



Respiratory Deposition Dose of PM_{2.5} and PM₁₀ Before, During and After COVID-19 Lockdown Phases in Megacity-Delhi, India

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Abstract: Considerable changes in particulate matter (PM) during COVID-19 lockdown in major cities around the World demand changes in exposure assessment studies of PM. The present study shows variations in respiratory deposition dose (RDD) of both fine (PM_{2.5}) and coarse (PM₁₀) particles before, during and after Covid-19 lockdown phases at three sites (with different pollution signatures) in Delhi—Alipur, Okhla and Pusa Road. Exposure assessment study showed mean PM_{2.5} RDD (\pm S.D.) ($\mu\text{g}/\text{min}$) for walk and sit mode during before lockdown (BL) as 2.41(\pm 1.20) and 0.84(\pm 0.42) for Alipur, 2.71(\pm 1.60) and 0.94(\pm 0.56) for Okhla, and 2.54(\pm 1.28) and 0.88(\pm 0.44) for Pusa road, which decreased drastically during Lockdown 1(L1) as 0.85(\pm 0.35) and 0.30(\pm 0.12) for Alipur, 0.83(\pm 0.33) and 0.29(\pm 0.11) for Okhla, and 0.68(\pm 0.28) and 0.23(\pm 0.10) for Pusa road, respectively. Mean PM₁₀ RDD (\pm S.D.) ($\mu\text{g}/\text{min}$) for walk and sit mode during before lockdown (BL) as 3.90 (\pm 1.73) and 1.36 (\pm 0.60) for Alipur, 4.74 (\pm 2.04) and 1.65 (\pm 0.71) for Okhla, and 4.25 (\pm 1.69) and 1.48 (\pm 0.59) for Pusa Road, respectively which decreased drastically during Lockdown 1(L1) as 2.19 (\pm 0.95) and 0.76 (\pm 0.33) for Alipur, 1.73 (\pm 0.67) and 0.60 (\pm 0.23) for Okhla and, 1.45 (\pm 0.50) and 0.50 (\pm 0.17) for Pusa Road, respectively. Significant decrease in RDD concentrations (Both PM_{2.5} and PM₁₀) than that of BL phase have been found during Lockdown 1(L1) phase and other successive lockdown and unlock phases—Lockdown 2(L2), Lockdown 3(L3), Lockdown 4(L4) and Unlock1 (UL1) phases. Changes in RDD values during lockdown phases were affected by lesser traffic emission, minimized industrial activities, biomass burning activities, precipitation activities, etc. Seasonal variations of RDD showed Delhites are found exposed to more fine and coarse particles' RDD (walk and sit modes) before and after lockdown, i.e. during normal days than during lockdown phases showing potential health effects. People in sit condition found less exposed to fine and coarse RDD comparison to those in walk condition both during normal and lockdown days.

Keywords: PM_{2.5}; PM₁₀; RDD; COVID-19

1. Introduction

The presence of atmospheric particulate matter (PM) causes seven million of premature deaths worldwide per year [1]. Ambient PM is classified under group 1 carcinogen to humans [2] and considered one among most significant environmental risks to human health [3, 4]. As per previous

empirical and experimental studies, PM is known to cause negative impacts on human health [5–7], climatic effects [8], adverse effects on ambient air quality and reduction in visibility [9, 10]. PM_{2.5} is found to be linked with more health risks than PM₁₀ [11] mainly due to its smaller size and longer residence time in the atmosphere [12]. According to a study, PM_{2.5} exposures for long-term is estimated to increase mortality caused by cardiopulmonary diseases by 6–13% per 10 $\mu\text{g}/\text{m}^3$ of PM_{2.5} [13, 14],

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whereas all-cause daily mortality associated with PM_{10} increases by 0.2–0.6% per 10 $\mu\text{g}/\text{m}^3$ of PM_{10} (WHO, [15]).

The novel Coronavirus (2019-nCoV) first reported in December 2019 at Wuhan, China [16] and found to be linked with causing over 12.80 million cases and over 570 thousands deaths worldwide till 12 July 2020 [17]. On 11 March 2020, COVID-19 has been declared a pandemic by the World Health Organization (WHO) as it spread rapidly to several countries including India [18]. COVID-19 caused more than 855 thousands cases and more than 22 thousands deaths in India till 11 July 2020 [19, 20]. COVID-19 virus which belongs to Coronaviridae family has known to cause a variety of respiratory diseases in human beings [16, 21]. A strong correlation has been studied between PM's exposure and ecotoxicity, genotoxicity, and oxidative potential of PM, and morbidity from respiratory infections [22]. Also, PM exposures for a longer period have a potential impact on exacerbation of the severity of emerging diseases due to the presence of viruses like COVID-19 and SARS variants in the ambient atmosphere [23, 24]. According to previous studies, a small increase of 1 g/m^3 of $PM_{2.5}$ exposure (fine particulates) may increase the risk of mortality due to COVID-19 virus by 20 times if human body exposed over a time period of 15–20 years [25]. The impact of PM on the increase in COVID-19 cases was studied during a research carried out in Italy where “air pollution-to-human transmission” was found as a major cause for COVID-19 transmissions than “human-to-human transmission” [25]. Studies have reported that apart from direct human-to-human contacts, high outdoor and indoor PM levels contributing to urban air pollution may have a significant impact on the increased rates of confirmed COVID-19-total number, daily new and total deaths cases [26].

The exposure assessment studies are useful for impact assessment of particulate matter and takes an account of ambient PM concentrations as well as the frequency and duration for which the PM comes in contact with human body. The air pollutants exposure varies with gender, age, socio-economic status, and pre-existing health conditions of the subject, time spent by human beings in the different micro-environment as well as different meteorological conditions [27, 28]. Both epidemiological and toxicological studies confirmed the association between air pollutants emitted from traffic-related sources and human health effects due to air pollution exposure [29, 30]. Significant associations have been confirmed by various epidemiological studies between ambient PM exposure and various cardiovascular and respiratory diseases along with the increase in their respective mortality rates [31–33]. Respiratory lung deposition dose (RDD) of PM is used for health risk assessment studies during different physical activities and modes of commuting. In an urban

environment, different modes of commuting causes different exposures to various air pollutants [34]. Commuting exposure studies for pollutants also consider the transport mode types for commuting, route opted, time of the day and fuel type used [35, 36]. Studies revealed that significant variations in PM concentrations inside the commuting vehicle are found that are closely linked with the changes taking place in the PM concentrations found outside the vehicle [37, 38]. Previous studies show deposition under moderate physical activities like walking to be three or more times greater than when at rest [39]. Therefore, there is an importance of studying RDD of ambient particulate matter to study deposition potential of PM into our lungs in different micro-environments and outdoor-environment during different physical activities (walk/sit) and modes of communication.

Studies have reported that fine PM fraction is greatly deposited in the airways during physical activities like exercise/walking as compared to rest [40]. With more intense level of physical activities, total respiratory deposition of PM will be higher which may enhance the adverse effects of PM. For example, people who exercise in a polluted urban environment are more vulnerable to air pollution effects than those who are at rest. The adverse health effects become more prominent for sensitive people like asthmatics, children, and elderly people [1]. More exposure to PM during physical activities can cause inflammation in the airways and worsen asthmatic responses [41] and may also trigger other respiratory health problems [42], cardiovascular disease [43–46], and cancer [47], leading to premature death [48]. The health risk from exposure depends on pollutants concentration entering into body as well as frequency and duration of contact of air pollutants with the body [49]. Since breathing rates define the amount of PM concentration inhaled into respiratory system, breathing rates during physical activities like walking or cycling are found to be greater than those sitting inside homes, car or bus leading to more exposure of human body to PM's RDD during physical activities than while sitting [50]. Therefore, RDDs are better indices for health risk assessment studies of PM during different modes of physical activities.

Higher levels of PM including fine ($PM_{2.5}$) and coarse (PM_{10}) have been found in the ambient atmosphere of Delhi, India [51, 52]. COVID-19 pandemic in India led to a nationwide lockdown starting from 24 March till 31 May 2020 which has caused lesser transportation activities and decreased industrial activities during lockdown event causing sudden decrease in air pollutants concentration in various parts of India [52–58]. Delhi is one of the hotspots of air pollution around the globe, and due to sudden changes in PM concentration during lockdown the change in exposure assessment due to PM during lockdown was studied. Since RDD concentration is directly linked to PM

concentration, the present work explores the changes in RDD concentrations of both fine and coarse particulate matter (PM) at different sites of Delhi having different pollution signatures before, during and after lockdown phases in Delhi, India.

2. Material and Methods

2.1. Site Description

In the present study, Megacity-Delhi has been chosen as study area where three sites have been selected including Alipur as rural, Okhla as Industrial and Pusa Road as Traffic with residential area whose details are provided in Table 1. The study site Alipur is the administrative headquarters and a sub-division of North Delhi district and lies in Northern part of Delhi. Alipur area is rural, has residential colonies and lies nearby national highway. Pusa Road site is a representative of traffic pollution in Central Delhi region with compact residential colonies in nearby region. Okhla site lies in the south-east district of Delhi nearby Okhla industrial area and connected to the border of Uttar Pradesh, State of India. Okhla industrial estate has three phases including readymade and leather garment exporters, plastic and packaging industries, pharmaceutical manufacturing units, printing presses, machinery manufacturers, call centres, MNCs Office, Bank, etc. Apart from industrial estate, Okhla also has an extension of a residential area including various residential colonies.

2.2. Data Collection and Analysis

Real-time data for PM_{2.5} and PM₁₀ have been collected from air quality monitoring stations (for three sampling sites—Alipur, Okhla, and Pusa road) which are installed and being monitored in collaboration with Central Pollution Control Board (CPCB) [continuous ambient air quality monitoring, CAAQM & manual ambient air quality monitoring, MAAQM]; DPCC (Delhi Pollution Control Committee); IITM (Indian Institute of Tropical Meteorology),

Pune; and, SAFAR (System of Air Quality and Weather Forecasting and Research). The collected data are converted into daily averaged data for concentrations ($\mu\text{g}/\text{m}^3$) of PM_{2.5} and PM₁₀ which have been used for RDD calculations. The raw data used here is available at CPCB online portal for air quality data dissemination [59] which has been analysed for air quality assessment studies for before and during the different lockdown periods of COVID-19 and unlock phases. The data analysis is in compliance with quality assurance or quality control (QA/QC) protocols which are carried out by CPCB such as timely calibration of the instruments. Data from January to June 2020 have been studied for analysing air quality, for before and during different COVID-19 lockdown and unlock phases as mentioned below:

2.2.1. Different Phases of Before Lockdown, Lockdown and Unlock during COVID-19

Before lockdown (BL): 1 January–24 March 2020 (~ 4 months);

Lockdown phase-1 (L1): 25 March–14 April 2020 (21 days);

Lockdown phase-2 (L2): 15 April–3 May 2020 (19 days).

Lockdown phase-3 (L3): 4 May–17 May 2020 (14 days);

Lockdown phase-4 (L4): 18 May–31 May 2020 (14 days).

Unlock phase-1.0 (UL1): 1 June–30 June 2020 (30 days);

For studying seasonal variations of RDD, PM_{2.5} and PM₁₀ concentration data have been analysed for the year 2019 and 2020. Seasonal variations have been studied for following four seasons—Winter (January–February), Pre-monsoon (March–May), Monsoon (June–September), Post-monsoon (October–December) classified as per India Meteorological Department [60].

2.3. Data Analysis

2.3.1. Equations for Respiratory Deposition Dose (RDD) Estimation for Fine (PM_{2.5}) and Coarse Particles (PM₁₀)

The RDD estimation method used in this study is based on the International Commission on Radiological Protection [61] method of calculation. As per previous studies, RDD has been calculated using Eqs. 1, 2 and 3, respectively [49, 62–65] as mentioned below:

$$\text{RDD of PM (size, } i; \text{ in } \mu\text{g}/\text{min}) = (V_T * f) * \text{DF}_i * \text{PM}_i \quad (1)$$

Table 1 Site description

Site name	Site details	Latitude	Longitude	Altitude
Pusa road	Traffic area	28°38'36.10"N	77°11'9.94"E	230 m
Okhla	Industrial area	28°33'44.96"N	77°17'28.17"E	204 m
Alipur	Rural-residential	28°47'49.97"N	77° 8'0.12"E	214 m

where V_T stands for the tidal volume (m^3 per breath), f represents typical breathing frequency (breath per minute), and DF_i shows deposition fraction of a size fraction i .

(Here, for $PM_{2.5}$, size fraction (i) or particle diameter (d_p) = $2.5 \mu m$; and, for PM_{10} , $d_p = 10 \mu m$.)

PM_i = Mass concentration in different size ranges (Here, mass concentrations (in $\mu g/m^3$) of $PM_{2.5}$ and PM_{10} is used for calculations of RDD for fine and coarse mode particles, respectively.)

$$DF = IF \left(0.058 + \frac{0.911}{1 + \exp(4.77 + 1.485 \ln dp)} + \frac{0.943}{1 + \exp(0.508 - 2.58 \ln dp)} \right) \quad (2)$$

$$IF = 1 - 0.5 \left(1 - \frac{1}{1 + 0.00076d_p^{2.8}} \right) \quad (3)$$

The V_T and f values used in this study mainly depend on the physical activity and gender of a person. As similar trends of RDD have been reported for both male and female [49, 65], the RDD calculation used in this study has been done using values available for males only for brevity reasons and to understand a general trend of RDD for residents of Delhi. V_T and f values used here for walk mode are $12.5 \times 10^{-4} m^3$ per breath and 20 breaths per minute, respectively, whereas for sit mode, $7.5 \times 10^{-4} m^3$ per breath and 12 breaths per minute, respectively, for male [50]. DF_i for both fine and coarse size particles used here has been calculated using particle diameter (represented as d_p) of $PM_{2.5}$ and PM_{10} , respectively [50].

3. Results and Discussion

In the present study, RDD is calculated for both fine ($PM_{2.5}$) and coarse (PM_{10}) particulate matter as discussed below:

3.1. Variations in $PM_{2.5}$ RDD for Walk and Sit Cases Before, During and After COVID-19 Lockdown

$PM_{2.5}$ RDD values have been calculated for fine mode particle diameter (d_p) = $2.5 \mu m$ (Fig. 1a–c & Table 2). The daily variation reveals higher respiratory deposition dose (RDD) for $PM_{2.5}$ before lockdown for both walk and sit cases than during lockdown and unlock conditions at all the three sites (Fig. 1a–c). Lower $PM_{2.5}$ RDD was reported during L1 phase for both walk and sit conditions with gradual increase during L2, L3, L4 and UL1 phases with slight variations among RDD for the selected sites (Fig. 1a–c). Daily, $PM_{2.5}$ RDD values were lowest during L1 comparison with BL but among different sites, reported

higher for Pusa Road site during L1 due to lower but ongoing vehicular activities as a result from lenient lockdown rules at this particular site. For Alipur site, higher daily $PM_{2.5}$ RDD values were reported during L4 due to increase in crop burning activities in nearby agricultural fields during L4 phase [70]. Mean $PM_{2.5}$ RDD (\pm Standard Deviation (S.D.)) for both walk and sit mode during BL is calculated (in $\mu g/min$) as $2.41 (\pm 1.20)$ and $0.84 (\pm 0.42)$ for Alipur, $2.71 (\pm 1.60)$ and $0.94 (\pm 0.56)$ for Okhla, and $2.54 (\pm 1.28)$ and $0.88 (\pm 0.44)$ for Pusa Road, respectively (Table 2). The values for mean $PM_{2.5}$ RDD (\pm S.D.) for both walk and sit mode decreased drastically during L1 as $0.85 (\pm 0.35)$ and $0.30 (\pm 0.12)$ for Alipur, $0.83 (\pm 0.33)$ and $0.29 (\pm 0.11)$ for Okhla, and $0.68 (\pm 0.28)$ and $0.23 (\pm 0.10)$ for Pusa Road, respectively, with increasing values in successive lockdown phases (Table 2). This signifies that RDD values for $PM_{2.5}$ were more for moving/walking persons than those sitting indoors and less for L1 phase than BL phase. Gupta and Elumalai, 2019, have reported that RDD of fine particles mostly affects alveolar region during physical activities like exercising [65]. Variations in mean $PM_{2.5}$ RDD concentrations (\pm S.D.) (in $\mu g/min$) during different phases (BL, L1, L2, L3, L4 and UL1) for both walk and sit modes are shown in Table 2.

Exposure assessment study using RDD is a major tool for health impact assessment studies, which considers the factors like physical activity of individuals, particle size and concentrations of fine and coarse particles as well as the time spent in a specific micro-environment [66, 67]. People's mobility reports also provide better indices for calculation of the time spent in various micro-environments. Google and Apple mobility reports [68, 69] confirmed that residents of Delhi have mostly resided inside their homes (outdoor activities showed 70–80% decrease in L1 from BL phase), while some people moved for groceries, retails and transit [70]. Although the RDD calculated here considers ambient $PM_{2.5}$ concentrations, the people living inside their homes may also be exposed to outdoor concentrations with windows open condition. According to a study conducted in warm season, indoor exposures to particles were found similar to that of outdoor particle exposures as a result of opening of windows for a longer periods [71]. The study shows that during lockdown (L1 phase), Delhites were exposed to lower RDD values ($PM_{2.5}$ and PM_{10}) of sit conditions, while majorly being at home (as per google/apple mobility reports) [68–70], whereas those walking outside were exposed with more RDD values than when in sit mode.

Kumar et al. [49] reported the RDD for fine particles for walk mode as $5\text{--}6 \mu g h^{-1}$ and for car mode as $1.8\text{--}2 \mu g h^{-1}$. According to a study conducted at Barcelona, higher RDD values with almost similar values were

Table 2 Mean values of PM_{2.5} and PM₁₀ RDD (µg/min) for walk and sit mode before lockdown, during different lockdown and unlock phases (BL, L1, L2, L3, L4 and UL1) at three sites in Delhi

Sampling Locations	Phases/Events	PM _{2.5} RDD for walk mode (µg/min)	PM _{2.5} RDD for sit mode (µg/min)	PM ₁₀ RDD for walk mode (µg/min)	PM ₁₀ RDD for sit mode (µg/min)
Alipur	Before	2.41 (± 1.20)	0.84 (± 0.42)	3.90 (± 1.73)	1.36 (± 0.60)
Okhla	Lockdown (BL)	2.71 (± 1.60)	0.94 (± 0.56)	4.74 (± 2.04)	1.65 (± 0.71)
Pusa Road		2.54 (± 1.28)	0.88 (± 0.44)	4.25 (± 1.69)	1.48 (± 0.59)
Alipur		0.85 (± 0.35)	0.30 (± 0.12)	2.19 (± 0.95)	0.76 (± 0.33)
Okhla	Lockdown 1 (L1)	0.83 (± 0.33)	0.29 (± 0.11)	1.73 (± 0.67)	0.60 (± 0.23)
Pusa Road		0.68 (± 0.28)	0.23 (± 0.10)	1.45 (± 0.50)	0.50 (± 0.17)
Alipur		1.10 (± 0.40)	0.38 (± 0.14)	3.20 (± 1.04)	1.11 (± 0.36)
Okhla	Lockdown 2 (L2)	0.94 (± 0.32)	0.33 (± 0.11)	2.35 (± 0.79)	0.82 (± 0.27)
Pusa road		0.97 (± 0.80)	0.34 (± 0.28)	1.08 (± 1.19)	0.72 (± 0.41)
Alipur		1.41 (± 0.44)	0.49 (± 0.15)	2.96 (± 0.93)	1.03 (± 0.32)
Okhla	Lockdown 3 (L3)	1.01 (± 0.31)	0.35 (± 0.11)	2.49 (± 0.67)	0.86 (± 0.23)
Pusa Road		1.24 (± 0.30)	0.43 (± 0.11)	2.52 (± 0.47)	0.87 (± 0.16)
Alipur		1.49 (± 0.76)	0.52 (± 0.27)	3.93 (± 1.67)	1.37 (± 0.58)
Okhla	Lockdown 4 (L4)	1.03 (± 0.53)	0.36 (± 0.18)	3.33 (± 1.37)	1.16 (± 0.48)
Pusa Road		1.00 (± 0.32)	0.35 (± 0.18)	2.87 (± 1.12)	1.00 (± 0.39)
Alipur		1.07 (± 0.22)	0.36 (± 0.10)	2.78 (± 1.33)	0.97 (± 0.46)
Okhla	Unlock 1 (UL1)	0.93 (± 0.22)	0.32 (± 0.08)	2.96 (± 1.38)	1.03 (± 0.48)
Pusa Road		0.81 (± 0.22)	0.28 (± 0.08)	2.12 (± 1.02)	0.74 (± 0.35)

observed for walk and cycle modes as $6.8 \mu\text{g h}^{-1}$ and $6.7 \mu\text{g h}^{-1}$, respectively, whereas lower RDD values were found for commuting modes like bus and car modes as $5.4 \mu\text{g h}^{-1}$ and $5.6 \mu\text{g h}^{-1}$, respectively [35]. Other studies reported highest RDD (mean ± S.D., Range $\mu\text{g h}^{-1}$) for PM_{2.5} for walk mode as $4.9 \pm 1.0, 3.7\text{--}6.1 \mu\text{g h}^{-1}$, lower RDD for bus mode as $2.7 \pm 1.1; 1.9\text{--}4. \mu\text{g h}^{-1}$ and the lowest for car mode as $1.0 \pm 0.2; 0.7\text{--}12 \mu\text{g h}^{-1}$, among various mode of commuting [72]. Kumar et al. [49] suggested that commuters have been exposed to the lowest RDD for fine particles in cars (sit mode) in comparison with other commuting modes like bus, cycling and walk despite the presence of higher concentrations of fine particles inside the cars. The reason behind the fact observed as the physical activity (sitting) of commuters inside cars is modest in comparison with those who are walking and cycling. Our study revealed the highest daily PM_{2.5} RDD before lockdown comparative to during lockdown phases (L1, L2, L3, L4, UL1), and L1 showed the lowest daily RDD both for walk and sit mode (Fig. 1a–c). Values for daily and mean RDD values (in $\mu\text{g}/\text{min}$) reported here are quite high in before lockdown conditions than the previous reported studies (in $\mu\text{g}/\text{hr}$) which shows that Delhi residents are exposed to quite higher RDD for fine particles present in the ambient concentrations during normal days. Apart from outdoor concentrations of PM, indoor conditions further increases fine particle concentrations which can enter into human lungs,

bronchiole and finally may penetrate into alveolar regions of the respiratory system. Also, use of vehicles like buses or public modes of transport are characterized by the entrance of outdoor particles and high PM concentrations as a result of natural ventilation while opening of doors during inflow and outflow of travellers at bus stations/stop points which causes re-suspension of dust particles at higher rates [72–74]. Gulliver and Briggs [75] reported that inside vehicles PM RDD increases with the related health effects.

3.2. Variations in PM₁₀ RDD for Walk and Sit Cases Before, During and After COVID-19 Lockdown

Respiratory deposition dose (RDD) values for PM₁₀ concentrations were calculated for coarse mode particle diameter (d_p) = $10 \mu\text{m}$ (Fig. 1d–f and Table 2). The study shows higher daily RDDs for PM₁₀ before lockdown for both walk and sit cases than during lockdown and unlock conditions (Fig. 1d–f). Mean PM₁₀ RDD values (in $\mu\text{g}/\text{min}$) for walk and sit mode during BL are found as $3.90 (\pm 1.73)$ and $1.36 (\pm 0.60)$ for Alipur, $4.74 (\pm 2.04)$ and $1.65 (\pm 0.71)$ for Okhla, and $4.25 (\pm 1.69)$ and $1.48 (\pm 0.59)$ for Pusa Road, respectively (Table 2). PM₁₀ mean RDD ((S.D.) in $\mu\text{g}/\text{min}$) for both walk and sit mode decreased during L1 as $2.19 (\pm 0.95)$ and $0.76 (\pm 0.33)$ for Alipur, $1.73 (\pm 0.67)$ and $0.60 (\pm 0.23)$ for Okhla, and $1.45 (\pm 0.50)$ and $0.50 (\pm 0.17)$ for Pusa Road,

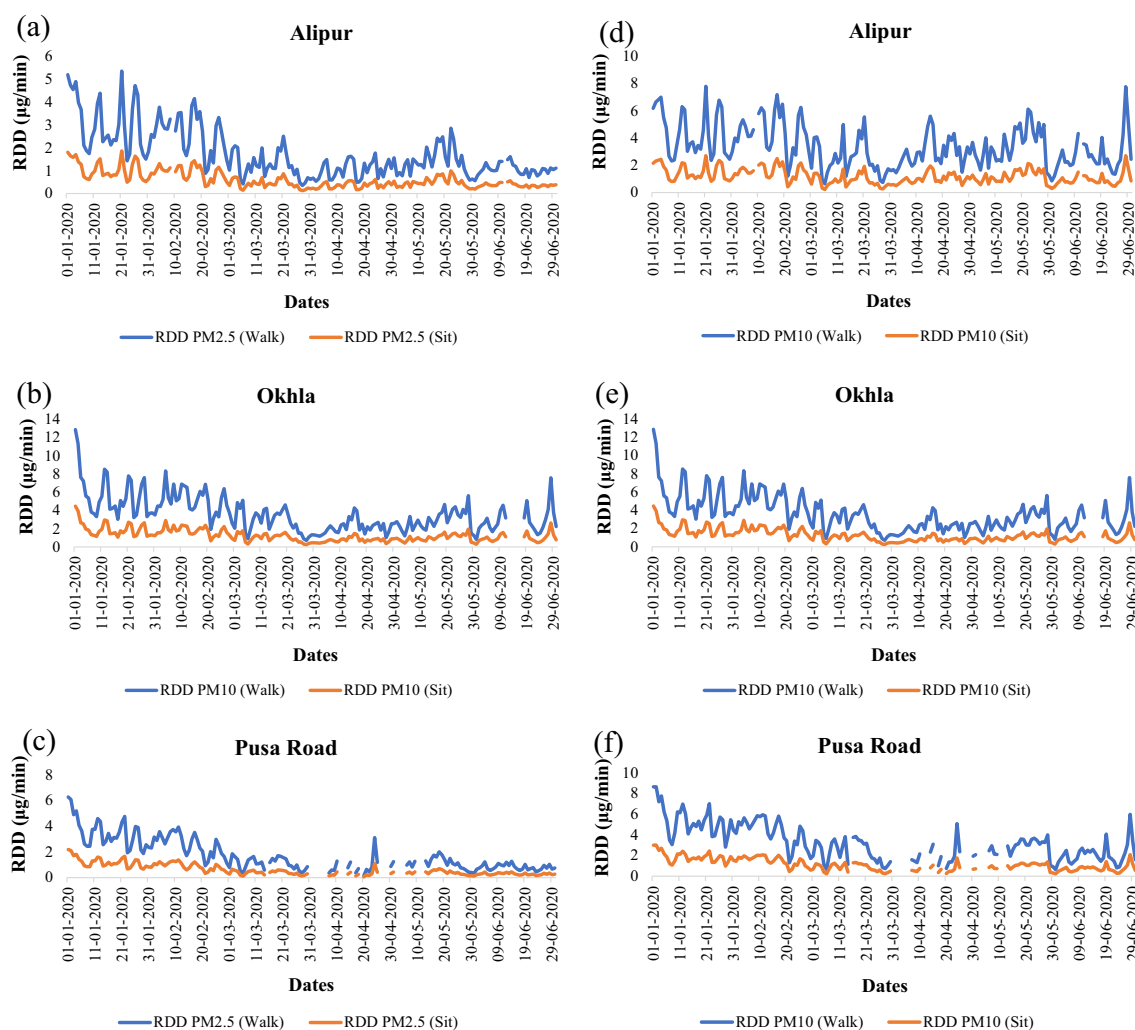


Fig. 1 Day-wise variations in RDD ($\mu\text{g}/\text{min}$) for $\text{PM}_{2.5}$ and PM_{10} (walk/sit mode) for three different sites identified by their region-specific signatures **a** Alipur **b** Okhla **c** Pusa Road before, during and after lockdown phases

respectively (Table 2). During L4 phase, Delhi (specifically Alipur site) was majorly contributed by biomass burning activities [70]. Both fine and coarse mode RDD values found to decrease during UL1 phase against L4 due to lesser biomass burning activities and more precipitation events (Table 2). Among site-wise variations, higher RDD values for both walk and sit mode were found for Okhla before lockdown, whereas for Alipur, higher RDD values found during whole lockdown period than the other sites (Table 2). The reason may be attributed to higher industrial operations at Okhla during BL phase which was minimized during lockdown phase (L1). Alipur being a rural area with sources like biomass burning from agricultural fields, vehicular emissions from highways, local waste burning, emissions from landfill dumping site present nearby shows comparatively lesser effect of lockdown conditions than the other sites in terms of PM_{10} RDD. Site-wise variations in mean PM_{10} RDD

concentrations (\pm S.D.) (in $\mu\text{g}/\text{min}$) during different phases (BL, L1, L2, L3, L4 and UL1) for both walk and sit modes have been shown in Table 2. Among site-wise variations during BL phase, Okhla being an industrial area and thus having higher PM_{10} concentrations has higher RDD values followed by Pusa Road (Traffic contribution) and the lowest RDD for Alipur (Rural area). The study shows higher RDD values of coarse PM than the fine PM for both walk and sit mode during all phases (BL, L1, L2, L3, L4 and UL1) (Table 2).

Kumar et al. [49] reported the highest mean RDD for coarse particles (PM_{10}) during the walk and car mode as $40\text{--}66 \mu\text{g h}^{-1}$ and $0.8\text{--}1.3 \mu\text{g h}^{-1}$, respectively. Emissions from traffic and re-suspension of dust particles during walking can increase coarse particle exposure during walk mode to the pedestrians [49]. The highest value of daily RDD (in $\mu\text{g}/\text{min}$) during before lockdown conditions for both walk and sit mode (Fig. 1d–f) is quite high in

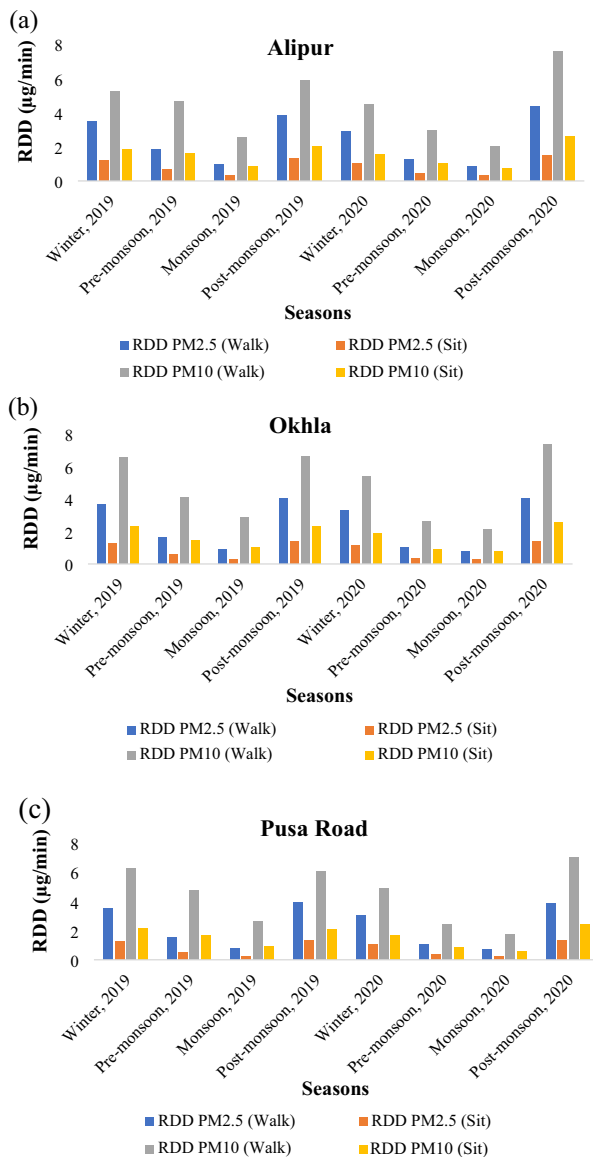


Fig. 2 Seasonal variations in RDD ($\mu\text{g}/\text{min}$) for PM_{2.5} and PM₁₀ (walk/sit mode) for three different sites **a** Alipur **b** Okhla **c** Pusa Road

comparison with the other reported studies (in $\mu\text{g}/\text{hr}$) showing higher exposure of Delhi residents towards coarse particle pollution during normal days.

3.3. Seasonal Variation of PM_{2.5} RDD and PM₁₀ RDD at Delhi

Figure 2 shows seasonal variations in RDD ($\mu\text{g}/\text{min}$) for PM_{2.5} and PM₁₀ (walk/sit mode) for three different sites—Alipur, Okhla and Pusa Road. The highest PM_{2.5} RDD and PM₁₀ RDD concentrations ($\mu\text{g}/\text{min}$) were reported for Post-monsoon and Winter season during year 2019 and 2020 at all the three sites for both walk/sit modes (Fig. 2a–c). In Delhi, Post-monsoon season is mainly characterized by

high biomass burning and lower temperature with high relative humidity contents in atmosphere leading to higher PM concentrations thus affecting PM's RDD. Lower concentrations of both PM_{2.5} and PM₁₀ RDD ($\mu\text{g}/\text{min}$) during Pre-monsoon season, 2020 in comparison with that during 2019, are attributed to the effect of COVID-19 lockdown for both walk/sit modes at all the three sites (Fig. 2a–c). This shows significant reduction in PM's RDD values (despite similar meteorological conditions due to same seasons) as a result on stringent rules during lockdown conditions.

4. Conclusion

The study shows higher daily and mean respiratory deposition dose values (RDD) for PM_{2.5} and PM₁₀ before COVID-19 lockdown for both walk and sit cases than during lockdown and unlock conditions. This signifies that Delhites are exposed to higher particulate matter associated RDD values during normal days as in before lockdown conditions, whereas the RDD exposure decreased during COVID-19 lockdown conditions. Also, RDD values for walk mode found higher than the sit mode showing extent of RDD inhaled during physical activities like walking and commuting outside were higher than sitting inside homes. Google and Apple mobility reports also confirmed that Delhites were mostly remained inside their homes thus affected by PM's RDD in sit mode because of less outdoor activities during lockdown conditions. Values for RDD ($\mu\text{g}/\text{min}$) reported here are quite high in before lockdown condition than the previous reported studies ($\mu\text{g}/\text{hr}$) which shows that Delhi residents are exposed to higher RDD values for both fine and coarse particles present in the ambient concentrations during normal days. During COVID-19 lockdown, RDD decreased drastically at all the three sites, whereas site-wise comparison during lockdown phases showed higher RDD values for Pusa Road site during L1 due to lesser but ongoing vehicular activities at the site and for Alipur site during L4 due to crop burning activities in nearby fields. Okhla site was found to have lowest RDD values during L1 due to minimized industrial activities. Both fine and coarse mode RDD values found to decrease during UL1 phase against L4 due to less biomass burning activities and precipitation events leading to suppression of ambient particulate matter concentrations (PM_{2.5} and PM₁₀) and hence lowering-related RDD. Seasonal variation in PM_{2.5} RDD and PM₁₀ RDD values showed higher concentration during Post-monsoon and Winter seasons (2019, 2020) for all the three sites as a result from source contribution and meteorological conditions. Pre-monsoon season, 2020, was characterized with lower RDD in comparison with that during 2019 due to

stringent rules on source emissions during lockdown conditions. Since, changes in RDD are directly linked to PM concentrations and physical activities, this study shows significant role of RDD to determine potential health effects related to PM exposure during different physical activities (walk/sit).

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Declarations

Conflict of interest The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

References

- [1] WHO. 2018. Managing epidemics, key facts about the major deadly disease. Report by the World Health Organization.
- [2] D. Loomis, Y. Grosse, B. Lauby-Secretan, F. El Ghissassi, V. Bouvard, L. Benbrahim-Tallaa, N. Guha, R. Baan, H. Mattock and K. Straif, International agency for research on cancer monograph working group IARC. *Lancet Oncology*, *14* (2013) 1262–1263.
- [3] Greenbaum, D. 2010. HEI panel on the health effects of traffic-related air pollution. in traffic-related air pollution: a critical review of the literature on emissions, exposure, and health effects; HEI special report 17. 386; Health Effects Institute: Boston, MA, USA, 2010.
- [4] P. Landrigan, Air Pollution and Health. *Lancet Public Health*, *2* (2017) e4–e5.
- [5] J. Guo, Y. Miao, Y. Zhang et al., The climatology of Planetary Boundary Layer height in China derived from radiosonde and reanalysis data. *Atmospheric Chemistry & Physics*, *16* (2016) 13309–13319.
- [6] L. Prieto-Parra, K. Johannessen, C. Brea, D. Vidal, C.A. Ubilla and P. Ruiz-Rudolph. Air pollution, PM_{2.5} composition, source factors, and respiratory symptoms in asthmatic and nonasthmatic children in Santiago, Chile. *Environment International*, *101* (2017) 190–200.
- [7] C.F. Wu, F.H. Shen, Y.R. Li, T.M. Tsao, M.J. Tsai, C.C. Chen, Hwang et al., Association of short-term exposure to fine particulate matter and nitrogen dioxide with acute cardiovascular effects. *Science of the Total Environment*, *569–570* (2016) 300–305.
- [8] C. Fang, Z. Zhang, M. Jin, P. Zou and J. Wang, Pollution characteristics of PM_{2.5} aerosol during haze periods in Changchun, China. *Aerosol and Air Quality Research*, *17* (2017) 888–895.
- [9] J. Wang, B. Zhao, S. Wang et al., Particulate matter pollution over China and the effects of control policies. *Science of the Total Environment*, *584–585* (2017) 426–447.
- [10] X.J. Zhao, P.S. Zhao, J. Xu et al., Analysis of a winter regional haze event and its formation mechanism in the North China Plain. *Atmospheric Chemistry and Physics*, *13* (2013) 5685–5696.
- [11] M.S. Hassanvand, K. Naddafi, H. Kashani et al., Short-term effects of particle size fractions on circulating biomarkers of inflammation in a panel of elderly subjects and healthy young adults. *Environmental Pollution*, *223* (2017) 695–704.
- [12] F. Dominici, M. Greenstone and C.R. Sunstein, Particulate matter matters. *Science*, *344* (2014) 257–259.
- [13] D. Krewski, M. Jerrett, R.T. Burnett, R. Ma, E. Hughes, Y. Shi, M.C. Turner, C.A. Pope, G. Thurston, E.E. Calle, M.J. Thun, B. Beckerman, P. DeLuca, N. Finkelstein, K. Ito, D.K. Moore, K.B. Newbold, T. Ramsay, Z. Ross, H. Shin and B. Tempalski, Extended follow-up and spatial analysis of the American cancer society linking particulate air pollution and mortality. *Research Report* (health Effect Institute), *140* (2009) 5–114.
- [14] C.A. Pope, R.T. Burnett III., M.J. Thun, E.E. Calle, D. Krewski, K. Ito and G.D. Thurston, Lung cancer, cardiopulmonary mortality, and long-term exposure to fine particulate air pollution. *JAMA*, *287* (2002) 1132–1141. <https://doi.org/10.1001/jama.287.9.1132>.
- [15] World Health Organization (WHO). 2006. Air Quality Guidelines. Global update 2005. Particulate Matter, Ozone, Nitrogen Dioxide and Sulfur Dioxide. 22. (http://apps.who.int/iris/bitstream/10665/69477/1/WHO_SDE_PHE_OEH_06.02_eng.pdf)
- [16] Q. Bukhari and Y. Jameel, Will coronavirus pandemic diminish by summer? *SSRN Electronic Journal* (2020). <https://doi.org/10.2139/ssrn.3556998>.
- [17] <https://worldmeters.info/coronavirus>
- [18] WHO, 2020. (<https://www.who.int/emergencies/diseases/novel-coronavirus-2019>).
- [19] COVID/Tracker (<https://www.covid19india.org/>).
- [20] S. Kumar, Effect of meteorological parameters on spread of COVID-19 in India and air quality during lockdown. *Science of the Total Environment*, *745* (2020) 141021.
- [21] S.R. Weiss and S. Navas-Martin, Coronavirus pathogenesis and the emerging pathogen severe acute respiratory syndrome coronavirus. *Microbiology and Molecular Biology Reviews.*, *69* (2005) 635–664.
- [22] S. Romano, M.R. Perrone, S. Becagli, M.C. Pietrogrande, M. Russo, R. Caricato and M.G. Lionetto, Ecotoxicity, genotoxicity, and oxidative potential tests of atmospheric PM10 particles. *Atmospheric Environment*, *221* (2020) 117085.
- [23] Y. Cui, Z. Zhang, J.R. Froines, J. Zhao, H. Wang, S.-Z. Yu and R. Detels, Air pollution and case fatality of SARS in the People's Republic of China: an ecologic study. *Environmental Health*, *2* (2003) 15.
- [24] Wu, X., Nethery, R.C., Sabath, B.M., Braun, D., & Dominici, F. 2020. Exposure to air pollution and COVID-19 mortality in the United States: A nationwide cross-sectional study. *MedRxiv*.
- [25] M. Coccia, Factors determining the diffusion of COVID-19 and suggested strategy to prevent future accelerated viral infectivity similar to COVID. *Science of the Total Environment*, *729* (2020) 138474.
- [26] M.A. Zoran, R.S. Savastru, D.M. Savastru and M.N. Tautan, Assessing the relationship between surface levels of PM_{2.5} and PM10 particulate matter impact on COVID-19 in Milan, Italy. *Science of the Total Environment*, *738* (2020) 139825.
- [27] P. Pant, S.K. Guttikunda and R.E. Peltier, Exposure to particulate matter in India: A synthesis of findings and future directions. *Environmental Research*, *147* (2016) 480–496. <https://doi.org/10.1016/j.envres.2016.03.011>.
- [28] C.A. Pope III., Epidemiology of fine particulate air pollution and human health: Biologic mechanisms and who's at risk? *Environmental Health Perspectives*, *108* (2000) 713–723.
- [29] WHO. 2005. Air Quality Guidelines for Particulate Matter, Ozone, Nitrogen Dioxide and Sulfur Dioxide. 22. (http://apps.who.int/iris/bitstream/10665/69477/1/WHO_SDE_PHE_OEH_06.02_eng.pdf).
- [30] Review of evidence on health aspects of air pollution—REVI-HAAP. 2013. World Health Organisation, Regional Office for

- Europe.33 (http://www.euro.who.int/__data/assets/pdf_file/0020/182432/e96762-final.pdf).
- [31] A. Bhatnagar, Environmental cardiology: studying mechanistic links between pollution and heart disease. *Circulation Research*, 99 (2006) 692–705.
- [32] I. Bos, L. Jacobs, T.S. Nawrot, B. de Geus, R. Torfs, L. Int Panis, B. Degraeuwe and R. Meeusen, No exercise-induced increase in serum BDNF after cycling near a major traffic road. *Neuroscience Letters*, 500 (2011) 129–132.
- [33] L. Jacobs, T.S. Nawrot, B. de Geus, R. Meeusen, B. Degraeuwe, A. Bernard, M. Sughis, B. Nemery and L.I. Panis, Subclinical responses in healthy cyclists briefly exposed to traffic-related air pollution: an intervention study. *Environmental Health*, 9 (2010) 64.
- [34] A. Goel and P. Kumar, A review of fundamental drivers governing the emissions, dispersion and exposure to vehicle-emitted nanoparticles at signalised traffic intersections. *Atmospheric Environment*, 97 (2014) 316–331.
- [35] A. de Nazelle, S. Fruin, D. Westerdahl, D. Martinez, A. Ripoll, N. Kubesch and M. Nieuwenhuijsen, A travel mode comparison of commuters' exposures to air pollutants in Barcelona. *Atmospheric Environment*, 59 (2012) 151–159.
- [36] M. Zuurbier, G. Hoek, M. Oldenwening, V. Lenters, K. Meliefste, P. van den Hazel and B. Brunekreef, Commuters' exposure to particulate matter air pollution is affected by mode of transport, fuel type, and route. *Environmental Health Perspectives*, 118 (2010) 783–789.
- [37] L.I. Panis, B. de Geus, G. Vandenbulcke, H. Willems, B. Degraeuwe, N. Bleux, V. Mishra, I. Thomas and R. Meeusen, Exposure to particulate matter in traffic: A comparison of cyclists and car passengers. *Atmospheric Environment*, 44 (2010) 2263–2270.
- [38] A. Goel and P. Kumar, Characterisation of nanoparticle emissions and exposure at traffic intersections through fast-response mobile and sequential measurements. *Atmospheric Environment*, 107 (2015) 374–390.
- [39] D.J. Briggs, K. de Hoogh, C. Morris and J. Gulliver, Effects of travel mode on exposures to particulate air pollution. *Environ. Int.*, 34 (2008) 12–22.
- [40] C.C. Daigle, D.C. Chalupa, F.R. Gibb, P.E. Morrow, G. Oberdorster, M.J. Utell and M.W. Frampton, Ultrafine particle deposition in humans during rest and exercise. *Inhalation Toxicology*, 15 (2003) 539–552.
- [41] Z. Zhao, R. Chen, Z. Lin, J. Cai, Y. Yang, D. Yang, D. Norback and H. Kan, Ambient carbon monoxide associated with alleviated respiratory inflammation in healthy young adults. *Environmental Pollution*, 208 (2016) 294–298.
- [42] R. McConnell, K. Berhane, F. Gilliland, S.J. London, T. Islam, W.J. Gauderman, E. Avol, H.G. Margolis and J.M. Peters, Asthma in exercising children exposed to ozone: A cohort study. *Lancet*, 359 (2002) 386–391.
- [43] A. Berger, W. Zareba, A. Schneider, R. Ruckerl, A. Ibaldueli, J. Cyrys, H.-E. Wichmann and A. Peters, Runs of ventricular and supraventricular tachycardia triggered by air pollution in patients with coronary heart disease. *Journal of Occupational and Environmental Medicine*, 48 (2006) 1149–1158.
- [44] R.J. Delfino, C. Sioutas and S. Malik, Potential role of ultrafine particles in associations between airborne particle mass and cardiovascular health. *Environmental Health Perspectives*, 113 (2005) 934–946.
- [45] M.R. Heal, P. Kumar and R.M. Harrison, Particles, air quality, policy and health. *Chemical Society Reviews*, 41 (2012) 6606–6630.
- [46] T.S. Nawrot, L. Perez, N. Kunzli, E. Munters and B. Nemery, Public health importance of triggers of myocardial infarction: A comparative risk assessment. *Lancet*, 377 (2011) 732–740.
- [47] L.V. Giles and M.S. Koehle, The health effects of exercising in air pollution. *Sports Medicine*, 44 (2014) 223–249.
- [48] R. Beelen, O. Raaschou-Nielsen, M. Stafoggia, Z.J. Andersen, G. Weinmayr, B. Hoffmann, K. Wolf, E. Samoli, P. Fischer, M. Nieuwenhuijsen, P. Vineis, W.W. Xun, K. Katsouyanni, K. Dimakopoulou, A. Oudin, B. Forsberg, L. Modig, A.S. Havulinna, T. Lanki, A. Turunen, B. Oftedal, W. Nystad, P. Nafstad, U. De Faire, N.L. Pedersen, C.G. Ostenson, L. Fratiglioni, J. Penell, M. Korek, G. Pershagen, K.T. Eriksen, K. Overvad, T. Ellermann, M. Eeftens, P.H. Peeters, K. Meliefste, M. Wang, B. Bueno-De-Mesquita, D. Sugiri, U. Kramer, J. Heinrich, K. De Hoogh, T. Key, A. Peters, R. Hampel, H. Concin, G. Nagel, A. Ineichen, E. Schaffner, N. Probst-Hensch, N. Kunzli, C. Schindler, T. Schikowski, M. Adam, H. Phuleria, A. Vilier, F. Clavel-Chapelon, C. Declercq, S. Gironi, V. Krogh, M.Y. Tsai, F. Ricceri, C. Sacerdote, C. Galassi, E. Migliore, A. Ranzani, G. Cesaroni, C. Badaloni, F. Forastiere, I. Tamayo, P. Amiano, M. Dorronsoro, M. Katsoulis, A. Trichopoulos, B. Brunekreef and G. Hoek, Effects of long-term exposure to air pollution on natural-cause mortality: An analysis of 22 European cohorts within the multicentre ESCAPE project. *The Lancet*, 383 (2014) 785–795.
- [49] P. Kumar, I. Rivas, A.P. Singh, V.J. Ganesh, M. Ananya and H.C. Frey, Dynamics of coarse and fine particle exposure in transport Microenvironments. *Npj-Climate and Atmospheric Science*, 1 (2018) 11.
- [50] Hinds, W.C. 1999. *Aerosol Technology: Properties, Behaviour and Measurement of Airborne Particles* 483 (John Wiley & Sons, UK, 1999).
- [51] P. Kumar, I. Rivas and L. Sachdeva, Exposure of in-pram babies to airborne particles during morning drop-in and afternoon pick-up of school children. *Environmental Pollution*, 224 (2017) 407–420.
- [52] S. Sharma, M. Zhang, J. Gao, H. Zhang and S.H. Kota, Effect of restricted emissions during COVID-19 on air quality in India. *Science of the Total Environment*, 728 (2020) 138878.
- [53] B. Ambade, S. Kurwadkar, T.K. Sankar and A. Kumar, Emission reduction of black carbon and polycyclic aromatic hydrocarbons during COVID-19 pandemic lockdown. *Air Quality, Atmosphere & Health*, 14 (2021a) 1081–1095.
- [54] B. Ambade, T.K. Sankar, A. Kumar, A.K. Gautam and S. Gautam, COVID-19 lockdowns reduce the Black carbon and polycyclic aromatic hydrocarbons of the Asian atmosphere: source apportionment and health hazard evaluation. *Environment, Development and Sustainability*, 3 (2021b) 1–20.
- [55] A. Chauhan and R.P. Singh, Decline in PM_{2.5} concentrations over major cities around the world associated with COVID-19. *Environmental Research*, 187 (2020) 109634.
- [56] A.K. Srivastava, P.D. Bhojwar, V.P. Kanawade, P.C.S. Devara, A. Thomas and V.K. Soni, Improved air quality during COVID-19 at an urban megacity over the Indo-Gangetic Basin: From stringent to relaxed lockdown phases. *Urban Climate*, 36 (2021) 100791.
- [57] A. Thomas, V.P. Kanawade, C. Sarangi and A.K. Srivastava, Effect of COVID-Shutdown on aerosol direct radiative forcing over the Indo-Gangetic Plain out flow region of the Bay of Bengal. *Science of the Total Environment* (2021). <https://doi.org/10.1016/j.scitotenv.2021.146918>.
- [58] M. Zhang, A. Katiyar, S. Zhu, J. Shen, M. Xia, J. Ma, S.H. Kota, P. Wang and H. Zhang, Impact of reduced anthropogenic emissions during COVID-19 on air quality in India. *Atmospheric Chemistry Physics*, 21 (2021) 4025–4037.
- [59] <https://app.cpcbcr.com/cct/#/caaqm-dashboard-all/caaqm-lan ding>
- [60] <https://www.imdpune.gov.in/Weather/Reports/glossary.pdf>

- [61] ICRP publication 66: human respiratory tract model for radiological protection. A Report of a Task Group of the International Commission on Radiological Protection. 1–482. (<http://www.icrp.org/publication.asp?id=icrp%20publication%2066>) (1994).
- [62] F. Azarmi and P. Kumar, Ambient exposure to coarse and fine particle emissions from building demolition. *Atmospheric Environment*, *137* (2016) 62–79.
- [63] P. Kumar and A. Goel, Concentration dynamics of coarse and fine particulate matter at and around the signalised traffic intersections. *Environmental Science: Processes & Impacts*, *18* (2016) 1220–1235.
- [64] B. Segalin, P. Kumar, K. Micadei, A. Fornaro and F.L. Gonçalves, Size-segregated particulate matter inside residences of elderly in the Metropolitan Area of Sao Paulo, Braz. *Atmospheric Environment*, *148* (2017) 139–151. <https://doi.org/10.1016/j.atmosenv.2016.10.004>.
- [65] S.K. Gupta and S.P. Elumalai, Size-segregated particulate matter and its association with respiratory deposition doses among outdoor exercisers in Dhanbad City, India. *Journal of the Air & Waste Management Association*, *67* (2017) 1137–1145.
- [66] A. Franzen, C. Van Landingham, T. Greene, K. Plotzke and R. Gentry, A global human health risk assessment for Decamethylcyclpentasiloxane (D5). *Regulatory Toxicology and Pharmacology*, *74* (2016) S25–S43.
- [67] B.J. Tunno, R. Dalton, L. Cambal, F. Holguin, P. Liroy and J.E. Clougherty, Indoor source apportionment in urban communities near industrial sites. *Atmospheric Environment*, *139* (2016) 30–36.
- [68] <https://www.apple.com/covid19/mobility>
- [69] <https://www.google.com/covid19/mobility>
- [70] S. Fatima, A. Ahlawat, S.K. Mishra, M. Maheshwari and V.K. Soni, Variations and source apportionment of PM_{2.5} and PM₁₀, before and during COVID-19 lockdown phases in Delhi, India. *MAPAN-J. Metrol. Soc India* (2021). <https://doi.org/10.1007/s12647-021-00506-5>.
- [71] Y. Zhu, J. Xie, F. Huang and L. Cao, Association between short-term exposure to air pollution and COVID-19 infection: Evidence from China. *Science of the Total Environment*. (2020). <https://doi.org/10.1016/j.scitotenv.2020.138704>.
- [72] I. Rivas, P. Kumar and A. Hagen-Zanker, Exposure to air pollutants during commuting in London: are there inequalities among different socio-economic groups? *Environ. Int.*, *101* (2017) 143–157.
- [73] H.S. Adams, M.J. Nieuwenhuijsen, R.N. Colvile, M.J. Older and M. Kendall, Assessment of road users' elemental carbon personal exposure levels, London, UK. *Atmospheric Environment*, *36* (2002) 5335–5342.
- [74] T. Moreno, C. Reche, I. Rivas, M.C. Minguillón, V. Martins, C. Vargas, G. Buonanno, J. Parga, M. Pandolfi, M. Brines, M. Ealo, A.S. Fonseca, F. Amato, G. Sosa, M. Capdevila, E. de Miguel, X. Querol and W. Gibbons, Urban air quality comparison for bus, tram, subway and pedestrian commutes in Barcelona. *Environmental Research*, *142* (2015) 495–510.
- [75] J. Gulliver and D.J. Briggs, Personal exposure to particulate air pollution in transport Micro environments. *Atmospheric Environment*, *38* (2004) 1–8.

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