

RESEARCH ARTICLE

Production of Bio-Energy from Pig Manure: A Focus on the Dynamics Change of Four Parameters under Sunlight-Dark Conditions

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Abstract

This study investigated the effect of sunlight-dark conditions on volatile fatty acids (VFAs), total ammonium nitrogen (TAN), total alkalinity (TA) and pH during pig manure (PM) digestion and then the subsequent influence on biogas yield of PM. PM₁ and PM₂ were performed in a transparent reactor and a non-transparent reactor, respectively. Two sets of experiments were conducted with a temperature of 35.0±2.0 °C and a total solid concentration of 8.0% to the digestion material. The dynamic change of the four parameters in response to sunlight-dark conditions resulted in variations of the physiological properties in the digester and affected the cumulative biogas production (CBP). PM₁ obtained higher CBP (15020.0 mL) with a more stable pH and a lower TAN concentration (1414.5 mg/L) compared to PM₂ (2675.0 mL and 1670.0 mg/L, respectively). The direct path coefficients and indirect path coefficients between the four parameters and CBP were also analyzed.

Introduction

With the increasing market demand for pork, the growth of swine herds leads to a large increase in swine manure worldwide [1]. The pollution impact of swine waste on water, soil and air caused is a growing concern in many countries [2, 3]. The sustainability of an efficient disposal mechanism for manure becomes a key factor in the expansion of pig industry in China [4].

Biogas production with PM is a suitable method for the treatment of this organic waste, yielding biogas as a useful by-product. This process could also produce renewable energy (cheap and clean methane), soil conditioner, and liquid fertilizer that are valuable for crop production [2, 5–10]. However, the complex anaerobic digestion processes consisting of a series of microbial reactions are vulnerable to inhibition by many factors, such as sunlight-dark conditions. Recently, a few studies focused on sunlight-dark conditions as an external artificial factor. It was suggested that dark fermentation of organic biomass is a promising technology for producing renewable bio-hydrogen [11, 12]. Research also suggested that bio-hydrogen

production by waste materials would be enhanced by sequential dark and light anaerobic fermentations [13]. Rittmann and Herwig and Levin *et al.* showed that dark fermentation can improve the hydrogen evolution rate of bio-hydrogen production and concomitantly produced carbon rich metabolites, like CO₂ would store in biomass or be converted to other substances, such as CH₄ [14, 15]. Chandra and Mohan suggested that co-culturing photosynthetic bacteria with acidogenic microflora could reduce VFAs accumulation by 40% which could overcome induced fatty acid inhibition during dark-fermentative hydrogen production process [16]. A study by Yin *et al.* showed that sunlight-dark conditions can increase the biogas yield from PM [3]. However, the promoting influence of sunlight-dark conditions on physiological properties of digester, such as the VFAs, TAN, TA and pH, important parameters to be monitored in anaerobic digestion [17–20], and their effects on biogas production are unclear.

Therefore, the present study emphatically evaluated dynamic changes of the four parameters in the fermentation process of PM under sunlight-dark conditions in order to reveal their effects on biogas production of PM.

Materials and Methods

Ethics statement of substrate and inoculum

PM was collected from “Besun Group” swine farm in industrial park, Changqing Road, Yangling, with the permission of the managers. The inoculum was obtained from household biogas digester in 13 North 2nd Street, Cuixigou, which is the model village of biogas utilization and more than 85% households installed biogas digesters. Collection was permitted by the owner Quanyou Cui. Both PM and inoculum were stored in a refrigerator (4.0°C) until use [5]. The experimental procedures were approved by the Ethics Committee of the Research Center of Recycle Agricultural Engineering and Technology of Shaanxi Province in China. Table 1 shows the chemical characteristics of the PM.

Experimental design and set-up

Fig 1 shows the desire of this study. Anaerobic fermentation of PM was carried out in triplicate at 35.0±2.0°C with Total solids (TS) of 8.0% for 53 days. The 1-L digestion reactor with 700.0 g of total liquid, including 140.0 g of inoculum, was conducted under a controlled and constant temperature using an anaerobic fermentation device (Fig 2).

Two sets of experiments were conducted: one was performed with sunlight-dark fermentation in transparent reactor with nature sunlight (PM₁), and the other was conducted in total dark in a non-transparent reactor (PM₂). This work lasted from September 17th to November 11th in 2012. The sunlight duration data (Fig 3) were gathered from the Yangling meteorological information network (<http://www.ylqx.gov.cn>). The gas volume was measured daily, and the VFA, TAN, TA and pH were measured every 7 days. All fermentation reactors were tested by sealing detection and flushed with nitrogen gas for approximately 3 min to assure anaerobic conditions before measuring [21].

Table 1. Chemical characterization of substrates used in the digestion experiments.

Material	TS (%)	VS (%)	Organic carbon a (g/kg VS)	Total kjeldahl nitrogen ^a (g/kg VS)	Carbon-to-nitrogen ratio	pH	TA (mg/L)	TAN (mg/L)	VFAs (mg/L)
PM	27.7	79.2	78.3	6.1	12.8	6.4	5093.0	1328.7	5569.5

^a Dry basis

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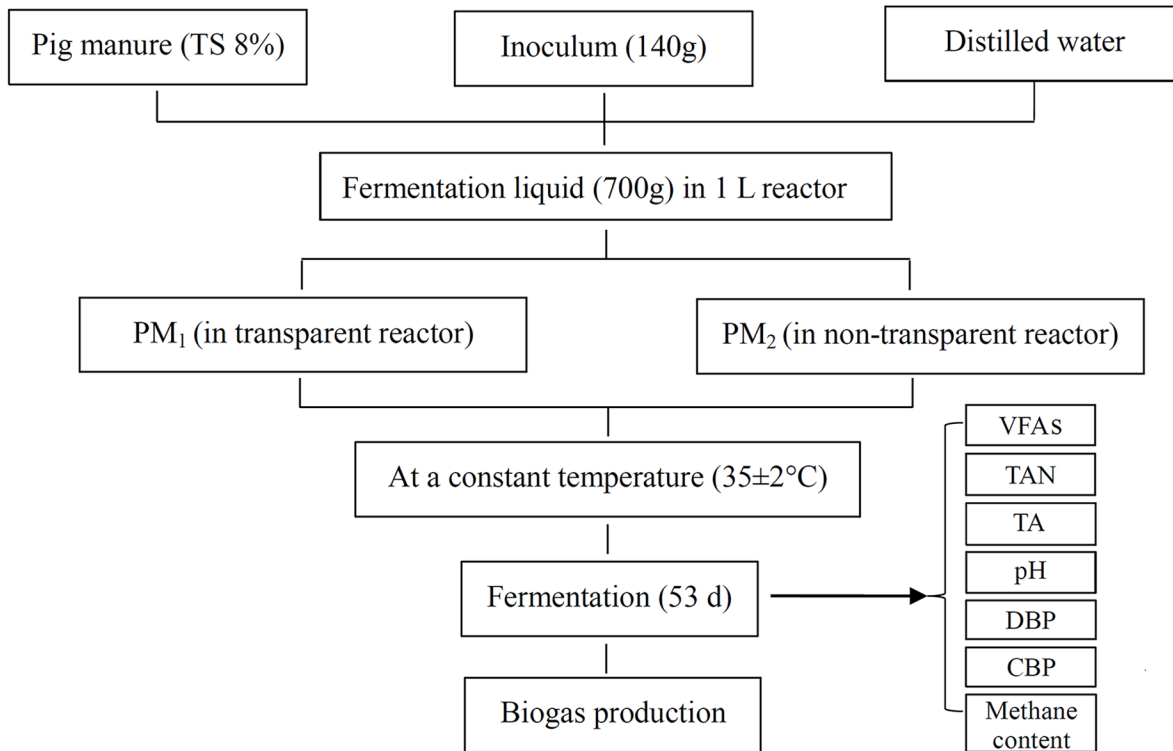


Fig 1. Flowchart of experiment including raw material, the experimental conditions and method.

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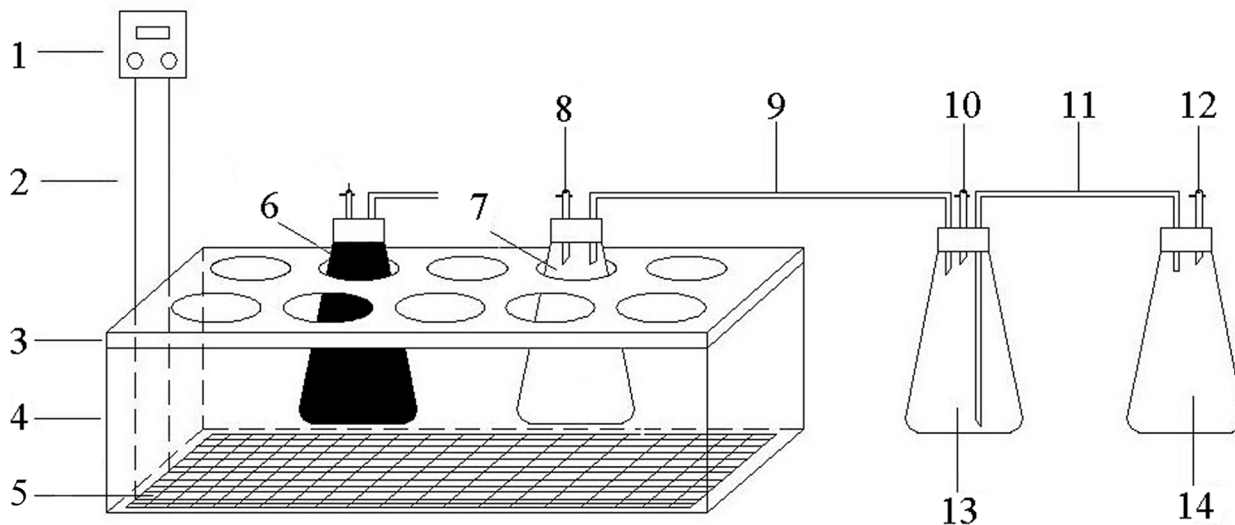


Fig 2. Controlled and constant temperature anaerobic fermentation device. 1. Temperature controlling box; 2. Temperature sensor; 3. Insulated cover; 4. Thermostatic water tank; 5. Strip heater; 6. None transparent digester; 7. Transparent digester; 8. Taking sampling; 9. Airway tube; 10. Taking biogas; 11. Aqueduct; 12. Air pipe; 13. Biogas collecting bottle; 14. Water collecting bottle.

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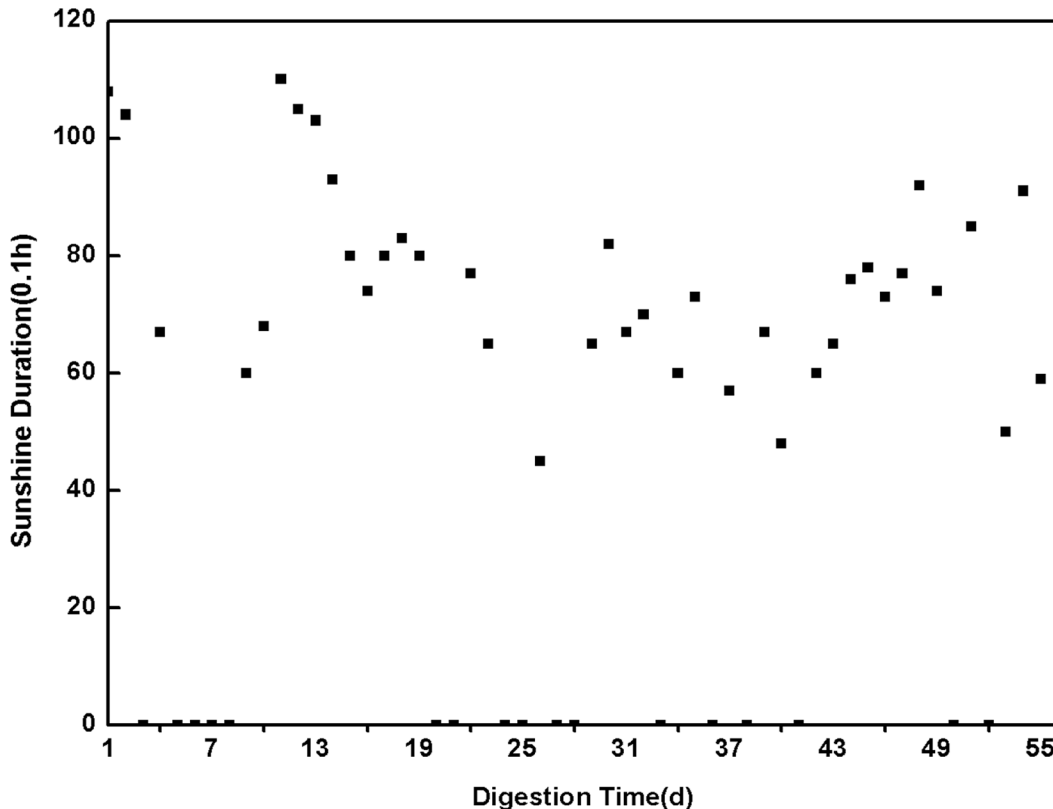


Fig 3. Sunlight duration in Yangling during digestion of PM₁ and PM₂.

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Analytical techniques

The daily biogas production (DBP) was monitored daily using a drainage gas-collecting method. The content of methane in biogas digester was analyzed by a fast methane analyzer (Model DLGA-1000, Infrared Analyzer, Dafang, Beijing, China). Total organic carbon was determined by the method described in Cuetos *et al.* [22]. The determination of TS and volatile solid (VS) composition was performed according to the APHA Standard Methods [23]. The VFAs concentration was determined using a754P UV spectrophotometer (adding 1.7 mL glycol into 0.5 mL sample before heating for 8 minutes at 90°C; when cooled, transferring this mixed solution to a 25-mL volumetric flask and adding 2.5 mL Hydroxylamine reagent, then diluting with distilled water to volume and mixing it). TAN was analyzed by KDN-08C type semiautomatic azotometer and then titrated with 0.02 N H₂SO₄. TA analysis was conducted by titrations with 0.02 M H₂SO₄ [18]. pH value was determined by Phs-3ct type pH meters and all titrations were performed in duplicate.

Statistical analysis

Fig 4 describes path analysis between independent variables (X_i) and dependent variable (Y). The two arrow lines in Fig 4(a) between the independent variables and dependent variable represent the path where $X_1 \rightarrow Y$ and $X_2 \rightarrow Y$ are independent of each other. Fig 4(b) shows four arrow lines that comprise the path network where a correlation exists between X_1 and X_2 . In addition to the two direct paths ($X_1 \rightarrow Y$ and $X_2 \rightarrow Y$), the path network has two indirect paths attributed to r_{12} . One path is generated by the effect of X_1 on Y via X_2 ($X_1 \rightarrow X_2 \rightarrow Y$), and

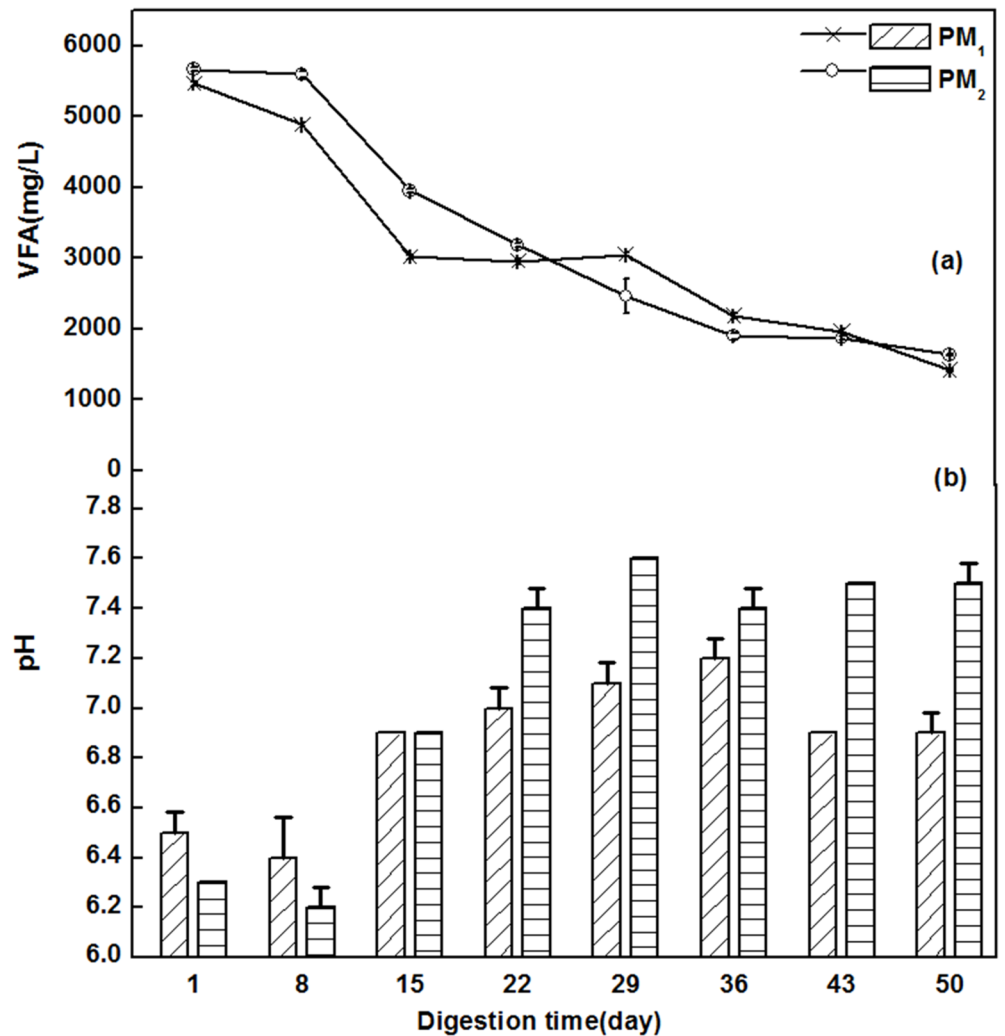


Fig 5. The dynamic changes of VFAs (a) and pH value (b) of PM₁ and PM₂.

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fermentation [26, 27]. In digester, methane bacteria mainly use VFAs to produce methane. However, it does not mean the more VFAs the better, because high concentrations could result in a decrease of pH and increase of non-dissociated fatty acids, which further intensifies inhibition [28, 29]. Therefore, the concentration of VFAs is an important consideration for good performance of a digester. It reflects the imbalance between the microbial groups involved in the degradation. The further degradation of these compounds can proceed only after the removal of hydrogen from the process [30].

Fig 5(a) shows dynamic changes of VFAs in PM₁ and PM₂. VFAs in PM₁ and PM₂ had a decreasing trend in the process of fermentation, which is consistent with the theory of anaerobic fermentation that VFAs were oxidized into substrates slowly by methanogenic bacteria [26]. At the beginning, VFAs decreased sharply, especially in PM₁, whose VFAs decreased from 5467.5 mg/L to 3018.5 mg/L in the first 15 days of fermentation. After that, the content of VFAs steeply decreased until the 29th day. The final VFAs value was 1418.5 mg/L. The VFAs of PM₂ had a gentle downtrend in the first 8 days from 5671.5 mg/L to 5598.5 mg/L, and then, this value showed a rapid downward trend until the 36th day of fermentation. Finally, the VFAs in PM₂

had a similar gradual decrease as that in PM₁. Comparison the initial and final of fermentation process, the VFAs contents of PM₁ decreased from 5467.5 mg/L to 1418.5 mg/L and PM₂ decreased from 5671.5 mg/L to 1633.5 mg/L, respectively.

The stability of the pH in an anaerobic reactor is extremely important because it influences enzymatic activity and the rate of methane production may decrease if the pH is lower than 6.3 or higher than 7.8 [31]. Therefore, a feasible way of improving stability for fermentation would be to monitor and to analyze. As shown in Fig 5(b), on the 8th day of digestion, the pH of PM₁ and PM₂ decreased from 6.5 to 6.4 and 6.3 to 6.2, respectively, which could be attributed to hydrolysis acidification. Astals *et al.* showed that large amounts of protein and carbohydrates but small amounts of lipids in PM probably led to hydrolysis acidification [32]. Along with the fermentation process, both sets showed an increasing trend in pH because the acids were rapidly consumed by methanogens, thus increased the pH and stabilized the digester performance [21]. The peak values of PM₁ and PM₂ were 7.2 and 7.6 on the 36th and 29th day, respectively. The pH of each group ranged from 6.3 to 7.8 at the end of fermentation and was higher than that of the initial fermentation.

Dynamic changes of TAN and TA. Ammonium is an essential nutrient for bacterial growth, but undesirably high concentrations could breakdown the proteins available in the substrate [33]. TAN is also an important parameter influencing methane production by providing buffering capacity [34].

Fig 6(a) shows dynamic changes of TAN in PM₁ and PM₂. The average amount of TAN in PM₁ (1385.0 mg/L) was lower than that in PM₂ (1665.1 mg/L). The value of TAN in PM₁ increased from 1289.5 mg/L to 1397.2 mg/L and then experienced a slight declined to 1214.1 mg/L on the 22nd day, when the value rebounded and reached a peak of 1602.7 mg/L on the 43rd day. In contrast, PM₂ had a faster increasing rate of TAN compared to PM₁ during the first 22 days, increasing from 1367.8 mg/L to a peak value of 1866.5 mg/L, which exceeded the

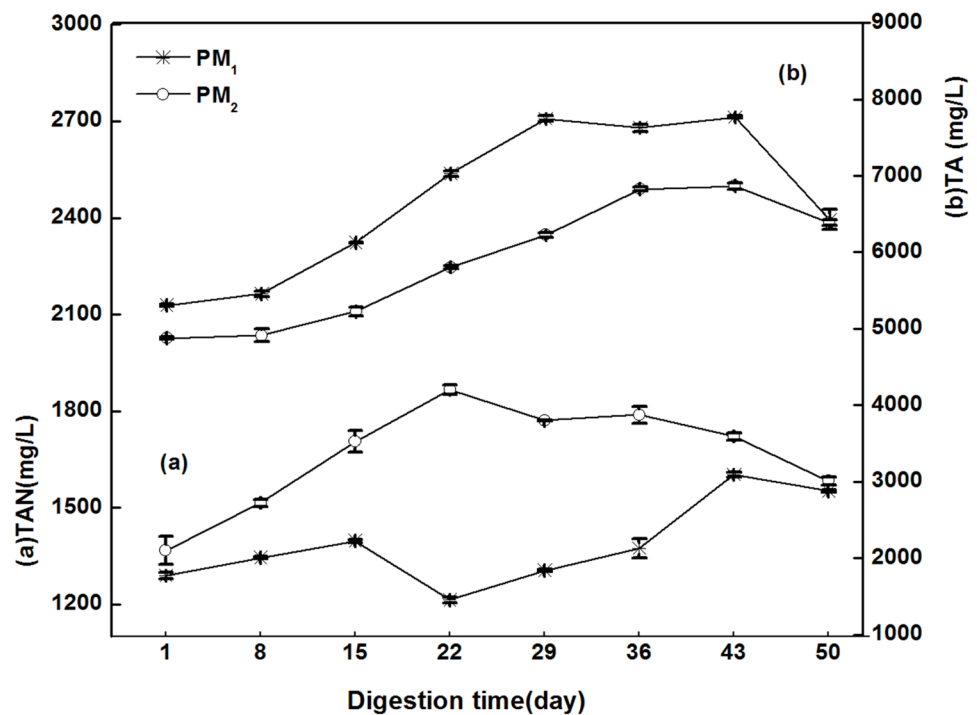


Fig 6. The dynamic changes of TAN (a) and TA (b) of PM₁ and PM₂.

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range of 1500.0 mg/L, and the pH reached 7.6. Thus, the biogas production of PM₂ (2675.0 mL) was significantly less than that of PM₁ (15020.0 mL). The same conclusion was reported on Calli *et al.* who suggested that ammonia inhibition usually occurs when the pH is above 7.4 and TAN is within the range of 1500.0 mg/L to 3000.0 mg/L [35].

In addition, TA is an ideal parameter to monitor the anaerobic digestion process because of the prevention of pH changes in the reactor and support of buffering capacity [18]. Fig 6(b) shows that TA in PM₁ and PM₂ first increased and then decreased until the 43rd day, which probably contributed to the downward trend of VFAs in the digester. The TA value in PM₁ increased from 5308.0 mg/L to 7746.0 mg/L in the first 29 days of fermentation, then reached a peak of 7766.0 mg/L on the 43rd day, and finally the amount was gradually reduced to 6426.0 mg/L. Similarly to PM₁, the TA of PM₂ also peaked at 6866.0 mg/L on the 43rd day, but the peak value of the TA in PM₂ was less than that of PM₁ (7766.0 mg/L). Then number decreased to 6386.0 mg/L. As seen in Fig 6(b), the average TA concentration in PM₁ (6684.8 mg/L) was higher than that of PM₂ (5891.8 mg/L), but the amount of TA in each group returned to close to the initial value by the end of the experiment.

Direct and indirect path coefficients between the four parameters and CBP

Path analysis was used to study whether the effects of the four parameters (X_i) on biogas production (Y) are significant and to test the indirect effects of each parameter on CBP by other parameters (X_i→X_j→Y, i ≠ j). The correlation coefficient (r_{ij}; Eq (2)) were then obtained. The results of path analysis show that the p-values of the four parameters were significant under different treatment conditions and the p-value of X^{pH}_{PM1} and X^{pH}_{PM2} were 0.0032 and 0.0026, respectively. According to Eq (2), the direct path coefficients (b_i) added to the indirect path coefficients (r_{ij}b_j) were equal to the correlation coefficients (r_{iy}).

Table 2 describes path analysis between VFAs, TAN, TA and pH and CBP of PM₁ and PM₂. For PM₁, X^{TA}_{PM1} obtained the maximum b_i on CBP (-0.6327). However, it had the lowest r_{iy} (-0.2190; p < 0.05) with CBP because its b_i was counterbalanced by the r_{ij}b_j (0.4137) generated from the interaction among X^{TA}_{PM1} and X^{VFAs}_{PM1}, X^{TAN}_{PM1} and X^{pH}_{PM1} included in the sum

Table 2. Path analysis between VFAs, TAN, TA and pH and CBP of PM₁ and PM₂.

Parameters	P-value	Direct path coefficients (b _i)	Indirect path coefficients (r _{ij} b _j)				Correlation coefficients (r _{iy})	
			X ^{VFAs} _{PM1}	X ^{TAN} _{PM1}	X ^{TA} _{PM1}	X ^{pH} _{PM1}	Total	
X ^{VFAs} _{PM1}	0.0253*	0.2227		-0.2766	0.7613	-0.1395	0.3452	0.5679
X ^{TAN} _{PM1}	0.0056**	0.3417	0.0344		-0.3355	0.583	0.2819	0.6236
X ^{TA} _{PM1}	0.0341*	-0.6327	0.3529	-0.2879		0.3487	0.4137	-0.219
X ^{pH} _{PM1}	0.0032**	0.5163	-0.1286	0.2749	0.1761		0.3224	0.8387
			X ^{VFAs} _{PM2}	X ^{TAN} _{PM2}	X ^{TA} _{PM2}	X ^{pH} _{PM2}	r _{ij} b _j	r _{iy}
X ^{VFAs} _{PM2}	0.0456*	-0.5013		-0.1766	0.6796	0.3395	0.8425	0.3412
X ^{TAN} _{PM2}	0.0371*	0.8335	0.0265		-0.2373	-0.1513	-0.3621	0.4714
X ^{TA} _{PM2}	0.0044**	0.4764	0.0344	-0.2355		0.6198	0.4187	0.8951
X ^{pH} _{PM2}	0.0026**	0.7247	-0.0286	0.0651	0.1761		0.2126	0.9373
Correlation coefficients: X ^{pH} _{PM1} > X ^{TAN} _{PM1} > X ^{VFAs} _{PM1} > X ^{TA} _{PM1} X ^{pH} _{PM2} > X ^{TA} _{PM2} > X ^{TAN} _{PM2} > X ^{VFAs} _{PM2}								

Note:

* P<0.05;

** P<0.01.

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of $X_{PM1}^{TA} \rightarrow X_{PM1}^{VFAs} \rightarrow CBP$, $X_{PM1}^{TA} \rightarrow X_{PM1}^{TAN} \rightarrow CBP$, $X_{PM1}^{TA} \rightarrow X_{PM1}^{pH} \rightarrow CBP$. Under the total dark condition, X_{PM2}^{TAN} had the same result as X_{PM1}^{TA} . X_{PM2}^{TAN} achieved the maximum b_i values on CBP (0.8355) and the $r_{ij}b_j$ value was -0.3621. The value of r_{ij} was 0.4714 and lower than that of X_{PM2}^{TA} (0.8951) and X_{PM2}^{pH} (0.9373). Furthermore, for PM_1 and PM_2 , the maximum r_{ij} was the pH value.

The complicated path network relationships between the four parameters and DBP indicate that a single large direct effect (b_i) does not always imply a strong correlation between X_i and Y . Therefore, further analyses were conducted to take into account the effect of interactions on biogas production and the influence on the physiological properties of the fermentation process [17–20, 36].

Effects of the four parameters' dynamic changes and interactions

As shown in Table 2 and Fig 5(b), the maximum r_{ij} for the two sets was the pH value, but PM_1 had a similar pH as PM_2 , which was achieved with higher biogas and methane potentials under sunlight-dark conditions. The changes of CBP were shown in Fig 7(a). CBP of PM_1 during the first half of anaerobic fermentation grew faster than that during the second half, whereas the CBP of PM_2 gradually increased. The total CBP of PM_1 (15020.0 mL) was 5.6 times as much as that of PM_2 (2675.0 mL). This result indicates that the difference of CBP was not due to pH alone. Therefore, further investigation is needed to evaluate the indirect effects of different parameters on biogas production.

Fig 7(b) shows that the DBP of PM_2 was lower than that of PM_1 and resulted in reduced methanogen reactivity, which was caused by higher average accumulations of VFAs in PM_2 (3286.4 mg/L) and higher amount of TAN in PM_2 (1866.5 mg/L) with a pH 7.4 on the 22nd day. Calliet *et al.* explained that ammonia inhibition usually occurs when the pH is above 7.4 and TAN is within the range of 1500.0 mg/L to 3000.0 mg/L [35]. Moreover, hydrolysis acidification easily occurs in PM digestion that has large amounts of protein and carbohydrates and low levels of lipids. Thus, the low average total alkalinity in PM_2 (5891.0 mg/L) resulted in a low buffer capacity and reduced ability to prevent the acidification of fermentation [37]. According to Table 2, the r_{ij} between VFAs and CBP for PM_2 had a minimum value of 0.3412 but the indirect effect generated by VFAs \rightarrow TA \rightarrow CBP ($X_{PM2}^{VFAs} \rightarrow X_{PM2}^{TA} \rightarrow CBP$) reached 0.6796 and the indirect effect generated by TAN \rightarrow pH \rightarrow CBP ($X_{PM2}^{TAN} \rightarrow X_{PM2}^{pH} \rightarrow CBP$) reached 0.1513, which may be plausible reasons for the shorter fermentation time and lower biogas production in PM_2 than in PM_1 . Along with fermentation, the DBP of PM_1 and PM_2 gradually increased and peaked with increasing pH levels and the r_{ij} between pH and DBP was largest. The maximum DBP of PM_1 was 740.0 mL/d on the 21st day, whereas that of PM_2 was 419 mL/d on the 14th day. The maximal biogas yield occurred at a pH of 6.5 to 7.5 [38], which is consistent with the findings of the current study. After the peak, the DBP began to slide.

Fig 7(c) shows the changes of CH₄ content. PM_1 had higher CH₄ potentials than PM_2 . PM_1 showed the highest methane content of 51.9% on the 29th day, followed by a sharp decrease to 8.8% at the end of the experiment. This trend confirmed the change of DBP in PM_1 (Fig 7(b)). For PM_2 , the low biogas yield in the short fermentation time (20 days) caused the CH₄ content to be lower than that of PM_1 and rapidly dropped after reaching the maximum value (27.8%) on the 22nd day.

Conclusion

The differences in four parameters caused by sunlight-dark conditions significantly affected the CBP. PM_1 achieved 15020.0 mL of CBP, which was 5.6 times as much as PM_2 . Direct ($X_i \rightarrow Y$) and indirect effects ($X_i \rightarrow X_j \rightarrow Y$) among four parameters on CBP determined the values of r_{ij} .

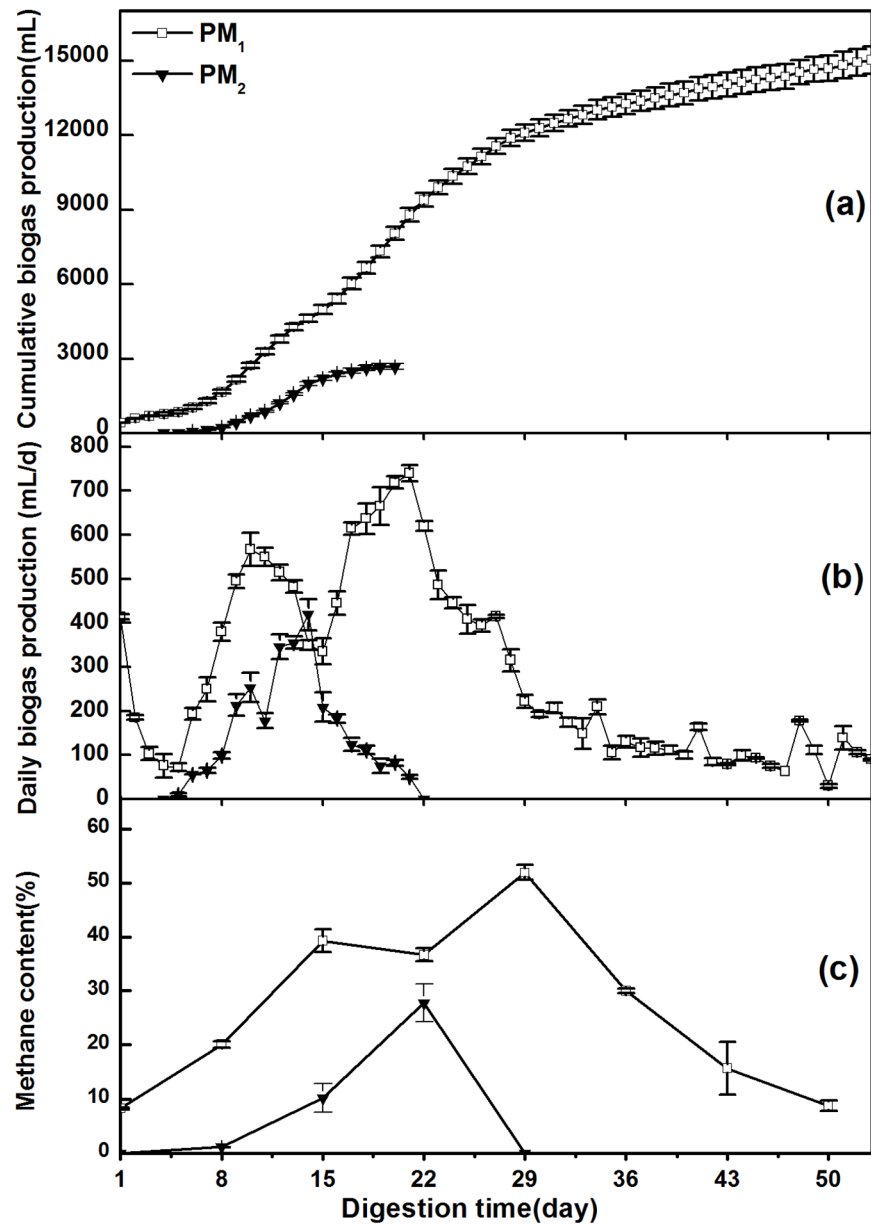


Fig 7. Cumulative biogas production (a), Daily biogas production (b) and Methane content (c) of PM₁ and PM₂.

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that, which were different in PM₁ ($X^{pH}_{PM1} > X^{TAN}_{PM1} > X^{VFAs}_{PM1} > X^{TA}_{PM1}$) and PM₂ ($X^{pH}_{PM2} > X^{TA}_{PM2} > X^{TAN}_{PM2} > X^{VFAs}_{PM2}$). It was suggested that the dynamic change of pH had the most dramatic effect on the fermentation performance of PM₁ and PM₂ and TA and VFA had the weakest influence on the fermentation performance of PM₁ and PM₂, respectively.

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Author Contributions

Conceived and designed the experiments: DXY YZF GHY XJW XHH. Performed the experiments: DXY NNZ. Analyzed the data: DXY WL. Contributed reagents/materials/analysis tools: DXY WL NNZ. Wrote the paper: DXY GHY.

References

1. Song M, Shin SG, Hwang S. Methanogenic population dynamics assessed by real-time quantitative PCR in sludge granule in upflow anaerobic sludge blanket treating swine wastewater. *Bioresour. Technol.* 2010; 101: S23–S28. doi: [10.1016/j.biortech.2009.03.054](https://doi.org/10.1016/j.biortech.2009.03.054) PMID: [19369072](https://pubmed.ncbi.nlm.nih.gov/19369072/)
2. Zhang DD, Yuan XF, Guo P, Suo Y, Wang XF, Wang WD, et al. Microbial population dynamics and changes in main nutrients during the acidification process of pig manures. *Journal of Environmental Sciences(China)*. 2011; 23: 497–505. PMID: [21520820](https://pubmed.ncbi.nlm.nih.gov/21520820/)
3. Yin DX, Liu W, Zhai NN, Yang GH, Wang XJ, Feng YZ, et al. Anaerobic digestion of pig and dairy manure under photo-dark fermentation condition. *Bioresour. Technol.* 2014; 166: 373–380. doi: [10.1016/j.biortech.2014.05.037](https://doi.org/10.1016/j.biortech.2014.05.037) PMID: [24929281](https://pubmed.ncbi.nlm.nih.gov/24929281/)
4. Shi JC, Liao XD, Wu YB, Liang JB. Effect of antibiotics on methane arising from anaerobic digestion of pig manure. *Anim. Feed Sci. Technol.* 2011; 166: 457–463.
5. Wang XJ, Yang GH, Feng YZ, Ren GX, Han XH. Optimizing feeding composition and carbon—nitrogen ratios for improved methane yield during anaerobic co-digestion of dairy, chicken manure and wheat straw. *Bioresour. Technol.* 2012; 120: 78–83. doi: [10.1016/j.biortech.2012.06.058](https://doi.org/10.1016/j.biortech.2012.06.058) PMID: [22784956](https://pubmed.ncbi.nlm.nih.gov/22784956/)
6. Angelidaki I, Ahring B. Anaerobic thermophilic digestion of manure at different ammonia loads: effect of temperature. *Water Res.* 1994; 28: 727–731.
7. Angelidaki I, Ahring BK. Methods for increasing the biogas potential from the recalcitrant organic matter contained in manure. *Water Sci. Technol.* 2000; 41: 189–194. PMID: [11381991](https://pubmed.ncbi.nlm.nih.gov/11381991/)
8. Hansen KH, Angelidaki I, Ahring BK. Improving thermophilic anaerobic digestion of swine manure. *Water Res.* 1999; 33: 1805–1810.
9. Bonmat A, Flotats X, Mateu L, Campos E. Study of thermal hydrolysis as a pretreatment to mesophilic anaerobic digestion of pig slurry. *Water Sci. Technol.* 2001; 44: 109–116. PMID: [11804081](https://pubmed.ncbi.nlm.nih.gov/11804081/)
10. Motte J- C, Trably E, Escudie R, Hamelin J, Steyer J- P, Bernet N, et al. Total solids content: a key parameter of metabolic pathways in dry anaerobic digestion. *Biotechnology for biofuels*. 2013; 6: 164. doi: [10.1186/1754-6834-6-164](https://doi.org/10.1186/1754-6834-6-164) PMID: [24261971](https://pubmed.ncbi.nlm.nih.gov/24261971/)
11. Azwar M, Hussain M, Abdul-Wahab A. Development of biohydrogen production by photobiological, fermentation and electrochemical processes: A review. *Renewable and Sustainable Energy Reviews*. 2014; 31: 158–173.
12. Ghimire A, Frunzo L, Pirozzi F, Trably E, Escudie R, Lens PN, et al. A review on dark fermentative biohydrogen production from organic biomass: Process parameters and use of by-products. *Applied Energy*. 2015; 144: 73–95.
13. Argun H, Kargi F. Effects of light source, intensity and lighting regime on bio-hydrogen production from ground wheat starch by combined dark and photo-fermentations. *Int J Hydrogen Energ.* 2010; 35: 1604–1612.
14. Rittmann S, Herwig C. A comprehensive and quantitative review of dark fermentative biohydrogen production. *Microb Cell Fact.* 2012; 11: 115. doi: [10.1186/1475-2859-11-115](https://doi.org/10.1186/1475-2859-11-115) PMID: [22925149](https://pubmed.ncbi.nlm.nih.gov/22925149/)
15. Levin DB, Zhu H, Beland M, Cicek N, Holbein BE. Potential for hydrogen and methane production from biomass residues in Canada. *Bioresour. Technol.* 2007; 98: 654–660. PMID: [16580198](https://pubmed.ncbi.nlm.nih.gov/16580198/)
16. Chandra R, Mohan SV. Enhanced bio-hydrogenesis by co-culturing photosynthetic bacteria with acidogenic process: Augmented dark-photo fermentative hybrid system to regulate volatile fatty acid inhibition. *Int J Hydrogen Energ.* 2014; 39: 7604–7615.
17. Ahring BK, Sandberg M, Angelidaki I. Volatile fatty acids as indicators of process imbalance in anaerobic digestors. *Appl. Microbiol. Biotechnol.* 1995; 43: 559–565.
18. Ward AJ, Hobbs PJ, Holliman PJ, Jones DL. Evaluation of near infrared spectroscopy and software sensor methods for determination of total alkalinity in anaerobic digesters. *Bioresour. Technol.* 2011; 102: 4083–4090. doi: [10.1016/j.biortech.2010.12.046](https://doi.org/10.1016/j.biortech.2010.12.046) PMID: [21227685](https://pubmed.ncbi.nlm.nih.gov/21227685/)
19. Lins P, Illmer P. Effects of volatile fatty acids, ammonium and agitation on thermophilic methane production from biogas plant sludge in lab-scale experiments. *Folia Microbiol.* 2012; 57: 313–316. doi: [10.1007/s12223-012-0132-7](https://doi.org/10.1007/s12223-012-0132-7) PMID: [22588621](https://pubmed.ncbi.nlm.nih.gov/22588621/)

20. Alcaraz-Gonzalez V, Azael Fregoso-Sanchez F, Oscar Mendez-Acosta H, Gonzalez-Alvarez V. Robust Regulation of Alkalinity in Highly Uncertain Continuous Anaerobic Digestion Processes. *Clean-Soil Air Water*. 2013; 41: 1157–1164.
21. Wang XJ, Yang GH, Li F, Feng YZ, Ren GX, Han XH. Evaluation of two statistical methods for optimizing the feeding composition in anaerobic co-digestion: Mixture design and central composite design. *Bioresour. Technol*. 2013; 131: 172–178. doi: [10.1016/j.biortech.2012.12.174](https://doi.org/10.1016/j.biortech.2012.12.174) PMID: [23347932](https://pubmed.ncbi.nlm.nih.gov/23347932/)
22. Cuetos MJ, Fernández C, Gómez X, Morán A. Anaerobic co-digestion of swine manure with energy crop residues. *Biotechnol Bioproc E*. 2011; 16: 1044–1052.
23. Apha A (1995) WPCF, Standard methods for the examination of water and wastewater: American Public Health Association/American Water Works Association/Water Environment Federation, Washington DC, USA.
24. Mo H. Path Coefficient and its Application. *Journal of Jiangsu Agricultural College* 1983 4: 45–51.
25. Yuan ZF, Zhou JY (2002) *Multivariate Statistical Analysis*. Beijing: Science Press. (In Chinese). Beijing: Science Press. (In Chinese).
26. Wang QH, Kuninobu M, Ogawa HI, Kato Y. Degradation of volatile fatty acids in highly efficient anaerobic digestion. *Biomass & Bioenergy*. 1999; 16: 407–416.
27. Page LH, Ni J-Q, Heber AJ, Mosier NS, Liu X, Joo H-S, et al. Characteristics of volatile fatty acids in stored dairy manure before and after anaerobic digestion. *Biosys. Eng*. 2014; 118: 16–28.
28. Harper SR, Pohland FG. Recent developments in hydrogen management during anaerobic biological wastewater treatment. *Biotechnol. Bioeng*. 1986; 28: 585–602. PMID: [18555364](https://pubmed.ncbi.nlm.nih.gov/18555364/)
29. Ev Münch, Greenfield PF. Estimating VFA concentrations in prefermenters by measuring pH. *Water Res*. 1998; 32: 2431–2441.
30. Kus F, Wiesmann U. Degradation kinetics of acetate and propionate by immobilized anaerobic mixed cultures. *Water Res*. 1995; 29: 1437–1443. PMID: [8582912](https://pubmed.ncbi.nlm.nih.gov/8582912/)
31. Lay JJ, Li YY, Noike T. Influences of pH and moisture content on the methane production in high-solids sludge digestion. *Water Res*. 1997; 31: 1518–1524.
32. Astals S, Nolla-Ardèvol V, Mata-Alvarez J. Anaerobic co-digestion of pig manure and crude glycerol at mesophilic conditions: Biogas and digestate. *Bioresour. Technol*. 2012; 110: 63–70. doi: [10.1016/j.biortech.2012.01.080](https://doi.org/10.1016/j.biortech.2012.01.080) PMID: [22341889](https://pubmed.ncbi.nlm.nih.gov/22341889/)
33. Yenigün O, Demirel B. Ammonia inhibition in anaerobic digestion: a review. *Process Biochem*. 2013; 48: 901–911.
34. Walker M, Iyer K, Heaven S, Banks C. Ammonia removal in anaerobic digestion by biogas stripping: an evaluation of process alternatives using a first order rate model based on experimental findings. *Chem. Eng. J*. 2011; 178: 138–145.
35. Calli B, Mertoglu B, Inanc B, Yenigun O. Effects of high free ammonia concentrations on the performances of anaerobic bioreactors. *Process Biochem*. 2005; 40: 1285–1292.
36. Björnsson L, Murto M, Mattiasson B. Evaluation of parameters for monitoring an anaerobic co-digestion process. *Appl. Microbiol. Biotechnol*. 2000; 54: 844–849. PMID: [11152079](https://pubmed.ncbi.nlm.nih.gov/11152079/)
37. Fernandez N, Montalvo S, Borja R, Guerrero L, Sanchez E, Cortes I, et al. Performance evaluation of an anaerobic fluidized bed reactor with natural zeolite as support material when treating high-strength distillery wastewater. *Renew Energ*. 2008; 33: 2458–2466.
38. Danesh S, Oleszkiewicz JA. Volatile fatty acid production and uptake in biological nutrient removal systems with process separation. *Water Environ. Res*. 1997; 69: 1106–1111.