



Nickel mine soil is a potential source for soybean plant growth promoting and heavy metal tolerant rhizobia

Han Liu^{1,*}, Yongliang Cui^{2,*}, Jie Zhou¹, Petri Penttinen¹, Jiahao Liu¹, Lan Zeng¹, Qiang Chen¹, Yunfu Gu¹, Likou Zou¹, Ke Zhao¹, Quanju Xiang¹ and Xiumei Yu¹

¹ College of Resources, Sichuan Agricultural University, Chengdu, Sichuan, China

² Sichuan Provincial Academy of Natural Resource and Sciences, Chengdu, Sichuan, China

* These authors contributed equally to this work.

ABSTRACT

Mine soil is not only barren but also contaminated by some heavy metals. It is unclear whether some rhizobia survived under extreme conditions in the nickel mine soil. Therefore, this study tries to isolate some effective soybean plant growth promoting and heavy metal resistant rhizobia from nickel mine soil, and to analyze their diversity. Soybean plants were used to trap rhizobia from the nickel mine soil. A total of 21 isolates were preliminarily identified as rhizobia, which were clustered into eight groups at 87% similarity level using BOXA1R-PCR fingerprinting technique. Four out of the eight representative isolates formed nodules on soybean roots with effectively symbiotic nitrogen-fixing and plant growth promoting abilities in the soybean pot experiment. Phylogenetic analysis of 16S rRNA, four housekeeping genes (*atpD-recA-glnII-rpoB*) and *nifH* genes assigned the symbiotic isolates YN5, YN8 and YN10 into *Ensifer xinjiangense* and YN11 into *Rhizobium radiobacter*, respectively. They also showed different tolerance levels to the heavy metals including cadmium, chromium, copper, nickel, and zinc. It was concluded that there were some plant growth promoting and heavy metal resistant rhizobia with the potential to facilitate phytoremediation and alleviate the effects of heavy metals on soybean cultivation in nickel mine soil, indicating a novel evidence for further exploring more functional microbes from the nickel mine soil.

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Corresponding author

Xiumei Yu, xiumeiyu@sicau.edu.cn

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page 14

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INTRODUCTION

Heavy metal contamination in mining related soils affects both the mining site and the surrounding environment. Heavy metals in soil can originate from natural minerals, yet anthropogenic activities are the main source (Lebrazi & Fikri-Benbrahim, 2018). Heavy metal contamination is a risk to food security, ecological environment, and even to human health through bioaccumulation in the food chain (Zhang et al., 2012; Zhou et al., 2013; Long et al., 2021; Qin et al., 2021; Tauqeer, Turan & Iqbal, 2022). Therefore, areas with

severe heavy metal-contaminations need to be remediated before being used for the cultivation of crops, and the selected crops should not accumulate contaminants when growing on the slightly-contaminated areas. Remediating the contaminated soils requires efficient and economical methods such as bioremediation, which is considered eco-friendly, secondary contamination-free, and suitable for non-point source contamination (Shao *et al.*, 2020; Zhang *et al.*, 2020; Yu *et al.*, 2021).

Phytoremediation, especially *in-situ* enhanced phytoremediation, is widely used for the bioremediation of heavy metal-contaminated soil (Thakare *et al.*, 2021). The growth of plants in contaminated soil can be facilitated by utilizing the biological nitrogen fixation (BNF) ability of legume-rhizobia symbionts (Hao *et al.*, 2014; Yu *et al.*, 2017; Wang *et al.*, 2019; Salmi & Boulila, 2021). For example, soybean (*Glycine max* L. Merrill) is applicable in remediating heavy metal-contaminated sites (Li *et al.*, 2019). In the symbiosis, rhizobia induce the formation of nodules on the roots of the host plant. Inside the nodules, rhizobia fix atmospheric nitrogen into ammonia which serves as a N source for the legume (Lindstrom & Mousavi, 2020; Wang *et al.*, 2020). Inoculation of effective N fixing rhizobial strains leads to the growth promotion of legumes (Catroux, Hartmann & Revellin, 2001). It has been proposed that strains suitable for legume-rhizobia phytoremediation can be isolated from the contaminated sites (Limcharoensuk *et al.*, 2015; Fan *et al.*, 2016; Balakrishnan *et al.*, 2017; Dhuldhaj & Pandya, 2020). Rhizobia include strains with heavy metal resistance and are capable to promote plant growth under heavy metal stress (Grandlic, Palmer & Maier, 2009; Fan *et al.*, 2018b). It showed that a copper-resistant *S. meliloti* strain promoted the growth of alfalfa under copper stress (Duan *et al.*, 2019). Rhizobial strains resistant against several heavy metals have the potential to be applied in the *in-situ* bioremediation of soils contaminated with multiple heavy metals (Grandlic, Palmer & Maier, 2009; Abd-Alla *et al.*, 2012; Yu *et al.*, 2014; Hao *et al.*, 2015; Yu *et al.*, 2017; Ke *et al.*, 2021). However, indigenous rhizobia resources that could be applied in *in situ* phytoremediation are still scarce in Southwest China.

We hypothesized that heavy metal-contaminated soil could be a putative source for such strains. Thus, soybean plants were used to trap rhizobia from nickel mine soil in Xichang, Sichuan Province, China. The isolates were identified using molecular methods, and soybean growth-promoting abilities and heavy metal resistance of these strains were tested. The results provide better understanding of the potential of using indigenous microbial resources for the alleviation of heavy metal contamination.

MATERIALS AND METHODS

Sampling and soil analysis

Soil samples were collected from a nickel mine (26°46'54.2"N, 102°06'53.2"E) in Xichang, Sichuan Province, China. Three sampling sites spaced 50 to 100 m apart were randomly selected within the area. From the sampling points, five topsoil (0–20 cm) subsamples spaced at least 5 m apart were collected and mixed to make one composite sample per site. The composite samples were quartered and stored on ice before being taken to the laboratory. The soil samples were ground, passed through a 2 mm nylon sieve, and air-dried.

Soil water content in the fresh soil was determined by measuring the weight difference after soil samples had been dried at 105 °C for eight hours. Soil pH was measured using a PHS-3C pH meter (Shanghai Yoke, Shanghai, China) in a 2.5:1 water-soil slurry which had been left to settle overnight. Soil organic carbon content was determined using the $K_2Cr_2O_7$ - H_2SO_4 method (Schumacher, 2002). Total nitrogen, phosphorus and potassium contents in the soil samples were determined using Kjeldahl method (Kjeltec 8400, FOSS, Sweden), Mo-Sb colorimetric method (WFJ2100, UNICO, China) and flame spectrophotometry (FP6410, Shanghai Precision & Scientific, China), respectively (Murphy & Riley, 1962; Page, Miller & Keeney, 1982; Yu et al., 2021). Available nitrogen content was determined using the alkali N-proliferation method; soil available phosphorus and potassium were extracted with sodium bicarbonate solution and NH_4AC solution, respectively, and measured using the previously described method (Wu et al., 2021). The contents of heavy metals were determined after digestion using mixed acid (HNO_3 : $HClO_4$ = 3: 1) using inductively coupled plasma optical emission spectrometry (ICP-OES; IRIS Intrepid II; Thermo Electron Corporation, Waltham, MA, USA).

Trapping and isolation of rhizobia

The indigenous rhizobial strains in the nickel mine soil were trapped using soybean cultivar Nandou No.12 bred by Nanchong Institution of Agricultural Sciences, Sichuan, China. The soybean seeds were sterilized by dipping into 95% alcohol for 3 min and 1% $HgCl_2$ for 5 min, followed by rinsing with sterile water (Yu et al., 2017). The soybean seeds were germinated in a pot filled with sterilized moist vermiculite in the dark for 24 h, and transplanted into pots filled with soil collected from the nickel mine site (S_N). The soybean plants were harvested after 90 days. Three root nodules per soybean plant were selected and sterilized using the above-mentioned methods. The nodules were incubated on beef extract peptone agar at 28 °C, and the surface sterilization was considered successful when no colonies appeared in 24 h. After that, nodules were crushed in plastic tubes and inoculated on yeast-extract mannitol agar (yeast extract 1.5 g L^{-1} , mannitol 1.0 g L^{-1} , K_2HPO_4 0.5 g L^{-1} , $MgSO_4 \cdot 7H_2O$ 0.2 g L^{-1} , NaCl 0.1 g L^{-1} , Congo red 0.04 g L^{-1} , agar 20 g L^{-1}). The plates were maintained at 28 °C for 7 to 10 days. During the incubation period, single colonies were selected based on colony morphology and purified by repeated streaking (Yu et al., 2017). Purified isolates with round, plump, milky white, mucilaginous and smooth-edged colonies were examined using light microscopy after Gram staining. Gram-negative isolates with rod-shaped cells were preserved in 20% (w/v) glycerol at -80 °C.

Phylogenetic diversity analysis

The isolates were grown in YM liquid medium (yeast extract 1.5 g L^{-1} , mannitol 1.0 g L^{-1} , K_2HPO_4 0.5 g L^{-1} , $MgSO_4 \cdot 7H_2O$ 0.2 g L^{-1} , NaCl 0.1 g L^{-1} , Congo red 0.04 g L^{-1}) for 24 h at 28 °C in a shaking incubator, and DNA was extracted using TIANamp Bacteria DNA Kit (TIANGEN, China). The genetic diversity of the isolates was assessed using BOX-A1R PCR fingerprinting with the primer 5'-CCTCGGCAAGGACGCTGACG-3' (Chen et al., 2014). Amplification was done in a 10 μ L volume system containing 5 μ L of 2 \times PCR

mix, 0.2 μL of the primer ($10 \mu\text{mol L}^{-1}$), 1 μL of template DNA (50 ng mL^{-1}), and 3.8 μL double distilled water (ddH_2O). The BOX-A1R PCR thermal profile included initial denaturation at 94°C for 3 min, 30 cycles of denaturation at 94°C for 1 min, annealing at 52°C for 1 min and extension at 65°C for 8 min, and a final extension at 65°C for 16 min. The amplified fragments (8 μL) from the amplification mixture and the 200 bp DNA ladder were separated in 2% (w/v) agarose gel with ethidium bromide at 80 V for 2.5 h, and were visualized under UV light with the patterns recorded. Based on the patterns, a BOX-A1R PCR cluster tree diagram was created using the unweighted pair group method with arithmetic averages (UPGMA) in NTSYSpc 2.1 (Yu et al., 2014).

Plant growth promotion ability test

Based on the BOX-A1R PCR fingerprints, eight representative rhizobial isolates were selected for the test of plant growth promotion ability using Leonard jars. Leonard jar and nitrogen-free nutrient solution were prepared as previously described (Kang et al., 2018; Yu et al., 2017). Soybean seeds were surface-sterilized and germinated as described above. Three sterilized seeds were transferred into the topsoil of a Leonard jar, which was then covered with a layer of approximately 3 cm moist vermiculite. When the soybean seedlings were 2–3 cm tall, the rhizobia culture in exponential phase ($\text{OD}_{600 \text{ nm}} 0.6\sim 0.8$) was inoculated around the roots, and then the topsoil was covered with a layer of 1 cm autoclaved quartz sand. The non-inoculated soybean supplemented with nitrogen-free nutrient solution was set as the negative control, while the N^+ treatment with $1 \text{ g L}^{-1}\text{KNO}_3$ as the nitrogen source in the nutrient solution was the positive control. Each treatment and control included three replicates. The soybean plants were kept in a greenhouse with artificial light-dark cycles: 17 h light, temperature 25°C and humidity 80% (to simulate daytime); 7 h dark, temperature 17°C and humidity 85% (to simulate nighttime). The jars were replenished with nutrient solution when needed. The soybean plants were harvested after 55 days. The number of nodules, root length, and plant height and weight were measured. The plant samples were dried at 105°C for 30 min to deactivate enzymes, and then at 55°C until constant weight. The nitrogen, phosphorus, and potassium contents of the dried roots and shoots were measured.

Sequence analysis

The isolates that nodulated soybean plants were further characterized using sequence analyses. The almost full length 16S rRNA gene was amplified using the primer pair 27F/1492R (Yu et al., 2017), the housekeeping genes *atpD*, *recA*, *glnII*, and *rpoB* using primer pairs *atpD* 255F/ *atpD* 782R, *recA* 63F/*recA* 555R, *glnII* 12F/*glnII* tsR, and *rpoB* 454F/ *rpoB* 1364R, respectively (Tang et al., 2012; Zhao et al., 2014), and the nitrogen fixation gene *nifH* using the primer pair *nifHF*/*nifHI* (Laguerre et al., 2001). All of the primers we used in this study were listed in Table S2. Amplification was done in a 30 μL volume system containing 15 μL of $2\times$ PCR mix, 0.15 μL of each primer, one μL of template DNA (50 ng mL^{-1}), and 13.7 μL ddH_2O . For the amplification of 16S rRNA, *atpD*, *recA*, *glnII* and *rpoB* genes, the thermal cycling conditions included an initial denaturation at 94°C for 3 min, 30 cycles of denaturation at 94°C for 1 min, annealing at 55°C for 1 min,

extension at 72 °C for 2 min, and a final extension at 72 °C for 10 min. For the amplification of *nifH*, the thermal cycling conditions included an initial denaturation at 95 °C for 3 min, 30 cycles of denaturation at 94 °C for 1 min, annealing at 59 °C for 1 min, extension at 72 °C for 5 min, and a final extension at 72 °C for 6 min. The amplified products were sequenced by Sangon Biotech (Shanghai, China). The fragments of *atpD* (274 bp), *recA* (245 bp), *glnIII* (413 bp) and *rpoB* (534 bp) were concatenated for multilocus sequence analysis (MLSA). The sequences were matched against reference sequences of rhizobia in the GenBank of NCBI database using BLAST. The reference sequences of different rhizobia species we used for the phylogenetic analysis of the three genes (16SrRNA, housekeeping genes, and *nifH*) were separately listed in [Tables S3, S4, and S5](#). The sequences of the isolates and the reference sequences were subjected to phylogenetic analysis using neighbor joining method in MEGA7.0. The phylogenetic trees were bootstrapped with 1,000 replications for each sequence to evaluate the reliability of the tree topologies ([Saitou & Nei, 1987](#)).

Heavy metal resistance test

For assessing the heavy metal-resistance ability of the isolates, 10 g L⁻¹ stock solutions of Cd²⁺ (CdCl₂ 2.5H₂O), Cr⁶⁺ (K₂Cr₂O₇), Cu²⁺ (CuCl₂ 2H₂O), Ni²⁺ (NiSO₄), and Zn²⁺ (ZnSO₄ 7H₂O) were prepared by dissolving the salts in ultrapure water. The isolates were grown in 5 mL YM liquid medium with 0, 4, 8, 12, 16, 20, 40, 60, 80, 100 mg L⁻¹ of each metal in an orbital shaker at 28 °C for 72 h, after which the optical density (OD_{600 nm}) of the cultures were measured using a spectrophotometer (UV-3300; Shanghai MAPADA, Shanghai, China) ([Abd-Alla et al., 2012](#)).

The minimum inhibitory concentrations (MIC) were determined by comparing the OD_{600 nm} value of the cultures spiked with metals to those without metals. MIC was defined as the lowest metal concentration resulting in a visually observable decrease in the growth curve of the isolates. Minimum lethal concentration (MLC) was defined as the lowest metal concentration resulting in an OD_{600 nm} value lower than 0.1.

Statistical analysis

Statistical differences in soil properties and soybean growth parameters were tested using one-way ANOVA and Fisher's protected LSD (least significant difference) test at $P \leq 0.05$ in IBM SPSS statistics 22.0. The differences in MIC and MLC values were not tested due to their zero variance.

RESULTS

Isolation of rhizobia from nickel mine soil

The soil in the nickel mining area was slightly acidic (pH 6.35), and the organic matter content was low (1.07%). The contents of total N, P and K were 261.35, 479.85, and 5869.73 mg kg⁻¹, respectively, and 15.08%, 2.93%, and 0.38% of them were available fragments ([Table 1](#)). Both nickel and lead contents were approximately 20 mg kg⁻¹, and cadmium, chromium, copper, iron, manganese and zinc were also detected ([Table 1](#)).

As a result of trapping rhizobial strains from the nickel mine soil using soybean as the host plant, 21 isolates were preliminarily identified as rhizobia based on colony

Table 1 The physicochemical properties and heavy metal contents of the nickel mine soil collected in Xichang, Sichuan Province, China.

| | | | | | | | | | |
|---------------|---------------------------|---------------------------|---------------------------|---------------------------|---------------------------|---------------------------|---------------------------|---------------------------|---------------------------|
| Properties | pH | WC (%) | OM (%) | TN (mg kg ⁻¹) | AN (mg kg ⁻¹) | TP (mg kg ⁻¹) | AP (mg kg ⁻¹) | TK (mg kg ⁻¹) | AK (mg kg ⁻¹) |
| Average value | 6.35 ± 0.01 | 7.71 ± 0.00 | 1.07 ± 0.06 | 261.35 ± 22.74 | 39.21 ± 3.04 | 479.85 ± 0.00 | 140.69 ± 1.44 | 5869.73 ± 0.05 | 22.68 ± 2.91 |
| Properties | Ni (mg kg ⁻¹) | Cu (mg kg ⁻¹) | Cr (mg kg ⁻¹) | Pb (mg kg ⁻¹) | Mn (mg kg ⁻¹) | Fe (mg kg ⁻¹) | Zn (mg kg ⁻¹) | Cd (mg kg ⁻¹) | |
| Average value | 19.69 ± 0.44 | 9.79 ± 0.10 | 12.78 ± 0.08 | 20.13 ± 0.03 | 8.57 ± 0.04 | 744.20 ± 4.10 | 80.07 ± 1.04 | 0.46 ± 0.01 | |

Notes.

WC, water content; OM, soil organic matter content; TN, soil total nitrogen content; AN, soil available nitrogen content; TP, soil total phosphorus content; AP, soil available phosphorus content; TK, soil total potassium content; AK, soil available potassium content; Ni, nickel; Cu, copper; Cr, chromium; Pb, lead; Mn, manganese; Fe, iron; Zn, zinc; Cd, cadmium. The values are expressed as mean standard deviation ($n = 3$).

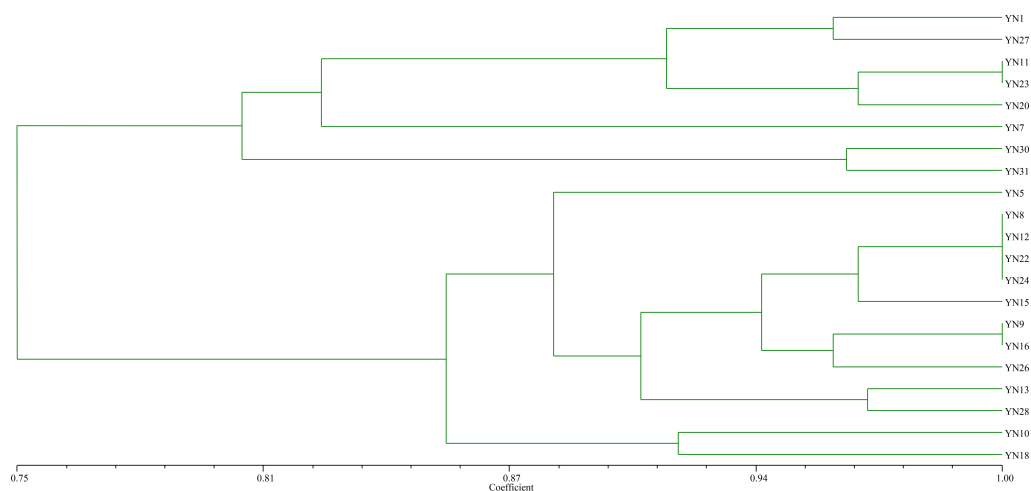


Figure 1 Dendrogram of the BOX-A1R PCR fingerprints of 21 isolates trapped by soybean from the nickel mine soil collected in Xichang, Sichuan Province, China.

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morphology, microscopic examination of cell shape, and Gram staining. The BOXA1R-PCR fingerprinting (Fig. 1) of the isolates showed that sixteen distinct fingerprint patterns were found, indicating that these isolates from nickel mine soil were genetically diverse. The isolates were clustered into 8 groups at 87% similarity level.

Plant growth promotion ability of selected isolates

Based on the fingerprinting, we selected 8 isolates (YN1, YN5, YN8, YN10, YN11, YN13, YN27, YN30) for the plant growth promotion ability test. Only four isolates (YN5, YN8, YN10, YN11) nodulated soybean plants, with nodule numbers ranging from 50 to 54 per plant (mean value of six plants). The biomass of soybeans inoculated with the four symbiotic isolates (YN5, YN8, YN10, YN11) and isolate YN13 was significantly higher than that in the N^- treatment ($P = 0.00$) (Fig. 2A). The biomass of soybeans inoculated with isolate YN8 was on the same level as in the N^+ treatment. The shoots of soybeans inoculated with the 4 symbiotic isolates (YN5, YN8, YN10, YN11) and isolate YN1 were significantly longer than those in the N^- treatment ($P = 0.00, 0.01, 0.00, 0.00, 0.00$). The roots of soybeans inoculated with the symbiotic isolates YN8, YN10, and YN11 were significantly longer than those in the N^- treatment ($P = 0.00$) (Fig. 2B). The shoot N content of soybean plants inoculated with the symbiotic strains was at the same level as in the N^+ treatment, and the shoot N content of soybeans inoculated with the symbiotic strains of YN5, YN8, YN10, and YN11 as well as the isolate YN1 was significantly higher than that in the N^- treatment ($P = 0.00$) (Fig. 3A). The root N content of inoculated soybean plants was higher than that in the N^- treatment (Fig. 3A). The root P content of soybeans inoculated with the symbiotic isolates and isolate YN30 was lower than that in the N^- treatment (Fig. 3B). The shoot K content of soybeans inoculated with the isolate YN10 was higher than that in the N^- treatment (Fig. 3C). The root K content of soybeans

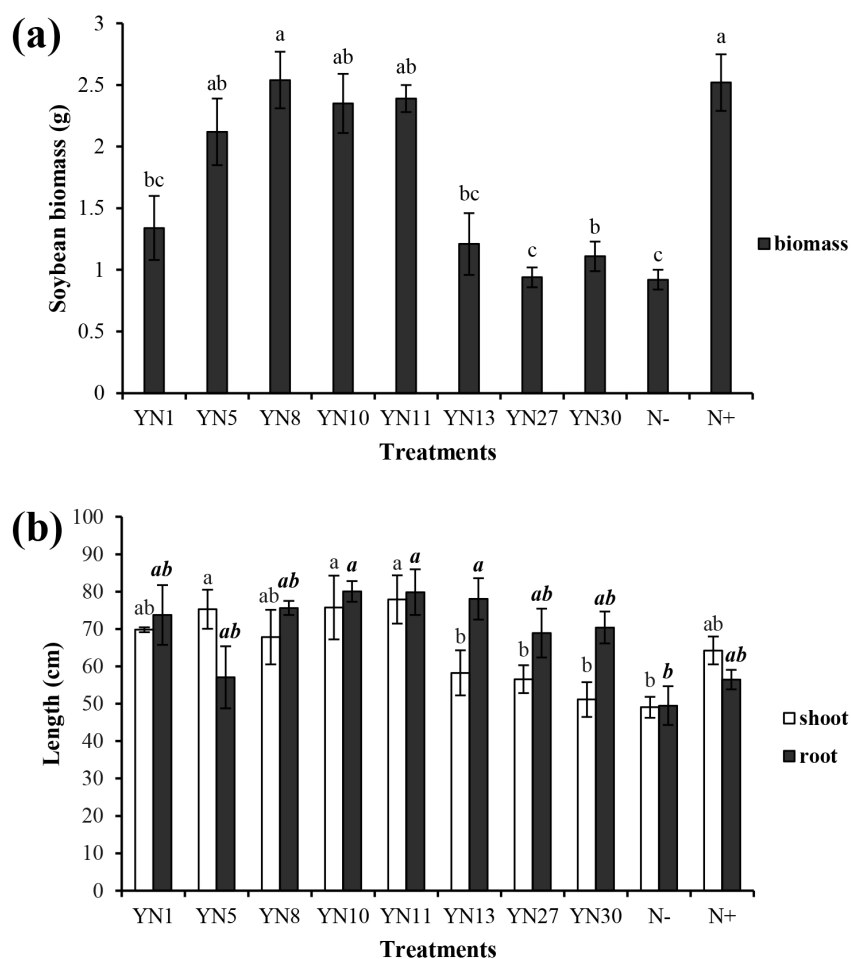


Figure 2 Biomass (A) and shoot and root lengths (B) of soybean plants inoculated with isolates trapped from the nickel mine soil collected in Xichang, Sichuan Province, China. Different letters above the bars indicate significant statistical difference at $p < 0.05$ ($n = 3$). Comparison between shoot lengths is in normal font and comparison between root lengths is in bold.

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inoculated with the isolates YN5, YN8, YN10, YN13, YN27 and YN30 was higher than that in the N^- treatment (Fig. 3C).

Sequence analyses of symbiotic isolates

As a result of the sequence analysis of the almost full length 16S rRNA gene, the symbiotic isolates were assigned to the genera *Ensifer* (formerly designated as *Sinorhizobium*) and *Rhizobium* (Fig. 4). The isolates YN5, YN8, and YN10 were respectively grouped together with the type strains *E. fredii* USDA205, *E. americanum* CFNEI156 and *E. xinjiangense* CCBAU110, and isolate YN11 with *R. radiobacter* ICMP 5856 (formerly *Agrobacterium tumefaciens* ICMP 5856). Based on the multilocus sequence analysis (MLSA) of the concatenated fragments of *atpD* (274 bp), *recA* (245 bp), *glnII* (413 bp) and *rpoB* (534 bp), the isolates YN5, YN8, and YN10 were identified as *E. xinjiangense* strains, and YN11 as *Rhizobium radiobacter* strain (Fig. 5). The *nifH* sequences from YN5, YN8, and YN10 were

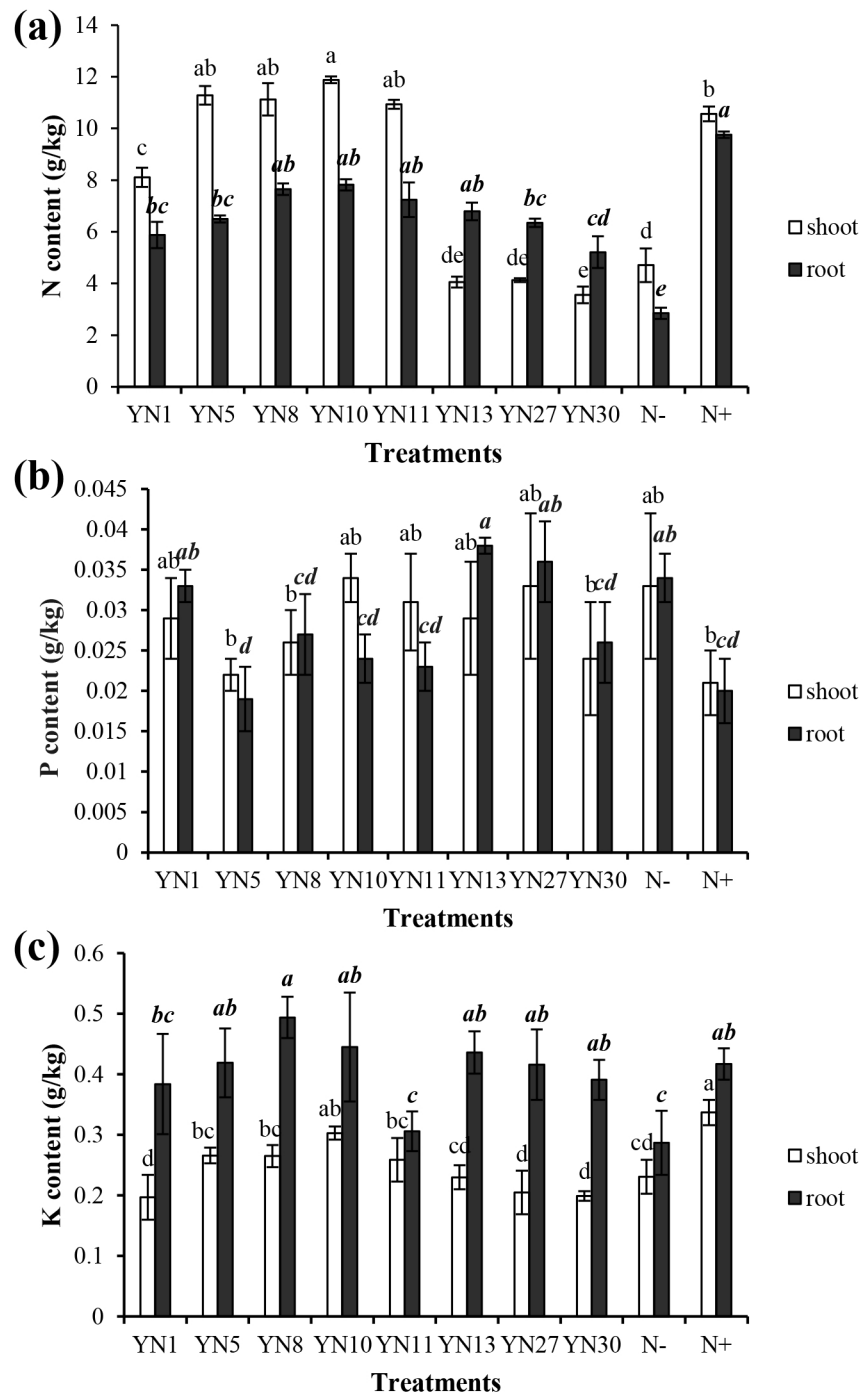


Figure 3 Nitrogen (A), phosphorus (B) and potassium (C) contents in the shoots and roots of soybean plants inoculated with the isolates trapped from the nickel mine soil collected in Xichang, Sichuan Province, China. Different letters above bars indicate significant statistical difference at $p < 0.05$ ($n = 3$). Comparison between shoot lengths is in normal font and comparison between root lengths is in bold.

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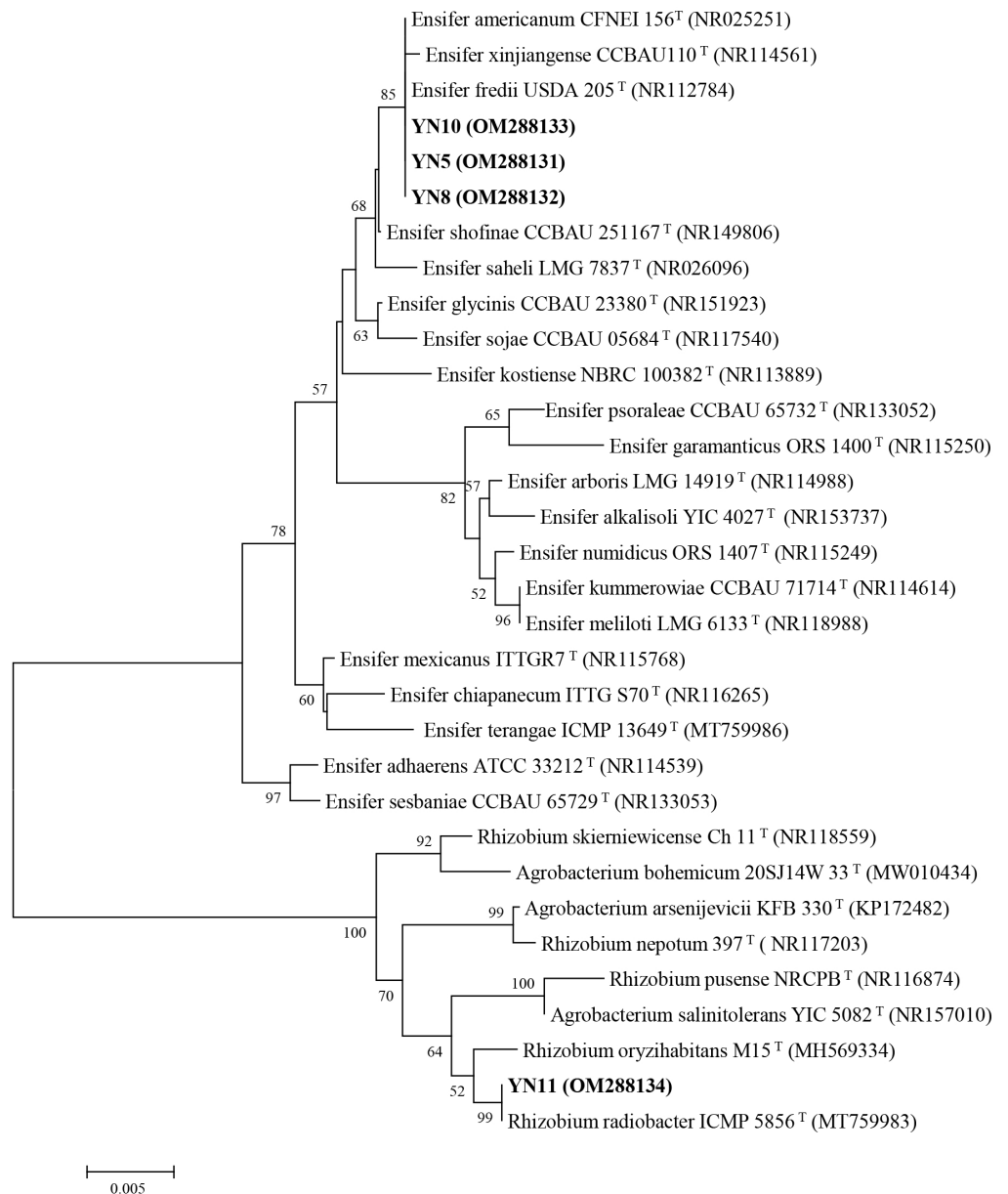


Figure 4 16S rRNA gene phylogeny of the soybean symbiotic isolates from the nickel mine soil collected in Xichang, Sichuan Province, China (in bold), and reference strains. GenBank accession numbers are in parentheses. Bootstrap values >50% are shown at the branch points. The scale bar denotes 5% substitutions per site.

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100% similar with those from *E. fredii* CCBAU23314 and *E. xinjiangense* CCBAU110, and the *nifH* sequence from YN11 with that from *R. radiobacter* (Fig. 6).

Heavy metal resistance ability of symbiotic isolates

In general, *E. xinjiangense* YN5 and *R. radiobacter* YN11 tolerated higher levels of heavy metals compared to *E. xinjiangense* YN8 and YN10 (Fig. 7, Table S1). For Cd^{2+} , the MIC

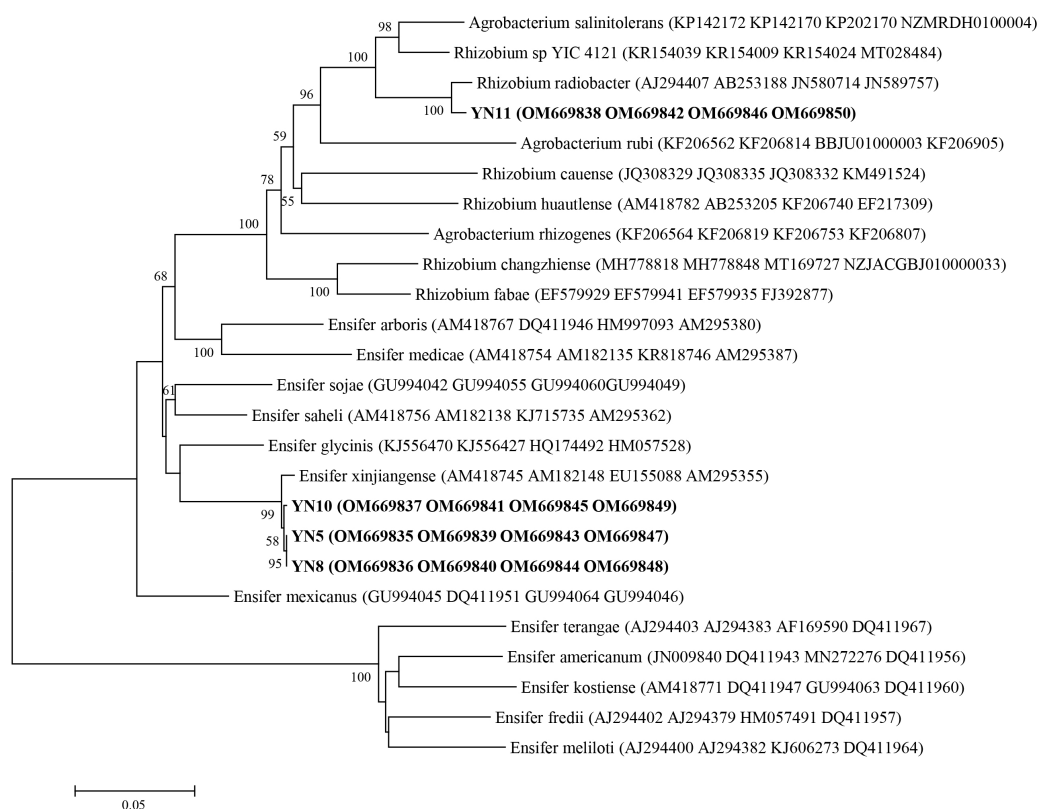


Figure 5 Multilocus sequence analysis (MLSA) of the soybean symbiotic isolates from the nickel mine soil collected in Xichang, Sichuan Province, China (in bold), and reference strains. MLSA was done using the concatenated fragments of *atpD* (274 bp), *recA* (245 bp), *glnII* (413 bp) and *rpoB* (534 bp). GenBank accession numbers are in parentheses. Bootstrap values > 50% are shown at the branch points. The scale bar denotes 5% substitutions per site.

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and MLC values of *E. xinjiangense* YN5 and *R. radiobacter* YN11 were the highest. For Cr^{6+} , the MLC of all the strains was 16 mg L^{-1} . For Cu^{2+} , the MIC and MLC values of *E. xinjiangense* YN8 were the lowest. For Ni^{2+} , the MLC values of YN5 and YN11 were higher than those of YN8 and YN10. For Zn^{2+} , the MLC value of *E. xinjiangense* YN5 was 300 mg L^{-1} , *i.e.*, over three times higher than that of *R. radiobacter* YN11 and over 18 to 75 times higher than those of YN11 and YN8, respectively.

DISCUSSION

We trapped rhizobia from nickel mine soil using soybean plants in Xichang, Sichuan Province, China, to find effectively plant growth promoting and heavy metal resistant strains for the enhancement of phytoremediation of heavy metal contaminated soil and for the promotion of soybean growth on slightly contaminated farmland. The low organic matter, N, P and K implied that the soil was barren (Zhang *et al.*, 2012; Wu *et al.*, 2021), yet the rhizobia-trapping plants were nodulated, and the isolates from nodules were diverse based on the BOXA1R-PCR fingerprints. However, when inoculated on soybeans, only

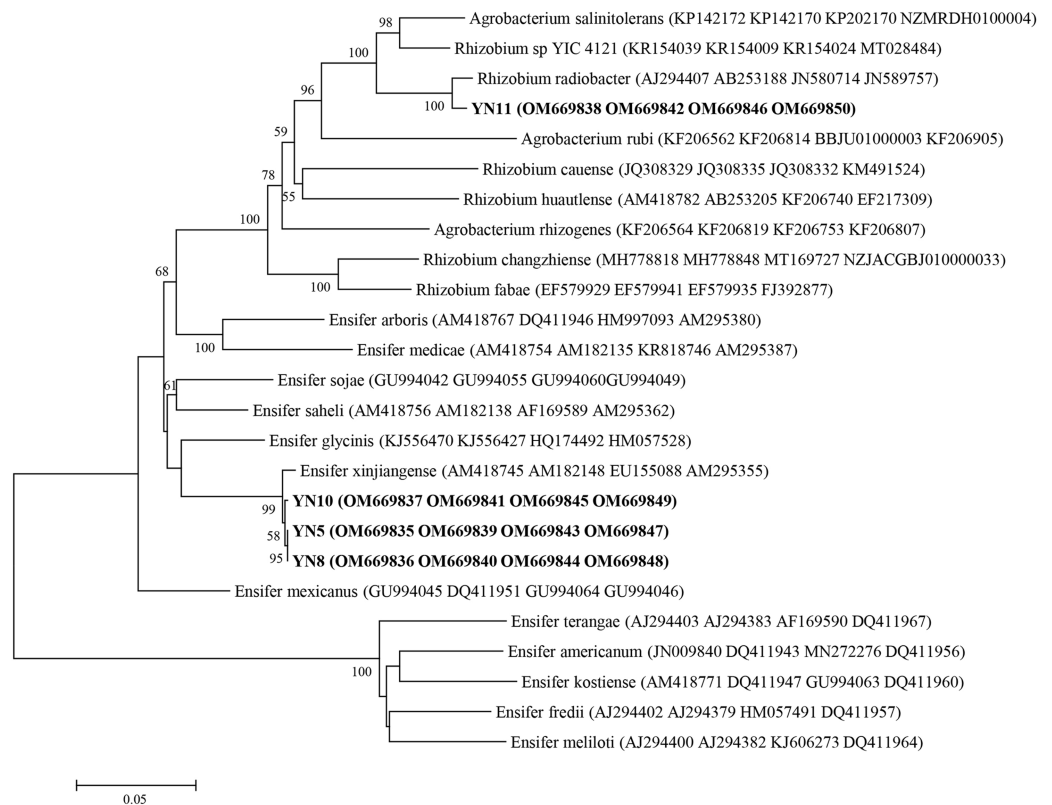


Figure 6 *nifH* phylogeny of the soybean symbiotic isolates from the nickel mine soil collected at Xi-chang, Sichuan Province, China (in bold), and reference strains. GenBank accession numbers are in parentheses. Bootstrap values >50% are shown at the branch points. The scale bar denotes 5% substitutions per site.

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four out of the eight representative isolates formed nodules on the roots. Similar to rhizobia isolated from *Glycyrrhiza* spp. (Li et al., 2012), the four non-nodulating isolates may have been sporadic symbionts or other endophytes that had entered the nodules along with a genuine symbiont. Similar to the model inoculant of soybean, *Bradyrhizobium diazoefficiens* USDA110 (Sibponkrung et al., 2020), inoculation with the symbiotic isolates resulted in over two times higher biomass than in the non-inoculated control; the higher biomass was accompanied with higher shoot nitrogen content. In addition, even the non-nodulating isolates showed some plant growth promoting abilities. Especially, inoculation with all the representative isolates resulted in higher root N content than in the nitrogen-free control. In most of the inoculated plants, the increase in root N content was accompanied with lower P content.

As a host plant, soybean is promiscuous and may be nodulated with both fast- and slow-growing rhizobia (Chen et al., 2021). Based on the 16S rRNA gene analysis, three of our isolates were identified as *Ensifer americanum*, *E. fredii* or *E. xinjiangense*, i.e., species with closely related 16S rRNA genes (Peng et al., 2002; Wang et al., 2013). Further analysis using MLSA of the four housekeeping genes assigned the symbiotic isolates into the fast-growing

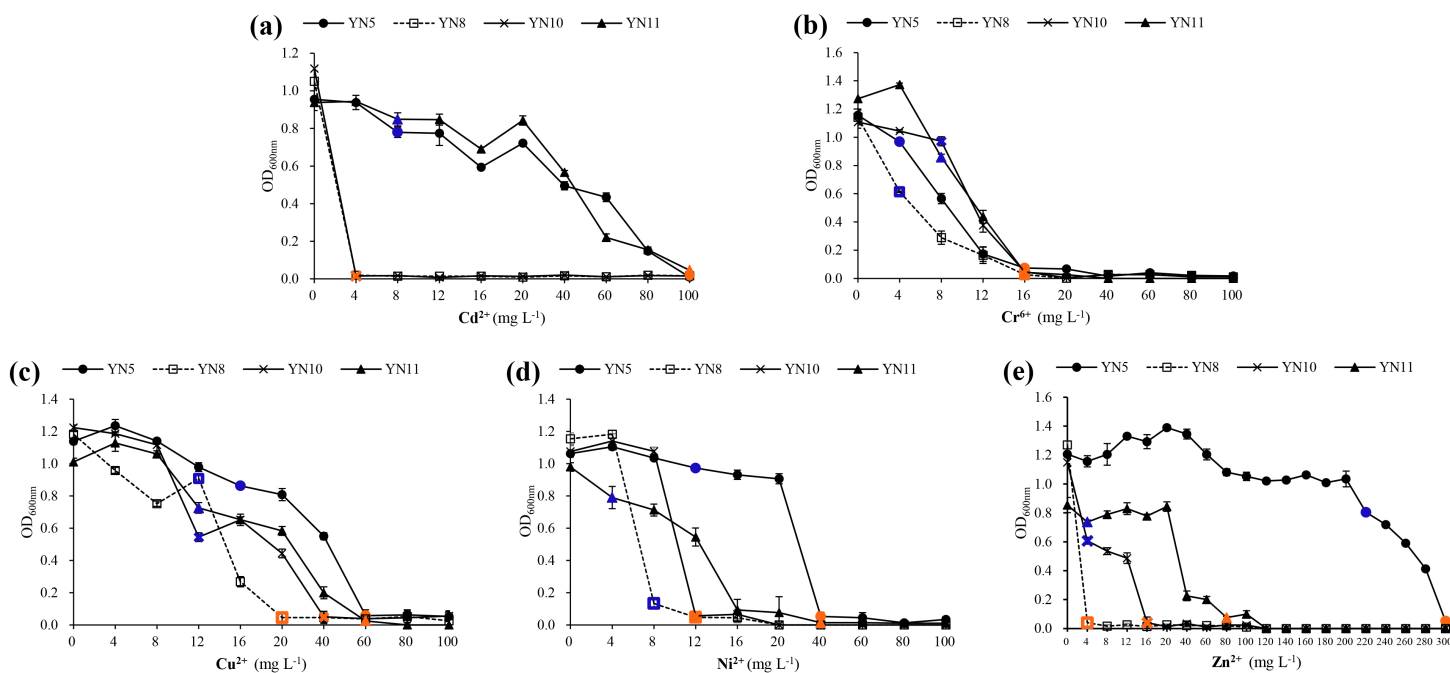


Figure 7 The growth of the soybean symbiotic isolates from the nickel mine soil collected in Xichang, Sichuan Province, China, at increasing concentrations of Cd²⁺ (A), Cr⁶⁺ (B), Cu²⁺ (C), Ni²⁺ (D), and (E) Zn²⁺. Minimum inhibitory concentrations (MIC) are indicated with green and minimum lethal concentrations (MLC) with red.

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rhizobial species *Ensifer xinjiangense* and *Rhizobium radiobacter*, strains of which have been identified as plant growth promoting symbionts of soybean plants (Peng et al., 2002; Iturralde et al., 2019). To our knowledge, neither *E. xinjiangense* (formerly *Sinorhizobium xinjiangense*) nor *R. radiobacter* (formerly *Agrobacterium tumefaciens*) strains have been applied as a single-inoculant plant growth promoter in bioremediation, yet co-inoculation of an *A. tumefaciens* strain with *S. meliloti* promoted the growth of *Medicago lupulina* under Cu and Zn stresses (Jian et al., 2019). *R. radiobacter* is traditionally considered as a plant pathogen and is a free-living nitrogen fixer (Kanvinde, Sastry & Microbiology, 1990). In our study, the amplification and sequencing of the *nifH* gene, which encodes nitrogenase iron protein, showed that both the *Ensifer* strains and *R. radiobacter* YN11 had the genetic potential for nitrogen fixation. The nodulation and plant growth promotion capacity of *R. radiobacter* YN11 added to the growing body of evidence that when carrying the nodulation genes, *Rhizobium* (*Agrobacterium*) clade strains can be legume-nodulating symbionts (Cummings et al., 2009).

In the soil from mining areas, the concentrations of heavy metals vary considerably from below the background values to hundreds of times higher than the average values for all soils (Li et al., 2014). Bacteria inhabiting the heavy metal-contaminated sites include legume nodulating strains with high tolerance against the metals (Mohamad et al., 2017). In our study, the symbiotic strains showed varied heavy metal resistance; *E. xinjiangense* YN5 outperformed the other *E. xinjiangense* isolates and the resistance levels of *R. radiobacter*

YN11 fell in-between. Compared to the rhizobia isolated directly from V-Ti magnetite mine tailing soil and those from the nodules *Robinia pseudoacacia* in a Pb-Zn mining area (Yu et al., 2014; Fan et al., 2018a; Fan et al., 2018b), our isolates tolerated lower levels of Cd²⁺ and Cu²⁺. One possible explanation is the level of contamination on the sites where the strains were isolated; the V-Ti magnetite and Pb-Zn mining areas were seriously contaminated (Yu et al., 2014; Fan et al., 2018a), but only Zn content in the nickel mine soil was higher than in the background value for soils in China (Li et al., 2014). The levels of heavy metals tolerated are approximately 10 to 100 times lower in liquid medium than on solid medium (Hassen et al., 1998). It is also important to take into account the different testing methods for the Zn tolerance of *E. xinjiangense* YN5. The V-Ti magnetite and Pb-Zn mining area isolates were tested on solid media (Yu et al., 2014; Fan et al., 2018a) but our isolates in liquid medium, yet *E. xinjiangense* YN5 tolerated a higher level of Zn²⁺.

CONCLUSIONS

Our study used soybean pot experiment to trap 21 rhizobia strains from nickel mine soil. As a result, we selected three *Ensifer xinjiangense* strains (YN5, YN8, and YN10) and one *Rhizobium radiobacter* (YN11) with good nitrogen fixing ability, which can significantly improve the soybean plant height, root length, and biomass yield. Moreover, these four strains carried the symbiotic gene *nifH* that can encode dinitrogenase reductase enzyme, which further confirmed their abilities to form root nodules and fix nitrogen. *E. xinjiangense* YN5 and *R. radiobacter* YN11 tolerated higher levels of heavy metals than *E. xinjiangense* YN8 and YN10. Taken together, the results showed that the nickel mine soil is a potential source for plant growth promoting rhizobia strains, which could be applied as indigenous inoculants in the phytoremediation of slightly contaminated farmland and in the alleviation of the adverse effects of heavy metals on soybean cultivation.

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Competing Interests

The authors declare there are no competing interests.

Author Contributions

- Han Liu conceived and designed the experiments, performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.
- Yongliang Cui and Xiumei Yu conceived and designed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.
- Jie Zhou performed the experiments, prepared figures and/or tables, and approved the final draft.
- Petri Penttinen analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.
- Jiahao Liu and Lan Zeng performed the experiments, prepared figures and/or tables, and approved the final draft.
- Qiang Chen, Yunfu Gu, Likou Zou, Ke Zhao and Quanju Xiang analyzed the data, authored or reviewed drafts of the paper, and approved the final draft.

Data Availability

The following information was supplied regarding data availability:

The sequences are available at NCBI GenBank: [OM288131–OM288134](#), [OM669835–OM669838](#), [OM669843–OM669846](#), [OM669851–OM669834](#), [OM669839–OM669842](#), [OM669847–OM669850](#).

The raw data, nickel mine soil basic properties, soybean growth, and heavy metal resistance ability of rhizobia isolates are available in the [Supplementary File](#).

Supplemental Information

Supplemental information for this article can be found online at <http://dx.doi.org/10.7717/peerj.13215#supplemental-information>.

REFERENCES

- Abd-Alla MH, Morsy FM, El-Enany A-WE, Ohyama T. 2012.** Isolation and characterization of a heavy-metal-resistant isolate of *Rhizobium leguminosarum* bv. viciae potentially applicable for biosorption of Cd^{2+} and Co^{2+} . *International Biodeterioration & Biodegradation* **67**:48–55 DOI [10.1016/j.ibiod.2011.10.008](#).
- Balakrishnan B, Sahu BK, Kothilmozhan Ranishree J, Lourduraj AV, Nithyanandam M, Packiriswamy N, Panchatcharam P. 2017.** Assessment of heavy metal concentrations and associated resistant bacterial communities in bulk and rhizosphere soil of *Avicennia marina* of Pichavaram mangrove, India. *Environmental Earth Sciences* **76**(1):1 DOI [10.1007/s12665-016-6304-z](#).
- Catroux G, Hartmann A, Revellin C. 2001.** Trends in rhizobial inoculant production and use. *Plant and Soil* **230**:21–30 DOI [10.1023/A:1004777115628](#).

- Chen JY, Gu J, Wang ET, Ma XX, Kang ST, Huang LZ, Cao XP, Li LB, Wu YL. 2014.** Wild peanut *Arachis duranensis* are nodulated by diverse and novel *Bradyrhizobium* species in acid soils. *Systematic Applied Microbiology* 37:525–532 DOI 10.1016/j.syapm.2014.05.004.
- Chen WF, Wang ET, Ji ZJ, Zhang JJ. 2021.** Recent development and new insight of diversification and symbiosis specificity of legume rhizobia: mechanism and application. *Journal of Applied Microbiology* 131:553–563 DOI 10.1111/jam.14960.
- Cummings SP, Gyaneshwar P, Vinuesa P, Farruggia FT, Andrews M, Humphry D, Elliott GN, Nelson A, Orr C, Pettitt D, Shah GR, Santos SR, Krishnan HB, Odee D, Moreira FM, Sprent JI, Young JP, James EK. 2009.** Nodulation of Sesbania species by *Rhizobium* (*Agrobacterium*) strain IRBG74 and other rhizobia. *Environmental Microbiology* 11:2510–2525 DOI 10.1111/j.1462-2920.2009.01975.x.
- Dhuldhaj U, Pandya U. 2020.** Combinatorial study of heavy metal and microbe interaction and resistance mechanism consort to microbial system. *Geomicrobiology Journal* 38:181–189.
- Duan C, Razavi BS, Shen G, Cui Y, Ju W, Li S, Fang L. 2019.** Deciphering the rhizobium inoculation effect on spatial distribution of phosphatase activity in the rhizosphere of alfalfa under copper stress. *Soil Biology and Biochemistry* 137:107574 DOI 10.1016/j.soilbio.2019.107574.
- Fan M, Lin Y, Huo H, Liu Y, Zhao L, Wang E, Chen W, Wei G. 2016.** Microbial communities in riparian soils of a settling pond for mine drainage treatment. *Water Research* 96:198–207 DOI 10.1016/j.watres.2016.03.061.
- Fan M, Liu Z, Nan L, Wang E, Chen W, Lin Y, Wei G. 2018a.** Isolation, characterization, and selection of heavy metal-resistant and plant growth-promoting endophytic bacteria from root nodules of *Robinia pseudoacacia* in a Pb/Zn mining area. *Microbiological Research* 217:51–59 DOI 10.1016/j.micres.2018.09.002.
- Fan M, Xiao X, Guo Y, Zhang J, Wang E, Chen W, Lin Y, Wei G. 2018b.** Enhanced phytoremediation of *Robinia pseudoacacia* in heavy metal-contaminated soils with rhizobia and the associated bacterial community structure and function. *Chemosphere* 197:729–740 DOI 10.1016/j.chemosphere.2018.01.102.
- Grandlic CJ, Palmer MW, Maier RM. 2009.** Optimization of plant growth-promoting bacteria-assisted phytostabilization of mine tailings. *Soil Biology and Biochemistry* 41:1734–1740 DOI 10.1016/j.soilbio.2009.05.017.
- Hao X, Taghavi S, Xie P, Orbach MJ, Alwathnani HA, Rensing C, Wei G. 2014.** Phytoremediation of heavy and transition metals aided by legume-rhizobia symbiosis. *International Journal of Phytoremediation* 16:179–202 DOI 10.1080/15226514.2013.773273.
- Hao X, Xie P, Zhu YG, Taghavi S, Wei G, Rensing C. 2015.** Copper tolerance mechanisms of *Mesorhizobium amorphae* and its role in aiding phytostabilization by *Robinia pseudoacacia* in copper contaminated soil. *Environmental Science and Technology* 49:2328–2340 DOI 10.1021/es504956a.

- Hassen A, Saidi N, Cherif M, Boudabous AJBT. 1998. Resistance of environmental bacteria to heavy metals. *Bioresource Technology* 64:7–15
DOI 10.1016/S0960-8524(97)00161-2.
- Iturralde ET, Covelli JM, Alvarez F, Pérez-Giménez J, Arrese-Igor C, Lodeiro AR. 2019. Soybean-nodulating strains with low intrinsic competitiveness for nodulation, good symbiotic performance, and stress-tolerance isolated from soybean-cropped soils in Argentina. *Frontiers in Microbiology* 10:1061 DOI 10.3389/fmicb.2019.01061.
- Jian L, Bai X, Zhang H, Song X, Li Z. 2019. Promotion of growth and metal accumulation of alfalfa by coinoculation with *Sinorhizobium* and *Agrobacterium* under copper and zinc stress. *PeerJ* 7:e6875 DOI 10.7717/peerj.6875.
- Kang X, Yu X, Zhang Y, Cui Y, Tu W, Wang Q, Li Y, Hu L, Gu Y, Zhao K, Xiang Q, Chen Q, Ma M, Li Zou, Zhang X, Kang J. 2018. Inoculation of *Sinorhizobium saheli* YH1 leads to reduced metal uptake for *Leucaena leucocephala* grown in mine tailings and metal-polluted soils. *Frontiers in Microbiology* 9:1853
DOI 10.3389/fmicb.2018.01853.
- Kanvinde L, Sastry GJA, Microbiology E. 1990. *Agrobacterium tumefaciens* is a diazotrophic bacterium. *Applied & Environmental Microbiology* 56:2087–2092
DOI 10.1128/aem.56.7.2087-2092.1990.
- Ke T, Guo G, Liu J, Zhang C, Tao Y, Wang P, Xu Y, Chen L. 2021. Improvement of the Cu and Cd phytostabilization efficiency of perennial ryegrass through the inoculation of three metal-resistant PGPR strains. *Environmental Pollution* 271:116314
DOI 10.1016/j.envpol.2020.116314.
- Laguerre G, Nour SM, Macheret V, Sanjuan J, Drouin P, Amarger N. 2001. Classification of rhizobia based on nodC and *nifH* gene analysis reveals a close phylogenetic relationship among *Phaseolus vulgaris* symbionts. *Microbiology* 147(Pt 4):981–993
DOI 10.1099/00221287-147-4-981.
- Lebrazi S, Fikri-Benbrahim K. 2018. Rhizobium-legume symbioses: heavy metal effects and principal approaches for bioremediation of contaminated soil. In: *Legumes for soil health and sustainable management*. Springer Singapore: Springer, 205–233.
- Li Z, Ma Z, Van der Kuijp TJ, Yuan Z, Huang L. 2014. A review of soil heavy metal pollution from mines in China: pollution and health risk assessment. *Science of Total Environment* 468–469:843–853 DOI 10.1016/j.scitotenv.2013.08.090.
- Li L, Sinkko H, Montonen L, Wei G, Lindström K, Räsänen LA. 2012. Biogeography of symbiotic and other endophytic bacteria isolated from medicinal *Glycyrrhiza* species in China. *FEMS Microbiology Ecology* 79:46–68
DOI 10.1111/j.1574-6941.2011.01198.x.
- Li X, Wang X, Chen Y, Yang X, Cui Z. 2019. Optimization of combined phytoremediation for heavy metal contaminated mine tailings by a field-scale orthogonal experiment. *Ecotoxicology and Environmental Safety* 168:1–8
DOI 10.1016/j.ecoenv.2018.10.012.
- Limcharoensuk T, Sooksawat N, Sumarnrote A, Awutpet T, Kruatrachue M, Pokethitiyook P, Auesukaree C. 2015. Bioaccumulation and biosorption of Cd²⁺ and Zn²⁺

- by bacteria isolated from a zinc mine in Thailand. *Ecotoxicology and Environmental Safety* **122**:322–330 DOI [10.1016/j.ecoenv.2015.08.013](https://doi.org/10.1016/j.ecoenv.2015.08.013).
- Lindstrom K, Mousavi SA. 2020.** Effectiveness of nitrogen fixation in rhizobia. *Microbial Biotechnology* **13**:1314–1335 DOI [10.1111/1751-7915.13517](https://doi.org/10.1111/1751-7915.13517).
- Long Z, Huang Y, Zhang W, Shi Z, Yu D, Chen Y, Liu C, Wang R. 2021.** Effect of different industrial activities on soil heavy metal pollution, ecological risk, and health risk. *Environmental Monitoring and Assessment* **193**:1–12 DOI [10.1007/s10661-020-08746-9](https://doi.org/10.1007/s10661-020-08746-9).
- Mohamad R, Maynaud G, Le Quere A, Vidal C, Klonowska A, Yashiro E, Cleyet-Marel JC, Brunel B. 2017.** Ancient heavy metal contamination in soils as a driver of tolerant *Anthyllis vulneraria* rhizobial communities. *Applied & Environmental Microbiology* **83**(2):e01735-16 DOI [10.1128/AEM.01735-16](https://doi.org/10.1128/AEM.01735-16).
- Murphy J, Riley J. 1962.** A modified single solution method for the determination of phosphate in natural waters. *Analytica Chimica Acta* **27**:31–36 DOI [10.1016/S0003-2670\(00\)88444-5](https://doi.org/10.1016/S0003-2670(00)88444-5).
- Page A, Miller R, Keeney D. 1982.** *Methods of soil analysis, part II*. 15. Madison: ASoA, 99–100.
- Peng G, Tan Z, Wang E, Reinhold-Hurek B, Chen WF, Chen WX. 2002.** Identification of isolates from soybean nodules in Xinjiang Region as *Sinorhizobium xinjiangense* and genetic differentiation of *S. xinjiangense* from *Sinorhizobium fredii*. *International Journal of Systematic and Evolutionary Microbiology* **52**:457–462 DOI [10.1099/00207713-52-2-457](https://doi.org/10.1099/00207713-52-2-457).
- Qin G, Niu Z, Yu J, Li Z, Ma J, Xiang P. 2021.** Soil heavy metal pollution and food safety in China: effects, sources and removing technology. *Chemosphere* **267**:129205 DOI [10.1016/j.chemosphere.2020.129205](https://doi.org/10.1016/j.chemosphere.2020.129205).
- Saitou N, Nei M. 1987.** The neighbor-joining method a new method for reconstructing phylogenetic trees. *Molecular Biology & Evolution* **4**:406.
- Salmi A, Boulila F. 2021.** Heavy metals multi-tolerant *Bradyrhizobium* isolated from mercury mining region in Algeria. *Journal of Environmental Management* **289**:112547 DOI [10.1016/j.jenvman.2021.112547](https://doi.org/10.1016/j.jenvman.2021.112547).
- Schumacher BA. 2002.** Methods for the determination of Total Organic Carbon (TOC) in Soils and Sediments. *Ecological Risk Assessment Support Center*.
- Shao Y, Yan T, Wang K, Huang S, Yuan W, Qin FGF. 2020.** Soil heavy metal lead pollution and its stabilization remediation technology. *Energy Reports* **6**:122–127 DOI [10.1016/j.egy.2020.11.074](https://doi.org/10.1016/j.egy.2020.11.074).
- Sibponkrung S, Kondo T, Tanaka K, Tittabutr P, Boonkerd N, Yoshida KI, Teamroong N. 2020.** Co-inoculation of *Bacillus velezensis* strain S141 and *Bradyrhizobium* strains promotes nodule growth and nitrogen fixation. *Microorganisms* **8**(5):678 DOI [10.3390/microorganisms8050678](https://doi.org/10.3390/microorganisms8050678).
- Tang J, Bromfield ES, Rodrigue N, Cloutier S, Tambong JT. 2012.** Microevolution of symbiotic *Bradyrhizobium* populations associated with soybeans in east North America. *Ecology & Evolution* **2**:2943–2961 DOI [10.1002/ece3.404](https://doi.org/10.1002/ece3.404).

- Tauqeer HM, Turan V, Iqbal M. 2022.** Production of safer vegetables from heavy metals contaminated soils: the current situation, concerns associated with human health and novel management strategies. In: Malik JA, ed. *Advances in bioremediation and phytoremediation for sustainable soil management: principles, monitoring and remediation*. Cham: Springer International Publishing, 301–312.
- Thakare M, Sarma H, Datar S, Roy A, Pawar P, Gupta K, Pandit S, Prasad R. 2021.** Understanding the holistic approach to plant-microbe remediation technologies for removing heavy metals and radionuclides from soil. *Current Research in Biotechnology* 3:84–98 DOI 10.1016/j.crbiot.2021.02.004.
- Wang H, Gu C, Liu X, Yang C, Li W, Wang S. 2020.** Impact of soybean nodulation phenotypes and nitrogen fertilizer levels on the rhizosphere bacterial community. *Frontiers in Microbiology* 11:750 DOI 10.3389/fmicb.2020.00750.
- Wang Q, Ma L, Zhou Q, Chen B, Zhang X, Wu Y, Pan F, Huang L, Yang X, Feng Y. 2019.** Inoculation of plant growth promoting bacteria from hyperaccumulator facilitated non-host root development and provided promising agents for elevated phytoremediation efficiency. *Chemosphere* 234:769–776 DOI 10.1016/j.chemosphere.2019.06.132.
- Wang Y, Wang F, Hou B, Wang E, Chen W, Sui X, Chen W, Li Y, Zhang Y. 2013.** Proposal of *Ensifer psoraleae* sp. nov. *Ensifer sesbaniae* sp. nov. *Ensifer morelense* comb. nov. and *Ensifer americanum* comb. nov. *Systematic and Applied Microbiology* 36:467–473 DOI 10.1016/j.syapm.2013.05.001.
- Wu B, Peng H, Sheng M, Luo H, Wang X, Zhang R, Xu F, Xu H. 2021.** Evaluation of phytoremediation potential of native dominant plants and spatial distribution of heavy metals in abandoned mining area in Southwest China. *Ecotoxicology and Environmental Safety* 220:112368 DOI 10.1016/j.ecoenv.2021.112368.
- Yu X, Li Y, Li Y, Xu C, Cui Y, Xiang Q, Gu Y, Zhao K, Zhang X, Penttinen P. 2017.** *Pongamia pinnata* inoculated with *Bradyrhizobium liaoningense* PZHK1 shows potential for phytoremediation of mine tailings. *Applied Microbiology and Biotechnology* 101:1739–1751 DOI 10.1007/s00253-016-7996-4.
- Yu X, Li Y, Zhang C, Liu H, Liu J, Zheng W, Kang X, Leng X, Zhao K, Gu Y, Zhang X, Xiang Q, Chen Q. 2014.** Culturable heavy metal-resistant and plant growth promoting bacteria in V-Ti magnetite mine tailing soil from Panzhihua, China. *PLOS ONE* 9:e106618 DOI 10.1371/journal.pone.0106618.
- Yu X, Shen T, Kang X, Cui Y, Chen Q, Shoaib M, Liu H, Zhang F, Hussain S, Xiang Q, Zhao K, Gu Y, Ma M, Li S, Zou L, Liang Y. 2021.** Long-term phytoremediation using the symbiotic *Pongamia pinnata* reshaped soil micro-ecological environment. *Science of the Total Environment* 774:145112 DOI 10.1016/j.scitotenv.2021.145112.
- Zhang X, Yang L, Li Y, Li H, Wang W, Ye B. EM and Assessment. 2012.** Impacts of lead/zinc mining and smelting on the environment and human health in China. *Environmental Monitoring & Assessment* 184:2261–2273 DOI 10.1007/s10661-011-2115-6.
- Zhang H, Yuan X, Xiong T, Wang H, Jiang L. 2020.** Bioremediation of co-contaminated soil with heavy metals and pesticides: influence factors, mechanisms and evaluation methods. *Chemical Engineering Journal* 398:125659.

Zhao L, Fan M, Zhang D, Yang R, Zhang F, Xu L, Wei X, Shen Y, Wei G. 2014. Distribution and diversity of rhizobia associated with wild soybean (*Glycine soja* Sieb. & Zucc.) in Northwest China. *Systematic & Applied Microbiology* **37**:449–456 DOI [10.1016/j.syapm.2014.05.011](https://doi.org/10.1016/j.syapm.2014.05.011).

Zhou H, Zeng M, Zhou X, Liao BH, Liu J, Lei M, Zhong QY, Zeng H. 2013. Assessment of heavy metal contamination and bioaccumulation in soybean plants from mining and smelting areas of southern Hunan Province, China. *Environmental Toxicology & Chemistry* **32**:2719–2727 DOI [10.1002/etc.2389](https://doi.org/10.1002/etc.2389).