Mercury and Selenium in Stranded Indo-Pacific Humpback Dolphins and Implications for Their Trophic Transfer in Food Chains



Duan Gui¹, Ri-Qing Yu², Yong Sun¹, Laiguo Chen³, Qin Tu¹, Hui Mo⁴, Yuping Wu¹*

1 Guangdong Provincial Key Laboratory of Marine Resources and Coastal Engineering, School of Marine Sciences, Sun Yat-Sen University, Guangzhou, China, 2 Department of Biology, University of Texas at Tyler, Tyler, Texas, United States of America, **3** Urban Environment and Ecology Research Center, South China Institute of Environmental Sciences (SCIES), Ministry of Environmental Protection, Guangzhou, China, **4** South China Botanical Garden, Chinese Academy of Sciences, Guangzhou, China

Abstract

As top predators in the Pearl River Estuary (PRE) of China, Indo-Pacific humpback dolphins (*Sousa chinensis*) are bioindicators for examining regional trends of environmental contaminants in the PRE. We examined samples from stranded *S. chinensis* in the PRE, collected since 2004, to study the distribution and fate of total mercury (THg), methylmercury (MeHg) and selenium (Se) in the major tissues, in individuals at different ages and their prey fishes from the PRE. This study also investigated the potential protective effects of Se against the toxicities of accumulated THg. Dolphin livers contained the highest concentrations of THg (32.34±58.98 μ g g⁻¹ dw) and Se (15.16±3.66 μ g g⁻¹ dw), which were significantly different from those found in kidneys and muscles, whereas the highest residue of MeHg (1.02±1.11 μ g g⁻¹ dw) was found in dolphin muscles. Concentrations of both THg and MeHg in the liver, kidney and muscle of dolphins showed a significantly positive correlation with age. The biomagnification factors (BMFs) of inorganic mercury (Hg_{inorg}) in dolphin livers (350×) and MeHg in muscles (18.7×) through the prey fishes were the highest among all three dolphin tissues, whereas the BMFs of Se were much lower in all dolphin tissues. The lower proportion of MeHg in THg and higher Se/THg ratios in tissues were demonstrated. Our studies suggested that *S. chinensis* might have the potential to detoxify Hg via the demethylation of MeHg and the formation of tiemannite (HgSe) in the liver and kidney. The lower threshold of hepatic THg concentrations for the equimolar accumulation of Se and Hg in *S. chinensis* suggests that this species has a greater sensitivity to THg concentrations than is found in striped dolphins and Dall's porpoises.

Citation: Gui D, Yu R-Q, Sun Y, Chen L, Tu Q, et al. (2014) Mercury and Selenium in Stranded Indo-Pacific Humpback Dolphins and Implications for Their Trophic Transfer in Food Chains. PLoS ONE 9(10): e110336. doi:10.1371/journal.pone.0110336

Editor: Dwayne Elias, Oak Ridge National Laboratory, United States of America

Received March 25, 2014; Accepted September 18, 2014; Published October 13, 2014

Copyright: © 2014 Gui et al. This is an open-access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Data Availability: The authors confirm that all data underlying the findings are fully available without restriction. All relevant data are within the paper and its Supporting Information files.

Funding: This research was supported by the National Natural Science Foundation of China (Grant no. 41276147); a Science and Research Project of the Marine Non-profit Industry (Grant no. 201105011-5); the Ocean Park Conservation Foundation, Hong Kong; the National Key Technology R & D Program (Grant no. 2011BAG07B05-3); and the Sousa chinensis Conservation Action Project from the Administrator of Ocean and Fisheries of Guangdong Province, China. The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing Interests: The authors have declared that no competing interests exist.

* Email: exwyp@mail.sysu.edu.cn

Introduction

Mercury (Hg) in its inorganic form is a ubiquitous pollutant that is globally distributed by atmospheric transportation. Less toxic Hg(II) in environments can easily be converted into the highly toxic methylmercury (MeHg), primarily by sulfate- and ironreducing bacteria and methanogens [1–7] through the putative Hg methylation genes hgcA and hgcB via the acetyl CoA pathway [5]. MeHg is a strong neurotoxic substance; once it is bioavailable in aquatic ecosystems, it can be bioaccumulated and biomagnified quickly through aquatic food webs, which creates a health threat to humans and aquatic mammals such as dolphins [3].

The estuary of the Pearl River, which is the third longest river in China, is a traditional nursery for fisheries and provides an ideal habitat for Indo-Pacific humpback dolphins (*Sousa chinensis*). The Pearl River Estuary (PRE) region contains a group of cities that include Guangzhou, Shenzhen and Hong Kong, forming one of the largest local and global economic hubs in southern China. Rapid industrial development and urbanization in recent decades have undermined the habitats of local fish and dolphins in the estuary. S. chinensis is considered one of the most endangered species (the National Key Species for Protection, Grade 1) in China. The Chinese government has delimitated the PRE as a national nature reserve for S. chinensis. Recent studies have indicated a decreasing trend of total mercury (THg) concentration in sediments with distance away from the PRE and toward the South China Sea [8] and showed an accelerated input of THg in sediment cores in recent decades [9]. These results suggest that THg contamination in this region has been strongly correlated with industrial development. Contamination by Hg(II) and MeHg was also observed in surface water in the tributary (e.g., Dong River) of the Pearl River Delta. However, THg and MeHg contamination in Indo-Pacific humpback dolphins and the interaction between THg toxicity and Se accumulation in their

bodies have not been systematically studied in this ecosystem [10,11].

In aquatic ecosystems, dolphins are top predators that have long lifespans, which increases their potential to accumulate Hg(II) and MeHg [12]. MeHg is easily accumulated in the livers of cetaceans, likely as a result of the ability of the liver to store and biotransform toxic contaminants [13]. The accumulation of Hg(II) and MeHg might cause detrimental effects on the reproduction system, immune responses, central nervous system and organs such as the liver and kidneys [14,15].

Previous studies have reported that selenium (Se) can mitigate the toxicity of MeHg in the liver of cetaceans by forming a highly insoluble Se-Hg compound after demethylating MeHg [16,17]. The presence of the nontoxic Se-Hg compound was confirmed in the cytoplasm of hepatic cells in Stenella coeruleoalba, and an equimolar ratio between Se and THg was reported in liver tissue with a high THg concentration [18,19]. This phenomenon likely explains the ability of cetaceans to tolerate high THg concentrations without directly observable adverse effects and exhibit a low MeHg residue in the brain tissue. Se is a micronutrient that plays an important role in enzymes (e.g., glutathione peroxidase) in the maintenance of normal organ functions. Adverse biological effects occur when there is a deficiency of bioavailable Se, and an Hginduced Se deficiency can lead to MeHg or THg toxicity [20]. The present investigation examined samples from dolphins stranded in the PRE, which have been collected since 2004, to study the distribution and fate of toxic pollutants (THg, MeHg and Se) in the major organs, in individuals at different ages and in their prey fishes from the PRE. The aims of the study were also to investigate the potential protective effects of Se against the toxicities of accumulated THg, and to provide evidence for the conservation of this endangered species.

Materials and Methods

Ethics Statement

This study on the Indo-Pacific humpback dolphins was approved by the Ministry of Agriculture of Chinese government under permit number 2003-54. The protocol was specifically verified by the Administration of Ocean and Fisheries of Guangdong Province, China under permit number 1999-583. No issue on ethics was concerned in this study.

Samples and chemical analysis

The tissue samples of S. chinensis were collected from dead and stranded animals along the PRE of the South China Sea from 2004 to 2012 (Fig. 1). Before performing the necropsies, the biological parameters were measured and tooth specimens were acquired to determine the age of specific individuals according to the method described by Jefferson [21] and Myrck et al. [22]. According to the report from Jefferson et al. [23], male dolphins inhabiting the PRE reach sexual maturity at 12-14 years, whereas females generally reach maturity at 9-10 years. Accordingly, the specimen assigned ID numbers from 21 to 28 in the present study were considered adults, and those with assigned numbers 1 to 20 were considered juveniles (Table S1). The liver, kidney and muscle tissue samples were placed in clean and acid-washed plastic bags and stored at -20° C immediately after collection. Information based on previous studies [21,24] indicated that the Indo-Pacific humpback dolphins have specific preferences for prey fishes. Based on that, 13 species of fish were collected from the PRE. Whole-body fish samples were smashed for metal analysis. All samples (dolphin and fish) were kept frozen until processing.

Samples were processed and prepared by a method similar to that of previous studies [25], in which portions of the different tissue samples were freeze-dried (Freeze-drying system, Labconco, Kansas City, Missouri, USA) for 48 h at 40–133×10⁻³ mBar and -49°C. The dried samples were then ground with an automatic agate mortar (Retsch, Germany) for 10 min. The concentration of THg in the dried tissue samples was measured without pretreatment or digestion by using the Hydra-C Automated Direct Hg Analyzer (Teledyne Instruments, Leeman Labs, USA). All specimens were analyzed in batches that included a procedural blank and standard reference material DORM-3 (National Research Council of Canada, Canada). The procedural blank and the reference material were treated and measured in the same way as the tissue samples. To analyze the MeHg concentration, tissue samples were digested in a KOH-methanol solution at 65°C for 4 h. The extractants were subjected to aqueous ethylation, separation by gas chromatography and detection by cold vapor atomic fluorescence spectrometry (CVAFS) (TEKRAN Model 2700 with an automated MERX purge and trap system, Brooks Rand Labs, USA) modified from USEPA method 1630 [26].

The digestion and preparation of dried tissue samples for Se analysis followed the methods described by Hung et al. [10]. Approximately 0.2 g of tissue samples was weighed in Teflon digestion tubes and soaked overnight in a mixture of 2 ml of double-distilled deionized water (3-D water) and 5 ml of 70% nitric acid (Merck, Germany). The digestion tubes were then sealed, placed in a microwave oven (Xin Tuo, model XT-9912, China) and subjected to a pressure increase of 65 psi for 15 m. After cooling, 2 ml H₂O₂ was added to the sample solution. The pressure was then increased to 65 psi for 15 m. The resulting digests were cooled and filtered through disposable syringe filter discs (0.45 µm pores, 25 mm diameter, with a mixed celluloseester filtering material, Jing Teng, China) equipped with 50-ml plastic syringes. The filtrates were transferred to 25-ml volumetric flasks and diluted with 3-D water. The samples were kept in acidwashed PVC tubes at 4°C prior to trace element analysis. The concentration of Se was measured by an inductively coupled plasma mass spectrometer (ICP-MS) (Agilent 7700, USA) with the procedural blank and reference TORT-2 included with every batch of tissue samples.

Quality assurance/quality control (QA/QC)

The precision and accuracy of the analytical methods were determined and monitored using the certified material TORT-2 (lobster hepatopancreas) and DORM-3 (fish protein) from the National Research Council of Canada. The recovery rates for THg, MeHg and Se were approximately 95.3%, 94.9% and 94.3%, respectively.

Statistical analysis

Statistical analyses were performed using SPSS (Statistical Package for the Social Sciences) software (StatSoft, ver. 22, USA), and the level of statistical significance was defined as p < 0.05. Grubbs' test was used to identify outliers, which were removed before further calculations. The Kolmogorov–Smirnoff test was used to assess the normality of the data distribution; if the data were not in normality, all the data were \log_{10} -transformed to improve normality prior to analysis to best fit the underlying assumptions of the analysis of variance. The correlation between measured parameters was assessed by coefficient of determination (\mathbb{R}^2). Differences in the concentrations of MeHg, THg and Se were analyzed among the groups (adult males, adult females, juvenile males and juvenile females) by using one-way ANOVA followed by Tukey's post-hoc tests. The contaminant concentrations of the



Figure 1. Sampling sites in the Pearl River Estuary (PRE) where the stranded Indo-Pacific humpback dolphins (n = 28) were collected from 2004 to 2012.

doi:10.1371/journal.pone.0110336.g001

stranded dolphins that exhibited a large dispersion were analyzed by non-parametric tests.

Results

The total Hg, Se and MeHg concentrations and their related ratios were determined in the tissue samples from the liver (n = 28), kidney (n = 22) and muscle (n = 15) of Indo-Pacific humpback dolphins (Table 1). The mean concentrations of THg, Se and MeHg in 13 fish species from the PRE were summarized in Table 2. The average THg concentration in the prey fish species was 0.146 $\mu g g^{-1}$ dw, with a range from 0.062 to 0.303 $\mu g g^{-1}$ dw, and the average Se concentration was 2.23 μ g g⁻¹ dw, with a range from 1.93 to 3.34 $\mu g g^{-1}$ dw. No significant differences in concentrations were found for the different fish species except the predatory species Arius sinensis and Pampus argenteus, which showed high THg concentrations. The potential biomagnification factor of inorganic mercury (Hg_{inorg}) in the dolphin tissue from the prey fishes was the highest in the liver tissue (350-fold) and the lowest in the muscle tissue (4.78-fold) (Table 2). For MeHg, dolphin muscle tissue had the highest biomagnification factor (18.7-fold), followed by liver tissue (14.3-fold) and kidney tissue (9.6-fold). Overall, Se in dolphin tissues showed the lowest potential for biomagnification of the three contaminants. For all dolphin individuals, muscle represents around 30% of total body mass, while other tissues (e.g., liver and kidney) contribute much less to the total body weight (<5%).

The highest mean concentration of THg was found in dolphin liver tissue $(32.3\pm59.0 \ \mu g \ g^{-1} \ dw)$, whereas a lower concentration

was found in the kidneys $(4.52\pm5.53 \ \mu g \ g^{-1} \ dw)$, with the smallest residue level shown in the muscle tissue $(1.45\pm1.62 \ \mu g \ g^{-1} \ dw)$. The difference in THg accumulation among the three tissues was significant (p < 0.05) (**Table 1**). Conversely, a significant difference in MeHg accumulation among different tissues was not found. Dolphin muscle tissue contained the highest MeHg concentration $(1.02\pm1.11 \ \mu g \ g^{-1} \ dw)$, followed by liver $(0.79\pm0.61 \ \mu g \ g^{-1} \ dw)$ and kidney $(0.53\pm0.49 \ \mu g \ g^{-1} \ dw)$. Se concentrations displayed a similar profile as THg, with the highest mean value found in the liver tissue $(15.2\pm19.4 \ \mu g \ sg^{-1} \ dw)$, followed by kidney $(7.57\pm3.47 \ \mu g \ g^{-1} \ dw)$ and muscle tissue $(1.89\pm1.69 \ \mu g \ g^{-1} \ dw)$.

The log₁₀-transformed concentrations of both THg and MeHg in the liver, kidney and muscle tissue of dolphins showed a significantly positive correlation with the log₁₀-transformed values of age (Fig. 2A, B). However, the slopes between tissue MeHg with age were much lower than those between THg and age for the three tissues. The MeHg/THg ratio in the livers and kidneys significantly decreased with age, while the MeHg/THg ratio in the muscles showed no trend with age (Fig. 2C). The mean percentage of MeHg/THg in the liver, kidney and muscle tissue was 18 ± 15 , 29 ± 22 and 77 ± 23 , respectively (**Table 1**), indicating that the MeHg concentrations in the liver and kidney tissue represented 30% or less of the THg, which was generally lower than the MeHg/THg fraction that appeared in the prey fish (ranging from 18% to 83%, with an average of 42%, Table 2). The dolphin muscle tissue showed the highest MeHg/THg ratio. Se accumulation in the liver and kidney tissue significantly increased with age, whereas the concentration of Se in muscle tissue was not obviously affected by age (Fig. 2D).

Table 1. The mean concentrations and standard deviations (SD), in μg g⁻¹ dry weight, of total mercury (THg), selenium (Se) and methyl mercury (MeHg); the molar ratio (%) of Se to THg; and the percentage (%) of MeHg/THg in liver, kidney and muscle tissue of Indo-Pacific humpback dolphins from the Pearl River Estuary (PRE), China.

		Live	r (n=28				Kidney ((n = 22)				Muscle (n	i=15)			
		THg	Se	MeHg	Se:THg	MeHg:T Hg	THg	Se	MeHg	Se:THg	MeHg:THg	THg	Se	MeHg	Se:THg	MeHg:THg
JM (n = 13)	Mean	4.24	5.63	0.49	8.51	25	1.27	5.67	0.26	37.7	43.4	0.58	1.52	0.48	10.6	78.7
	SD	5.12	3.54	0.35	5.90	14.5	1.65	2.86	0.21	27.8	22.2	0.40	0.42	0.36	6.67	12.9
JF (n = 7)	Mean	2.27	5.09	0.47	7.14	23.4	1.73 (6.95	0.36	15.46	27.5	1.29	2.78	0.43	7.82	66.9
	SD	1.36	1.20	0.15	2.59	6.66	1.14	1.83	0.16	8.6	11.9	1.91	2.67	0.34	2.64	32.3
	p values	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.
AM (n=3)	Mean	84.3	35.6	1.56	1.07	1.88	8.21	10.3	0.98	3.13	11	2.27	1.61	2.03	0.90	89.4
	SD	8.19	6.93	0.44	0.14	0.64	1.19	2.68	0.35	0.35	3.39	0	0	0	0	0
AF (n=5)	Mean	116	41.8	1.56	0.99	1.89	13.4	11.1	1.06	2.29	7.42	3.18	1.43	2.72	0.79	85.6
	SD	81.8	24.9	0.46	0.11	0.91	5.16	2.45	0.60	0.47	2.18	1.05	0.24	0.94	0.21	6.47
	p values	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.
M (n = 16)	Mean	3.19	5.63	0.60	9.08	20.7	2.87	6.73	0.43	29.7	36.1	0.82	1.53	0.70	9.01	80.4
	SD	3.76	3.54	0.44	5.76	15.9	3.32	3.42	0.39	28.4	23.7	0.70	0.39	0.64	7.07	12.5
F (n = 12)	Mean	31.6	20.4	0.93	4.92	14.4	6.90	8.79	0.67	9.61	18.6	2	2.2	1.29	4.81	73.9
	SD	50.3	24.2	0.62	3.90	11.8	6.77	2.96	0.54	9.17	13.4	1.88	2.13	1.28	4.01	27.4
	p values	n.s.	0.016	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.	n.s.
All (n = 28)	Mean	32.3	15.2	0.79	6.03	18.5	4.52	7.57	0.53	21.5	29.0	1.45	1.89	1.02	6.75	76.9
	SD	59.0	19.4	0.61	5.40	15.3	5.53	3.47	0.49	25.3	22.5	1.62	1.69	1.11	6.26	23.3
JM: juvenile r n.s.: not signi doi:10.1371/jc	nale (<12 yea ficant. urnal.pone.01	ırs); JF: 10336.	juvenile t001	female (<9	years); AM: a	dult male (>12 yea	ırs); AF: ad	lult fema	le (>9 years	Ġ						

Table 2. The mean concentrations and standard deviations (SD), in µg g⁻¹ dry weight, of total mercury (THg), inorganic mercury (Hg_{inorg}), methylmercury (MeHg), selenium (Se), and the percentage (%) of MeHg/THg in the prey fishes (whole body) for Indo-Pacific humpback dolphins and their average biomagnification factors (BMFs) in the dolphin tissues collected from the Pearl River Estuary (PRE), China.

Species	Family	Sample number	тнց		Hg _{inorg}		MeHg		Se		MeHg/THg
			Mean	SD	Mean	SD	Mean	SD	Mean	SD	Mean
Johnius belengerii	Sciaenidae	e	0.146	0.075	0.100	0.079	0.046	0.01	2.51	0.41	25
Collichthys lucidus	Sciaenidae	4	0.134	0.106	0.073	0.082	0.060	0.026	2.61	0.47	45
Clupanodon thrissa	Clupeidae	2	0.182	0.129	0.147	0.127	0.036	0.002	1.87	0.58	20
Coilia mystus	Engraulidae	5	0.158	0.152	0.121	0.144	0.037	0.008	1.99	0.43	23
Pampus argenteus	Stromateidae	3	0.221	0.111	0.189	0.119	0.032	0.008	2.22	0.22	15
Harpadon nehereus	Synodontidae	2	0.079	0.024	0.030	0.016	0.048	0.009	2.06	0.63	61
Cynoglossus bilineatus	Cynoglossidae	3	0.077	0.024	0.039	0.015	0.038	0.015	2.21	0.33	50
Sillago sihama	Sillaginidae	3	0.166	0.036	0.029	0.017	0.137	0.044	1.99	0.27	83
Arius sinensis	Ariidae	2	0.303	0.007	0.247	0.006	0.056	0.001	1.93	0.08	19
Selaroides leptolepis	Carangidae	2	0.118	0.065	0.079	0.058	0.039	0.007	1.99	0.40	33
Mugil cephalus	Mugilidae	2	0.062	0.002	0.033	0.005	0:030	0.003	1.94	0.32	48
Odontamblyopus rubicundus	Taenioididae	2	0.161	0.027	0.053	0.002	0.107	0.028	3.34	0.35	67
Odontamblyopus lacepedii	Taenioididae	З	0.077	0.000	0.033	0.009	0.044	0.009	2.39	0.53	57
Mean biomagnification factor	Dolphin liver				350		14.3		6.8		
Mean biomagnification factor	Dolphin kidney				44.3		9.6		3.4		
Mean biomagnification factor	Dolphin muscle				4.78		18.7		0.8		

doi:10.1371/journal.pone.0110336.t002



Figure 2. Relationships of dolphin age with the concentrations of THg, MeHg, Se and MeHg/THg ratio in the liver, kidney and muscle tissue, respectively, in the Indo-Pacific humpback dolphins stranded in the Pearl River Estuary (PRE) region. doi:10.1371/journal.pone.0110336.g002

The mean molar ratio of Se to THg was highest in the kidney tissue (21.5±25.3) and the lowest in the liver (6.04 ± 5.40), with a medium ratio observed in the muscle tissue (6.75 ± 6.26) (**Table 1**). The molar ratios of Se/THg in the livers and kidneys of the juveniles were eight-fold higher than in the adult livers and 14-fold higher than in the adult kidneys, respectively. A significant positive correlation was found between the concentrations of Se and THg in both the livers and kidneys (**Fig. 3**A, B). The log₁₀ Se/THg molar ratios showed strongly negative regressions with the log₁₀ MeHg levels in the livers (p < 0.05) and kidneys (p < 0.05) (**Fig. 3**C, D). A significantly positive relationship between THg and MeHg was observed in the livers, with an inflection range of 8.4–16.9 µg g⁻¹ dw of THg (**Fig. 4**).

No significant differences were found in the THg and MeHg concentrations in the liver and muscle tissues between males and females. Only the mean Se concentrations in the liver tissue were significantly different between males and females (p < 0.05) (**Table 1**).

Discussion

Compared with the kidneys and muscles, extremely high concentrations of THg were found in the livers of *S. chinensis* $(32.3\pm59.0 \ \mu g \ g^{-1} \ dw)$, although the upper limits of the THg concentrations in the livers were lower than those found in dolphins from the waters of Hong Kong (906 $\mu g \ g^{-1} \ dw)$ [27].

This phenomenon may be related to the detoxification function of marine mammal livers in terms of storage and biotransformation. The tolerance limit of THg in mammalian hepatic tissues appears to be within the range of 100 to 400 μ g g⁻¹ wet weight (ww) [28]. The THg concentrations shown in **Table 1** are below this range; however, a dolphin analyzed in our laboratory was found with a liver THg concentration of 1374 μ g g⁻¹ dw. Albeit not as high as the results from Japan (1600 μ g g⁻¹ dw) [29] and the Mediterranean (13150 μ g g⁻¹ dw) [30], the THg concentrations detected in the liver samples of *S. chinensis* in the PRE were high enough to cause damage to the internal organs of the contaminated individuals.

The accumulation of MeHg tended to increase with age (**Fig. 2**) but at a lower rate than that of THg. No obvious trend existed between the MeHg/THg ratio (%) and age in the muscle samples, whereas the liver and kidney samples showed a decreasing trend of MeHg with age, which could be ascribed to a slow demethylation process that occurs in livers and kidneys but not in muscles [29,31]. Among the three tissue types, muscle tissue accumulated the highest MeHg concentrations, which is consistent with the hypothesis that lack of demethylation mechanisms occur in muscles [32]. Based on studies of several small mammals [33], the lethal level of MeHg in brain tissue was proposed to be in the range of 12–30 μ g g⁻¹ ww (equivalent to 60–150 μ g g⁻¹ dw). However, the highest MeHg concentration in the brain tissue



Figure 3. Regression analysis of Se with THg, and of log₁₀ (Se/Hg) with log₁₀ MeHg in the liver and kidney of Indo-Pacific humpback dolphins (n = 28) stranded in the Pearl River Estuary (PRE) region. doi:10.1371/journal.pone.0110336.g003

 $(0.41~\mu g~g^{-1})$ from our previous study (unpublished data) was significantly lower than the lethal level.

The relationships between age and THg concentration in livers and kidneys of different species of marine mammal have been extensively examined [12,30,34], and indicate that the hepatic and renal THg concentrations increase with age. This correlation implies a higher capability for the bioaccumulation of toxic elements than for their elimination throughout these animals' life span. In addition, as the size of the prey and the quantities of food tends to increase in proportion with the growth of the dolphins, the trophic transfer of the toxic metals may also progressively increase [35]. In the present study, the THg concentration increased with age in the livers, kidneys and muscles (**Fig. 2**).

Because of the difficulties in tracking down the prey fish species of dolphins in natural habitats to analyze their tissue contaminant residues, the trophic transfer of THg, MeHg and Se from prey fish to dolphin has scarcely been studied, even though dolphins are considered one of the top predators in the ocean. Through a comparison of the mean concentrations of THg, MeHg and Se from the prey fish to dolphin tissues, our studies outline the first direct evidence of the biomagnification processes of contaminants in this cetacean (**Table 2**). For Hg_{inorg}, the average concentrations



Figure 4. MeHg concentration (μ g g⁻¹ dw) (A) and concentration ratio of MeHg to THg (B) against THg concentration (μ g g⁻¹ dw) in the liver samples of the stranded Indo-Pacific humpback dolphins in the Pearl River Estuary region (n = 28). doi:10.1371/journal.pone.0110336.q004

in the three dolphin tissues were 4.78- to 350-fold greater than the concentrations detected in their prey fish, suggesting that a significant biomagnification process could occur for Hg, especially in the liver. The present study also showed that concentrations of MeHg in the cetacean tissue were 9- to 18-fold higher than the concentrations in the prey fish because of the high assimilation and bioaccumulation capability of MeHg across trophic levels. Conversely, trophic transfer of Se was weak except in dolphin livers. The concentrations of Se in the dolphin muscle tissue were even lower than the concentrations found in the prey fish (BMF< 1), which is likely due to the lack of accumulation or rapid elimination of Se in this tissue. The comparison of THg, Se and MeHg between the dolphin organ tissues and prev fish indicated that THg and MeHg could significantly accumulate in dolphin organs. However, whether THg and MeHg can threaten the health of this cetacean is best determined by the molar ratio of Se/ THg, which will be discussed below.

In the marine environment, the dominant portion of Hg in fish and squid is present in the form of MeHg [36]. However, the majority of Hg accumulated in the internal organs of cetaceans appears as inorganic Hg(II), especially in the liver, indicating that a demethylation of MeHg occurs in the liver [34,37]. The formation of the compound Se-Hg, resulting from the combination of a demethylation product (i.e., Hg(II)) and Se, appears to be the last step in the demethylation processes, leading to the fossilization of THg and Se in the form of an inert compound [17,19,38]. In the present study, the relationship between the \log_{10} Se/THg value and log₁₀ MeHg concentration in the livers and kidneys of Indo-Pacific humpback dolphins showed a significant negative correlation (Fig. 3). This finding strongly suggests that the Se-Hg compound was formed, causing the fraction of Se/THg in the form of the Se-Hg compound to decrease with respect to the increased MeHg levels [39]. The significant positive relationship between Se and THg and the significant negative correlation between the proportion of MeHg in THg and the THg concentration further confirms that detoxification processes act on MeHg in the livers and kidneys of the dolphins (Fig. 3 and Fig. S1), similar to the results reported for other cetaceans [18,37]

The molar ratio of Se/THg can be considered an indicator of the extent of toxicity caused by Hg contamination because the toxic effects are alleviated after the formation of the Se-Hg compound, which is stable and has no relevant biological impact. Se can be combined with other elements to form various compounds. A molar ratio between Se and THg higher than 1 implies that Se is providing potential protection against Hg toxicity. However, a ratio below 1 suggests limited Se protection against Hg toxicity [40,41]. In the present study, the average molar ratios of Se to THg in the liver, kidney and muscle tissue were 6.03, 21.48 and 6.75, respectively, which are all significantly higher than 1. The high Se/THg ratios in tissues were due to the fact that juveniles comprised most of the analyzed samples and low THg concentrations were found in these samples. However, certain adults had molar ratios below 1 and were in danger of suffering Hg toxicity. According to Caurant et al. [38], young animals were still unable to demethylate MeHg efficiently, leading to higher mean percentages of MeHg/THg concentrations in juveniles than adults. We concluded that a molar ratio of Se/THg of 1 in the organs of adults can be used as an indicator of demethylation processes in the organs. The mean molar ratios of Se/THg in the liver samples of adult dolphins were approximately 1, whereas the mean value in the kidneys was 2.65. However, this

pattern does not preclude the possibility of the demethylation process occurring in the kidneys. Se, as the constituent of selenoprotein P and selenoenzymes, also plays an important role in maintaining the normal functions of the kidneys in addition to its detoxification function against Hg.

The relationship between THg and MeHg in dolphin livers may indirectly reveal how demethylation occurs. In this study (**Fig. 4**A), when the THg concentration was lower than 8.4 μ g g^{-1} dw, the MeHg concentration increased significantly, indicating that no demethylation occurred when the body burden of THg was low. However, when the THg concentration increased from 8.4 μ g g⁻¹ dw to 16.9 μ g g⁻¹ dw, the increase of MeHg in the livers slowed down because the demethylation process was presumably occurring, although we lack the direct chemical and biochemical evidence to confirm this assumption. The concentration ratio of MeHg to THg was $1.86 \pm 0.8\%$ (Mean \pm SD, n = 9) when the THg concentration exceeded 16.9 $\mu g \ g^{-1}$ dw, whereas the concentration ratio was $27 \pm 11\%$ (Mean $\pm SD$, n = 16) when the THg concentration was below 8.4 μ g g⁻¹ dw (**Fig. 4**B). Therefore, it is assumed that the demethylation processes could be activated when the THg concentration reaches a threshold range (i.e. 8.4–16.9 μ g g⁻¹ dw). According to Palmisano et al. [42], mercury is generally stored as MeHg in the livers of dolphins initially, and after a threshold value (100 $\mu g g^{-1}$ ww) is reached, demethylation takes place with the co-accumulation of Se/THg at a molar ratio of 1:1. In the livers of S. chinensis, Se was significantly related to THg ($R^2 = 0.9740$, p < 0.05), and the molar ratios of Se/THg in most our samples were far higher than 1 (Fig. **S2**), suggesting that Se was involved in the detoxification process of MeHg.

The co-accumulation of Se/THg with a molar ratio of 1:1 in the livers of cetaceans after a threshold value is reached has been reported by several authors in striped dolphins and Dall's porpoises [42,43]. Together, these results indicated that exceeding the threshold value of THg was a necessarily initial step before triggering the demethylation processes. Compared with the threshold value of striped dolphins (100 μ g g⁻¹ ww) and Dall's porpoises (20–30 μ g g⁻¹ dw), a lower THg concentration was found in the present study for S. chinensis. These differences could be attributed to the metabolism and varying foraging habits among the cetacean species, which might also influence the accumulation of Hg and Se, as reflected in the Se/THg molar ratios [12,44]. The lower threshold of hepatic THg concentrations for the equimolar accumulation of Se and Hg in S. chinensis suggests that this species has a greater sensitivity to THg concentrations than is found in striped dolphins and Dall's porpoises.

Conclusions

This study is the first comprehensive investigation on the distribution and fate of THg, MeHg and Se in livers, kidneys and muscles of Indo-Pacific Humpback Dolphins collected for almost 10 years from the PRE of China, and in the prey fishes from the PRE. The results clearly showed that THg and Se were mainly accumulated in dolphin livers while the highest residue of MeHg was found in dolphin muscles. This study contains the first experimental evidence for the potential trophic transfer from the fish to Indo-Pacific Humpback dolphins, showing a high biomagnification factor of Hg_{inorg} in liver ($350 \times$) and MeHg in muscle ($18.7 \times$). Our studies suggest that the Indo-Pacific humpback dolphins may have the potential to detoxify Hg via

the demethylation of MeHg and the formation of tiemannite (HgSe) in the liver and kidney.

Supporting Information

Figure S1 Relationship between log THg (μ g g⁻¹ dw) and percentage of MeHg/THg in the liver, kidney and muscle of Sousa chinesis stranded in the PRE. (TIF)

Figure S2 (A) Individuals with molar ratio of Se/Hg higher than the threshold range (8.4–16.9 μ g g⁻¹ dw). (B) Individuals with molar ratio of Se/Hg lower than the threshold range (8.4–16.9 μ g g⁻¹ dw). (TIF)

References

- 1. Compeau G. Bartha R (1985) Sulfate-reducing bacteria: principal methylators of mercury in anoxic estuarine sediment, Appl Environ Microbiol 50: 498-502.
- 2. Fleming EJ, Mack EE, Green PG, Nelson DC (2006) Mercury methylation from unexpected sources: molybdate-inhibited freshwater sediments and an ironreducing bacterium. Appl Environ Microbiol 72: 457-464.
- 3. Hong YS, Hunter S, Clayton LA, Rifkin E, Bouwer EJ (2012) Assessment of mercury and selenium concentrations in captive bottlenose dolphin's (Tursiops truncatus) diet fish, blood, and tissue. Sci Total Environ 414: 220-226
- Gilmour CC, Podar M, Bullock AL, Graham AM, Brown SD, et al. (2013) Mercury Methylation by Novel Microorganisms from New Environments. Environ Sci Technol 47: 11810-11820.
- 5. Parks JM, Johs A, Podar M, Bridou R, Hurt RA, et al. (2013) The genetic basis for bacterial mercury methylation. Science 339: 1332-1335.
- 6. Yu RQ, Flanders J, Mack EE, Turner R, Mirza MB, et al. (2012) Contribution of coexisting sulfate and iron reducing bacteria to methylmercury production in freshwater river sediments. Environ Sci Technol 46: 2684-2691
- 7. Yu RQ, Reinfelder JR, Hines ME, Barkay T (2013) Mercury methylation by the methanogen Methanospirillum hungatei. Appl Environ Microbiol 79: 6325-6330
- 8. Shi JB, Ip C, Zhang G, Jiang G, Li X (2010) Mercury profiles in sediments of the Pearl River Estuary and the surrounding coastal area of South China. Environ Pollut 158: 1974-1979.
- Yu X, Li H, Pan K, Yan Y, Wang WX (2012) Mercury distribution, speciation and bioavailability in sediments from the Pearl River Estuary, Southern China. Mar Pollut Bull 64: 1699-1704.
- 10. Hung CL, So MK, Connell DW, Fung CN, Lam MH, et al. (2004) A preliminary risk assessment of trace elements accumulated in fish to the Indo-Pacific Humpback dolphin (Sousa chinensis) in the northwestern waters of Hong Kong. Chemosphere 56: 643-651.
- 11. Wu Y, Shi J, Zheng GJ, Li P, Liang B, et al. (2013) Evaluation of organochlorine contamination in Indo-Pacific humpback dolphins (Sousa chinensis) from the Pearl River Estuary, China. Sci Total Environ 444: 423-429.
- 12. Kunito T, Nakamura S, Ikemoto T, Anan Y, Kubota R, et al. (2004) Concentration and subcellular distribution of trace elements in liver of small cetaceans incidentally caught along the Brazilian coast. Mar Pollut Bull 49: 574-587.
- Thompson DR (1990) Metal levels in marine vertebrates. CRC PRESS, BOCA RATÓN, FL(USA): 143-182.
- 14. Endo T, Hisamichi Y, Kimura O, Haraguchi K, Baker CS (2008) Contamination levels of mercury and cadmium in melon-headed whales (Peponocephala electra) from a mass stranding on the Japanese coast. Sci Total Environ 401: 73-80.
- 15. Dietz R, Riget F, Born E (2000) An assessment of selenium to mercury in Greenland marine animals. Sci Total Environ 245: 15-24.
- 16. Moreira I, Seixas T, Kehrig H, Fillmann G, Di Beneditto A, et al. (2009) Selenium and mercury (total and organic) in tissues of a coastal small cetacean, Pontoporia blainvillei. J Coast Res: 866-870.
- 17. Lailson-Brito J, Cruz R, Dorneles PR, Andrade L, Azevedo AdF, et al. (2012) Mercury-Selenium Relationships in Liver of Guiana Dolphin: The Possible Role of Kupffer Cells in the Detoxification Process by Tiemannite Formation. PLoS ONE 7: e42162.
- 18. Nakazawa E, Ikemoto T, Hokura A, Terada Y, Kunito T, et al. (2011) The presence of mercury selenide in various tissues of the striped dolphin: evidence from µ-XRF-XRD and XAFS analyses. Metallomics 3: 719-725
- 19. Nigro M (1994) Mercury and selenium localization in macrophages of the striped dolphin, Stenella coeruleoalba. Journal of the Marine Biological Association of the United Kingdom Plymouth 74: 975–978.
- 20 Khan MAK, Wang F (2009) Mercury-selenium compounds and their toxicological significance: Toward a molecular understanding of the mercuryselenium antagonism. Environ Toxicol Chem 28: 1567-1577.

Table S1 Sampling information of Sousa chinensis stranded in the Pearl River Estuary. (DOCX)

Acknowledgments

We thank the Guangdong Pearl River Estuary Chinese White Dolphin National Nature Reserve for the logistic service; Mr. Wenzhi Lin, Mr. Xi Chen and Mr. Yinku Wang for collecting samples.

Author Contributions

Conceived and designed the experiments: YW. Performed the experiments: YS QT HM. Analyzed the data: DG RY YS. Contributed reagents/materials/analysis tools: RY DG. Contributed to the writing of the manuscript: DG YW RY YS. Supervised the research: YW RY LC. Revised the manuscript: DG YR YW.

- 21. Jefferson TA (2000) Population biology of the Indo-Pacific hump-backed dolphin in Hong Kong waters. Wildlife monographs: 1-65.
- 22. Myrck AC Jr., Hohn AA, Sloan PA, Kimura M, Stanley DD (1983) Estimating age of spotted and spinner dolphins (Stenella attenuata and Stenella longirostris) from teeth.
- 23. Jefferson TA, Hung SK, Robertson KM, Archer FI (2012) Life history of the Indo-Pacific humpback dolphin in the Pearl River Estuary, southern China. Mar Mammal Sci 28: 84-104.
- Barros NB, Jefferson TA, Parsons E (2004) Feeding habits of Indo-Pacific 94 humpback dolphins (Sousa chinensis) stranded in Hong Kong. Aquat Mamm 30: 179 - 188
- 25. Ruelas-Inzunza JR, Horvat M, Pérez-Cortés H, Páez-Osuna F (2003) Methylmercury and total mercury distribution in tissues of gray whales (Eschrichtius robustus) and spinner dolphins (Stenella longirostris) stranded along the lower Gulf of California, Mexico. Cienc Mar 29: 1-8.
- U.S.EPA Method 1630 Methyl Mercury in Water by Distillation, Aqueous Ethylation, Purge and Trap, and Cold Vapor Atomic Fluorescence Spectrometry. US Environmental Protection Agency Office of Water Office of Science and Technology Engineering and Analysis Division (4303) 401 M Street SW Washington, DC 20460.
- 27. Parsons E (1998) Trace metal pollution in Hong Kong: Implications for the health of Hong Kong's Indo-Pacific hump-backed dolphins (Sousa chinensis). Science of the Total Environment 214: 175-184.
- 28. Wagemann R, Muir DCG (1984) Concentrations of heavy metals and organochlorines in marine mammals of northern waters: overview and evaluation. Western Region. Department of Fisheries and Oceans Canada.
- 29. Honda K, Tatsukawa R, Itano K, Miyazaki N, Fujiyama T (1983) Heavy metal concentrations in muscle, liver and kidney tissue of striped dolphin, Stenella coeruleoalba, and their variations with body length, weight, age and sex. Agric Biol Chem 47: 1219-1228.
- 30. Leonzio C, Focardi S, Fossi C (1992) Heavy metals and selenium in stranded dolphins of the Northern Tyrrhenian (NW Mediterranean). Sci Total Environ 119: 77-84.
- 31. Martoja R, Berry JP (1981) Identification of tiemannite as a probable product of demethylation of mercury by selenium in cetaceans. A complement to the scheme of the biological cycle of mercury [detoxification]. Vie Milieu.
- 32. Storelli M, Ceci E, Marcotrigiano G (1998) Comparison of total mercury, methylmercury, and selenium in muscle tissues and in the liver of Stenella coeruleoalba (Meyen) and Caretta caretta (Linnaeus). Bull Environ Contam Toxicol 61: 541-547.
- 33. Wren CD (1986) A review of metal accumulation and toxicity in wild mammals: I. Mercury, Environ Res 40: 210-244.
- 34. Meador J, Ernest D, Hohn A, Tilbury K, Gorzelany J, et al. (1999) Comparison of elements in bottlenose dolphins stranded on the beaches of Texas and Florida in the Gulf of Mexico over a one-year period. Arch Environ Contam Toxicol 36: 87-98.
- 35. André J, Ribeyre F, Boudou A (1990) Mercury contamination levels and distribution in tissues and organs of delphinids (Stenella attenuata) from the Eastern Tropical Pacific, in relation to biological and ecological factors. Mar Environ Res 30: 43-72.
- 36. Das K, Debacker V, Bouquegneau JM (2000) Metallothioneins in marine mammals, Cell Mol Biol 46.
- 37. Holsbeek L, Siebert U, Joiris CR (1998) Heavy metals in dolphins stranded on the French Atlantic coast. Sci Total Environ 217: 241-249.
- Caurant F, Navarro M, Amiard JC (1996) Mercury in pilot whales: possible 38. limits to the detoxification process. Sci Total Environ 186: 95-104.
- 39. Cáceres-Saez I, Dellabianca NA, Goodall RNP, Cappozzo HL, Guevara SR (2013) Mercury and Selenium in Subantarctic Commerson's Dolphins (Cephalorhynchus c. commersonii). Biol Trace Elem Res 151: 195-208.

- Mercury and Selenium in Stranded Indo-Pacific Humpback Dolphins
- Peterson SA, Ralston NV, Peck DV, Sickle JV, Robertson JD, et al. (2009) How might selenium moderate the toxic effects of mercury in stream fish of the western US? Environ Sci Technol 43: 3919–3925.
- Sormo EG, Ciesielski TM, Overjordet IB, Lierhagen S, Eggen GS, et al. (2011) Selenium Moderates Mercury Toxicity in Free-Ranging Freshwater Fish. Environ Sci Technol 45: 6561–6566.
- Palmisano F, Cardellicchio N, Zambonin P (1995) Speciation of mercury in dolphin liver: a two-stage mechanism for the demethylation accumulation process and role of selenium. Mar Environ Res 40: 109–121.
- Yang J, Kunito T, Tanabe S, Miyazaki N (2007) Mercury and its relation with selenium in the liver of Dall's porpoises (*Phocoenoides dalli*) off the Sanriku coast of Japan. Environ Pollut 148: 669–673.
- 44. Seixas TG, Kehrig HdA, Fillmann G, Di Beneditto APM, Souza CM, et al. (2007) Ecological and biological determinants of trace elements accumulation in liver and kidney of *Pontoporia blainvillei*. Sci Total Environ 385: 208–220.