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OPEN Prenatal and early life exposure to air pollution and the incidence of Kawasaki disease

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Kawasaki disease (KD) is the most common form of acquired pediatric cardiac disease in the developed world. However, its etiology is still unclear. Epidemiological studies have shown that air pollution is a plausible risk factor in stimulating oxidative stress, inducing inflammation and causing autoimmune diseases. This study aims to assess the connections between prenatal and early life air pollution exposure to the incidence of KD. The main data source of this nationwide longitudinal study was the National Health Insurance Research Database (NHIRD) of Taiwan. NHIRD was linked with Taiwan Maternal and Child Health Database to establish the link between mothers and children. In total, 4192 KD cases involving children under 6 years of age were identified between January 2004 and December 2010. Children in the control group were randomly selected at a 1:4 ratio and matched using their age and index year. Integrated data for the air pollutants were obtained from 71 Environmental Protection Agency monitoring stations across Taiwan. Patients who had main admission diagnosis of KD and subsequently received intravenous immunoglobulin treatment were defined as incidence cases. Ambient exposure, including pollutant standards index (PSI), carbon monoxide (CO), nitric oxide (NO), nitric dioxide (NO₂), and nitrogen oxide (NOx) during pregnancy were all positively associated with KD incidence. Conversely, ozone (O₃) exposure had a negative correlation. Exposure to CO, NO, NO₂ and NOx after childbirth remained consistent with regards to having a positive association with KD incidence. Exposure to PSI and O₃ after delivery displayed no significant association with KD. Both prenatal and postnatal cumulative CO, NO, NO₂, and NOx exposure had a dose dependent effect towards increasing KD incidence. Certain prenatal and early life air pollutant exposure may increase the incidence of KD.

Kawasaki disease (KD), an acute form of multi-systemic vasculitis, predominantly occurs amongst preschool children. KD is characterized by prolonged fever and systemic principal features including conjunctivitis, cervical lymphadenopathy, changes in extremities, polymorphous skin rashes, and oral mucosa changes. KD is also the most common form of acquired pediatric cardiac disease in the developed world due to the sequelae of coronary artery lesions if not treated promptly^{1,2}. The etiology of KD remains unclear. After decades of research and epidemiological observations, an environmental trigger causing an inflammatory process in genetically predisposed individuals is the most reasonable mechanism of pathogenesis³.

Epidemiological studies have shown air pollution to be a plausible risk factor in stimulating oxidative stress, inducing inflammation and causing autoimmune diseases^{4,5}. Air pollution has also been reported to be associated with Kawasaki disease⁶. Geographically evaluating exposure to air pollution parameters such as ozone (O₃), nitrogen oxides (NO_x), sulfur dioxide (SO₂), carbon monoxide (CO), particulate matter < 2.5 µm in diameter (PM2.5), and particulate matter < 10 µm in diameter (PM10) may influence the susceptibility of children with Kawasaki disease^{7,8}. However, studies regarding the association between Kawasaki disease and air pollution are sparse and their results are divergent^{6,9,10}. Furthermore, the effect of prenatal air pollution exposure on the incidence of KD has never been reported. This nationwide longitudinal study aims to assess the correlation between prenatal and early-life air pollution exposure to the development of Kawasaki disease.

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Figure 1. Composition of the study cohort.

Methods

Data source. A single-payer National Health Insurance (NHI) program with mandatory enrollment was launched in Taiwan in 1995, with a coverage rate of 99.99% of Taiwan's population. The National Health Insurance Research Database (NHIRD) was established based upon the claim data of NHI in 2002^{11,12}. Since 2015, the NHIRD has been further integrated with other health-related databases located at the Health and Welfare Data Center (HWDC) of Taiwan's Ministry of Health and Welfare (MOHW)^{11,13}. The main source of data for analysis is found in the inpatient expenditures by admission (DD) files from the NHIRD. The NHIRD was further linked with the Taiwan Maternal and Child Health Database (MCHD) maintained by Taiwan's Health Promotion Administration (HPA) to help establish the link between mothers and children. Investigators are required to conduct on-site analysis for epidemiological studies, as Taiwan's Ministry of Health and Welfare (MOHW) endeavors to protect privacy and validate its database reliability. All diagnoses in the NHIRD are coded through use of the International Classification of Diseases, Ninth Revision, Clinical Modification (ICD-9-CM) format.

Study design. Figure 1 presents the flowchart of the patient enrollment. Our study identified a cohort of 4192 Kawasaki patients who were hospitalized at an age under 6 years between January 2004 to December 2010. Patients were regarded as incidence cases of KD when their main admission diagnosis was KD (ICD-9-CM 446.1), and subsequently received treatment with intravenous immunoglobulin (IVIG, ATC code: J06BA02)¹⁴. Control groups matched through both age and index month were randomly chosen at a case-to-control ratio of 1:4. To test the robustness of our analysis, we chose another control group matched by age, index month, and gender. Patients' gender, date of diagnosis, age, maternal information and ambient air pollution where they were living were all retrieved. All subjects were followed from the index date until either the diagnosis of KD or the end of the 6-year follow-up. This study protocol was approved, and the need of informed consent was waived by the institutional review board of Taichung Veterans General Hospital (CE17178A-3). All methods were performed in accordance with the relevant guidelines and regulations of Scientific Reports.

Assessment of exposure to air pollutants. Taiwan Environmental Protection Agency (TEPA) established the Taiwan Air Quality Monitoring Network in 1994. This network continuously monitored the air quality of Taiwan with a total of 76 stations and reported the concentrations of SO_2 , NO_2 , O_3 , CO, and PM10 hourly. The selection of the site of air quality monitoring stations is established after careful planning and design based on emission of pollutants, meteorological status and distribution of air quality concentration in each area, etc. According to different monitoring purposes, monitoring stations are divided into the following 6 types: (1) general stations, containing 60 sites, were installed in populous sites that can represent the distribution of air quality of a larger area and reflect the air quality status of people's daily life; (2) traffic air quality monitoring stations, containing 6 sites, were installed at areas with heavy traffic so to reflect the air quality to which the pedestrians are exposed; (3) industrial air quality monitoring stations, containing 5 sites, were installed at areas, so to understand the influence of industrial pollution; (4) National park monitoring stations, were installed at appropriate sites in reserved areas; (5) background air quality monitoring

stations, were installed in areas with less pollution or in the windward side of the prevailing wind in the total quantity control zones. Taiwan is divided into 22 primary administrative divisions and is further divided into 368 townships. In our study, National Park monitoring stations and background air quality monitoring stations were excluded due to low density dwellings. The remaining 71 stations provided the exposure level of air pollutants in our study population. Residential postal codes of our study population scattered in 368 subdivisions. The data of each air pollutant exposure were determined by the residential postal codes of 368 townships in accordance with one air quality monitoring station. Taiwan Air Quality Monitoring Network publishes air quality report annually and provide real-time air quality indicator on the website¹⁵. In our study, the calculation results represented for each pollutants exposure were monthly average concentrations corresponding to 1-h SO2, 8-h CO, 8-h CO2, 8-h O3, 24-h PM10, 24-h PM2.5, 1-h NOx, 1-h NO and 1-h NO2 monitoring continuously. PSI, which was calculated according to five sub-indexes (SO2, NO2, O3, CO, and PM10), was presented as a scale extending from 0 (healthy) to 500 (extremely unhealthy). The air pollutant measurements, which focus on the consistency in air quality assessment, are derived from Taiwan EPA monitoring stations. As an indicator of average prenatal exposure, we used monthly average concentrations of air pollutants assigned to mother's exposure for 10 consecutive months (40 weeks) before birth and the mother's residential postal code at time of delivery. As an indicator of average early childhood, we used monthly average concentrations of air pollutants derived from the child's exposure until KD was diagnosed (index date) and the child's residential postal code at time of admission to hospital.

Statistics. We performed all analyses using the SAS statistical package (version 9.3; SAS Institute, Cary, North Carolina, USA). All quantitative data were expressed as either frequency and percentage or mean and standard deviation. Continuous variables were compared using the Student's t-test, while Pearson's chi-square test was applied for categorical data. A multiple logistic regression model was applied for adjusting the potential confounding factors between the correlation of air pollution and KD incidence. The risks for both KD and air pollution exposure amount per interquartile range were calculated to determine the dose dependent effect. To test the robustness of our analysis, we further used conditional logistic regression models to analyze the second group of controls who were matched by index month, age, and gender.

Ethics approval and consent to participate. This study protocol was approved by the institutional review board of Taichung Veterans General Hospital (CE17178A-3). Consent to participate in this study was waived by the review board.

Results

Baseline characteristics of KD and age-matched controls. A total of 4192 KD admissions below the age of 6 years were identified from our study population between January 2004 to December 2010. A total of 16,768 children (1:4) without KD, matched by age and index date were randomly selected as the control group. KD occurred mostly at the age of 0–1 year, with a significant male predominance (62%). The KD group tended to have more mothers at an advanced age, underwent a vaginal delivery, and experienced maternal allergic rhinitis as well as maternal atopic dermatitis. There were no clear disproportionate differences between birth weight and preterm delivery (Table 1). The demographic data of the second control group was summarized in the Supplement File. Regarding air pollution exposure during pregnancy, levels of exposure to PSI, CO, O₃, and NOx were higher in patients with KD than in the matched controls (p < 0.05). Interestingly, levels of exposure to PSI and O₃ were inconsistent with the previously mentioned results regarding air pollution exposure after pregnancy (Table 2).

Multiple logistic regression analysis for air pollutant and risk of KD. Table 3 summarizes the results of multiple logistic regression models of PSI, CO, O_3 , NOx, NO and NO₂ exposure both during pregnancy and after delivery. After adjustment for possible confounders, ambient exposure to PSI, CO and NOx was positively associated with KD during pregnancy; however, O_3 had a negative correlation. Additionally, ambient exposure to CO, NOx, NO and NO₂ after delivery remained a consistent positive association with KD. In contrast, exposure to PSI and O_3 after delivery revealed no significant association with KD occurrence.

Conditional logistic regression analysis of factors associated with KD admissions. To test the robustness of our analysis, we further selected another control group by matching age, index month, and gender. Alternatively, we adapted a conditional logistic regression model for confounders adjustment. Table 4 summarizes the results of conditional logistic regression models of PSI, CO, O_3 , NOx, NO and NO₂ exposure both during pregnancy and after delivery. After adjustment for possible confounders, exposure measured by PSI changed from borderline significance to non-significance. Ambient exposure to CO and NOx was still positively associated with KD during pregnancy. However, O_3 had a non-significant correlation. Interestingly, PSI after delivery became significant when analyzed by conditional logistic regression models. Ambient exposure to CO, NOx, NO and NO₂ after delivery remained a consistent positive association with KD.

Dose dependent effect of air pollution on risk of KD. The dose dependent effect of air pollution was further analyzed according to the quartiles of the cumulative amounts of air pollutants. For exposure during pregnancy, odds ratios of KD became larger as increase in cumulative amounts of pollutants. Liner trends can be observed for CO, NO, NO₂, and NOx (Fig. 2). For early life exposure, similar trends could also be observed

	Non-KD group (n = 16,768)	KD group (n=4192)	p-value		
Characteristic	n (%)	n (%)			
Neonatal age			1.000		
0-1	6868 (41)	1717 (41)			
1-2	5028 (30)	1257 (30)			
2-3	2332 (13.9)	583 (13.9)			
3-4	1180 (7)	295 (7)			
4-5	824 (4.9)	206 (4.9)			
5-6	536 (3.2)	134 (3.2)			
Neonatal gender			< 0.001		
Female	8029 (47.9)	1593 (38)			
Male	8739 (52.1)	2599 (62)			
Birth weight (g)		1	0.78		
≥2500	15,680 (93.5)	3925 (93.6)			
<2500	1088 (6.5)	267 (6.4)			
Maternal age					
< 35	14,564 (86.9)	3567 (85.1)			
≥35	2204 (13.1)	625 (14.9)			
Mode of delivery					
Vaginal delivery	10,940 (65.2)	2631 (62.8)			
Cesarean section	5828 (34.8)	1561 (37.2)			
Preterm delivery			0.46		
≥37 weeks	15,462 (92.2)	3851 (91.9)			
< 37 weeks	1306 (7.8)	341 (8.1)			
Maternal comorbidity					
Asthma	301 (1.8)	84 (2)	0.37		
Allergic rhinitis	1614 (9.6)	458 (10.9)	0.012		
Atopic dermatitis	578 (3.4)	195 (4.7)	< 0.001		

Table 1. Baseline characteristics of Kawasaki disease and control groups. *KD* Kawasaki Disease; *Non-KD*non-Kawasaki Disease.

for CO, NO, NO₂, and NOx (Fig. 3). On the contrary, both prenatal and postnatal cumulative O_3 exposure had a negative linear trend towards decreasing KD incidence (Figs. 2, 3).

Discussion

This population-based nationwide longitudinal cohort study provides preliminary evidence that exposure to certain types of air pollution, both prenatally and during early-life, may contribute to the development of KD in children.

KD has long been considered an immune disorder. The strong association between KD and allergic diseases can be demonstrated by elevated serum IgE levels in KD patients¹⁶. Activation of Th1 immune reactions such as interferon-Gamma, tumor necrosis factor-alpha, IL-1 and IL-10, as well as Th2-mediated immune reactions, such as IL-4, IL-5, and IL-13 have also been proven during the acute phase of KD^{17,18}. Furthermore, immune complexes are now considered to be playing an important role in the pathogenesis of KD¹⁹. Attention has been given to the possibility of ambient air pollutions triggering KD^{20,21}. Both NOx and CO are predominantly trafficrelated emissions^{22,23}. NOx includes odorless NO and pungently odorous NO₂. NOx is a known traffic pollutant which affects the nervous system, thus diminishing forced expiratory flow, and increasing respiratory symptoms in children^{23,24}. Human hemoglobin (Hb) has 218 times the affinity to CO than oxygen. Once carboxyhemoglobin (COHb) is formed, it reduces oxygen capacity in blood (hypoxemia), possibly causing low oxygen in tissues (hypoxia)²⁵. CO-associated ambient air pollution is shown to cause a hypoxia status resulting in cell damage in the heart, lung, muscle, and brain, along with the immune and nervous systems^{26,27}. Prenatal and/or postnatal exposures to air pollution may induce immune dysregulation, systemic inflammation and oxidative stress²⁸. Those effects hold biological plausibility for the development of KD. Exposure to traffic-related pollutants could cause higher allergic susceptibility for children early in life²⁹. Increased allergic disease susceptibility due to traffic-related air pollutants may also contribute to the risk of KD²⁹⁻³¹.

Another interesting finding of this study is that O_3 acts as a protective factor for KD in those females who are exposed during pregnancy. Ozone pollution is mainly caused by traffic-related emissions and their interaction with sunlight. Exposure to ozone results in a secondary increase in reactive oxygen species, causing an oxidative stress state in the organism which triggers an inflammatory state³². Hazards to epithelial injury in the tracheobronchial tree have been identified in those who have experienced short-term exposure, as well as exposure levels as low as 0.15 ppm ozone in monkeys³³. As described in the time-stratified case-crossover study in Taiwan, exposure to ground-level ozone may increase the risk of KD during childhood¹⁰. Maternal exposures to ozone

	Non KD group (n = 16,768)	KD group (n = 4192)	<i>p</i> value				
Level of exposure to air pollutants during pregnancy, mean (SD)							
PSI	57.16 (7.44)	57.41 (7.38)	0.05				
SO ₂ (ppb)	4.43 (1.49)	4.45 (1.49)	0.40				
CO (ppm)	0.51 (0.11)	0.52 (0.11)	< 0.001				
CO ₂ (ppm)	917.13 (815.62)	887.54 (578.42)	0.12				
O ₃ (ppb)	28.71 (3.04)	28.59 (3)	0.018				
PM10 (µg/m ³)	59.18 (13.57)	59.37 (13.37)	0.40				
PM2.5 (µg/m ³)	52.26 (66.82)	51.6 (61.98)	0.56				
NOx (ppb)	24.67 (6.76)	25.17 (7.01)	< 0.001				
NO (ppb)	6.09 (2.52)	6.28 (2.62)	< 0.001				
NO ₂ (ppb)	18.59 (4.5)	18.89 (4.65)	< 0.001				
Level of exposure to air pollutants in early life, mean (SD)							
PSI	56.49 (8.16)	56.75 (7.60)	0.05				
SO ₂ (ppb)	4.39 (1.50)	4.40 (1.45)	0.70				
CO (ppm)	0.49 (0.10)	0.50 (0.10)	< 0.001				
CO ₂ (ppm)	915.6 (1377.3)	924.7 (1474.2)	0.83				
O ₃ (ppb)	28.79 (3.80)	28.72 (2.93)	0.25				
PM10 (μg/m ³)	58.71 (15.52)	58.90(14.84)	0.48				
PM2.5 (µg/m ³)	43.47 (77.50)	44.91 (95.42)	0.37				
NOx (ppb)	23.95 (7.36)	24.47 (8.25)	< 0.001				
NO (ppb)	5.89 (2.64)	6.06 (2.65)	< 0.001				
NO ₂ (ppb)	18.06 (4.95)	18.41 (5.89)	< 0.001				

Table 2. Level of exposure to air pollutants during pregnancy and in early life. *CO* carbon monoxide, CO_2 carbon dioxide, *KD groups* Kawasaki disease, *Non-KD* non-Kawasaki disease, *NO* nitric oxide, *NO*₂ nitric dioxide, *NO*_x nitrogen oxide, *O*₃ ozone, *PM2.5* particulate matter 2.5 uM, *PM10* particulate matter 10 uM, *PSI* pollutant standards index, *SO*₂ sulfur dioxide.

	Air pollution exposure during pregnancy				Air pollution exposure after delivery			
Characteristic	OR	95% CI		p-value	OR	95% CI		p-value
PSI	1.01	1.00	1.01	0.05	1.00	1.00	1.01	0.12
CO (ppm)	1.81	1.34	2.46	< 0.001	1.64	1.18	2.27	0.003
O ₃ (ppb)	0.99	0.98	1.00	0.019	1.00	0.99	1.01	0.34
NOx (ppb)	1.01	1.01	1.02	< 0.001	1.01	1.00	1.01	0.001
NO (ppb)	1.03	1.02	1.04	< 0.001	1.02	1.01	1.03	0.003
NO ₂ (ppb)	1.01	1.01	1.02	< 0.001	1.01	1.00	1.02	0.001

Table 3. Multiple logistic regression analysis for air pollutant and risk of Kawasaki disease. Model adjusted for age, gender, birth weight, mother age, mode of delivery, preterm delivery and maternal comorbidity. *PSI* pollutant standards index, *CO* carbon monoxide, *O3* ozone, NO_x nitrogen oxide, *NO* nitric oxide, NO_2 nitric dioxide, *OR* odds ratio, *CI* confidence interval.

	Air pollution exposure during pregnancy				Air pollution exposure after delivery			
Characteristic	OR	95% CI		p-value	OR	95% CI		p-value
PSI	1.00	1.00	1.01	0.07	1.01	1.00	1.01	0.02
CO (ppm)	1.67	1.23	2.28	0.001	1.61	1.16	2.22	0.004
O ₃ (ppb)	0.99	0.98	1.01	0.26	1.00	0.99	1.01	0.61
NOx (ppb)	1.01	1.00	1.01	0.002	1.01	1.00	1.01	0.001
NO (ppb)	1.02	1.01	1.03	0.002	1.02	1.01	1.03	0.001
NO ₂ (ppb)	1.01	1.00	1.02	0.004	1.01	1.00	1.02	0.002

Table 4. Conditional logistic regression analysis for air pollutant and risk of Kawasaki disease. Model adjusted for birth weight, mother age, mode of delivery, preterm delivery and maternal comorbidity. *PSI* pollutant standards index, *CO* carbon monoxide, *O3* ozone, NO_x nitrogen oxide, *NO* nitric oxide, NO_2 nitric dioxide, *OR* odds ratio, *CI* confidence interval.



Figure 2. Quartiles of prenatal cumulative air pollutant exposure to the risk of Kawasaki disease.



Figure 3. Quartiles of early life cumulative air pollutant exposure to the risk of Kawasaki disease.

concentrations may attribute to allergic immune responses in their offspring³⁴. However, maternal exposure to ozone was negatively associated with the incidence of KD in this study. The possible explanation for this is that the effect of O_3 exposure is dose-dependent: high dosages stimulate severe oxidative stress resulting in inflammatory response and tissue injury, whereas low O_3 concentrations induce a moderate oxidative stress activating protective antioxidant pathways⁸. The reason for maternal exposure to O_3 during pregnancy reducing KD may be a result of the threshold effect, in that the hazardous effects of O_3 may not be obvious below its harmful cut-off value. Instead, low O_3 concentrations trigger maternal anti-oxidative system, leading to the protective role of the antioxidant. Further clinical and experimental studies on this relationship are still needed in order to identify any direct role O_3 may play.

Previous reports have suggested that exposure to Particulate Matter (PM) air pollution contributes to cardiovascular and pulmonary morbidity, as well as mortality, by stimulating the onset of the systemic inflammatory response^{35,36}. However, available literature has shown there are no significant associations between fine and coarse particulate matter exposure and KD occurrence^{6,9,10,37}. SO₂, PM2.5 and PM10 are mainly derived from stationary fossil combustion processes. In this study, fossil fuel combustion-related air pollution brought upon no significant effects on the incidence of KD.

Male gender is a well-known risk factor for KD³⁸. Because misclassification of gender in the database is very rare and it is easy to control in the statistical model, we did not apply "sex" as a matching factor. Moreover, we would like to observe whether the gender factor took a role in the relationship of air pollution exposure and KD incidence. Other potential confounders may also be considered for the matching process to control the confounding effects. As the same reason, we would like to control it in the statistical model rather than matching them. In the contrary, because KD incidence waxes and wanes as seasons change, it is hard to compare if we did not choose children in the control group by the same birth month¹⁴. Furthermore, those who developed KD earlier tend to have a higher exposure level (air pollution concentrations are likely to decrease along time) and a greater variation. These are reasons why we match case and control groups by index month. Although air pollution level presenting by absolute values cannot tell significant differences clearly by the number itself and standardized values convey obvious concrete concepts, continuous variable values provide more information than the categorical values that classified air pollutants level into scales. Moreover, continuous variables own more statistical power to detect differences among groups. Previous literature also presented the pollution level as original values³⁹. Only the data of PSI were presented as a scale extending from 0 (healthy) to 500 (extremely unhealthy) since PSI is calculated according to five sub-index (SO2, NO2, O3, CO, and PM10). Previous articles regarding prenatal and early life exposure to air pollution and KD demonstrated the confounding factors such as age, gender, birth weight, mother age, mode of delivery and preterm delivery, which were then applied in our study^{21,28}. Maternal comorbidity including asthma, allergic rhinitis and atopic dermatitis were selected in the list of confounders since previous literature illustrated that children born from mothers with asthma and allergic rhinitis had a higher risk of developing KD⁴⁰. In our study, we control those potential confounders in multiple logistic regression models to see the independent effect of air pollution itself.

Our study has certain limitations. Firstly, the average exposure level of air pollutants was calculated according to residential postal codes. The residential postal codes typically matched along with one block-face in urban areas but was larger in rural areas due to low population density. The study cannot determine the actual distance between the residential postal codes and air quality monitoring stations nor distinguished difference in coverage from urban to rural areas. The air quality monitoring station that assigned to each postal code was based on administrative districts, which may not be the closest to each participant. Also, migration from one city to another could lead to misclassifications in levels of exposure. Secondly, seasonal weather patterns such as temperature, humidity, and wind were not considered in our study. Seasonal weather patterns may influence the composition of environmental pollutants due to fluctuating stratified atmospheric structures⁴¹. KD incidence in Taiwan also follows a significant seasonal trend^{10,14}. Under the influence of the Asian continental anticyclone system, a strong northeasterly monsoonal flow causes more favorable conditions for disease onset, which could possibly be associated with transboundary air pollution and agents⁴². Thirdly, urbanization might be associated with easier access to healthcare, and therefore possessed a higher positive diagnosis rate. The National Health Insurance, characterized by good accessibility and comprehensive population coverage, was launched in Taiwan since 1995. In addition, Taiwan is a small island. As a result, urban-rural gap regarding medical accessibility is relatively narrow. Although higher positive diagnosis rate may be existed in urban area, it should not be a major concern. And finally, the present study has considered only relatively important air pollutants. However, some other airborne agents such as spores and microorganisms have also been reported to be associated with the occurrence of KD⁴². This study used outdoor concentration with fixed ambient monitoring system to analyze the cumulative exposure. However, the real exposure can be in indoors because most people have spent their time in indoors. Temperature and humidity should also be considered as potential confounders in multiple logistic regression models. However, the data were not recorded in those monitoring stations. Furthermore, the biological mechanism of how prenatal air pollution affects the incidence of Kawasaki disease is still unclear.

Conclusions

Certain types of prenatal and early life air pollutant exposure may increase the incidence of KD. Further prospective studies and additional data taken from countries other than Taiwan are still needed in order to elucidate on this causal relationship.

Data availability

The datasets presented in this article are not readily available because data release is not allowed by the National Health Insurance Research Database. Requests to access the datasets should be directed to Dr. Chien-Heng Lin/epid@ms39.hinet.net.

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Author contributions

N.-C.K. conceptualized and designed the study, drafted the initial manuscript, and revised the manuscript. C.-H.L. designed the data collection instruments, collected data, carried out the initial analyses, and reviewed and revised the manuscript. M.-C.L. conceptualized and designed the study, reviewed and revised the manuscript. All authors approved the final manuscript as submitted and agree to be accountable for all aspects of the work.

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Competing interests

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Additional information

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