves (LD12). *ISME J* 5:1242–1252. <https://doi.org/10.1038/ismej.2011.8>.
51. Šimek K, Hornák K, Jezbera J, Nedoma J, Znachor P, Hejzlar J, Sedá J. 2008. Spatio-temporal patterns of bacterioplankton production and community composition related to phytoplankton composition and protistan bacterivory in a dam reservoir. *Aquat Microb Ecol* 51:249–262. <https://doi.org/10.3354/ame01193>.
52. Yurkov VV, Beatty JT. 1998. Aerobic anoxygenic phototrophic bacteria. *Microbiol Mol Biol Rev* 62:695–724. <https://doi.org/10.1128/MMBR.62.3.695-724.1998>.
53. Goecke F, Thiel V, Wiese J, Labes A, Imhoff JF. 2013. Algae as an important environment for bacteria – phylogenetic relationships among new bacterial species isolated from algae. *Phycologia* 52:14–24. <https://doi.org/10.2216/12-24.1>.
54. Piwosz K, Vrdoljak A, Frenken T, González-Olalla JM, Šantić D, McKay RM, Spilling K, Guttman L, Znachor P, Mujakić I, Fecsková LK, Zoccarato L, Hanusová M, Pessina A, Reich T, Grossart H-P, Koblížek M. 2020. Light and primary production shape bacterial activity and community composition of aerobic anoxygenic phototrophic bacteria in a microcosm experiment. *mSphere* 5:e00673-20. <https://doi.org/10.1128/mSphere.00673-20>.
55. Zheng Q, Zhang R, Koblížek M, Boldareva EN, Yurkov V, Yan S, Jiao N. 2011. Diverse arrangement of photosynthetic gene clusters in aerobic anoxygenic phototrophic bacteria. *PLoS One* 6:e25050. <https://doi.org/10.1371/journal.pone.0025050>.
56. Hendrickson EL, Leigh JA. 2008. Roles of coenzyme F420-reducing hydrogenases and hydrogen- and F420-dependent methylenetetrahydromethanopterin dehydrogenases in reduction of F420 and production of hydrogen during methanogenesis. *J Bacteriol* 190:4818–4821. <https://doi.org/10.1128/JB.00255-08>.
57. Vavourakis CD, Mehrshad M, Balkema C, Van Hall R, Andrei AŞ, Ghai R, Sorokin DY, Muyzer G. 2019. Metagenomes and metatranscriptomes shed new light on the microbial-mediated sulfur cycle in a Siberian soda lake. *BMC Biol* 17:69. <https://doi.org/10.1186/s12915-019-0688-7>.
58. Zorz JK, Sharp C, Kleiner M, Gordon PMK, Pon RT, Dong X, Strous M. 2019. A shared core microbiome in soda lakes separated by large distances. *Nat Commun* 10:4230. <https://doi.org/10.1038/s41467-019-12195-5>.
59. Hanson TE, Tabita FR. 2001. A ribulose-1,5-bisphosphate carboxylase/oxygenase (RuBisCO)-like protein from *Chlorobium tepidum* that is involved with sulfur metabolism and the response to oxidative stress. *Proc Natl Acad Sci U S A* 98:4397–4402. <https://doi.org/10.1073/pnas.081610398>.
60. Tabita FR, Hanson TE, Satagopan S, Witte BH, Kreel NE. 2008. Phylogenetic and evolutionary relationships of RuBisCO and the RuBisCO-like proteins and the functional lessons provided by diverse molecular forms. *Philos Trans R Soc Lond B Biol Sci* 363:2629–2640. <https://doi.org/10.1098/rstb.2008.0023>.
61. Ashida H, Danchin A, Yokota A. 2005. Was photosynthetic RuBisCO recruited by acquisitive evolution from RuBisCO-like proteins involved in sulfur metabolism? *Res Microbiol* 156:611–618. <https://doi.org/10.1016/j.resmic.2005.01.014>.
62. Ashida H, Saito Y, Kojima C, Kobayashi K, Ogasawara N, Yokota A. 2003. A functional link between RuBisCO-like protein of *Bacillus* and photosynthetic RuBisCO. *Science* 302:286–290. <https://doi.org/10.1126/science.1086997>.
63. Yurkov V, Hughes E. 2017. Aerobic anoxygenic phototrophs: four decades of mystery, p 193–214. *In* Modern topics in the phototrophic prokaryotes: environmental and applied aspects. Springer International Publishing, Cham, Switzerland.
64. Piwosz K, Kaftan D, Dean J, Šetlík J, Koblížek M. 2018. Nonlinear effect of irradiance on photoheterotrophic activity and growth of the aerobic anoxygenic phototrophic bacterium *Dinoroseobacter shibae*. *Environ Microbiol* 20:724–733. <https://doi.org/10.1111/1462-2920.14003>.
65. Znachor P, Hejzlar J, Vrba J, Nedoma J, Seda J, Šimek K, Komarková J, Kopáček J, Šorf M, Kubečka J, Matena J, Riha M, Peterka J, Cech M, Vasek M. 2016. Brief history of long-term ecological research into aquatic ecosystems and their catchments in the Czech Republic. Part I: manmade reservoirs. *Biology Centre CAS, v.v.i., Institute of Hydrobiology, Česká Budějovice, Czechia*.
66. Bushnell B. 2015. BBMap (version 35.14). <https://sourceforge.net/projects/bbmap/>.
67. Bushnell B, Rood J, Singer E. 2017. BBMerge – accurate paired shotgun read merging via overlap. *PLoS One* 12:e0185056. <https://doi.org/10.1371/journal.pone.0185056>.
68. Bushnell B. 2016. Reformat. <https://github.com/BioInfoTools/BBMap/blob/master/sh/reformat.sh>.
69. Li D, Luo R, Liu C-M, Leung C-M, Ting H-F, Sadakane K, Yamashita H, Lam T-W. 2016. MEGAHIT v1.0: a fast and scalable metagenome assembler driven by advanced methodologies and community practices. *Methods* 102:3–11. <https://doi.org/10.1016/j.ymeth.2016.02.020>.
70. Pruesse E, Quast C, Knittel K, Fuchs BM, Ludwig W, Peplies J, Glöckner FO. 2007. SILVA: a comprehensive online resource for quality checked and aligned ribosomal RNA sequence data compatible with ARB. *Nucleic Acids Res* 35:7188–7196. <https://doi.org/10.1093/nar/gkm864>.
71. Steingegger M, Söding J. 2017. MMseqs2 enables sensitive protein sequence searching for the analysis of massive data sets. *Nat Biotechnol* 35:1026–1028. <https://doi.org/10.1038/nbt.3988>.
72. Nawrocki EP. 2009. Structural RNA homology search and alignment using covariance models. PhD thesis. Washington University in St. Louis, St. Louis, MO.
73. Anderson MJ, Legendre P. 1999. An empirical comparison of permutation methods for tests of partial regression coefficients in a linear model. *J Stat Comput Simul* 62:271–303. <https://doi.org/10.1080/00949659908811936>.
74. Legendre P, Anderson MJ. 1999. Distance-based redundancy analysis: testing multispecies responses in multifactorial ecological experiments. *Ecol Monogr* 69:1–24. [https://doi.org/10.1890/0012-9615\(1999\)069\[0001:DBRATM\]2.0.CO;2](https://doi.org/10.1890/0012-9615(1999)069[0001:DBRATM]2.0.CO;2).
75. Bushnell B. 2016. BBMap short-read aligner, and other bioinformatic tools. <https://sourceforge.net/projects/bbmap>.
76. Kang DD, Froula J, Egan R, Wang Z. 2015. MetaBAT, an efficient tool for accurately reconstructing single genomes from complex microbial communities. *PeerJ* 3:e1165. <https://doi.org/10.7717/peerj.1165>.
77. Parks DH, Imelfort M, Skennerton CT, Hugenholtz P, Tyson GW. 2015. CheckM: assessing the quality of microbial genomes recovered from isolates, single cells, and metagenomes. *Genome Res* 25:1043–1055. <https://doi.org/10.1101/gr.186072.114>.
78. Haft DH, Selengut JD, White O. 2003. The TIGRFAMs database of protein families. *Nucleic Acids Res* 31:371–373. <https://doi.org/10.1093/nar/gkg128>.
79. Löytynoja A. 2014. Phylogeny-aware alignment with PRANK. *Methods Mol Biol* 1079:155–170. https://doi.org/10.1007/978-1-62703-646-7_10.
80. Criscuolo A, Gribaldo S. 2010. BMGE (Block Mapping and Gathering with Entropy): a new software for selection of phylogenetic informative regions from multiple sequence alignments. *BMC Evol Biol* 10:210. <https://doi.org/10.1186/1471-2148-10-210>.
81. Nguyen LT, Schmidt HA, Von Haeseler A, Minh BQ. 2015. IQ-TREE: a fast and effective stochastic algorithm for estimating maximum-likelihood phylogenies. *Mol Biol Evol* 32:268–274. <https://doi.org/10.1093/molbev/msu300>.
82. Kalyaanamoorthy S, Minh BQ, Wong TKF, Von Haeseler A, Jermin LS. 2017. ModelFinder: fast model selection for accurate phylogenetic estimates. *Nat Methods* 14:587–589. <https://doi.org/10.1038/nmeth.4285>.
83. Konstantinidis KT, Tiedje JM. 2005. Genomic insights that advance the species definition for prokaryotes. *Proc Natl Acad Sci U S A* 102:2567–2572. <https://doi.org/10.1073/pnas.0409727102>.
84. Larkin MA, Blackshields G, Brown NP, Chenna R, McGettigan PA, McWilliam H, Valentin F, Wallace IM, Wilm A, Lopez R, Thompson JD, Gibson TJ, Higgins DG. 2007. Clustal W and Clustal X version 2.0. *Bioinformatics* 23:2947–2948. <https://doi.org/10.1093/bioinformatics/btm404>.
85. Mirarab S, Nguyen N, Guo S, Wang LS, Kim J, Warnow T. 2015. PASTA: ultra-large multiple sequence alignment for nucleotide and amino-acid sequences. *J Comput Biol* 22:377–386. <https://doi.org/10.1089/cmb.2014.0156>.
86. Bulzu PA, Andrei AŞ, Salcher MM, Mehrshad M, Inoue K, Kandori H, Beja O, Ghai R, Banciu HL. 2019. Casting light on Asgardarchaeota metabolism in a sunlit microoxic niche. *Nat Microbiol* 4:1129–1137. <https://doi.org/10.1038/s41564-019-0404-y>.

87. Whelan S, Irisarri I, Burki F. 2018. PREQUAL: detecting non-homologous characters in sets of unaligned homologous sequences. *Bioinformatics* 34:3929–3930. <https://doi.org/10.1093/bioinformatics/bty448>.
88. Weese D, Holtgrewe M, Reinert K. 2012. RazerS 3: faster, fully sensitive read mapping. *Bioinformatics* 28:2592–2599. <https://doi.org/10.1093/bioinformatics/bts505>.
89. Huang Y, Gilna P, Li W. 2009. Identification of ribosomal RNA genes in metagenomic fragments. *Bioinformatics* 25:1338–1340. <https://doi.org/10.1093/bioinformatics/btp161>.
90. Kumar S, Stecher G, Tamura K. 2016. MEGA7: Molecular Evolutionary Genetics Analysis version 7.0 for bigger datasets. *Mol Biol Evol* 33:1870–1874. <https://doi.org/10.1093/molbev/msw054>.
91. Martin M. 2011. Cutadapt removes adapter sequences from high-throughput sequencing reads. *EMBnetjournal* 17:10. <https://doi.org/10.14806/ej.17.1.200>.
92. Behrens S, Rühlend C, Inácio J, Huber H, Fonseca A, Spencer-Martins I, Fuchs BM, Amann R. 2003. In situ accessibility of small-subunit rRNA of members of the domains Bacteria, Archaea, and Eucarya to Cy3-labeled oligonucleotide probes. *Appl Environ Microbiol* 69:1748–1758. <https://doi.org/10.1128/aem.69.3.1748-1758.2003>.
93. Fuchs BM, Glockner FO, Wulf J, Amann R. 2000. Unlabeled helper oligonucleotides increase the in situ accessibility to 16S rRNA of fluorescently labeled oligonucleotide probes. *Appl Environ Microbiol* 66:3603–3607. <https://doi.org/10.1128/aem.66.8.3603-3607.2000>.
94. Rychtecký P, Znachor P. 2011. Spatial heterogeneity and seasonal succession of phytoplankton along the longitudinal gradient in a eutrophic reservoir. *Hydrobiologia* 663:175–186. <https://doi.org/10.1007/s10750-010-0571-6>.
95. Porter KG, Feig YS. 1980. The use of DAPI for identifying and counting aquatic microflora. *Limnol Oceanogr* 25:943–948. <https://doi.org/10.4319/lo.1980.25.5.0943>.
96. Sekar R, Pernthaler A, Pernthaler J, Warnecke F, Posch T, Amann R. 2003. An improved protocol for quantification of freshwater Actinobacteria by fluorescence in situ hybridization. *Appl Environ Microbiol* 69:2928–2935. <https://doi.org/10.1128/aem.69.5.2928-2935.2003>.