

SCIENTIFIC REPORTS



OPEN

Catalytic activity of nickel nanoparticles stabilized by adsorbing polymers for enhanced carbon sequestration

Seokju Seo¹, Gabriela Alvarez Perez¹, Ketan Tewari², Xavier Comas³ & Myeongsub Kim¹ 

This work shows the potential of nickel (Ni) nanoparticles (NPs) stabilized by polymers for accelerating carbon dioxide (CO₂) dissolution into saline aquifers. The catalytic characteristics of Ni NPs were investigated by monitoring changes in diameter of CO₂ microbubbles. An increase in ionic strength considerably reduces an electrostatic repulsive force in pristine Ni NPs, thereby decreasing their catalytic potential. This study shows how cationic dextran (DEX), nonionic poly(vinyl pyrrolidone) (PVP), and anionic carboxy methylcellulose (CMC) polymers, the dispersive behaviors of Ni NPs can be used to overcome the negative impact of salinity on CO₂ dissolution. The cationic polymer, DEX was less adsorbed onto NPs surfaces, thereby limiting the Ni NPs' catalytic activity. This behavior is due to a competition for Ni NPs' surface sites between the cation and DEX under high salinity. On the other hand, the non/anionic polymers, PVP and CMC could be relatively easily adsorbed onto anchoring sites of Ni NPs by the monovalent cation, Na⁺. Considerable dispersion of Ni NPs by an optimal concentration of the anionic polymers improved their catalytic capabilities even under unfavorable conditions for CO₂ dissolution. This study has implications for enhancing geologic sequestration into deep saline aquifers for the purposes of mitigating atmospheric CO₂ levels.

Greenhouse gas emissions attributed to anthropogenic activities have increased significantly over the past century and continue to escalate. This increase causes measurable impacts on global warming, leading to climate-related concerns such as the unprecedented rising of sea levels along with more frequent and intense wildfires, floods, droughts, and tropical storms. Such adverse effects not only disrupt natural ecosystems, but they also pose serious risks to human populations¹. Carbon dioxide (CO₂) from fossil fuel combustion and industrial processes has accounted for 78% of the total greenhouse gas emissions². Although natural systems can absorb CO₂ back as part of the carbon cycle, these natural uptake processes are not enough to offset the escalating amount of anthropogenic CO₂ released into the atmosphere³. Mitigation of CO₂ emissions is an imminent global concern and must be urgently addressed by both political and technological implementations. As a result, many strategies have been proposed to mitigate global atmospheric CO₂ levels^{4–6}.

One promising strategy is to inject CO₂ into subsurface rock formations because they provide the largest storage potential^{7–10}. Among various geological storage options (e.g., saline aquifers, depleted oil or natural gas reservoirs, and coalbed reservoirs), deep saline aquifers provide the highest potential capacity, with approximately 10,000 billion metric tons of carbon in total¹¹. However, safe storage through a solubility trapping process is controversial since most of the injected CO₂ remains as CO₂ molecules in the absence of a catalyst for enhancing the reaction rate of carbonic acid (H₂CO₃) formation¹². This reaction rate is further hindered due to a high salinity content in brine, as preferable saline concentrations for CO₂ storage range under 5% weight per volume while some aquifers have salinities ranging up to 25%¹³. As a result, the strategy of the accelerated CO₂ dissolution process should be investigated in order to achieve a more feasible means of storing CO₂ into deep saline aquifers.

¹Department of Ocean and Mechanical Engineering, Florida Atlantic University, 777 Glades Road, Boca Raton, FL33431, USA. ²Department of Mechanical Engineering, Nirma University, Sarkhej-Gandhinagar Highway, Chandlodia, Gota, Ahmedabad, Gujarat, 382481, India. ³Department of Geosciences, Florida Atlantic University, 777 Glades Road, Boca Raton, FL33431, USA. Correspondence and requests for materials should be addressed to M.K. (email: kimm@fau.edu)

Engineered nanoparticles (NPs) have attracted much attention in numerous applications due to their unique material and surface properties^{14–16}. In addition, a high surface to volume ratio of NPs can considerably enhance their interfacial reactivity with the surrounding medium¹⁷. In particular, nickel (Ni) NPs has the potential to accelerate the rate of CO₂ dissolution, especially under unfavorable water chemistry^{18–21}. Major catalysts for the hydration of CO₂ include the enzyme carbonic anhydrase^{22,23}, hypobromous acid (HOBr), hypochlorous acid (HClO)²⁴ and boric acid (B(OH)₃)^{25,26}. However, these catalysts are effective primarily in the alkaline range. Especially, enzyme carbonic anhydrase at a pH less than 7 catalyzes the dehydration of the bicarbonate ion (HCO₃⁻)²⁷. Unlike these catalysts, the Ni NPs' catalytic activity is pH independent^{18,21}. Since many saline aquifers are in an acidic condition (e.g., pH 2.3–2.24 in Indiana County²⁸, Pennsylvania, pH 4.2 in Indiana County, Pennsylvania²⁹), the feasibility of metallic Ni NPs as a catalytic additive for the hydration of CO₂ was tested in a wide range of pH to enhance carbon sequestration in deep saline aquifers²¹. Furthermore, Ni NPs can be an effective catalyst in deep saline aquifers due to superior corrosion resistance and resistance to high-temperature oxidation of Ni^{30–32} under reservoir-specific temperature conditions (e.g., 40 °C¹⁰ and 62 °C³³). Although high ionic strengths by salinity contents in brine lead to a substantial decrease in Ni NPs' catalytic activity²¹, we observed the possibility of Ni NPs to accelerate CO₂ dissolution even in acidic brine at a moderate salinity. However, many reservoirs are characterized by extreme salinities up to 25% and the upper limit of the Ni NPs' catalytic activity at various salinities is still uncertain. To maximize the CO₂ storage capacity in geologic formations at high salinity and to test the feasibility of Ni NPs-assisted CO₂ dissolution in the field scale, quantification of the catalytic behaviors of Ni NPs in wide-ranging ionic strengths should be accomplished.

In addition, because NPs are easily aggregated due to the strong potential of van der Waals force as salinity increases, finding an efficient way to disperse particles in a solution for their maximum performance is another critical step for fully utilizing the Ni NP's potential in brine with an extreme salinity content. One of the most promising solutions to improve the Ni NPs' catalytic activity is to stabilize them in polymer matrices. Without altering material properties of NPs, small molecule ligands in polymers are used to control stabilization of NPs against agglomeration^{34,35}. While polymer matrices retain the inherent properties of pristine NPs, NPs-polymer interactions lead to an increase in potential of NPs as catalysts and processability of the polymer matrix³⁶. As many publications have shown, NPs bonded by polymer ligands promote NPs' catalytic reactions at their surface^{37–39}. The most widely used host polymer materials include poly(vinyl pyrrolidone) (PVP)⁴⁰ and carboxy methylcellulose (CMC)^{41,42}, and the best performance of NPs' catalytic behaviors in these polymers are generally determined based on the mechanical and thermal properties of the polymeric supports. Due to large variations in polymer properties that control the NPs' catalytic activity, substantial uncertainties about the optimal combination of NPs and polymers remain. These uncertainties can be mitigated by a fast and simple approach, as shown in this study, that determines CO₂ dissolution phenomena accurately.

An experimental bubble-based microfluidic approach has been used as one of the most promising and accurate methods to study the CO₂ dissolution dynamics into aqueous solutions. CO₂ dissolution into an aqueous phase occurring at the interface has been quantified successfully by observing bubble size and morphology^{43–48}. Owing to the high surface area to volume ratio of microbubble, the mass transfer rate is considerably enhanced over very short time scales, resulting in fast changes in bubble size^{48–54}. The rapid change in CO₂ microbubbles facilitates accurate quantification of the CO₂ dissolution rate. The time-dependent size shrinkage or expansion of CO₂ bubbles is due mainly to a reversible reaction of absorption and desorption of CO₂ associated with the properties of the surrounding liquid. When gaseous CO₂ absorption into the aqueous phase dominates, the removal of CO_{2(g)} trapped as a bubble leads to shrinkage of CO₂ bubbles. On the other hand, as increasing desorption of CO_{2(aq)} from the aqueous phase to the gaseous phase, bubble expansion occurs. In a similar way, the role of catalytic activity controlled by Ni NPs-polymer on CO₂ dissolution can be also revealed by observation of microbubble morphology via a microfluidic technology.

In the present work, we introduce a simple and time efficient microfluidic technique that tests the feasibility of utilizing Ni NPs in polymer matrices to accelerate gaseous CO₂ dissolution into a continuous aqueous solution (Fig. 1). First, the time-dependent changes in CO₂ microbubble size in response to the salinity concentrations with and without Ni NPs was quantified using high-speed imaging. The results offered an evidence of a decrease in catalytic potential as the salinity increases because of a Ni NPs' aggregation behavior in brine. To further enhance Ni NPs' catalytic potential, the hydration of CO₂ was measured using Ni NPs stabilized with cationic dextran (DEX), nonionic PVP, and anionic CMC polymers. The results suggest that the Ni NPs-polymer matrix could enhance CO₂ dissolution up to three times when compared to dissolution into a pure aqueous solution. Optimal concentrations of these polymers for enhancing Ni NPs performance were also determined to enhance carbon sequestration processes.

Results

Salinity effects on CO₂ dissolution. Figure 2a shows the schematic of a microfluidic device for investigation of CO₂ dissolution into an aqueous solution. In a flow-focusing geometry, a gaseous CO₂ was focused by the injected continuous aqueous phase on both sides. A combination of a shear force by the two aqueous streams and an interfacial tension force between the liquid and the gas phases generates spherical CO₂ bubbles at the junction. Once the bubbles are generated, the mass transfer of gaseous CO₂ immediately occurs at the gas-liquid interface while traveling downstream and subsequently changing their sizes. To enhance measurement accuracy, the initial bubble diameter and generation frequency of CO₂ bubbles should be maintained constant (Fig. 2b). To achieve this goal, monodispersed bubbles were consistently generated in all experiments by fixing input flow rate and CO₂ gas pressure at 0.3 mL min⁻¹ and 1 psi, respectively. Figure 2d shows a representative distribution of the bubble size in one of our tests.

To evaluate the dissolution behaviors of CO₂ at salinity, a series of tests was performed at 0–30% (weight per volume) NaCl solutions with neutral pH. It should be noted that pH is an important parameter to change the

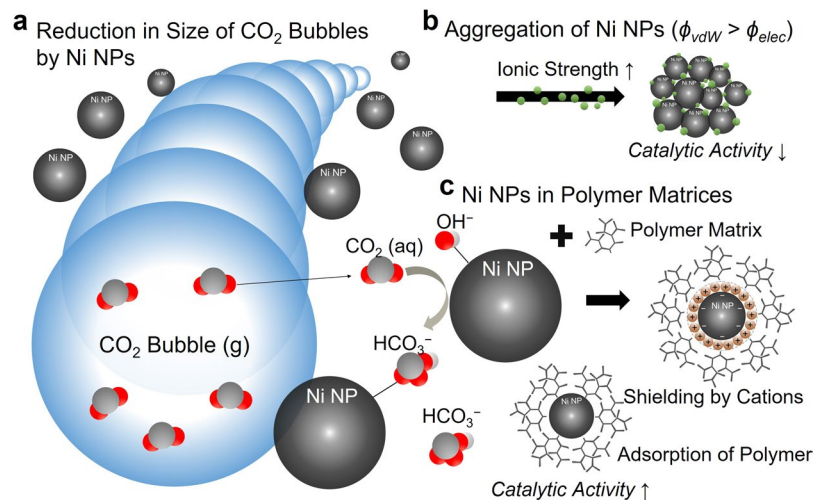


Figure 1. Conceptual illustration for the hydration of CO₂. **(a)** Accelerated CO₂ dissolution by Ni NPs catalytic potential. **(b)** The decreased catalytic potential of Ni NPs by the aggregation behavior of NPs in high ionic strengths. **(c)** Stabilized Ni NPs by polymers in brine at high salinity levels.

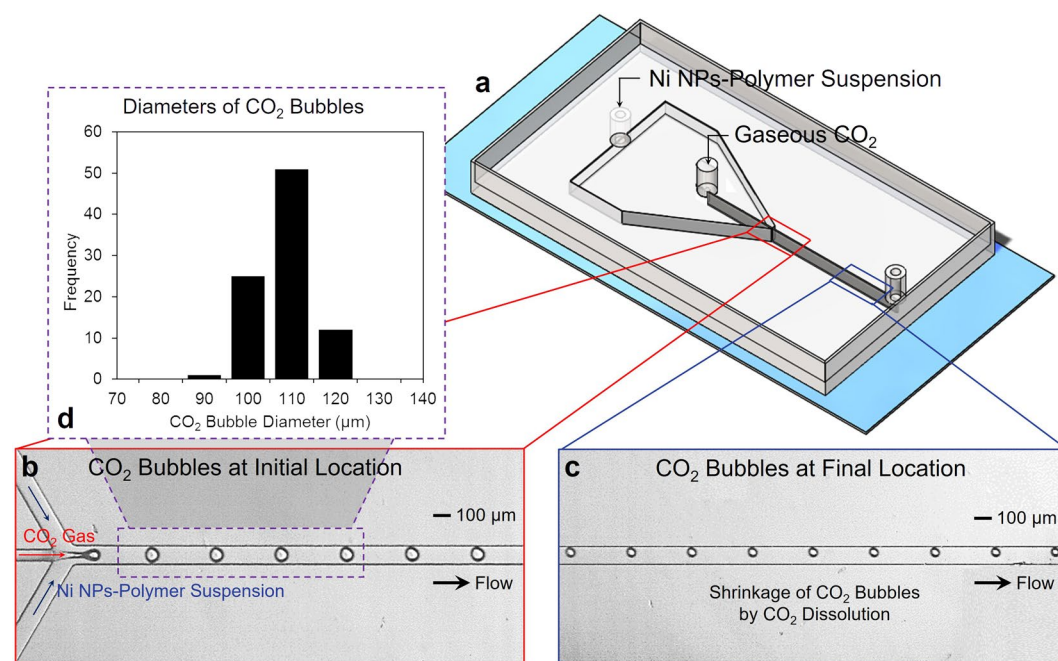


Figure 2. A microfluidic approach for evaluation of CO₂ dissolution into water. **(a)** A microfluidic platform with its configuration. **(b,c)** Representative micrographs of CO₂ bubbles at two different locations (near the junction and the outlet) in the flow-focusing geometry. **(d)** A histogram of estimated diameters of CO₂ bubbles near the junction.

hydration of CO₂ at different salinity levels^{21,55}. Figure 3a shows a time series of CO₂ bubbles near the junction in the microfluidic device. The velocity of CO₂ bubbles in the microchannel was approximately $5.926 \pm 0.92 \text{ mm s}^{-1}$ and their residence time between two different locations (near junction and outlet) was $3.46 \pm 0.54 \text{ s}$. Each CO₂ bubble was generated at a constant interval of approximately 20 ms. A distribution of CO₂ bubble diameters was determined for more than 50 CO₂ bubbles collected for each experiment (Fig. 3b). Applying Gaussian, Lognormal or Lorentzian fittings in these histograms, the average diameters of CO₂ bubbles near junction (the initial location) are determined to be 104.93, 103.96, 99.41, 101.8, 100.35, 99.08, 96.04, and 98.66 μm at 0%, 2%, 4%, 6%, 8%, 10%, 15%, and 30% salinity, respectively (Fig. S1). We observed no substantial changes in initial size at different salinities. However, due to an increase in ionic strength, the degree of shrinkage of CO₂ bubbles decreases proportionally (Fig. 3c and Supplementary Movie S1).

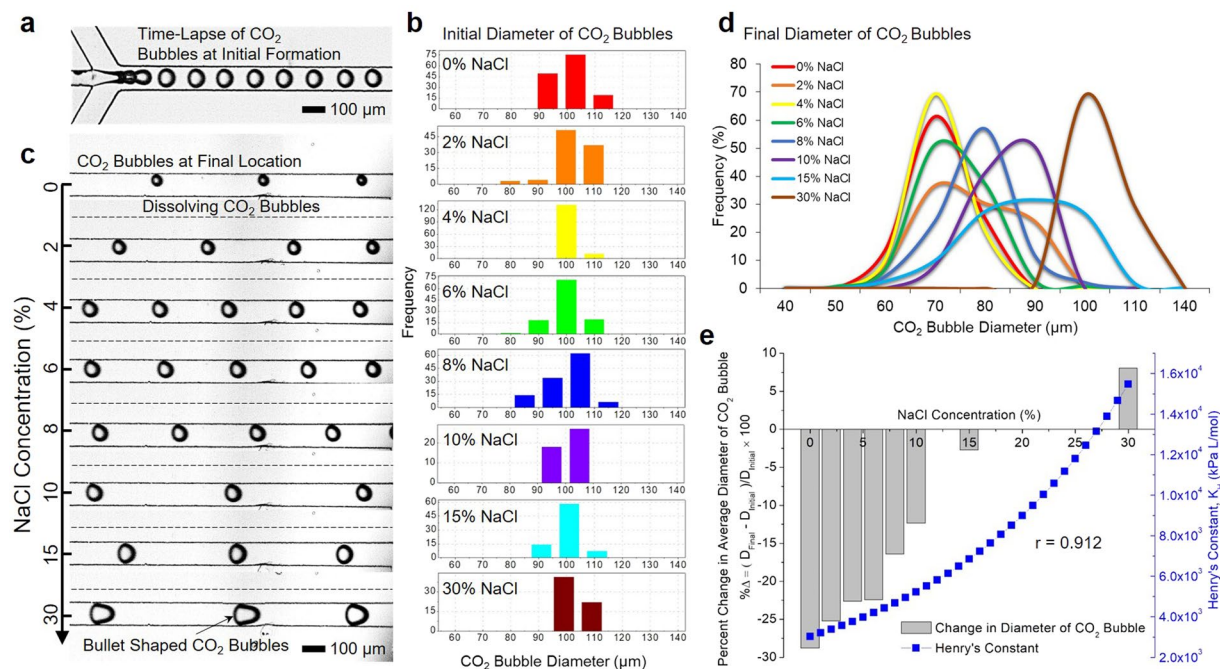


Figure 3. (a) A time series of CO₂ bubbles at the junction of the microchannel. (b) Histograms of the initial diameter of CO₂ bubbles. (c) Representative micrographs of CO₂ bubbles near the outlet at the concentration of salinity (0%, 2%, 4%, 6%, 8%, 10%, 15%, and 30% NaCl). (d) Variations in size of CO₂ bubbles near the outlet at different salinities. (e) A diagram of percent changes ($\% \Delta = (D_{final} - D_{initial})/D_{initial} \times 100$) in average diameter of CO₂ bubbles and the graph of Henry's constant (■) at different salinities.

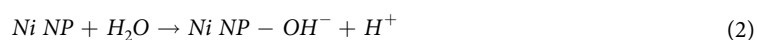
The average sizes of CO₂ bubbles near the outlet (the final location) decrease by 30.2, 26.21, 22.51, 22.81, 16.47, 12.22, and 2.63 μm from 0 to 15% NaCl, respectively (Fig. 3d). At 30% NaCl, the expansion of CO₂ bubbles occurs, increasing by 7.94 μm. Bulletlike CO₂ bubbles are observed since the width of expanded CO₂ bubbles is greater than that of the microchannel (100 μm). Gas exchange at the interface between the CO₂ bubble and the aqueous phase starts when the bubble is generated^{52,56}. Gaseous CO₂ molecules trapped in the bubble start to diffuse into the aqueous phase. Meanwhile, dissolved gas molecules including nitrogen and oxygen are simultaneously desorbed back to the CO₂ bubble. While the absorption rate of CO₂ gas molecules is decreased with an increase in salinity, the change in CO₂ bubble diameter was decreased since the desorption rate of dissolved gas molecules remains constant. Because desorption of dissolved gases dominates at 30% NaCl, the expansion of CO₂ bubbles was observed near the outlet (the final location). Collectively, the variation in size of CO₂ bubbles at different salinities indicates that the extent of CO₂ dissolution considerably decreases as increasing salinity concentration.

The effect of salinity on CO₂ dissolution could be explained by Henry's Law described as follows⁵⁷

$$k_H = k_H^0 \cdot 10^{0.138I} = \frac{P_{CO_2}}{[CO_2]_{aq}}, \quad (1)$$

where k_H^0 ($3.45 \times 10^3 \text{ kPa L mol}^{-1}$) is Henry's constant for CO₂ in water at 25 °C and I is the ionic strength ($I = 0.5 \sum_i C_i z_i^2$, C_i is the concentration of ion of type i , and z_i is its valence), $[CO_2]_{aq}$ is aqueous CO₂ concentration, and P_{CO_2} is the partial pressure of the gaseous CO₂. As the ionic strength in the aqueous phase increases, the value of Henry's constant increases as well, and thereby decreases CO₂ dissolution. To evaluate the influence of the ionic strength on the change in CO₂ bubble size, Henry's constants at 0 to 30% salinity were determined (Fig. 3e ■). Figure 3e also shows percent changes of CO₂ bubbles diameter at the initial and final locations ($\% \Delta = (D_{final} - D_{initial})/D_{initial} \times 100$). The Pearson's correlation coefficient ($r = 0.912$) suggests that a significant correlation exists between the variation in size of CO₂ bubbles and Henry's constants. It should be noted that a critical value for correlation coefficient r at the $\alpha = 0.01$ level is 0.834 (6 degrees of freedom).

Ni NPs-assisted CO₂ dissolution. Ni NPs can overcome the decreased dissolution of CO₂ under high saline concentrations through their excellent catalytic potential²¹. When the aqueous phase containing Ni NPs surrounds CO₂ bubbles that travel along the microchannel, the diameter of CO₂ bubbles becomes decreased thanks to the accelerated hydration of gaseous CO₂ by the Ni NPs' catalytic activity. The relevant chemical reactions occur subsequently as follows:



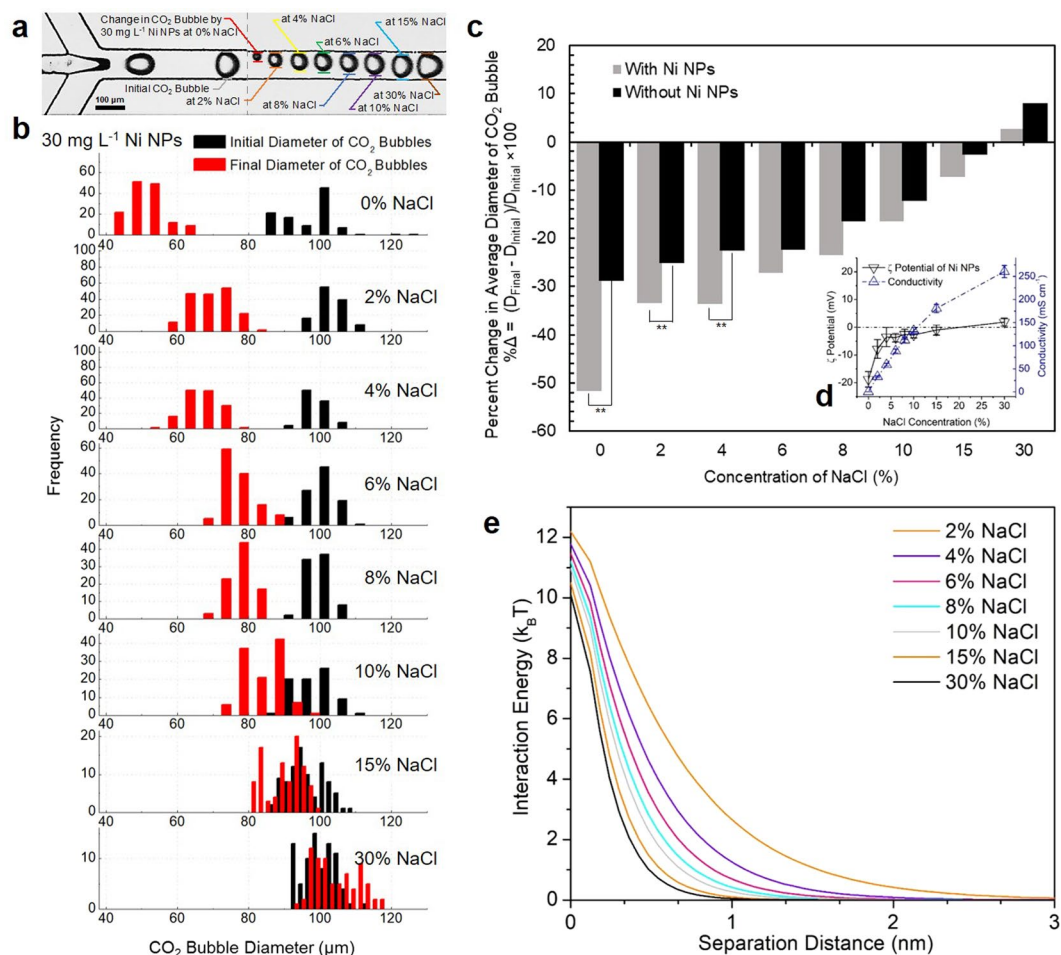
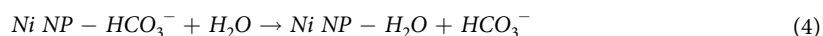
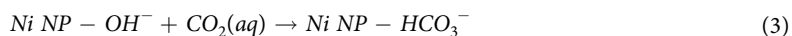


Figure 4. (a) A representative time-lapse microscopy image of CO₂ bubbles near the junction and the outlet in the presence of Ni NPs at different salinities (0%, 2%, 4%, 6%, 8%, 10%, 15%, and 30% NaCl). (b) Histograms of sizes of CO₂ bubbles near the junction (the initial location) and the outlet (the final location) to confirm the effect of salinity on Ni NPs' catalytic potential. (c) The percent changes in average diameter of CO₂ bubbles near the outlet in the presence and absence of Ni NPs at different salinities (***p* < 0.01). (d) The graphs of experimentally measured ζ potential and conductivity of Ni NPs solution at different salinities. (Error bars show mean \pm standard error of the mean from three independent measurements). (e) The electrostatic double layer potential (ϕ_{elec}) between two Ni NPs from the extended-DLVO theory at various ionic strengths.



When Ni NPs are put into an aqueous solution, they react with H₂O to generate hydroxyl (OH⁻) groups on NPs' surfaces. Then, OH⁻ groups react with CO₂(aq) to produce HCO₃⁻ ions attached to the surface of Ni NPs. These HCO₃⁻ ions are replaced by H₂O. A series of these catalytic chemical reactions associated with Ni NPs promotes accelerated removal of CO₂(aq) from the water that consequently catalyzes the hydration of CO₂. It should be noted that when metallic Ni NPs are exposed to an oxidative environment, their catalytic performance for the hydration of CO₂ can considerably decrease due to the formation of an oxide layer on Ni NPs' surfaces⁵⁸. However, according to previous studies^{58,59}, oxidation of Ni NPs did not occur at an ambient temperature for two months⁵⁹ and the initiation of oxidation occurred at approximately 300 °C⁵⁸. Taking this fact into consideration, the temperature range of deep saline aquifers (e.g., 40 °C¹⁰ and 62 °C³³) is unlikely to considerably inhibit the Ni NPs' catalytic activity by oxidation.

Figure 4a shows representative time-lapse micrographs of CO₂ bubbles in 30 mg L⁻¹ concentration of Ni NPs solution under varying ionic strengths with neutral pH. The concentration of 30 mg L⁻¹ Ni NPs that provides the maximum catalytic behavior of Ni NPs in pure water was derived from our previous research²¹. In the absence of salinity, a substantial decrease in diameter of CO₂ bubbles was observed as a consequence of the catalytic potential of Ni NPs. As salinity increases, the change in CO₂ bubble diameter was similar regardless of the presence of Ni NPs (Figs 4a and 3c). To understand the effect of salinity on Ni NPs' catalytic potential, the changes on CO₂

bubbles were determined at different salinities (Fig. 4b and Supplementary Movie S2). In the histogram, as salinity increases, the increase in diameter of CO₂ bubbles near the outlet yields a shift to the right (■). No considerable variation in initial size of CO₂ bubbles (■, 101.64, 103.9, 98.96, 102.44, 100.55, 99.51, 95.87, 99.8 μm) was observed at 0%, 2%, 4%, 6%, 8%, 10%, 15%, and 30% salinity, respectively (Fig. S2). Figure 4c shows the percent changes (%Δ = (D_{final} - D_{initial})/D_{initial} × 100) in CO₂ bubble diameter with -51.73 (-28.79), -33.45 (-25.21), -33.6 (-22.64), -27.13 (-22.42), -23.47 (-16.41), -16.41 (-12.33), -7.29 (-2.74), 2.71 (8.04)% in the presence (absence) of Ni NPs at the different ionic strengths, respectively. The difference in magnitude of these percent changes in CO₂ bubble diameter represents Ni NPs' catalytic potential for enhancement of the hydration of CO₂. In the absence of salt, the percent change of CO₂ bubbles in Ni NPs solution was approximately twofold higher than that without Ni NPs. However, an increase in NaCl concentrations results in a substantial decrease in Ni NPs' catalytic potential due to their aggregation in high ionic suspension. Up to 4% NaCl, significant differences of the CO₂ bubble size change with and without Ni NPs were confirmed by statistical estimations ($p = 7.62 \times 10^{-18}$, 6×10^{-3} , 1.91×10^{-5} between the presence and absence of Ni NPs at 0~4% NaCl); however, no significant differences were estimated at greater than 6% NaCl regardless of Ni NPs ($p = 6.5 \times 10^{-1}$, 2.5×10^{-1} , 6.57×10^{-1} , 6.62×10^{-2} , and 2.6×10^{-2} between the presence and absence of Ni NPs at 6~30% NaCl, respectively) (Fig. 4c).

The aggregation behavior of NPs at high salinity was verified by measurements of their zeta (ζ) potential at different salinities, as shown in Fig. 4d. According to Derjaguin, Landau, Verwey, and Overbeek (DLVO) theory^{60,61}, an electrical double layer (EDL) of each colloidal particle generates a repulsion force to prevent particle-particle aggregation when particles approach each other. It is known that, if the ζ potential, electrical potential in EDL is less than ± 20 mV, particles are considered unstable and consequently aggregate and settle out of suspension. The average ζ potential of pristine Ni NPs in the absence of NaCl was measured to be -18.8 ± 2.8 mV. This suggests that Ni NPs in pure aqueous solution are still unstable. An increase in concentration of NaCl considerably changes the ζ potential to -3.75 ± 1.52 mV at the ionic strength of 1.02 M (6% NaCl). A plateau value of ζ potential was observed at ionic strengths of 1.37 M (8% NaCl) or greater, indicated as an electrolytic conductivity (Fig. 4d). This implies that the increase in ionic strength of suspension diminishes considerably the NPs stability, thereby decreasing their catalytic potential.

An extended DLVO model that accounts for two Ni NPs' electrostatic double layer potential (ϕ_{elec}) can support these results (Fig. 4e).

$$\phi_{elec} = \frac{2\pi e_0 \varepsilon r \varphi^2 \ln(1 + e^{-h/\kappa^{-1}})}{k_B T}, \quad (5)$$

where e_0 is permittivity of free space ($8.85 \times 10^{-12} \text{ F m}^{-1}$), r is a radius of Ni NPs (50 nm), ε is a dimensionless dielectric constant of water (80.1), φ is the ζ potential of Ni NPs (-18.8 mV), e is an electron charge ($1.6 \times 10^{-19} \text{ coulomb}$), h is the separation distance between two Ni NPs, k_B is the Boltzmann's constant ($1.381 \times 10^{-23} \text{ J K}^{-1}$), T is the absolute temperature (298.15 K). κ^{-1} , which is the Debye length (or thickness of the double layer), and described as follows:

$$\kappa^{-1} = \frac{\sqrt{e_0 \varepsilon k_B T}}{\sqrt{2 N_A e^2 I}} \quad (6)$$

where e_0 is the permittivity of free space ($8.85 \times 10^{-12} \text{ F m}^{-1}$), ε is the dimensionless dielectric constant of water (80.1) at 20 °C, N_A is the Avogadro's number ($6.02 \times 10^{23} \text{ mol}^{-1}$), e is the electron charge ($1.6 \times 10^{-19} \text{ coulomb}$), and I is the ionic strength of solution (mol m^{-3}). As the ionic strength in Ni NPs suspension increases, an energy barrier induced by two double layers of Ni NPs to prevent their aggregation considerably decreases since the Debye length is inversely proportional to the ionic strength of the solution. As a result, the electrostatic repulsive force becomes weaker and van der Waals attraction dominates, thereby becoming a favorable condition for particle-particle aggregation.

Enhanced hydration of CO₂ in Ni NPs-polymer solution. Polymer matrices greatly stabilize NPs and therefore increase their catalytic potential³⁶. The variations in diameter of CO₂ bubbles were measured in 30 mg L⁻¹ Ni NPs solution with the different types of polymers including DEX, PVP, and CMC, respectively. The salinity was chosen at 10% NaCl where the absolute value of ζ potential near zero (Fig. 4d) to find the optimal combination for the enhancement of steric stabilization of Ni NPs. Steric stabilization is a process in which NPs are prevented from the aggregation through the adsorption of macromolecules derived from polymers on the NPs surface. Figure 5a and Supplementary Movie S3 show representative images and videos of CO₂ bubbles near the junction and the outlet. For each polymer, three different concentrations (0.01, 0.02, and 0.03%) were tested to evaluate the degree of bubble shrinkages.

Figure 5b shows the percent changes in CO₂ bubble diameter under aforementioned conditions. For DEX, PVP, and CMC at 0.01%, 0.02%, and 0.03% concentrations, the initial diameters of CO₂ bubbles were observed to be: 102.7, 106.37, 99.87, 103.9, 98.35, 105.57, 105.06, 97.5, and 104.6 μm, respectively (Fig. S3). The diameter changes for DEX, PVP, and CMC at 0.01%, 0.02%, and 0.03% were -15.65, -35.12, -27.99, -32.94, -40.85, -28.29, -49.31, -48.14, and -24.77%, respectively. With comparing the results at 10% NaCl without Ni NPs and polymer, Ni NPs at 10% NaCl (without polymers), Ni NPs with 0.01% DEX at 10% NaCl, no significant differences ($p = 1.56 \times 10^{-1}$ between Ni NPs and Ni NPs with 0.01% DEX at 10% NaCl) in the percent change in CO₂ bubbles (-12.33, -16.41, and -15.65%, respectively) were observed. The cationic polymer molecules derived from DEX lead to increase in net surface charge of Ni NPs to be positive. However, 0.01% DEX was not

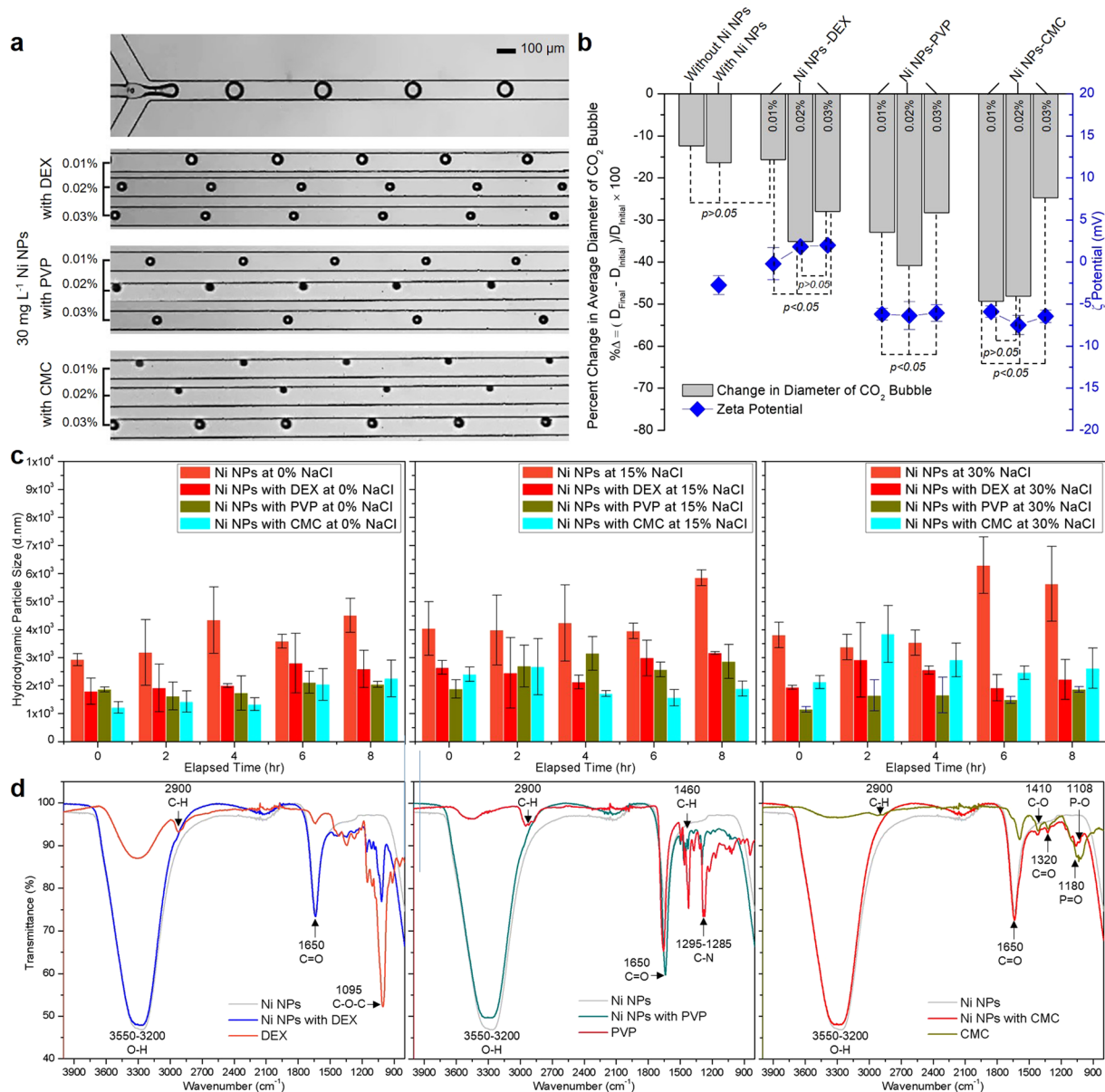


Figure 5. (a) Representative micrographs of CO₂ bubbles near the junction (top) and the outlet (bottom) at stabilized Ni NPs by 0.01%, 0.02%, and 0.03% of DEX, PVP, and CMC at 10% NaCl. (b) A diagram of percent changes in average diameter of CO₂ bubbles in Ni NPs solution with DEX, PVP, CMC and graphs (♦) of measured ζ potential of stabilized Ni NPs as a function of polymer concentration at 0.01%, 0.02%, and 0.03%. (c) Diagrams of changes in hydrodynamic particle size (d.nm) of Ni NPs with 0.03% of DEX, PVP, and CMC at different salinities (0%, 15%, and 30% NaCl). (d) FTIR spectra of Ni NPs, polymers, and Ni NPs with DEX, PVP, and CMC.

considerably effective in increasing the ζ potential of Ni NPs in the presence of salinity. Stabilized Ni NPs by 0.01% DEX at 10% NaCl still possess almost zero net electrical charge on the surface (−0.2 mV, Fig. 5b). This observation can be explained as follows. The monovalent cation, Na⁺ and cationic polymeric molecules from DEX deposit on negatively charged Ni NPs' surface due to the electrostatically mediated attraction. Under aqueous solutions enriched in the monovalent cation, there exists a competition for occupying Ni NPs' surface sites between the monovalent cation and the cationic polymer. A large amount of Na⁺ ion in a solution is responsible for the decreased deposition of cationic polymer molecules on Ni NPs. As a result, cationic polymer molecules from DEX are less adsorbed at low concentration of DEX due to this competition. However, as the percentage of DEX contents in the polymer-Ni NPs mixture increases from 0.01% to 0.03%, more DEX polymer molecules adsorbed onto Ni NPs' surface sites so that the ζ potential of Ni NPs was linearly increased from −0.2 to 2.0 mV (Fig. 5b). As a result, CO₂ bubbles noticeably shrank to a smaller size at 0.02% and 0.03%.

Unlike cationic DEX polymer, the ζ potential of Ni NPs effectively decreased to −6.2 and −5.91 mV at even 0.01% concentration of PVP and CMC, respectively. This is because higher ionic strengths produce a shielding

of negatively charged Ni NPs' surface by a monovalent cation (Na^+)^{62,63}. Na^+ ions attached on the NPs' surface serve as anchoring sites so that the electrostatic repulsion force between Ni NPs and nonionic/anionic polymer molecules is decreased effectively⁶⁴. Therefore, the stabilization of negatively charged Ni NPs in solution enriched by monovalent cations from NaCl, resulting in the accelerated hydration of CO_2 , attributes to the increase in ζ potential of Ni NPs. The increased catalytic potential of Ni NPs in PVP and CMC polymers reflects the large percent changes in CO_2 bubbles diameter at -32.94 and -49.31% , respectively.

When the concentration of PVP was increased from 0.01 to 0.02%, the catalytic potential of Ni NPs was enhanced. As a result, the percent change in CO_2 bubbles diameter was decreased from -32.94 to -40.85% , respectively. However, an increase in concentration of CMC from 0.01 to 0.02% is unlikely to be beneficial for steric stabilization of Ni NPs (Fig. 5b). There were insignificant differences ($p = 9.43 \times 10^{-2}$) in the effect of stabilized Ni NPs by 0.01 and 0.02% CMC on the percent change in CO_2 bubbles size (-49.31 and -48.14%). As shown in the figure, while the decrease in ζ potential of Ni NPs in PVP and CMC was observed with no consequences, the Ni NPs' catalytic activity was decreased considerably at 0.03% PVP and CMC. This is because the surplus polymer at 0.03% concentration has a tendency to create flocculation between NPs once the non/anionic polymer concentration is over the saturation of NPs dispersion. This result indicated that 0.02% PVP and 0.01% CMC are the optimal concentration for steric stabilization of Ni NPs.

To confirm the enhanced stability of Ni NPs by polymers, time-dependent changes in hydrodynamic Ni particle size were measured at different salinities (0%, 15%, 30% NaCl). Without polymers, initial average hydrodynamic sizes were 2929.33, 4040.5, 3810.5 nm at 0%, 15%, and 30% NaCl, respectively (Fig. 5c). However, Ni NPs with DEX, PVP and CMC showed a considerable decrease in size: 1800.8, 1866.33, 1216.2 nm at 0% NaCl, 2645, 1878, 2404.67 nm at 15% NaCl, and 1933.67, 1147.77, 2121.5 nm at 30% NaCl, respectively. Furthermore, while considerable changes in size were observed in pure Ni NPs after 8 hr (2929.33 \rightarrow 4513, 4040.5 \rightarrow 5855, 3810.5 \rightarrow 5634 nm in 0%, 15%, and 30% NaCl, respectively), these changes were negligible in Ni NPs with polymers, suggesting that polymers could effectively increase the stability of Ni NPs.

General characterization of changes in Ni NPs' properties associated with polymer adsorption and the specific interaction between Ni NPs and polymers could be further verified from the Fourier transform infrared spectroscopy (FTIR) spectra (Fig. 5d). Without polymers, a FTIR spectrum of Ni NPs has transmission bands at 3550–3200 cm^{-1} attributed to the O–H stretching region and 1650 cm^{-1} from the vibrations of C=O stretching. On the other hand, with adsorbed DEX polymer, C–O–C stretching at 1095 cm^{-1} appears in a FTIR spectrum in good agreement with the previous study⁶⁵. Characteristic transmission peaks in Ni NPs with PVP were 1460 cm^{-1} for C–H group and 1285–1295 cm^{-1} for C–N bond in PVP derivatives⁶⁶. Compared to pure PVP, slightly shifted C–N peak (from 1285 to 1295 cm^{-1}) in Ni NPs with PVP indicates that the formation of coordination bond between Ni and nitrogen atoms from PVP⁶⁶. An IR spectrum of Ni NPs with CMC verified the presence of C–O, C=O, P=O, and P–O stretching at 1410, 1320, 1180 and 1108 cm^{-1} ⁶⁷.

Discussion

This study successfully examined stabilized Ni NPs by polymers for enhancing the hydration of CO_2 in the application of geological carbon sequestration. We quantified the influence of ionic strengths on the aggregation of Ni NPs and variations in mass transfer of CO_2 into the aqueous solution. Although Ni NPs can improve CO_2 dissolution, their catalytic potential is considerably decreased as the salinity increases due to the aggregation of NPs at the high ionic strength of brine. We found that the performance of Ni NPs could be advanced through the utilization of cationic, nonionic, and anionic polymers. Quantitative comparison of Ni NPs' catalytic potential in different polymers enables us to determine the best performing polymer for Ni NPs dispersion. The stabilizing ability of nonionic and anionic polymer molecules derived from PVP and CMC at low concentration was found to be superior to the cation polymer for enhancing the Ni NPs' catalytic activity in brine at high salinity levels. The monovalent cation-enriched in the aqueous solutions could prove conducive to facilitating interactions between anionic polymer molecules from CMC and Ni NPs by providing anchoring sites. These adsorbed polymer layers could reduce the van der Waals attraction force. Collectively, the steric stabilization of Ni NPs by adsorbing anionic polymer has been proved effective in accelerating the hydration of CO_2 , resulting in improving the feasibility and stability of geologic CO_2 storage in saline aquifers.

Methods

Microfluidic devices. The microchannels were fabricated in Poly(dimethylsiloxane) (PDMS; Sylgard 184, Dow Corning Corp., USA) using a standard soft photolithography technique. The geometry designs were printed onto a transparency sheet (25,400 dpi, CAD/ART Services Inc., USA) to fabricate a patterned photomask. After exposure to ultraviolet light through the patterned photomask, patterns were transferred to a layer of negative photoresist (KMPR 1025, Microchem, Newton, MA, USA) coated on a silicon wafer, and a master mold was fabricated. PDMS and its cross-linker were mixed at a ratio of 10:1 (weight per weight) and then degassed using a vacuum pump. The mixture was poured onto the master molds. After curing in an oven at 70 °C for at least 1 hour, the PDMS layers were carefully peeled off from the molds. Two inlet and outlet holes were made using a 1.0 mm diameter biopsy punch (Integra Miltex, Inc., Germany) for connection to a pump (PHD ULTRA 4400, Harvard Apparatus, USA) and a regulator equipped to CO_2 cylinder (99.9% purity, Airgas, USA) through PTFE (polytetrafluoroethylene) tubing. The PDMS layer and a glass microscope slide (25 \times 75 \times 1.0 mm, Fisher Scientific, USA) were bonded through oxygen plasma treatment (Harrick Plasma, USA). In a microfluidic device, a 50 μm wide microchannel for gaseous CO_2 and 100 μm wide microchannels for the liquid and the main channel (where the CO_2 bubbles and a continuous liquid flow together) were fabricated for generating spherical CO_2 bubbles. The height of microchannels was 75 μm .

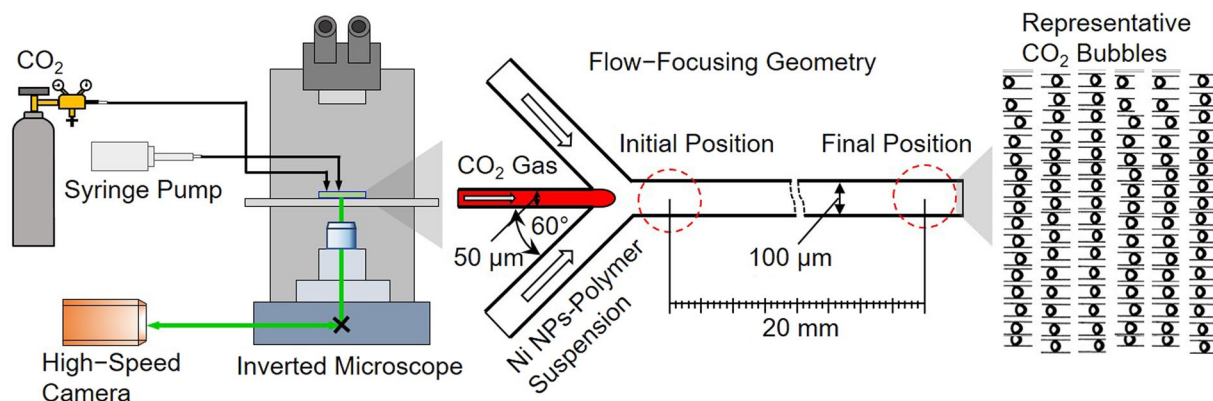


Figure 6. Schematic of test setup (not to scale) and the microfluidic chip of the flow-focusing configuration. An example of collected CO₂ bubbles is shown for image processing.

Experimental configuration. Figure 6 shows the overall configuration used for all experiments. The CO₂ gas and the solutions were injected to a microfluidic chip at a pressure and flow rate of 1 psi and 0.3 mL min⁻¹, respectively. A high-speed camera (Fastec IL5S, Fastec Imaging Corp., USA) attached to an inverted microscope (IX73, Olympus Corp., Japan) with a 4x objective lens was used to capture images of CO₂ bubbles. Since the gas-liquid interface scatters emission light from the microbubbles and appears considerably dark over the surrounding aqueous solution, a binary function in the ImageJ platform was used for estimating the diameter of CO₂ bubbles. For all experiments, at least fifty images of CO₂ bubbles were collected at the locations of observation in the microfluidic device. The histograms of CO₂ bubbles were made to calculate the average initial and final diameters with standard deviations. Then, the percent changes (%Δ = $(D_{final} - D_{initial})/D_{initial} \times 100$) of CO₂ bubble size caused by CO₂ dissolution were calculated using the average initial ($D_{initial}$) and final (D_{final}) diameters.

Sample preparation and analysis. Under 100 nm Ni NPs powder (≥99% purity), phosphate buffer powder, and sodium chloride (NaCl, ≥99.5% purity) were purchased from Sigma-Aldrich (USA). One mM phosphate buffer solution (PBS) was prepared using deionized water (18.2 MΩ·cm). Experiments for the effect of the ionic strength on the change in CO₂ bubble diameter were performed at different salinities in one mM PBS. The mixtures of 30 mg L⁻¹ of Ni NPs solutions with different salinities (ionic strengths) of 0%, 2% (0.34), 4% (0.68), 6% (1.03), 8% (1.37), 10% (1.71), 15% (2.57), and 30% (5.13 M) were made to study the salinity effect on the Ni NPs' catalytic activity. These suspensions were mixed well with a Voltex mixer for 1 min before being injected into the microfluidic device. To make the mixtures of Ni NPs with DEX (Dextran, M.W. 5.0×10^5 g mol⁻¹, Alfa Aesar, USA), PVP (Poly(vinyl pyrrolidone), M.W. 5.0×10^4 g mol⁻¹, Acros Organics, USA), and CMC (Sodium carboxymethyl cellulose, M.W. 9.0×10^4 g mol⁻¹, Acros Organics, USA) at 10% NaCl, Ni NPs suspension, polymeric dispersant solutions, and solution containing salinity were prepared separately at first. Then, they were compounded by the solution to the desired concentrations of 30 mg L⁻¹ Ni NPs suspensions with the concentration range of 0.01% to 0.03% of polymer at 10% NaCl. Vortexing these mixtures at 3200 rpm was performed for homogeneity. These mixtures were used to test how efficiently Ni NPs can be dispersed depending on the type and concentration of the polymer. In support of experimental results, the ζ potential of Ni NPs and time-dependent changes in hydrodynamic particle size (d.nm) of Ni NPs with 0.03% of DEX, PVP, and CMC at different salinities (0%, 15%, and 30% NaCl) were measured by a dynamic light scattering system, Zetasizer Nano ZS90 (Malvern Instruments Ltd., UK) using a transparent and disposable zeta cell (DTS 1070, Malvern Instruments Ltd., UK) and a UV-transparent disposable quartz cuvette (Sarstedt AG & Co., Germany), respectively. The mean values of ζ potential of these samples were analyzed from hundred determinations for triplicate samples. A FTIR (Nicolet iS10, Thermo Fisher Scientific, US) was employed to study the interactions between Ni NPs with different polymers. Each drop of 30 mg L⁻¹ of Ni NPs suspension with 0.03% of DEX, PVP, and CMC was deposited onto a substrate and dried at room temperature. Then, FTIR spectra of test samples were recorded in the spectral range of 800–4000 cm⁻¹.

Statistical Analysis. All experiments were conducted in triplicate to measure the average initial and final sizes of CO₂ bubbles. The Pearson's correlation coefficient was calculated to statistically analyze the dependence of change in CO₂ bubble diameter at different salinities. The two-sample student's t-test was used to confirm the statistical significance of the degree of catalytic activity of the Ni NPs-polymer.

References

1. IPCC Climate C 2014: Impacts, adaptation, and vulnerability in *Contribution of Working Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change* (eds Field, C. B. et al.) pp. 1123 (Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 2014).
2. IPCC Climate C 2014: Mitigation of climate change in *Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change* (eds Edenhofer, O. et al.) pp. 6 (Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 2014).
3. Sundquist, E. T. et al. Carbon sequestration to mitigate climate change. *U.S. Geological Survey, Fact Sheet 2008–3097* (2008).

4. Sakti, A. W., Nishimura, Y., Sato, H. & Nakai, H. Divide-and-Conquer Density-Functional Tight-Binding Molecular Dynamics Study on the Formation of Carbamate Ions during CO₂ Chemical Absorption in Aqueous Amine Solution. *Bull. Chem. Soc. Jpn.* **90**, 1230–1235 (2017).
5. Xia, P. *et al.* Ultra-thin nanosheet assemblies of graphitic carbon nitride for enhanced photocatalytic CO₂ reduction. *J. Mater. Chem. A* **5**, 3230–3238 (2017).
6. Sohn, Y., Huang, W. & Taghipour, F. Recent progress and perspectives in the photocatalytic CO₂ reduction of Ti-oxide-based nanomaterials. *Appl. Surf. Sci.* **396**, 1696–1711 (2017).
7. Gunter, W. D., Wong, S., Cheel, D. B. & Sjoström, G. Large CO₂ Sinks: Their role in the mitigation of greenhouse gases from an international, national (Canadian) and provincial (Alberta) perspective. *G. Appl. Energy* **61**, 209–227 (1998).
8. Herzog, H. J. Peer Reviewed: What Future for Carbon Capture and Sequestration? *Environ. Sci. Technol.* **35**, 148A–153A (2001).
9. Hasan, M. M. F., Baliban, R. C., Elia, J. A. & Floudas, C. A. Modeling, simulation, and optimization of postcombustion CO₂ capture for variable feed concentration and flow rate. 2. Pressure swing adsorption and vacuum swing adsorption processes. *Ind. Eng. Chem. Res.* **51**, 15642–15664 (2012).
10. Rathnaweera, T. D., Ranjith, P. G. & Perera, M. S. A. Experimental investigation of geochemical and mineralogical effects of CO₂ sequestration on flow characteristics of reservoir rock in deep saline aquifers. *Sci. Rep.* **6**, 19362 (2016).
11. Bennaceur, K. *et al.* CO₂ capture and storage – A solution within. *Oilfield Review* 44–60 (2004).
12. Ropp, R. C. *Encyclopedia of the alkaline earth compounds*. Elsevier: Amsterdam 352–353 (2013).
13. Navsi, M. C. M., Ranjith, P. G., Sanjayam, J., Haque, A. & Li, X. Mechanical behaviour of wellbore materials saturated in brine water with different salinity levels. *Energy* **66**, 239–249 (2014).
14. Jia Liu, J. *et al.* Sintering-resistant nanoparticles in wide-mouthed compartments for sustained catalytic performance. *Sci. Rep.* **7**, 41773 (2017).
15. Chakraborty, I. & Pradeep, T. Atomically precise clusters of noble metals: Emerging link between atoms and nanoparticles. *Chem. Rev.* **117**, 8208–8271 (2017).
16. Shirai, H. *et al.* Preparation of Au/Pd bimetallic nanoparticles by a microwave-induced plasma in liquid process. *Bull. Chem. Soc. Jpn.* **90**, 279–285 (2017).
17. Zhao, X. *et al.* Polymer-supported nanocomposites for environmental application: A review. *Chem. Eng. J.* **170**, 381–394 (2011).
18. Bhaduri, G. A. & Šiller, L. Nickel nanoparticles catalyze reversible hydration of carbon dioxide for mineralization carbon capture and storage. *Catal. Sci. Tech.* **3**, 1234–1239 (2013).
19. Kanold, J. M., Wang, J., Brümmer, F. & Šiller, L. Metallic nickel nanoparticles and their effect on the embryonic development of the sea urchin *Paracentrotus lividus*. *Environ. Pollut.* **212**, 224–229 (2016).
20. Baker, T. J., Tyler, C. R. & Galloway, T. S. Impacts of metal and metal oxide nanoparticles on marine organisms. *Environ. Pollut.* **186**, 257–271 (2014).
21. Seo, S. *et al.* Microbubbles loaded with nickel nanoparticles: A perspective for carbon sequestration. *Anal. Chem.* **89**, 10827–10833 (2017).
22. Lindskog, S. Structure and mechanism of carbonic anhydrase. *Pharmacol. Ther.* **74**, 1–20 (1997).
23. Bräuer, M., Pérez-Lustres, J. L., Weston, J. & Anders, E. Quantitative reactivity model for the hydration of carbon dioxide by biomimetic zinc complexes. *Inorg. Chem.* **41**, 1454–1463 (2002).
24. Kiese, M. & Hastings, A. B. The catalytic hydration of carbon dioxide. *J. Biol. Chem.* **132**, 267–280 (1940).
25. Ahmadi, M., Gomes, V. G. & Ngian, K. Advanced modelling in performance optimization for reactive separation in industrial CO₂ removal. *Sep. Purif. Technol.* **63**, 107–115 (2008).
26. Ghosh, U. K., Kentish, S. E. & Stevens, G. W. Absorption of carbon dioxide into aqueous potassium carbonate promoted by boric acid. *Energy Procedia* **1**, 1075–1081 (2009).
27. Gerdemann, S. J. *et al.* *Ex situ* aqueous mineral carbonation. *Environ. Sci. Technol.* **41**, 2587–2593 (2007).
28. Druckenmiller, M. L. & Maroto-Valer Parametric investigation of carbon sequestration using brine. M. M. *Prepr. Pap.-Am. Chem. Soc., Div. Fuel Chem.* **49**, 418–419 (2004).
29. Rosenbauer, R. J., Koksalan, T. & Palandri, J. L. Experimental investigation of CO₂-brine-rock interactions at elevated temperature and pressure: Implications for CO₂ sequestration in deep-saline aquifers. *Fuel Process. Technol.* **86**, 1581–1597 (2005).
30. Junker, B. Corrosion in bioprocessing applications. *Bioprocess Biosyst. Eng.* **32**, 1–29 (2009).
31. Mohammadi, M. *et al.* Cyclic oxidation behavior of CoNiCrAlY coatings produced by LVPS and HVOF processes. *Trans JWRI* **40**, 53–58 (2011).
32. Khakpour, I., Soltani, R. & Sohi, M. H. Microstructure and high temperature oxidation behaviour of ZrDoped aluminide coatings fabricated on nickel-based super alloy. *Procedia Mater. Sci.* **11**, 515–521 (2015).
33. Holt, T., Jensen, J. I. & Lindeberg, E. Underground storage of CO₂ in aquifers and oil reservoirs. *Energy Convers. Mgmt. WI.* **36**, 535–538 (1995).
34. Grubbs, R. B. Roles of polymer ligands in nanoparticle stabilization. *Polym. Rev.* **47**, 197–215 (2007).
35. Ehlert, S. *et al.* Polymer ligand exchange to control stabilization and compatibilization of nanocrystals. *ACS Nano* **8**, 6114–6122 (2014).
36. Xu, C., Ohno, K., Admiral, V. & Composto, R. J. Dispersion of polymer-grafted magnetic nanoparticles in homopolymers and block copolymers. *Polymer* **49**, 3568–3577 (2008).
37. Lewis, L. N. Chemical catalysis by colloids and clusters. *Chem. Rev.* **93**, 2693–2730 (1993).
38. Aiken, J. D. & Finke, R. G. A review of modern transition-metal nanoclusters: Their synthesis, characterization, and applications in catalysis. *J. Molec. Catal. A* **145**, 1–44 (1999).
39. Astruc, D., Lu, F. & Aranzas, J. R. Nanoparticles as recyclable catalysts: the frontier between homogeneous and heterogeneous catalysis. *Angew. Chem., Int. Ed.* **44**, 7852–7872 (2005).
40. Xu, X., Wang, Q., Choi, H. C. & Kim, Y. H. Encapsulation of iron nanoparticles with PVP nanofibrous membranes to maintain their catalytic activity. *J. Membr. Sci.* **348**, 231–237 (2010).
41. Wu, L., Shamsuzzoha, M. & Ritchie, S. M. C. Preparation of cellulose acetate supported zero-valent iron nanoparticles for the dechlorination of trichloroethylene in water. *J. Nanopart. Res.* **7**, 469–476 (2005).
42. Dong, T., Luo, H., Wang, Y., Hu, B. & Chen, H. Stabilization of Fe-Pd bimetallic nanoparticles with sodium carboxymethyl cellulose for catalytic reduction of para-nitrochlorobenzene in water. *Desalination* **271**, 11–19 (2011).
43. Tumarin, E. *et al.* Temperature-controlled ‘breathing’ of carbon dioxide bubbles. *Lab Chip* **11**, 3545–3550 (2011).
44. Park, J. I., Nie, Z., Kumachev, A. & Kumacheva, E. A microfluidic route to small CO₂ microbubbles with narrow size distribution. *Soft Matter* **6**, 630–634 (2010).
45. Shim, S., Wan, J., Hilgenfeldt, S., Panchal, P. D. & Stone, H. A. Dissolution without disappearing: multicomponent gas exchange for CO₂ bubbles in a microfluidic channel. *Lab Chip* **16**, 2428–2436 (2014).
46. Abolhasani, M., Singh, M., Kumacheva, E. & Günther, A. Automated microfluidic platform for studies of carbon dioxide dissolution and solubility in physical solvents. *Lab Chip* **12**, 1611–1618 (2012).
47. Sun, R. & Cubaud, T. Dissolution of carbon dioxide bubbles and microfluidic multiphase flows. *Lab Chip* **11**, 2924–2928 (2011).
48. Lefortier, S. G. R., Hamersma, P. J., Bardow, A. & Kreutzer, M. T. Rapid microfluidic screening of CO₂ solubility and diffusion in pure and mixed solvents. *Lab Chip* **18**, 3387–3391 (2012).
49. Park, J. I. *et al.* Microbubbles loaded with nanoparticles: A route to multiple imaging modalities. *ACS Nano* **4**, 6579–6586 (2010).

50. Sell, A., Fadaei, H., Kim, M. & Sinton, D. Measurement of CO₂ diffusivity for carbon sequestration: A microfluidic approach for reservoir-specific analysis. *Environ. Sci. Technol.* **47**, 71–78 (2013).
51. Kim, M., Sell, A. & Sinton, D. Aquifer-on-a-Chip: understanding pore-scale salt precipitation dynamics during CO₂ sequestration. *Lab Chip* **13**, 2508–2518 (2013).
52. Garstecki, P., Fuerstman, M. J., Stone, H. A. & Whitesides, G. M. Formation of droplets and bubbles in a microfluidic T-junction—scaling and mechanism of break-up. *Lab Chip* **6**, 437–446 (2006).
53. Mastiani, M., Seo, S., Jimenez, S. M., Petrozzi, N. & Kim, M. M. Flow regime mapping of aqueous two-phase system droplets in flow-focusing geometries. *Colloids Surf. A* **531**, 111–120 (2017).
54. Mastiani, M., Mosavati, B. & Kim, M. M. Numerical simulation of high inertial liquid-in-gas droplet in a T-junction microchannel. *RSC Adv.* **7**, 48512–48525 (2017).
55. Seinfeld, J. H. & Pandis, S. N. Atmospheric chemistry and Physics: From air pollution to climate change. 2nd ed.; Wiley: New York (2006).
56. Sauzade, M. & Cubaud, T. Initial microfluidic dissolution regime of CO₂ bubbles in viscous oils. *Phys. Rev. E* **88**, 051001 (2013).
57. Danckwerts, P. V. *Gas-liquid reactions*. McGraw-Hill Book Company: New York 17–20 (1970).
58. Song, P., Wen, D., Guo, Z. X. & Korakianitis, T. Oxidation investigation of nickel nanoparticles. *Phys. Chem. Chem. Phys.* **10**, 5057–5065 (2008).
59. Kuerbanjiang, B. *et al.* Exchange bias of Ni nanoparticles embedded in an antiferromagnetic IrMn matrix. *Nanotechnology* **24**, 455702 (2013).
60. Derjaguin, B. & Landau, L. D. Theory of the stability of strongly charged lyophobic sols and of the adhesion of strongly charged particles in solutions of electrolytes. *Acta Phys. Chim. URSS* **14**, 633–662 (1941).
61. Verwey, E. J. W. & Overbeek, J. Th. G. *Theory of the stability of lyophobic colloids: The interaction of sol particles having an electric double layer*. Elsevier: Amsterdam 66–133 (1948).
62. French, R. A. *et al.* Influence of ionic strength, pH, and cation valence on aggregation kinetics of titanium dioxide nanoparticles. *Environ. Sci. Technol.* **43**, 1354–1359 (2009).
63. Chen, K. L. & Elimelech, M. Aggregation and deposition kinetics of fullerene (C₆₀) nanoparticles. *Langmuir* **22**, 10994–11001 (2006).
64. Plach, J. M., Elliotte, A. V. C., Droppo, I. G. & Warren, L. A. Physical and ecological controls on freshwater floc trace metal dynamics. *Environ. Sci. Technol.* **45**, 2157–2164 (2011).
65. Damodaran, V. B., Place, I. W., Kipper, M. J. & Reynolds, M. M. Enzymatically degradable nitric oxide releasing S-nitrosated dextran thiomers for biomedical applications. *J. Mater. Chem.* **22**, 23038 (2012).
66. Bryaskova, R., Pencheva, D., Nikolov, S. & Kantardjiev, T. Synthesis and comparative study on the antimicrobial activity of hybrid materials based on silver nanoparticles (AgNps) stabilized by polyvinylpyrrolidone (PVP). *J. Chem. Biol.* **4**, 185–191 (2011).
67. Wang, H. *et al.* Oriented and ordered biomimetic remineralization of the surface of demineralized dental enamel using HAP@ACP nanoparticles guided by glycine. *Sci. Rep.* **7**, 40701 (2017).

Acknowledgements

This work was supported by Janke Research Fund at Florida Environmental Studies.

Author Contributions

S.S. and M.K. designed the experiments and optimized the method. S.S., G.A.P. and K.T. carried out all the experiments and processed the data reported here. S.S., X.C. and M.K. discussed and interpreted results and commented on the manuscript. S.S. wrote the manuscript. All authors participated in the preparation and review of the manuscript.

Additional Information

Supplementary information accompanies this paper at <https://doi.org/10.1038/s41598-018-29605-1>.

Competing Interests: The authors declare no competing interests.

Publisher's note: Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons license, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons license, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons license and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this license, visit <http://creativecommons.org/licenses/by/4.0/>.

© The Author(s) 2018