# Five decades of change in somatic growth of Pacific hake from Puget Sound and Strait of Georgia 

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#### Abstract

Declines in fish body size have been reported in many populations and these changes likely have important ramifications for the sustainability of harvested species and ecosystem function. Pacific hake, Merluccius productus, have shown declines in size over the last several decades for populations located in Puget Sound (PS), Washington, USA, and Strait of Georgia (SoG), British Columbia, Canada. To examine this decrease in size, we used archived otoliths from both populations to assess when the decrease in somatic growth occurred and explored what factors and processes might explain the decline, including otolith microchemistry to infer the environment experienced by fish at different ages. Results indicated that substantial changes in juvenile somatic growth have occurred across decades. The divergence in body size occurred in the second summer, whereby SoG fish grew, on average, $18 \%$ more than PS fish. Within the PS population, somatic growth differed significantly among fish that hatched in the 1980s, 1990s, and 2010s, such that the more recently hatched fish grew $26 \%$ more in their first summer and $71 \%$ less in their second summer relative to those that hatched in the 1980s. In comparison, growth of SoG fish did not differ between those that hatched in 1970s and 1990s. For both populations growth in the first and third summer was positively and negatively related, respectively, to the abundance of harbor seals, while growth in the first and second summer was negatively related to salinity. Overall, this study highlights the complicated nature of Pacific hake population recovery under dynamic, and typically uncontrollable, variation in biotic and abiotic conditions.


Subjects Aquaculture, Fisheries and Fish Science, Ecology, Marine Biology, Population Biology Keywords Species of concern, Trace elements, Fish body size, Pacific whiting, Population recovery

## INTRODUCTION

Some populations respond to mitigation actions, such as harvest regulation, while other populations do not. Atlantic cod (Gadus morhua), for example, has not recovered from decades of overfishing despite harvest being largely curtailed (Olsen et al., 2004). This is in contrast to species recovering and supporting sustainable fisheries around the world
${ }^{1}$ ESA allows for conservation of organisms at the subspecies level and thus defines "species" as "any subspecies of fish or wildlife or plants, and any distinct population segment of any species of vertebrate fish or wildlife which interbreeds when mature" USFWS-NMFS (1996) Policy regarding the recognition of distinct vertebrate population segments under the Endangered Species Act. Federal Register. 61(26):4722-4725.
(Hilborn et al., 2020) some of which are in U.S. waters (NOAA Fisheries, 2020). For instance, on the U.S. west coast the population of Pacific Hake (Merluccius productus), which ranges from southern California to northern British Columbia, Canada, has responded quite favorably to harvest regulation and is the largest fishery (by metric tons landed) on the U.S. West Coast (Grandin et al., 2020). Nonetheless, a closely related population, found in Puget Sound, Washington, USA has been subjected to fishing regulations (i.e., fishery closure) since 1991, but still has not recovered (Gustafson et al., 2000).

The Puget Sound (PS) population along with the Strait of Georgia (SoG) population, in British Columbia, Canada, are considered a distinct population segment (DPS) ${ }^{1}$ under the U.S. Endangered Species Act (ESA). This DPS, referred to as the Georgia Basin DPS (Fig. 1), was listed as a Species of Concern under the ESA by both U.S. federal and state agencies, in part due to the decades-long declining status of the PS population (Gustafson et al., 2000; WDFW, 2008).

In addition to a decrease in biomass of approximately $85 \%$ since at least the 1980s, the PS population has also experienced substantial declines in body size ( $26 \%$ for 5 - to 7 -year-olds) during this same time period (Gustafson et al., 2000), some of which is attributable to high harvest rates in the early and mid-1980s (Goñi, 1988). In contrast, the SoG population has experienced moderate reductions in size (approximately $12 \%$ reduction for 4 and 5 year olds) and biomass (35\%) from 1981 to 1997 (King \& McFarlane, 2006). Taken together, these declines in the Georgia Basin DPS biomass and body size highlight that changes to the prey and/or habitat resources or processes (e.g., predation) critical to a specific life stage (e.g., juveniles) have occurred and may have impacted the sustainability of the DPS.

In this study we evaluate Pacific hake resource quality and availability in both the PS and SoG populations by quantifying somatic growth in individuals that hatched over a 41-year period (1974-2015). We used somatic growth as an indicator of individual performance and habitat quality (Le Pape et al., 2003; Meng et al., 2000; Necaise, Ross e Miller, 2005) given the known positive relationships between body size and fitness (Kingsolver \& Huey, 2008), and survival (Juan-Jordá et al., 2015; Sogard, 1997; Zabel \& Achord, 2004). Measuring somatic growth offers a direct method of evaluating the physiological status (i.e., performance) of an individual because somatic growth is related to factors such as food quality (Steves \& Cowen, 2000) and availability (Graeb et al., 2004; Ligas et al., 2015), temperature (Baumann et al., 2006; Ligas et al., 2015), and processes like competition (Helser \& Almeida, 1997) and predation (Baumann et al., 2003). To reconstruct somatic growth of an individual we counted visible increments in their otoliths to estimate age (Jones, 1992; Stevenson \& Campana, 1992) and used body size-age relationships to estimate individual somatic growth during each summer and winter (Casselman, 1987). Using this measure of performance, we evaluated the extent to which size-at-age varied within and between populations, how size-at-age varied across years, and what factors and/or processes putatively explain variability in somatic growth.


Figure 1 Map of Strait of Georgia (Canada) and Puget Sound (USA) where Pacific hake (Merluccius productus) were collected. White ovals indicate location of known and historical spawning ground. Black circles indicate collection sites at (A) Lasqueti Island, (B) Gabriola Island, (C) Burrard Inlet, and (D) Port Susan.
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## MATERIALS AND METHODS

## Study species

Pacific hake (Merluccius productus) is a pelagic species distributed from Baja California, USA to northern B.C., Canada (Berger et al., 2019). We studied size-at-age patterns of Pacific hake within the Georgia Basin DPS, which includes individuals from populations in greater Puget Sound (PS) and the Strait of Georgia (SoG) (Fig. 1) for which genetic differences have been reported (Iwamoto, Ford \& Gustafson, 2004; Iwamoto et al., 2015). In PS, spawning

Table 1 Number of sampled Pacific hake per population, year, and age.

| Year\age | Strait of Georgia |  |  |  |  |  | Puget Sound |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 |  | 0 | 1 | 2 | 3 | 4 | 5 | 6 |  |  |
| 1979 |  | 7 | 16 | 6 |  |  |  |  |  |  |  |  |  |  | 29 |
| 1982 |  | 8 | 19 | 4 |  |  |  |  |  |  |  |  |  |  | 31 |
| 1987 |  |  |  |  |  |  |  |  | 5 | 18 | 2 |  |  |  | 25 |
| 1996 | 12 |  |  |  |  |  |  |  | 27 |  |  |  |  |  | 39 |
| 1997 | 13 |  |  |  |  |  |  |  | 35 | 1 |  |  |  |  | 49 |
| 1999 | 23 | 7 |  |  |  |  |  |  |  |  |  |  |  |  | 30 |
| 2000 |  |  |  |  |  |  |  |  | 4 | 72 | 7 | 1 |  |  | 84 |
| 2001 |  |  |  |  |  |  |  |  | 27 | 8 |  |  |  |  | 35 |
| 2002 |  |  | 2 | 12 | 4 |  | 19 | 8 | 2 | 8 | 8 | 10 | 1 |  | 74 |
| 2005 |  |  |  |  |  |  | 3 |  |  |  |  |  |  |  | 3 |
| 2008 |  |  |  |  |  |  | 2 |  |  |  |  |  |  |  | 2 |
| 2016 |  |  |  |  |  |  |  | 3 | 22 | 10 | 3 |  |  |  | 38 |
| Total | 48 | 22 | 37 | 22 | 4 | 133 | 24 | 11 | 122 | 117 | 20 | 11 | 1 | 306 | 439 |

${ }^{2}$ Reference to trade names does not imply endorsement by the National Marine Fisheries Service, NOAA.
aggregations occur in Port Susan (Pedersen, 1985) and Dabob Bay (Utter, Stormont e Hodgins, 1970). In SoG, spawning aggregations occur in Stuart Channel (McFarlane \& Beamish, 1985) and Saanich Inlet (Beamish, Smith \& Scarsbrook, 1978). Individuals mature by age 3-4 (McFarlane \& Beamish, 1985) and spawning occurs primarily from February through April, peaking in March (Goñi, 1988). Hatching takes place 4-6 days after fertilization, at which point larvae are 2-3 mm total length (Bailey, 1982b; Hollowed, 1992).

## Fish collections and otolith preparation

Using pelagic trawls, juvenile and adult Pacific hake were collected from PS and SoG spawning grounds by the Washington Department of Fish and Wildlife (WDFW) (Burger et al., 2019) and Fisheries and Oceans, Canada (DFO), respectively (Fig. 1) across several years (Table 1). Fish total length (TL) was measured, and otoliths were removed and stored dry. Left sagittal otoliths were mounted on glass microscope slides using thermoplastic cement; otoliths were then polished in a sagittal plane using a grinding wheel with slurries of 600 -grit silicon carbide, 5.0 alumina oxide, and 1.0 micropolish (Buehler ${ }^{2}$ ) until the core was visible. Images of each otolith were taken with transmitted light using a digital camera (Mediacybernetics, EvolutionMP) at $4 \times$ magnification and analyzed using image processing software (Image Pro Plus, version 7.0).

## Age, length, and growth

To estimate age and change in size over time, we examined the annual pattern of otolith growth, which consists of an opaque and translucent zone (Chilton \& Beamish, 1982). Under transmitted light, the opaque zone corresponds to a summertime period when fish do most of their growing, while the translucent zone is a period of reduced winter growth (Beamish, 1979; Chilton \& Beamish, 1982) (Fig. 2A). For each fish, we estimated age by counting the number of otolith translucent zones (Bailey, 1982b; Beamish, 1979) along


Figure 2 Pacific hake otolith microstructure and microchemistry. (A) Image of a Pacific hake otolith (individual WDFW-2000-194 under transmitted light) showing the opaque and translucent zones that correspond to summer and winter otolith growth, respectively for each of three years. (B) Calcium intensity (counts per second, cps; yrlS and yrlW corresponds to year one summer and year one winter, respectively) and ratios of magnesium, manganese, strontium, barium, and lead ( $\mathrm{Mg}, \mathrm{Mn}, \mathrm{Sr}, \mathrm{Ba}, \mathrm{Pb}$, respectively) to calcium (Ca) $\left(\mathrm{mm} \mathrm{mol}^{-1}\right)$ with respect to distance from the otolith core (microns) as measured by laser ablation inductively coupled plasma mass spectrometry.

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a transect perpendicular to its longest axis on the dorsal side. Age was estimated for PS and SoG fish by WDFW in Olympia, Washington, and by DFO at the Pacific Biological Station in Nanaimo, BC, respectively. A second age was estimated by the U.S. National Marine Fisheries Service, Northwest Fisheries Science Center in Seattle, Washington. If the two ages differed for an individual, then age was estimated a third and final time by the Northwest Fisheries Science Center. Fish were assigned an age and the otolith included for statistical analysis only if two age estimates were the same.

Because somatic and otolith growth are positively correlated and consistent between both populations ( $R^{2}=0.83, n=532$, Fig. S1) we used the otolith radius at opaque and translucent zones to estimate fish length at the end of each summer and winter, respectively. Specifically, to back-calculate summer and winter body sizes we measured the otolith radius to each opaque $\left(O_{s}\right)$ and translucent $\left(O_{w}\right)$ zone, as well as the distance from the otolith core to the edge (i.e., otolith radius at time of capture, $O_{c}$ ) (Beamish, 1979) (Fig. 2A). For each individual, fish length after each summer $\left(L_{s}\right)$ and winter $\left(L_{w}\right)$ was estimated using the Biological Intercept equation (Campana, 2011; Campana \& Jones, 1992):
$L_{s}=L_{c}+\frac{\left(O_{s}-O_{c}\right) x\left(L_{c}-L_{i}\right)}{\left(O_{c}-O_{i}\right)}$
where $L_{i}$ and $O_{i}$ were the biological intercepts and were defined as 3.9 mm and $14.1 \mu \mathrm{~m}$, respectively and $L_{c}$ corresponds to the length at capture. The biological intercepts corresponded to the size at first feeding and were based on work by Bailey (1982a) and Butler \& Nishimoto (1997). In addition, we calculated the estimated amount an individual grew ( mm ) each summer and winter as the difference between its length at any summer or winter from its length estimated for the previous winter or summer, respectively.

Our statistical analyses of length and growth data had two objectives. Our first objective was to identify when size differences arose between PS and SoG fish. We used a repeated measures ANOVA to compare length and growth between fish collected from PS and SoG, regardless of hatch year. Repeated measures ANOVA was appropriate because our otolith-derived estimates of length and growth represent repeated measures that were calculated for each age-season of a fish's life (i.e., first summer, first winter, second summer, etc.) (Table 2).

The second objective was to determine the extent to which length and growth have changed across hatch years, grouped by decade, within a population. We used a repeated measures ANOVA to investigate if length and growth differed among each year/season and among hatch years. We grouped fish according to the decade in which they hatched because low sample sizes for certain hatch years precluded year-specific analysis. If significant differences were detected we performed a Tukey's post-hoc test to determine which decades were significantly different in terms of length and growth. We used a mixed effects model to describe the repeated measures analysis using the lme function in the nlme package of RStudio (version 1.1.463) ( $R$ Core Team, 2021). 'Individual' was treated as a random variable in the model and we included an autocorrelation structure for lags in the time variable (i.e., year/season) via the $A C F$ function in the nlme package.

| Age Season | 1 <br> Summer |  | 1 <br> Winter |  | $2$ <br> Summer |  | 2 <br> Winter |  | $3$ <br> Summer |  | 3 <br> Winter |  | 4 <br> Summer |  | 4 Winter |  | $5$ <br> Summer |  | 5 <br> Winter |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Hatch year | SoG | PS | SoG | PS | SoG | PS | SoG | PS | SoG | PS | SoG | PS | SoG | PS | SoG | PS | SoG | PS | SoG | PS |
| 1974 | 6 |  | 6 |  | 6 |  | 6 |  | 6 |  | 6 |  | 6 |  | 6 |  | 6 |  | 6 |  |
| 1975 | 16 |  | 16 |  | 16 |  | 16 |  | 16 |  | 16 |  | 16 |  | 16 |  |  |  |  |  |
| 1976 | 7 |  | 7 |  | 7 |  | 7 |  | 7 |  | 7 |  |  |  |  |  |  |  |  |  |
| 1977 | 4 |  | 4 |  | 4 |  | 4 |  | 4 |  | 4 |  | 4 |  | 4 |  | 4 |  | 4 |  |
| 1978 | 19 |  | 19 |  | 19 |  | 19 |  | 19 |  | 19 |  | 19 |  | 18 |  |  |  |  |  |
| 1979 | 8 |  | 8 |  | 8 |  | 8 |  | 8 |  | 8 |  |  |  |  |  |  |  |  |  |
| 1983 |  | 2 |  | 2 |  | 2 |  | 2 |  | 2 |  | 2 |  | 2 |  | 2 |  |  |  |  |
| 1984 |  | 21 |  | 18 |  | 18 |  | 18 |  | 18 |  | 18 |  |  |  |  |  |  |  |  |
| 1985 |  | 10 |  | 5 |  | 5 |  | 5 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1994 | 12 | 30 | 12 | 28 | 12 | 28 | 12 | 28 |  | 1 |  | 1 |  |  |  |  |  |  |  |  |
| 1995 | 13 | 36 | 13 | 36 | 13 | 36 | 13 | 36 |  | 1 |  | 1 |  | 1 |  | 1 |  | 1 |  | 1 |
| 1996 | 11 | 8 | 11 | 8 | 11 | 8 | 11 | 8 | 11 | 8 | 11 | 8 | 4 | 7 | 4 | 7 | 4 | 1 | 4 | 1 |
| 1997 | 35 | 83 | 35 | 82 | 35 | 82 | 35 | 82 | 12 | 82 | 12 | 82 | 12 | 1 | 12 | 1 | 12 | 1 | 12 | 1 |
| 1998 | 2 | 21 | 2 | 20 | 2 | 20 | 2 | 20 | 2 | 16 | 2 | 16 | 2 | 8 | 2 | 8 |  |  |  |  |
| 1999 |  | 36 |  | 35 |  | 35 |  | 35 |  | 8 |  | 8 |  |  |  |  |  |  |  |  |
| 2000 |  | 3 |  | 2 |  | 2 |  | 2 |  |  |  |  |  |  |  |  |  |  |  |  |
| 2001 |  | 8 |  | 8 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2002 |  | 25 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2005 |  | 3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2008 |  | 2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2012 |  | 3 |  | 3 |  | 3 |  | 3 |  | 3 |  | 3 |  | 3 |  | 3 |  |  |  |  |
| 2013 |  | 10 |  | 10 |  | 10 |  | 10 |  | 10 |  | 10 |  |  |  |  |  |  |  |  |
| 2014 |  | 22 |  | 22 |  | 22 |  | 22 |  |  |  |  |  |  |  |  |  |  |  |  |
| 2015 |  | 3 |  | 3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 133 | 306 | 135 | 284 | 134 | 272 | 134 | 272 | 85 | 149 | 85 | 149 | 63 | 31 | 62 | 31 | 26 | 12 | 26 | 12 |

## Otolith chemistry

We chemically analyzed the same otoliths from which we estimated age and seasonal growth since ratios of certain isotopes are known to reflect the environment in which the fish resides (Elsdon et al., 2008). Prior to their chemical analysis, otoliths mounted to microscope slides were cleaned by rinsing with $95 \%$ ethanol. We used laser ablation inductively coupled plasma mass spectrometry (LA-ICP-MS) at the GeoAnalytical Lab, Washington State University. Each laser ablation of an otolith obtained trace elemental concentrations from the otolith edge, which corresponds in space and time to their collection from a spawning ground, to the otolith core, which corresponds to hatching.

Data were collected as previously described in Chittaro et al. (2013). Specifically, isotope concentrations were determined using high resolution single collector inductively coupled plasma mass spectrometry (Finnigan Element2, with helium as the carrier gas) that was coupled with a laser ablation system (New Wave UP-213, frequency of $20-\mathrm{Hz}, 30-\mu \mathrm{m}$ spot size). Using an automated microscope stage, the laser beam was focused on the otolith, and a transect from edge to its core was ablated at a speed of $15 \mu \mathrm{~m} \mathrm{~s}$ s . Each transect was placed along the same axis used to estimate age and size, and formed an edge-to-core scan line corresponding to the entire life of the fish. Data acquisition of the LA-ICP-MS lasted $240 \mathrm{~s}, 20 \mathrm{~s}$ of which were designated for instrument calibration and gas background counts prior to the start of each ablation. To correct for instrument drift, we obtained a glass standard doped with trace elements from the National Institute of Standards and Technology (NIST 610). We analyzed this standard both at the beginning and end of each sample set (i.e., 16-20 otoliths). Calcium was used as an internal standard to compensate for signal variation caused by differences in the mass of ablated material. We calculated detection limits for each isotope as the average background plus three standard deviations.

From the laser ablation of each core to edge transect, six isotopes (magnesium- $25, \mathrm{Mg}$; calcium-43, Ca; manganese-55, Mn; strontium-86, Sr; barium-138, Ba; and lead-208, Pb ) were analyzed by ICP-MS. Counts per second were measured for each isotope along this transect, from which we calculated ratios of each isotope to calcium (e.g., $\mathrm{Mg}: \mathrm{Ca}, \mathrm{mm}$ $\mathrm{mol}^{-1}$ ) (Fig. 2B). Next, we calculated average isotope ratios for each period of summer and winter growth. Specifically, we used the otolith radius of each opaque and translucent zone (summer and winter growth, respectively) to designate the section of the core-to-edge transect for which to calculate the average for every isotope ratio (Fig. 2). When calculating the average isotopic ratio for the first summer of growth we excluded the yolk-sac stage because this stage has been shown in other species to be a product of maternal investment, and thus is not a site-specific signature (Brophy, Jeffries \& Danilowicz, 2004; Chittaro et al., 2006). The yolk-sac stage was excluded from our isotope ratios by omitting the chemical signature of the laser transect segment within $100-\mu \mathrm{m}$ of the otolith core. We were confident that this method excluded any maternal signal since yolk-sac absorption occurs 5-7 days after hatching (Bailey, 1982b) and 30-40 day old larvae are reported to have an otolith radius of $100-150 \mu \mathrm{~m}$ (Butler $\uplus$ Nishimoto, 1997).

Isotopes were included for statistical analyses if they met two criteria (Chittaro et al., 2013): concentrations of the isotope were greater than the detection limit in more than $80 \%$
of otoliths analyzed, and concentrations of the isotope in NIST samples were determined with satisfactory precision (coefficient of variation $<10 \%$ ).

## Generalized linear model

We used a generalized linear modeling (GLM) approach to investigate the extent to which variability in somatic growth (dependent variable) was explained by seven variables: year, harbor seal (Phoca vitulina) abundance, sea surface temperature, regional climate index, and several isotope ratios. This analysis was performed separately for each of three summers of growth, corresponding to the first, second, and third year of life. We used the $g l m$ function in the stats package of RStudio ( $R$ Core Team, 2021) and specified a gamma family distribution with a log link to account for the normally distributed, but positive, growth data.

We included year to account for interannual variation in marine environmental conditions not otherwise accounted for explicitly in the model. 'Year' corresponds to the year in which the summer growth occurred. For example, if a fish hatched in 1990 and we are investigating its second summer of growth, then the year associated with this growth would be 1992. Harbor seal abundance (Chasco et al., 2017) was included to account for the known predation pressure of harbor seals on Pacific hake (Gustafson et al., 2000). According to Saunders \& McFarlane (1999), SoG hake represented $42 \%$ of the diets in harbor seals. Because Pacific hake are ectothermic their metabolic demands are positively correlated to water temperature. We therefore included sea surface temperature, averaged from April to August, collected from a buoy located at Race Rocks on the southeastern end of Victoria Island, British Columbia, Canada. To account for regional climate patterns we included Pacific Northwest Index (PNI) (Ebbesmeyer \& Strickland 1995), which is a composite of location-specific air temperature, total precipitation, and snowpack data. Finally, because some aspects of water quality are recorded in the uptake of trace elements onto the growing otolith surface we used the chemical constituents of Pacific hake otoliths as a means to gain insight into the habitat where they resided during each summer. Specifically, we used average $\mathrm{Mn}: \mathrm{Ca}$ as an indicator of hypoxic conditions (Limburg e Casini, 2018; Limburg et al., 2011; Limburg et al., 2015; Mohan © Walther, 2016), average Ba:Ca as an indicator of nutrient rich upwelling (Bath et al., 2000; Mohan et al., 2018), and average Sr:Ca (Macdonald \& Crook, 2010; Martin, Thorrold \& Jones, 2004; Zimmerman, 2005) and Ba:Ca (Elsdon \& Gillanders, 2005; (Hamer, Jenkins \& Coutin, 2006; Miller, Gray $\leftrightarrow$ Merz, 2010) as an indicator of salinity. Lead was excluded from our analyses because little is understood about how it relates to water quality.

We ran GLMs only for periods of summer growth and separately for each of the first three years, and each population. We limited our analyses to summer growth because yearly changes in growth were more dramatic, relative to winter, (see results) and harbor seal abundance data were collected in spring and summer (Chasco et al., 2017). The most complex model was describes as:
$G \sim \mathrm{Yr}+\mathrm{Seal}+\mathrm{Temp}+\mathrm{PNI}+\mathrm{Mn}+\mathrm{Sr}+\mathrm{Ba}$
where $G$ is the summer growth estimated from otoliths, $Y r$ is the year the summer growth occurred, Seal is harbor seal abundance, Temp is the sea surface temperature, PNI is the

Pacific Northwest Index, and $\mathrm{Mn}, \mathrm{Sr}$, and Ba are the average $\mathrm{Mn}: \mathrm{Ca}, \mathrm{Sr}: \mathrm{Ca}$, and $\mathrm{Ba}: \mathrm{Ca}$, respectively, corresponding to the summer of otolith growth (e.g., first summer of growth).

All possible GLM model combinations $(n=127)$ were run and model parameters were estimated by maximizing the likelihood function. To compare models we calculated four values for each model; Akaike's Information Criterion (AIC), delta AIC, relative likelihood, and AIC weight. Smaller AIC values indicate "better" models and when comparing two models we calculated the difference in AIC values (delta AIC) (Akaike, 1973; Burnham \& Anderson, 2002). A delta AIC of less than 2 indicates little difference between competing models; a delta AIC of 2-10 indicates moderate support for a difference between the models, and a delta AIC of greater than 10 indicates strong support (Burnham \& Anderson, 2002). Relative likelihood represents the probability of a model given the data, whereas AIC weight is the discrete probability of each model (Burnham \& Anderson, 2002).

The best model from each dataset was defined as having a delta AIC of 0.00 , although preference was given to the model with the fewest variables if two or more models had a delta AIC of less than 2. This process was repeated for all six data sets: Puget Sound first, second, and third summer of growth and Strait of Georgia first, second, and third summer of growth.

## RESULTS

## Age, length, and growth

We estimated the age of 439 Pacific hake collected from PS and SoG in 1979-2016 (Table 1). These fish ranged in age from young-of-the-year to 6 , with 2 year-olds being most abundant and comprising $38 \%$ of all individuals (Fig. 3A). Based on these ages and when fish were collected, we estimated hatch year to range from 1974 to 2015 (Table 2). Median total length ( mm ) at capture was 243 mm and 410 mm for PS and SoG, respectively, with SoG fish showing a bi-modal size distribution (Fig. 3B).

Our reconstruction of Pacific hake length at each summer and winter together with our repeated measures ANOVA, across all hatch years, indicated a significant interaction between population and year/season of growth ( $F=339.7$; $d f=1,1950 ; p=0.00022$ ). By the second summer, individuals from SoG were, on average, $18 \%$ larger than fish from PS (mean size of 265 mm and 215 mm , and standard deviation of 42 mm and 28 mm , respectively) (Fig. 4A). The repeated measures ANOVA comparing estimates of length per year/season of SoG fish did not reveal significant differences between those that hatched in the 1970s versus 1990s (Fig. 4B). However, the same analysis performed on PS fish revealed a significant interaction between decade in which they hatched and year/season $(F=40.6$; $d f=3,1201 ; p=0.00022$ ). Tukey's post hoc test showed significant differences between fish that hatched in the 2010s to those that hatched in both 1980s and 1990s (Fig. 4C).

Repeated measures ANOVA of our growth estimates, across all hatch years, indicated a significant interaction between population and year/season $(F=45.4 ; d f=1,1950$; $p=0.00015)$. By the second summer individuals from SoG grew, on average, $42 \%$ more than fish from PS (mean growth 120 mm and 69 mm , and standard deviation of 28 mm and 31 mm , respectively) (Fig. 5A). The repeated measures ANOVA comparing estimates


Figure 3 Pacific hake age and size. Frequency histogram of Pacific hake (A) age and (B) total length $(\mathrm{mm})$ at capture from Puget Sound (white) and Strait of Georgia (black) populations. Grey color corresponds to overlap between populations.

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of growth per year/season of SoG fish did not reveal significant differences between those that hatched in 1970s versus 1990s (Fig. 5B). When we performed the same analysis on fish from PS it revealed a significant interaction $(F=15.7 ; d f=3,1201 ; p=0.00031)$ between decade in which they hatched and year/season. Tukey's post-hoc test showed significant differences among fish that hatched in the 1980s, 1990s, and 2010s, as well as between those that hatched in the 1980s and 2000s (Fig. 5C). The first and second summers showed the most dramatic increases and decreases, respectively, in growth across the decades in which fish hatched. Specifically, in their first summer fish that hatched in the 2010s grew, on average, 147 mm compared to 108 mm for those that hatched in the 1980s. However, fish that hatched in the 2010s grew 31 mm in their second summer while those that hatched in 1980s grew 107 mm (Fig. 5C).

## Generalized linear model

We obtained elemental concentrations from 214 individuals collected in PS and 132 from SoG. All isotopes analyzed were found in concentrations above the detection limit for $100 \%$ of the otoliths sampled. We observed an acceptable level of analytical precision based on NIST ( $n=87$ ) mean coefficient of variation and standard deviation: Mg:Ca (mean = 2.35 , standard deviation $=1.44$ ), $\mathrm{Mn}: \mathrm{Ca}($ mean $=2.42$, standard deviation $=1.09), \mathrm{Sr}: \mathrm{Ca}$


Figure 4 Otolith-derived estimates of total length at each year and season for Pacific hake collected from Strait of Georgia and Puget Sound. Otolith-derived estimates of total length ( mm ) at each year and season (i.e., summer and winter) for Pacific hake collected from (A) Strait of Georgia and Puget Sound with pooled hatch years, and (B) Strait of Georgia and (C) Puget Sound fish grouped by the decade in which they hatched. Repeated measures ANOVA indicated significant length differences between fish from Puget Sound and Strait of Georgia ( $F=303.2 ; d f=1,1952 ; p=0.00022$ ). Within populations, repeated measures ANOVA revealed only a significant interaction between decade and year/season for Puget Sound fish ( $F=39.6$; $d f=3,1204 ; p=0.00022$ ). In plot $C$, lower case letters in the legend correspond to results from the Tukey's post hoc test that revealed significant differences (at $p<0.05$; indicated as different letters) between fish that hatched in the 1990's and 2010's. Black horizontal line, box, and whiskers represent median, first and third quartile (i.e., 25th and 75th percentiles), and minimum and maximum values, respectively. To improve visualization of the length differences among decades for Puget Sound fish (C), the maximum value on the $y$-axis was reduced ( 500 mm ) relative to the other plots ( 600 mm ).

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Figure 5 Otolith-derived estimates of somatic growth at each year and season for Pacific hake collected from Strait of Georgia and Puget Sound. Otolith-derived estimates of somatic growth ( mm ) at each year and season (i.e., summer and winter) for Pacific hake collected from (A) Strait of Georgia and Puget Sound with pooled hatch years, and (B) Strait of Georgia and (C) Puget Sound fish grouped by the decade in which they hatched. Repeated measures ANOVA indicated significant growth differences ( $F=$ $35.9 ; d f=1,1952 ; p=0.00026$ ) between fish from Puget Sound and Strait of Georgia. Within populations, repeated measures ANOVA indicated a significant interaction only between decade and year/season for Puget Sound fish ( $F=16.9$; $d f=3,1204 ; p=0.00054$ ). In plot $C$, lower case letters in the legend correspond to results from the Tukey's post hoc test that revealed significant pairwise differences (at $p<0.05$; indicated as different letters) between fish that hatched among certain decades. Black horizontal line, box, and whiskers represent median, first and third quartile (i.e., 25th and 75th percentiles), and minimum and maximum values, respectively.

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$($ mean $=1.65$, standard deviation $=0.52)$, and $\mathrm{Ba}: \mathrm{Ca}($ mean $=2.29$, standard deviation $=$ 1.13).

We investigated variability in somatic growth separately for each population owing to the significant differences in growth and relatively little overlap in hatch years between populations. Overall, our GLM approach to understand what factors explain variability in somatic growth rate revealed a different suite of variables, more or less, with each summer and population. Further, AIC weights were relatively low $(<0.3)$ indicating that a substantial amount of variability in growth was unexplained.

For SoG fish, results of the GLM analysis indicated five models that explained the greatest amount of variability in the first summer of somatic growth and were indistinguishable (i.e., delta AIC < 2.0) (Table 3). Of these five models the simplest model was chosen as the best model and it showed a negative relationship between somatic growth and $\mathrm{Mn}: \mathrm{Ca}$ and $\mathrm{Sr}: \mathrm{Ca}$, and a positive relationship between somatic growth and $\mathrm{Ba}: \mathrm{Ca}$, 'Harbor seal abundance', and 'Sea surface temperature' (Figs. 6A-6E). Our GLM analysis of the second summer of growth revealed nine models that best explained variability and were indistinguishable (Table 3). Of these nine models, the one with the lowest AIC was also the simplest model. This model showed positive and negative relationships between somatic growth and $\mathrm{Mn}: \mathrm{Ca}$ and $\mathrm{Sr}: \mathrm{Ca}$, respectively (Figs. 6F-6G). Finally, the GLM analysis of the third summer of growth showed that nine models best explained variability and were indistinguishable (Table 3). Of these nine models, three each consisted of only one variable: 'Sea surface temperature', 'Harbor seal abundance', and 'Year'. Each of these three models showed a negative relationship to somatic growth (Figs. 6H-6J).

With respect to PS fish, the GLM analysis of the first summer of growth showed that three models best explained variability in somatic growth and were indistinguishable (Table 3). Of these, the best was also the simplest. This model indicated positive relationships between somatic growth and year, $\mathrm{Mn}: \mathrm{Ca}, \mathrm{Ba}: \mathrm{Ca}$, and 'Harbor seal abundance', and negative relationships with $\mathrm{Sr}: \mathrm{Ca}$ (Figs. 7A-7E). Results of the GLM analysis of the second summer of growth revealed that nine models best explained variability in somatic growth and were indistinguishable (Table 3). Of these, the best was also the simplest, in which a positive relationship was observed between somatic growth and $\mathrm{Mn}: \mathrm{Ca}$ and a negative relationship to Sr:Ca and 'Harbor seal abundance' (Figs. 7F-7H). Lastly, the GLM analysis of the third summer of growth showed that seven models best explained variability and were indistinguishable (Table 3). Of these, two were the simplest model, both consisted of four variables, and had $\mathrm{Mn}: \mathrm{Ca}, \mathrm{Ba}: \mathrm{Ca}$, and 'Harbor seal abundance' in common (Table 3). Both models showed a positive relationship between growth and $\mathrm{Mn}: \mathrm{Ca}$, and negative relationships with Ba:Ca and 'Harbor seal abundance' (Figs. 7I-7L). In one model there was a negative relationship between growth and year, and a lack of a relationship with PNI.

## DISCUSSION

The ability of a population to be sustainably managed depends upon its rate of harvest and recovery. With respect to Pacific hake in the Georgia Basin, there has been two

Table 3 Results of the generalized linear modeling approach that assessed what variables explained variability in somatic growth within each of the first three summers, and within each population. Relative likelihood (Like) is the likelihood of a model given the data, and AIC weight (AIC weight) is the discrete probability of each model. Only models that are indistinguishable (i.e., delta AIC (delta AIC) of $\leq 2.0$ ) are displayed. Best models (X) are identified as models with the lowest delta AIC and with fewest variables. Population (Pop) corresponds to Strait of Georgia (SoG) and Puget Sound (PS).

| Year | Pop | Model | Best | AIC | dAIC | Like | AICwt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | SoG | Year + Mn:Ca + Sr:Ca + Ba:Ca + Temp + PNI |  | 1292.9 | 0.00 | 1.00 | 0.20 |
| 1 | SoG | $\mathrm{Mn}: \mathrm{Ca}+\mathrm{Sr}: \mathrm{Ca}+\mathrm{Ba}: \mathrm{Ca}+$ Seals + Temp + PNI |  | 1293.0 | 0.11 | 0.95 | 0.19 |
| 1 | SoG | $\mathrm{Mn}: \mathrm{Ca}+\mathrm{Sr}: \mathrm{Ca}+\mathrm{Ba}: \mathrm{Ca}+$ Seals + Temp | X | 1294.5 | 1.66 | 0.44 | 0.09 |
| 1 | SoG | Year + Mn:Ca + Sr:Ca + Ba:Ca + Seals + Temp |  | 1294.7 | 1.80 | 0.41 | 0.08 |
| 1 | SoG | Year + Mn:Ca + Sr:Ca + Ba:Ca + Seals + Temp + PNI |  | 1294.9 | 2.00 | 0.37 | 0.08 |
| 2 | SoG | $\mathrm{Mn}: \mathrm{Ca}+\mathrm{Sr}: \mathrm{Ca}$ | X | 1263.3 | 0.00 | 1.00 | 0.13 |
| 2 | SoG | Year + Mn:Ca + Sr:Ca + PNI |  | 1264.4 | 1.10 | 0.58 | 0.07 |
| 2 | SoG | $\mathrm{Mn}: \mathrm{Ca}+\mathrm{Sr}: \mathrm{Ca}+\mathrm{Ba}: \mathrm{Ca}$ |  | 1264.7 | 1.40 | 0.50 | 0.06 |
| 2 | SoG | $\mathrm{Mn}: \mathrm{Ca}+\mathrm{Sr}: \mathrm{Ca}+\mathrm{Temp}$ |  | 1265.1 | 1.87 | 0.39 | 0.05 |
| 2 | SoG | Year + Mn:Ca + Sr:Ca |  | 1265.2 | 1.94 | 0.38 | 0.05 |
| 2 | SoG | $\mathrm{Mn}: \mathrm{Ca}+\mathrm{Sr}: \mathrm{Ca}+$ Seals |  | 1265.2 | 1.96 | 0.38 | 0.05 |
| 3 | SoG | Temp | X | 815.7 | 0.00 | 1.00 | 0.06 |
| 3 | SoG | Temp + PNI |  | 816.8 | 1.06 | 0.59 | 0.04 |
| 3 | SoG | Ba:Ca + Temp |  | 817.1 | 1.39 | 0.50 | 0.03 |
| 3 | SoG | Seals + Temp |  | 817.2 | 1.52 | 0.47 | 0.03 |
| 3 | SoG | Seals | X | 817.3 | 1.60 | 0.45 | 0.03 |
| 3 | SoG | Year + Temp |  | 817.3 | 1.60 | 0.45 | 0.03 |
| 3 | SoG | Year | X | 817.4 | 1.66 | 0.44 | 0.03 |
| 3 | SoG | Sr:Ca + Temp |  | 817.4 | 1.70 | 0.43 | 0.03 |
| 3 | SoG | $\mathrm{Mn}: \mathrm{Ca}+$ Temp |  | 817.6 | 1.87 | 0.39 | 0.02 |
| 1 | PS | Year + Mn:Ca + Sr:Ca + Ba:Ca + Seals | X | 1946.1 | 0.00 | 1.00 | 0.30 |
| 1 | PS | Year + Mn:Ca + Sr:Ca + Ba:Ca + Seals + Temp |  | 1947.2 | 1.12 | 0.57 | 0.17 |
| 1 | PS | Year + Mn:Ca + Sr:Ca + Ba:Ca + Seals + PNI |  | 1947.9 | 1.82 | 0.40 | 0.12 |
| 2 | PS | $\mathrm{Mn}: \mathrm{Ca}+\mathrm{Sr}: \mathrm{Ca}+$ Seals + Temp + PNI |  | 1954.6 | 0.00 | 1.00 | 0.14 |
| 2 | PS | $\mathrm{Mn}: \mathrm{Ca}+\mathrm{Sr}: \mathrm{Ca}+$ Seals | X | 1955.4 | 0.80 | 0.67 | 0.09 |
| 2 | PS | Year + Mn:Ca + Sr:Ca + Seals + Temp + PNI |  | 1955.6 | 0.99 | 0.61 | 0.09 |
| 2 | PS | $\mathrm{Mn}: \mathrm{Ca}+\mathrm{Sr}: \mathrm{Ca}+$ Seals + Temp |  | 1955.8 | 1.20 | 0.55 | 0.08 |
| 2 | PS | $\mathrm{Mn}: \mathrm{Ca}+\mathrm{Sr}: \mathrm{Ca}+\mathrm{Ba}: \mathrm{Ca}+$ Seals + Temp + PNI |  | 1956.0 | 1.37 | 0.50 | 0.07 |
| 2 | PS | Year + Mn:Ca + Sr:Ca + Seals |  | 1956.4 | 1.75 | 0.42 | 0.06 |
| 3 | PS | Year + Mn:Ca + Ba:Ca + Seals | X | 890.5 | 0.00 | 1.00 | 0.14 |
| 3 | PS | Year + Mn:Ca + Sr:Ca + Ba:Ca + Seals |  | 892.1 | 1.65 | 0.44 | 0.06 |
| 3 | PS | Year + Mn:Ca + Ba:Ca + Seals + PNI |  | 892.4 | 1.93 | 0.38 | 0.05 |
| 3 | PS | $\mathrm{Mn}: \mathrm{Ca}+\mathrm{Ba}: \mathrm{Ca}+$ Seals + PNI | X | 892.4 | 1.96 | 0.38 | 0.05 |
| 3 | PS | Year + Mn:Ca + Ba:Ca + Seals + Temp |  | 892.5 | 1.98 | 0.37 | 0.05 |

different responses to harvest with one population failing to recover despite decades of a closed fishery. In this study we evaluated what factors and processes may be involved in preventing Pacific hake recovery by examining otolith-derived estimates of somatic growth across several decades. We observed dramatic changes in seasonal growth and size-at-age across sampling years, especially for fish sampled from PS.


Figure 6 Model fits of Strait of Georgia somatic growth (mm) in the first (A-E), second (F-G), and third ( $\mathrm{H}-\mathrm{J}$ ) summers with respect to predictor variables that best explained variability in somatic growth (Table 3). $\mathrm{Mn}: \mathrm{Ca}, \mathrm{Sr}: \mathrm{Ca}$, and $\mathrm{Ba}: \mathrm{Ca}$ are average values that correspond to the summer of otolith growth for the corresponding year (e.g., first summer of growth). 95 percent confidence intervals of the predicted values are represented as the dash lines.

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## Size-at-age and growth differences between populations

Our findings of size-at-age differences between PS and SoG populations (Fig. 4A) were expected given that the listing of the Georgia Basin DPS as a Species of Concern was due, in part, to the reduced size-at-age of PS fish observed through the 1990s (Gustafson et al., 2000). For example, Pedersen (1985) and Goñi (1988) reported that PS fish collected in the mid-1960s to early 1980s grew approximately $2-5 \mathrm{~cm}$ shorter than those from SoG. What our study also revealed is that this divergence in body size between fish from PS and SoG occurred during their second summer of life (Fig. 4A). Size differences further increased by their third summer when PS fish grew, on average, $73 \%$ less than fish from SoG (i.e.,






$\mathrm{Ba}: \mathrm{Ca}$
$(\mathrm{mm} / \mathrm{mol})$


Figure 7 Model fits of Puget Sound somatic growth (mm) in the first (A-E), second (F-H), and third (I-L) summers with respect to predictor variables that best explained variability in somatic growth (Table 3). $\mathrm{Mn}: \mathrm{Ca}, \mathrm{Sr}: \mathrm{Ca}$, and Ba:Ca are average values that correspond to the summer of otolith growth for the corresponding year (e.g., first summer of growth). 95 percent confidence intervals of the predicted values are represented as the dash lines.

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23.2 mm compared to 88.7 mm , respectively; Fig. 5A). It should be noted, however, the limited temporal overlap of fish that hatched in PS and SoG. Specifically, only the hatch years from 1994 through 1998 included individuals that were sampled from both PS and SoG (Table 2).

From 1965-1974 most of the PS fish ranged in size from 32 cm to 45 cm (average length of 36.2 cm , with recorded maximum lengths of 45 cm and 73 cm for males and females, respectively) (Pedersen, 1985), yet since 2008, less than $5 \%$ of sampled PS fish ( $n=11,652$ ) had a length greater than 30 cm (Blaine, Lowry \& Pacunski, 2020) and Supplemental Table 1). Pedersen (1985) suggested that the declining body size in PS through time, and
the fact that PS fish mature at a smaller size, relative to SoG fish, may have been due to the intense commercial fishery in PS in the 1970s and 1980s. Specifically, the extensive removal of large individuals from a population has been shown to select for individuals that mature earlier and grow less (Allendorf \& Hard, 2009; Olsen et al., 2004; Sharpe \& Hendry, 2009). For instance, California sheephead (Semicossyphus pulcher) had significantly smaller body size, earlier maturation, and a reduced maximum lifespan following more than 20 years of high fishing pressure (Hamilton et al., 2007). Similarly, zebrafish (Danio rerio) invested more in reproduction and attained a smaller adult body size when exposed to intense sizeselective harvesting (Uusi-Heikkilä et al., 2015). Interestingly, models simulating zebrafish population dynamics indicated a slow population recovery even after a moratorium on harvest. Uusi-Heikkilä et al. (2015) suggested that this lack of population recovery was linked to changes in body size and reproductive investment, which placed the population at a disadvantage to manage natural selection pressures that often favor large body size. Similar harvest-related changes to life-history could explain why the smaller body size and the lack of recovery of the PS population persists today despite the closure of the fishery for thirty years.

## Size-at-age and growth differences through time and within populations

We observed significant differences in size-at-age of PS fish such that a shift was seen towards smaller fish through time (Fig. 4C), yet no differences in size-at-age were detected between SoG fish that hatched in the 1970s and 1990s (Fig. 4B). In terms of somatic growth, we found a similar pattern in both populations whereby in recent years fish tended to grow more in their first summer and less in subsequent summers. Specifically, SoG fish that hatched in the 1990s grew, on average, $11 \%$ more in the first summer, and $25 \%$ less in the third summer, relative to those that hatched in the 1970s (Fig. 5B). In comparison, PS fish that hatched in 2010s grew $26 \%$ more in their first summer than those that hatched in 1980s, yet $71 \%$ less in the second summer compared to those that hatched in 1980s (Fig. 5C). Beamish \& McFarlane (1999) suggest that this increased somatic growth during the first summer is linked to a shift in climate and ocean conditions that occurred in the late 1980 's. This regime shift is hypothesized to have caused a greater temporal overlap between young-of-year hake and an important prey item (e.g., copepods) in SoG (King e McFarlane, 2006). Because PS and SoG populations are part of the same basin and share similar characteristics it is possible that the same processes played a role in the size and growth patterns in SoG would be seen in PS.

Our observed decrease in growth during the third summer of SoG fish is consistent with findings by King $\&$ McFarlane (2006) whereby an increase in hake abundance following the aforementioned regime shift likely lead to reduced somatic growth later in life because of increased density dependent competition. The third year of growth is an important period in hake development in which they become reproductively mature and are suspected to move deeper to feed (Saunders $\&$ McFarlane, 1999). Therefore, if intraspecific competition is high during the time when surplus energy is allocated to reproductive development then it is expected that reduced energy will be available for growth. Regardless of the
mechanism(s), the declines in growth through time, in both populations, highlights increased concern for the sustainability of this Species of Concern.

Our investigation of what factors and/or processes explain variability in seasonal growth indicated some similarities between populations. For instance, growth in the first summer was best explained by models that included positive relationships to harbor seal abundance and to barium levels (i.e., Ba:Ca) (Table 4). The positive growth associated with harbor seal abundance could indicate that hake resources are density-mediated and therefore predation-induced reductions in hake abundance may have led to reductions in intraspecific competition (Heithaus et al., 2008). Ba:Ca is positively associated with areas of nutrient rich upwelling (Mohan et al., 2018; Wheeler et al., 2016) and freshwater input (Elsdon \& Gillanders, 2005; Hamer, Jenkins \& Coutin, 2006; Miller, Gray \& Merz, 2010), and the positive relationship between Ba:Ca and growth (Figs. 6C and 7D) might be due to increases in productivity, which could have increased prey (e.g., copepods) abundance due to nutrient enriched waters brought to the surface and/or from riverine sources. For instance, Takahashi et al. (2012) reported a positive relationship between otolith growth of northern anchovy (Engraulis mordax) and upwelling intensity, which they attributed to increased productivity and in turn improved prey nutritional value. Strontium (i.e., Sr:Ca) in otoliths is reported to be positively associated with salinity (Macdonald \& Crook, 2010; Martin \& Thorrold, 2005; Zimmerman, 2005). Thus the negative relationship we observed in both populations between growth (in the first and second summers) and $\mathrm{Sr}: \mathrm{Ca}$ (Table 4; Figs. 6B, 6G, 7C, and 7G), and the aforementioned relationship to Ba:Ca, suggests that the first two years of growth are sensitive to freshwater flow, such that growth declines with a reduction in freshwater. Main spawning aggregations for both PS and SoG populations occur in close proximity to major sources of freshwater (e.g., Stillaguamish and Snohomish rivers in PS and Fraser River in SoG). This freshwater input would cause a stratified layer of low-salinity water above the well-mixed marine layer. This is significant since Pacific hake larvae are known to aggregate in surface waters (Gustafson et al., 2000).

The third summer of growth for fish from both populations was negatively associated with harbor seal abundance, which is an opposite pattern observed for the first summer of growth. These positive and negative relationships may suggest that younger hake experience a competitive release in years with more harbor seals, while older hake experience reduced performance possibly due to energy expenditure through predatory avoidance. Heithaus et al. (2008) noted that aside from direct mortality, predators can strongly influence prey behavior such as foraging. It is likely that hake recovery, particularly in PS, is limited given that harbor seal abundances increased by 7\% (Saunders e McFarlane, 1999) and 3\% (Gustafson et al., 2000) in SoG and PS, respectively, in the early to mid-1990s and appear to be stable in both populations (Jeffries et al., 2003; Olesiuk, 1999).

## CONCLUSION

Declines in fish body size have been reported in many commercial stocks (Ohlberger et al., 2018; Sharpe \& Hendry, 2009) and these changes likely have important ramifications for the sustainability of the harvested species (Allendorf \& Hard, 2009;

Table 4 Results of the GLM analysis that identified independent variables that best explained variability in summer growth. Positive ( + ) and negative (-) relationships, and lack of a relationship ( $\phi$ ) are indicated for each population (Strait of Georgia, SoG, and Puget Sound, PS) with respect to each of three summers. The relationships indicated in the first and second summers correspond to the variables found in the model that best explains variability in growth for both populations (see Table 3). For the third summer for SoG, each of the three variables represents its own model. For the third summer for PS, two models each of which consisted of four variables and had three variables in common ( $\mathrm{Mn}: \mathrm{Ca}, \mathrm{Ba}: \mathrm{Ca}$, and Seals) and differed in the fourth variable (either PNI or Year).

| Variable | 1st summer |  | 2nd summer |  | 3rd summer |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SoG | PS | SoG | PS | SoG | PS |
| $\mathrm{Mn}: \mathrm{Ca}$ | - | + | + | + |  | + |
| Sr:Ca | - | - | - | - |  |  |
| Ba:Ca | + | + |  |  |  | - |
| Seals | + | + |  | - | - | - |
| Year |  | + |  |  | - | - |
| SST | + |  |  |  | - |  |
| PNI |  |  |  |  |  | $\varnothing$ |

Losee, Kendall \& Dufault, 2019) as well as ecosystem function and services (Oke et al., 2020). The Pacific hake Georgia Basin DPS are prey to top predators and are important to a healthy and functioning ecosystem (Gustafson et al., 2000). Unfortunately, the Georgia Basin DPS has experienced substantial changes in juvenile somatic growth over the last several decades, despite fishery regulations, which highlights increased concern for the sustainability of this Species of Concern. Further, our investigation of juvenile somatic growth indicated that a diverse suite of processes (e.g., competition and predation) and factors (e.g., salinity and upwelling) were important for explaining variability and that these relationships differed with respect to population and age. These results highlight the complicated nature of population recovery under dynamic, and typically uncontrollable, variation in biotic and abiotic conditions.

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## ADDITIONAL INFORMATION AND DECLARATIONS

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## Competing Interests

The authors declare there are no competing interests.

## Author Contributions

- Paul Chittaro conceived and designed the experiments, performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the article, and approved the final draft.
- Chris Grandin performed the experiments, authored or reviewed drafts of the article, collected and aged fish, and approved the final draft.
- Robert Pacunski performed the experiments, authored or reviewed drafts of the article, collected and aged fish, and approved the final draft.
- Rich Zabel analyzed the data, authored or reviewed drafts of the article, and approved the final draft.


## Animal Ethics

The following information was supplied relating to ethical approvals (i.e., approving body and any reference numbers):

Animal ethics and care approval was not required for this study because otoliths were collected as part of other studies.

## Field Study Permissions

The following information was supplied relating to field study approvals (i.e., approving body and any reference numbers):

Field work was conducted by Washington Department of Fish and Wildlife and Department of Fisheries and Ocean Canada.

## Data Availability

The following information was supplied regarding data availability:
The raw data measurements of Pacific hake are available in the Supplementary File.

## Supplemental Information

Supplemental information for this article can be found online at http://dx.doi.org/10.7717/ peerj.13577\#supplemental-information.

## REFERENCES

Akaike H. 1973. Information theory as an extension of the maximum likelihood principle. In: Petrov BN, Csaki F, eds. Second international symposium on information theory. Akademiai Kiado, Budapest, Hungary. 267-281.
Allendorf FW, Hard JJ. 2009. Human-induced evolution caused by unnatural selection through harvest of wild animals. Proceedings of the National Academy of Sciences of the United States of America 106:9987-9994 DOI 10.1073/pnas. 0901069106.
Bailey KM. 1982a. The early life history of the Pacific hake, Merluccius Productus. Fishery Bulletin 80(3):589-598.
Bailey KM. 1982b. The early life-history of the Pacific hake, Merluccius Productus. Fishery Bulletin 80:589-598.

Bath GE, Thorrold SR, Jones CM, Campana SE, McLaren JW, Lam JWH. 2000. Strontium and barium uptake in aragonitic otoliths of marine fish. Geochimica et Cosmochimica Acta 64:1705-1714 DOI 10.1016/S0016-7037(99)00419-6.
Baumann H, Hinrichsen H-H, Voss R, Stepputtis D, Grygiel W, Clausen LW, Temming A. 2006. Linking growth to environmental histories in central Baltic young-of-the-year sprat, Sprattus Sprattus: an approach based on otolith microstructure analysis and hydrodynamic modelling. Fisheries Oceanography 15:465-476 DOI 10.1111/j.1365-2419.2005.00395.x.
Baumann H, Pepin P, Davidson FJM, Mowbray F, Schnack D, Dower JF. 2003. Reconstruction of environmental histories to investigate patterns of larval radiated shanny (Ulvaria subbifurcata) growth and selective survival in a large bay of Newfoundland. Ices Journal of Marine Science 60:243-258 DOI 10.1016/S1054-3139(03)00019-5.
Beamish RJ. 1979. Differences in the age of Pacific hake (Merluccius productus) using whole otoliths and sections of otoliths. Journal of the Fisheries Research Board of Canada 36:141-151 DOI 10.1139/f79-023.
Beamish RJ, McFarlane GA. 1999. Applying ecosystem management to fisheries in the Strait of Georgia. In: Ecosystem approaches for fisheries management. Sea Grant, AK-SG-99-01. Fairbanks: University of Alaska, Fairbanks, Alaska, 637-664.
Beamish RJ, Smith M, Scarsbrook R. 1978. Hake and pollock study, Strait of Georgia Cruise, RV G.B. Reed January 6-February 21, 1975. Data Report 48. Naniamo: Department of Fisheries and Environment, Fisheries and Marine Services.
Berger AM, Edwards AM, Grandin CJ, Johnson KF. 2019. Status of the Pacific Hake (whiting) stock in U.S. and Canadian waters in 2019. Prepared by the joint technical committee of the U.S. and Canada Pacific Hake/Whiting Agreement, National Marine Fisheries Service and Fisheries and Oceans Canada. 249.
Blaine J, Lowry D, Pacunski R. 2020. 2002-2007 WDFW scientific bottom trawl surveys in the southern Salish Sea: species distributions, abundance, and population trends. Washington Department of Fish and Wildlife, Olympia, WA 252.
Brophy D, Jeffries TE, Danilowicz BS. 2004. Elevated manganese concentrations at the cores of clupeid otoliths: possible environmental, physiological, or structural origins. Marine Biology 144:779-786 DOI 10.1007/s00227-003-1240-3.
Burger M, Sandell T, Fanshier C, Lindquist A, Biondo P, Lowry D. 2019. 2016-17 Puget Sound Acoustic Mid-water Trawl Survey. Washington Department of Fish and Wildlife. Olympia, WA. FPT 20-40 47.
Burnham KP, Anderson DR. 2002. Model selection and multimodel inference: a practical information-theoretic approach. New York: Springer-Verlag.
Butler JL, Nishimoto RN. 1997. Growth and cohort dynamics of larval Pacific hake (Merluccius productus). Reports of California Cooperative Oceanic Fisheries Investigations 38:63-68.
Campana S. 2011. How reliable are growth back-calculations based on otoliths? Canadian Journal of Fisheries and Aquatic Sciences 47:2219-2227 DOI 10.1139/f90-246.

Campana SE, Jones CM. 1992. Analysis of otolith microstructure data. In Otolith microstructure examination and analysis. Edited by D.K. Stevenson and S.E. Campana. Canadian Special Publication of Fisheries and Aquatic Sciences 117:73-100.
Casselman JM. 1987. Determination of age and growth. In: Weatherley AH, Gill HS, eds. The biology of fish growth. London: Academic Press, 209-242.
Chasco B, Kaplan IC, Thomas A, Acevedo-Gutiérrez A, Noren D, Ford MJ, Hanson MB, Scordino J, Jeffries S, Pearson S, Marshall KN, Ward EJ. 2017. Estimates of Chinook salmon consumption in Washington State inland waters by four marine mammal predators from 1970 to 2015. Canadian Journal of Fisheries and Aquatic Sciences 74:1173-1194 DOI 10.1139/cjfas-2016-0203.
Chilton DE, Beamish RJ. 1982. Age determination methods for fishes studied by the groundfish program at the Pacific Biological Station. Canadian Special Publication of Fisheries and Aquatic Sciences 60:102-101.
Chittaro PM, Hogan JD, Gagnon J, Fryer BJ, Sale PF. 2006. In situ experiment of ontogenetic variability in the otolith chemistry of Stegastes partitus. Marine Biology 149:1227-1235 DOI 10.1007/s00227-006-0280-x.
Chittaro P, Zabel R, Palsson W, Grandin C. 2013. Population interconnectivity and implications for recovery of a species of concern, the Pacific hake of Georgia Basin. Marine Biology 160:1157-1170 DOI 10.1007/s00227-013-2168-x.
Ebbesmeyer CC, Strickland RM. 1995. Oyster Condition and Climate: evidence from Willapa Bay. Publication WSG-MR 95-02. Washington Sea Grant Program, University of Washington, Seattle, WA 11.
Elsdon TS, Gillanders BM. 2005. Alternative life-history patterns of estuarine fish: barium in otoliths elucidates freshwater residency. Canadian Journal of Fisheries and Aquatic Sciences 62:1143-1152 DOI 10.1139/f05-029.
Elsdon TS, Wells BK, Campana SE, Gillanders BM, Jones CM, Limburg KE, Secor DH, Thorrold SR, Walther BD. 2008. Otolith chemistry to describe movements and lifehistory parameters of fishes: hypotheses, assumptions, limitations and inferences. Oceanography and Marine Biology: An Annual Review 46:297-330.
Goñi R. 1988. Comparison of Pacific hake (Merluccius productus Ayres, 1855) stocks in inshore waters of the Pacific Ocean: puget Sound and Strait of Georgia. Master's Thesis, University of Washington, Seattle 104.
Graeb BDS, Dettmers JM, Wahl DH, Caceres CE. 2004. Fish size and prey availability affect growth, survival, prey selection, and foraging behavior of larval yellow perch. Transactions of the American Fisheries Society 133:504-514 DOI 10.1577/T03-050.1.
Grandin C, Johnson K, Edwards A, Berger A. 2020. Status of the Pacific Hake (whiting) stock in U.S. and Canadian waters in 2020. Prepared by the Joint Technical Committee of the U.S. and Canada Pacific hake/whiting agreement National Marine Fisheries Service and Fisheries and Oceans Canada. 273. Available at https://www. fisheries. noaa.gov/resource/document/2020-pacific-hake-whiting-stock-assessment.
Gustafson RG, Lenarz WH, McCain BB, Schmitt CC, Grant WS, Builder TL, Methot RD. 2000. Status review of Pacific Hake, Pacific Cod, and Walleye Pollock from

Puget Sound, Washington. U.S. Dept. Commer., NOAA Tech. Memo. NMFS-NWFSC- 44. 275.
Hamer PA, Jenkins GP, Coutin P. 2006. Barium variation in Pagrus auratus (Sparidae) otoliths: a potential indicator of migration between an embayment and ocean waters in south-eastern Australia: ecologiocal and management implications of seagrass landscapes. Estuarine, Coastal and Shelf Science 68:686-702 DOI 10.1016/j.ecss.2006.03.017.
Hamilton SL, Caselle JE, Standish JD, Schroeder DM, Love MS, Rosales-Casian JA, Sosa-Nishizaki O. 2007. Size-selective harvesting alters life histories of a temperate sex-changing fish. Ecological Applications 17:2268-2280 DOI 10.1890/06-1930.1.
Heithaus MR, Frid A, Wirsing AJ, Worm B. 2008. Predicting ecological consequences of marine top predator declines. Trends in Ecology \& Evolution 23:202-210 DOI 10.1016/j.tree.2008.01.003.
Helser TE, Almeida FP. 1997. Density-dependent growth and sexual maturity of silver hake in the north-west Atlantic. Journal of Fish Biology 51:607-623 DOI 10.1111/j.1095-8649.1997.tb01516.x.
Hilborn R, Amoroso RO, Anderson CM, Baum JK, Branch TA, Costello C, De Moor CL, Faraj A, Hively D, Jensen OP, Kurota H, Little LR, Mace P, McClanahan T, Melnychuk MC, Minto C, Osio GC, Parma AM, Pons M, Segurado S, Szuwalski CS, Wilson JR, Ye Y. 2020. Effective fisheries management instrumental in improving fish stock status. Proceedings of the National Academy of Sciences of the United States of America 117:2218 DOI 10.1073/pnas. 1909726116.
Hollowed AB. 1992. Spatial and temporal distributions of Pacific hake, Merluccius productus, larvae and estimates of survival during early life stages. California Cooperative Oceanic Fisheries Investigations Reports 33:100-123.
Iwamoto EM, Elz AE, García-De León FJ, Silva-Segundo CA, Ford MJ, Palsson WA, Gustafson RG. 2015. Microsatellite DNA analysis of Pacific hake Merluccius productus population structure in the Salish Sea. ICES Journal of Marine Science: Journal Du Conseil 72:2720-2731 DOI 10.1093/icesjms/fsv146.
Iwamoto E, Ford M, Gustafson R. 2004. Genetic population structure of Pacific hake, Merluccius productus, in the Pacific northwest. Environmental Biology of Fishes 69:187-199 DOI 10.1023/B:EBFI.0000022895.10683.c5.
Jeffries S, Huber H, Calambokidis J, Laake J. 2003. Trends and status of harbor seals in washington state: 1978-1999. The Journal of Wildlife Management 67:207-218 DOI 10.2307/3803076.
Jones CM. 1992. Development and application of the otolith increment technique. In: Stevenson DK, Campana SE, eds.In Otolith microstructure examination and analysis. Department of Fisheries and Oceans, Ottawa. Canadian Special Publication of Fisheries and Aquatic Sciences 117 1-12.
Juan-Jordá MJ, Mosqueira I, Freire J, Dulvy NK. 2015. Population declines of tuna and relatives depend on their speed of life. Proceedings of the Royal Society of London B: Biological Sciences 282:1-7.

King JR, McFarlane GA. 2006. Shift in size-at-age of the Strait of Georgia population of Pacific hake (Merluccius productus). California Cooperative Oceanic Fisheries Investigations Reports 47:111-118.
Kingsolver JG, Huey RB. 2008. Size, temperature, and fitness: three rules. Evolutionary Ecology Research 10:251-268.
Le Pape O, Holley J, Guerault D, Desaunay Y. 2003. Quality of coastal and estuarine essential fish habitats: estimations based on the size of juvenile common sole (Solea solea L.). Estuarine, Coastal and Shelf Science 58:793-803 DOI 10.1016/S0272-7714(03)00185-9.
Ligas A, Colloca F, Lundy MG, Mannini A, Sartor P, Sbrana M, Voliani A, Belcari P. 2015. Modeling the growth of recruits of European hake (Merluccius merluccius) in the northwestern Mediterranean Sea with generalized additive models. Fishery Bulletin 113:69-81.
Limburg KE, Casini M. 2018. Effect of marine hypoxia on baltic sea cod Gadus morhua: evidence from otolith chemical proxies. Frontiers in Marine Science 5:482 DOI 10.3389/fmars.2018.00482.
Limburg KE, Olson C, Walther Y, Dale D, Slomp CP, Høie H. 2011. Tracking Baltic hypoxia and cod migration over millennia with natural tags. Proceedings of the National Academy of Sciences of the United States of America 108:E177-E182.
Limburg KE, Walther BD, Lu Z, Jackman G, Mohan J, Walther Y, Nissling A, Weber PK, Schmitt AK. 2015. In search of the dead zone: use of otoliths for tracking fish exposure to hypoxia. Journal of Marine Systems 141:167-178 DOI 10.1016/j.jmarsys.2014.02.014.
Losee JP, Kendall NW, Dufault A. 2019. Changing salmon: an analysis of body mass, abundance, survival, and productivity trends across 45 years in Puget Sound. Fish and Fisheries 20:934-951 DOI 10.1111/faf. 12385.
Macdonald JI, Crook DA. 2010. Variability in Sr:Ca and Ba:Ca ratios in water and fish otoliths across an estuarine salinity gradient. Marine Ecology Progress Series 413:147-161 DOI 10.3354/meps08703.
Martin GB, Thorrold SR. 2005. Temperature and salinity effects on magnesium, manganese, and barium incorporation in otoliths of larval and early juvenile spot Leiostomus xanthurus. Marine Ecology Progress Series 293:223-232 DOI 10.3354/meps293223.
Martin GB, Thorrold SR, Jones CM. 2004. Temperature and salinity effects on strontium incorporation in otoliths of larval spot (Leiostomus xantharus). Canadian Journal of Fisheries and Aquatic Sciences 61:34-42 DOI 10.1139/f03-143.
McFarlane GA, Beamish RJ. 1985. Biology and fishery of Pacific hake (Merluccius productus) in the Strait of Georgia. Marine Fisheries Review 47:365-392.
Meng L, Gray C, Taplin B, Kupcha E. 2000. Using winter flounder growth rates to assess habitat quality in Rhode Island's coastal lagoons. Marine Ecology-Progress Series 201:287-299 DOI 10.3354/meps201287.

Miller JA, Gray A, Merz J. 2010. Quantifying the contribution of juvenile migratory phenotypes in a population of Chinook salmon Oncorhynchus tshawytscha. Marine Ecology Progress Series 408:227-240 DOI 10.3354/meps08613.
Mohan JA, Miller NR, Herzka SZ, Sosa-Nishizaki O, Kohin S, Dewar H, Kinney M, Snodgrass O, Wells RJD. 2018. Elements of time and place: manganese and barium in shark vertebrae reflect age and upwelling histories. Proceedings of the Royal Society B: Biological Sciences 285(1890):20181760 DOI 10.1098/rspb.2018.1760.
Mohan J, Walther B. 2016. Out of breath and hungry: natural tags reveal trophic resilience of Atlantic croaker to hypoxia exposure. Marine Ecology Progress Series 560:207-221 DOI 10.3354/meps11934.
Necaise AD, Ross SW, Miller JM. 2005. Estuarine habitat evaluation measured by growth of juvenile summer flounder Paralichthys dentatus in a North Carolina estuary. Marine Ecology Progress Series 285:157-168 DOI 10.3354/meps285157.
NOAA Fisheries. 2020. Status of stocks 2019: annual report to congress on the status of U.S. Fisheries. 11.

Ohlberger J, Ward EJ, Schindler DE, Lewis B. 2018. Demographic changes in Chinook salmon across the Northeast Pacific Ocean. Fish and Fisheries 19:533-546 DOI 10.1111/faf. 12272.
Oke KB, Cunningham CJ, Westley PAH, Baskett ML, Carlson SM, Clark J, Hendry AP, Karatayev VA, Kendall NW, Kibele J, Kindsvater HK, Kobayashi KM, Lewis B, Munch S, Reynolds JD, Vick GK, Palkovacs EP. 2020. Recent declines in salmon body size impact ecosystems and fisheries. Nature Communications 11:4155-4151 DOI 10.1038/s41467-020-17726-z.
Olesiuk PF. 1999. Status of harbour seals (Phoca vitulina) in British Columbia. Canadian Stock Assessment Secretariat Research Document 99/33.
Olsen EM, Heino M, Lilly GR, Morgan MJ, Brattey J, Ernande B, Dieckmann U. 2004. Maturation trends indicative of rapid evolution preceded the collapse of northern cod. Nature 428:932-935 DOI 10.1038/nature02430.
Pedersen M. 1985. Puget Sound Pacific whiting, Merluccius productus, resource and industry-an overview. Marine Fisheries Review 47:35-38.
R Core Team. 2021. R: a language and environment for statistical computing. Vienna: R Foundation for Statistical Computing. Available at https://www.R-project.org/.
Saunders MW, McFarlane GA. 1999. Pacific hake - Strait of Georgia stock assessment for 1999 and recommended yield options for 2000. Pacific Scientific Advice Review Committee (PSARC) Working Paper G:99-8. 10.
Sharpe DMT, Hendry AP. 2009. Life history change in commercially exploited fish stocks: an analysis of trends across studies. Evolutionary Applications 2:260-275 DOI 10.1111/j.1752-4571.2009.00080.x.
Sogard SM. 1997. Size-selective mortality in the juvenile stage of teleost fishes: a review. Bulletin of Marine Science 60:1129-1157.
Stevenson DK, Campana SE. 1992. Otolith microstructure examination and analysis. Canadian Special Publication of Fisheries and Aquatic Sciences 117:126.

Steves BP, Cowen RK. 2000. Settlement, growth, and movement of silver hake Merluccius bilinearis in nursery habitat on the New York Bight continental shelf. Marine Ecology Progress Series 196:279-290 DOI 10.3354/meps196279.
Takahashi M, Checkley Jr DM, Litz MNC, Brodeur RD, Peterson WT. 2012. Responses in growth rate of larval northern anchovy (Engraulis mordax) to anomalous upwelling in the northern California Current. Fisheries Oceanography 21:393-404 DOI 10.1111/j.1365-2419.2012.00633.x.
Utter FM, Stormont CJ, Hodgins HO. 1970. Esterase polymorphism in vitreous fluid of Pacific hake, Merluccius productus. Animal Blood Groups and Biochemical Genetics 1:69-82.
Uusi-Heikkilä S, Whiteley AR, Kuparinen A, Matsumura S, Venturelli PA, Wolter C, Slate J, Primmer CR, Meinelt T, Killen SS, Bierbach D, Polverino G, Ludwig A, Arlinghaus R. 2015. The evolutionary legacy of size-selective harvesting extends from genes to populations. Evolutionary Applications 8:597-620 DOI 10.1111/eva.12268.
WDFW. 2008. Priority habitat and species list, Olympia, Washington. Washington Department of Fish and Wildlife. 177.
Wheeler SG, Russell AD, Fehrenbacher JS, Morgan SG. 2016. Evaluating chemical signatures in a coastal upwelling region to reconstruct water mass associations of settlement-stage rockfishes. Marine Ecology Progress Series 550:191-206 DOI 10.3354/meps11704.
Zabel RW, Achord S. 2004. Relating size of juveniles to survival within and among populations of Chinook salmon. Ecology 85:795-806 DOI 10.1890/02-0719.
Zimmerman CE. 2005. Relationship of otolith strontium-to-calcium ratios and salinity: experimental validation for juvenile salmonids. Canadian Journal of Fisheries and Aquatic Sciences 62:88-97 DOI 10.1139/f04-182.

