

Article

Zinc Uptake, Photosynthetic Efficiency and Oxidative Stress in the Seagrass *Cymodocea nodosa* Exposed to ZnO Nanoparticles

Paraskevi Malea¹, Katerina Charitonidou^{1,2}, Ilektra Sperdouli^{1,3}, Zoi Mylona¹ and Michael Moustakas^{1,*}

- ¹ Department of Botany, Aristotle University of Thessaloniki, GR-54124 Thessaloniki, Greece
- ² School of Agricultural Sciences, University of Thessaly, GR-38446 Volos, Greece
- ³ Institute of Plant Breeding and Genetic Resources, Hellenic Agricultural Organisation-Demeter, Thermi, GR-57001 Thessaloniki, Greece
- * Correspondence: moustak@bio.auth.gr; Tel.: +30-2310-99-8335

Received: 22 May 2019; Accepted: 27 June 2019; Published: 29 June 2019



Abstract: We characterized zinc oxide nanoparticles (ZnO NPs) by dynamic light scattering (DLS) measurements, and transmission electron microscopy (TEM), while we evaluated photosystem II (PSII) responses, Zn uptake kinetics, and hydrogen peroxide (H₂O₂) accumulation, in *C. nodosa* exposed to 5 mg L⁻¹ and 10 mg L⁻¹ ZnO NPs for 4 h, 12 h, 24 h, 48 h and 72 h. Four h after exposure to 10 mg L⁻¹ ZnO NPs, we noticed a disturbance of PSII functioning that became more severe after 12 h. However, after a 24 h exposure to 10 mg L⁻¹ ZnO NPs, we observed a hormetic response, with both time and dose as the basal stress levels needed for induction of the adaptive response. This was achieved through the reduced plastoquinone (PQ) pool, at a 12 h exposure, which mediated the generation of chloroplastic H₂O₂; acting as a fast acclimation signaling molecule. Nevertheless, longer treatment (48 h and 72 h) resulted in decreasing the photoprotective mechanism to dissipate excess energy as heat (NPQ) and increasing the quantum yield of non-regulated energy loss (Φ_{NO}). This increased the formation of singlet oxygen (¹O₂), and decreased the fraction of open reaction centers, mostly after a 72-h exposure at 10 mg L⁻¹ ZnO NPs due to increased Zn uptake compared to 5 mg L⁻¹.

Keywords: adaptive response; hormetic response; hydrogen peroxide; marine angiosperms; non-photochemical quenching; photoprotective mechanism; plastoquinone pool; reactive oxygen species (ROS); redox state; zinc oxide nanoparticles

1. Introduction

The small size of nanoparticles (NPs) provides them with special physical and chemical properties that are not found in bulk materials allowing their utilization, among others, in agricultural products, catalysis, cosmetics, electronics, energy production, engineering, food industry, pharmaceutics and textiles [1–5].

Among the variety of metal NPs that are often used for marketable purposes, zinc oxide (ZnO) NPs are the most commonly used ones [6–8]. ZnO NPs, with their unique chemical and physical properties, such as high photostability, broad range of radiation absorption, high electrochemical coupling coefficient, and high chemical stability, are widely used in a diversity of applications, varying from paints to chemicals, from tires to ceramics, and from pharmaceuticals to agriculture [3,9]. ZnO NPs are specifically used in clothing, skin care products, anticancer medicines, sunscreens, coatings for solar cells, bottle coatings, and gas sensors [10].



The rapid expansion, due to their unique properties, and their release in the environment has raised considerable worries regarding manufactured NPs [11]. An essential aspect of the risk assessment of NPs is to understand their interactions with plants, a basic component of all ecosystems [11]. Thus, the extensive use of ZnO NPs has extended the requirements of research on their consequences on living organisms [6]. Environmental levels of ZnO-NPs were stated to be among $3.1–31 \ \mu g \ kg^{-1}$ soil and 76–760 $\mu g \ L^{-1}$ water [12]. Manufactured NPs are unavoidably released into the soil and through streams, rivers, and sewage treatment they finally reach the sea [5]. Since NPs finally end in aquatic ecosystems, aquatic plants may be at higher risks than terrestrial. Thus, there is a need to evaluate the risks related to NP presence in aquatic ecosystems [12].

In photosynthesis, electron transport is mediated by photosystem I (PSI), and photosystem II (PSII) that work coordinately in the thylakoid membranes [13–15]. The most prone constituent of the photosynthetic apparatus to environmental stress is thought to be PSII [15–18]. Perturbations of PSII functionality resulted in declining the photosynthetic capacity, limiting the growth and development of plants, and reducing crop production [19,20]. Measuring PSII function by chlorophyll fluorescence imaging analysis is the most appropriate methodology to detect NPs-induced stress on plants [21–23] Metal oxides NPs alter photosynthetic efficiency by reducing energy transfer efficiency and quantum yield [5,21,23].

Reactive oxygen species (ROS) generated as byproducts in chloroplasts by the light reactions of photosynthesis are responsible for NPs induced toxicity [24], and thus the impact of NPs toxicity on plants can be also estimated by ROS production [22,25]. However, ROS generation can activate the plant's defense mechanisms in order to cope with the oxidative stress damage [15,26–29].

C. nodosa (Ucria) Ascherson is a perennial, fast-growing seagrass that colonizes shallow waters and degraded environments [5,30]. It grows in coastal areas in vicinity to anthropogenic actions and has been proposed a suitable bio-indicator species [31]. In this species, cellular, physiological and biochemical measurable responses induced by various chemical stressors have been proposed to monitor environmental quality [32,33].

The objectives of our study were to investigate the relationship between ZnO NPs effects in *C. nodosa* with Zn uptake in order to understand their impact on seagrasses. We wanted to test whether exposure of seagrasses to NPs will be a dose dependent response (concentration- and time-dependent) or a hormetic response. We evaluated ZnO NPs effects on *C. nodosa* PSII photochemistry by chlorophyll fluorescence imaging analysis, and detected ROS generation as a byproduct of ZnO NP's effects, linking also the ROS generation to PSII functionality.

2. Materials and Methods

2.1. Plant Material

C. nodosa (Ucria) Ascherson plants were collected from the Gulf of Thessaloniki, Aegean Sea (40°33 N, 22°58 E), by their maximum leaf biomass production, at 0.7 m–1.0 m depth [34].

2.2. Experimental Conditions and Exposure to Zinc Oxide Nanoparticles

C. nodosa plants (leaves, orthotropic and plagiotropic rhizomes and roots) were kept in seawater aquaria that have a salinity of 36.9 psu, pH of 7.9, and dissolved oxygen of 5.9 mg L^{-1} , using continuously aerated aquarium pumps as previously described [5].

Zinc oxide NPs with less than 50 nm particle sizes were purchased from Sigma-Aldrich (St. Louis, MO, USA). Stock solution of ZnO NPs in millique water (50 mg L⁻¹), after sonication for 30 min, was stored in the dark at 4 °C [5]. ZnO NPs concentrations in natural waters stated to be among 76–760 μ g L⁻¹ [12] are below concentrations known to have environmental effects on aquatic organisms [35]. In preliminary experiments with 1 and 3 mg L⁻¹ ZnO NPs, no effect was detected on PSII functionality. Thus, we applied 5 mg L⁻¹ and 10 mg L⁻¹ ZnO NPs, which is 7–13 times more the maximum levels of ZnO-NPs reported for water environments [12].

C. nodosa plants were exposed to 5 and 10 mg L⁻¹ for 0 (control), 4 h, 12 h, 24 h, 48 h and 72 h. The two ZnO NPs concentrations were prepared with filtered (0.45 μ m GF/C Whatman) seawater immediately before use. Control and treatment solutions were changed every 24 h. *C. nodosa* intermediate leaf blades (about 300 mm length) were used for chlorophyll fluorescence imaging, H₂O₂ imaging, and for Zn uptake measurements.

2.3. Zinc Oxide Nanoparticles Characterization

Primary particle size, and the morphological and structural characteristics of ZnO NPs were investigated by transmission electron microscopy (TEM). Samples were prepared by drop-casting dispersions of the NPs onto carbon-coated Cu grids after a 3 min sonication with an ultrasonic (VibraCell 400 W, Sonics & Materials Inc., Newtown, CT, USA), applying a microtip probe under intensity settings 4 [36]. Finally, TEM images were obtained with a Jeol JEM 1010 microscope (Jeol, Tokyo, Japan) [37].

Dynamic light scattering (DLS) analysis was used to define the size-distribution profile of ZnO NPs (5 mg L⁻¹, 10 mg L⁻¹ and 50 mg L⁻¹). Zeta (ζ) potential measurements were conducted to assess the surface charge of the particles as described previously [5]. All measurements were performed in Milli-Q water after brief sonication at 25 °C [5]. Results are presented as means (±SD) of three measurements.

2.4. Zinc Determination

Intermediate blades after wet digestion were processed following the methodology described previously [38,39]. Zinc concentrations were determined by flame atomic absorption spectrophotometry (AAnalyst 400 FAAS, Perkin-Elmer, Waltham, MA, USA) with the procedure described in detail before [38,39].

2.5. Zinc Leaf Uptake Kinetics

Zinc leaf uptake kinetics was fitted to the Michaelis-Menten equation: $(C_{max} \times t)/(K_m + t)$, as described in detail previously [5]. Briefly, C represents Zn leaf concentration reached in time t, K_m the time taken to reach half of the value of C_{max} , and C_{max} the maximum or saturation Zn concentration. The rate of the initial uptake $(C_{max}/2 \times K_m)$, the time needed to get equilibrium (T_{eq}) , the equilibrium concentration (C_{eq}) and the mean rate of uptake (V_c) were also estimated. Equilibrium concentration (C_{eq}) is a concentration where the hourly increase is less than 1% compared to the previous hour [38–41]. The time required to reach equilibrium (T_{eq}) was assessed as the time needed to get the C_{eq} , and the mean rate of uptake (V_c) was assessed as C_{eq}/T_{eq} [38,39]. Bioconcentration factor (BCF) was estimated as $(C_{eq} - C_i)/C_w$, where C_i is the initial Zn tissue concentration and C_w is the Zn concentration in water [38,39].

2.6. Chlorophyll Fluorescence Imaging Analysis

An Imaging-PAM Chlorophyll Fluorometer (Walz, Effeltrich, Germany) was used for photosynthetic efficiency measurements as previously described [5]. In *C. nodosa* dark-adapted (15 min) leaf samples, we selected six areas of interest, and the allocation of absorbed light energy to photochemistry (Φ_{PSII}), non-photochemical energy loss as heat (Φ_{NPQ}), and non-regulated energy loss (Φ_{NO}), were calculated as described previously [29]. Relative PSII electron transport rate (ETR), non-photochemical quenching (NPQ), and photochemical quenching (q_P), were also measured [42].

Color-coded images, acquired with 200 µmol photons m⁻² s⁻¹, of Φ_{PSII} , Φ_{NPQ} , Φ_{NO} , and the redox state of plastoquinone (PQ) pool ($q_{\rm P}$), are also presented.

2.7. Imaging of Hydrogen Peroxide Generation

For the estimation of H_2O_2 production, *C. nodosa* leaves were treated with 25 μ M 2',7'-dichlorofluorescein diacetate (Sigma) in the dark for 30 min, as described previously [43,44].

2.8. Statistical Analyses

Zinc leaf uptake kinetics data were analyzed using IBM Statistics SPSS[®] 24 (New York, NY, USA). The significant differences on the fluorescence variables, between control and different incubation time in each concentration and between different concentrations at the same exposure time were tested at the 5% level of probability using t-test analysis (IBM Statistics SPSS[®] 24). Modal analysis by using NORMSEP computer program was employed to estimate particle size (nm) distribution of ZnO NPs [45].

3. Results

3.1. Characterization of ZnO NPs

The size and morphology of ZnO NPs was measured in stock solution by TEM (Figure 1). Modal analysis was used in data emerged from TEM micrographs, in order to determine the ZnO particle diameter distribution (Figure 2). A percentage of 92.8% of the particles was generally in agreement with manufacturer's characteristics (size < 50 nm). The computed mean (\pm SD) NPs size group with the highest frequency was 20.44 nm \pm 7.95 nm (Figure 2).



Figure 1. Transmission electron microscope (TEM) images of zinc oxide nanoparticles (ZnO NPs) stock solution (50 mg L^{-1}).

The hydrodynamic size of 5 mg L⁻¹, 10 mg L⁻¹ and 50 mg L⁻¹ (stock) ZnO NPs solutions, ranged from 220.6 nm to 225.0 nm (Table 1). The negative surface charge (ζ potential) ranged from -17.5 mV to -18.13 mV (Table 1). This similarity among the different NPs concentrations was indicative of the colloidal stability of the different populations. The size distribution by intensity of ZnO NPs is shown in Figure 3a,b.



Figure 2. Distribution pattern of ZnO NPs from TEM micrographs.



Table 1. Hydrodynamic size (\pm SD) (diameter in nm) and zeta potential (\pm SD) (mV) values of the ZnO NPs (n = 3).

Figure 3. Size distribution by intensity of 5 mg L^{-1} ZnO NPs (**a**); and 10 mg L^{-1} ZnO NPs (**b**). Different colours indicate the replications.

3.2. Zinc Leaf Uptake Kinetics

Leaf Zn uptake at both ZnO NPs concentration displayed a time dependent variation (Figure 4). The uptake kinetics at both exposure concentrations was fitted to the Michaelis-Menten equation (r^2 :0.744 for 5 mg L⁻¹, and 0.681 for 10 mg L⁻¹, p < 0.01; Figure 4, Table 2). Zinc uptake of *C. nodosa* leaves increased more rapidly at the beginning of the experiment in the lower ZnO NPs solution (5 mg L⁻¹), while it showed a higher and more than doubled initial rate [$C_{max}/(2 \times K_m)$], and a higher mean rate (V_c) in comparison to the 10 mg L⁻¹ solution (Figure 4, Table 2). At 5 mg L⁻¹ exposure concentration, the uptake reached the equilibrium concentration earlier ($T_{eq} = 28$ h) than in the higher solution (42 h) (Table 2). However, both the maximum concentration (C_{max}) and the equilibrium concentration (C_{eq}) displayed their higher values at the higher ZnO NPs concentration (10 mg L⁻¹) (Figure 4, Table 2). During the experiment, the highest uptake was observed after 72 h of exposure, at both ZnO NP treatments (748.7 ± 29.7 µg g⁻¹ dry wt at 5 mg L⁻¹ and 1086.0 ± 33.7 µg g⁻¹ dry wt at 10 mg L⁻¹) (Figure 4). Moreover, BCF value was higher at the lower exposure concentration (Table 2).



Figure 4. Kinetics of zinc uptake (μ g g⁻¹ dry weight) in *C. nodosa* leaf blades at 5 mg L⁻¹ and 10 mg L⁻¹ ZnO NPs \pm SD (n = 3); dashed and bold lines are the uptake kinetics calculated using Michaelis-Menten equation.

Parameter	5 mg L ⁻¹	10 mg L ⁻¹
C _{max}	795.4 (±178.9)	1316.7 (±614.7)
K_m	8.8 (±6.7)	32.1 (±34.2)
$C_{max}/(2 \times K_m)$	44.950	20.507
r ²	0.744 *	0.681 *
C_{eq}	573.272	761.263
T_{eq}	28	42
V_c	20.473	18.125
BCF	79.588	58.593
	* $p < 0.01$.	

Table 2. Kinetics of Zn accumulation in *C. nodosa* leaf blades exposed to 5 mg L^{-1} and 10 mg L^{-1} ZnO NPs.

The fits correspond to a Michaelis-Menten equation: $C = (C_{max} \times t)/(K_m + t)$; C, in µg g⁻¹ dry wt; K_m , in hours; t, in hours; C_{eq} , in µg g⁻¹ dry wt; T_{eq} , in hours; V_c , concentration/hours. In parentheses, standard errors are given.

3.3. Allocation of Absorbed Light Energy in Leaf Blades of Cymodocea nodosa Exposed to ZnO NPs

The allocation of absorbed light energy at PSII in leaf blades of *C. nodosa* exposed to ZnO NPs for 4 h, 24 h, 48 h and 72 h is shown in Figures 5 and 6. Exposure to 5 and 10 mg L⁻¹ ZnO NPs significantly decreased the quantum efficiency of PSII photochemistry (Φ_{PSII}) compared to control values, with the exception of the 24 h treatment that Φ_{PSII} increased at 10 mg L⁻¹ ZnO NPs and did not differ compared to control values at 5 mg L⁻¹ ZnO NPs (Figure 5a). Thus, at the 24-h treatment, Φ_{PSII} in 10 mg L⁻¹ ZnO NPs was higher than in 5 mg L⁻¹ (Figure 5a).



Figure 5. The quantum efficiency of PSII photochemistry (Φ_{PSII}) (**a**); and the quantum yield of regulated non-photochemical energy loss as heat (Φ_{NPQ}) (**b**), in control leaf blades of *C. nodosa* and in leaf blades exposed to 5 mg L⁻¹ and 10 mg L⁻¹ ZnO NPs for 4 h, 24 h, 48 h and 72 h. Columns with the same letter (lower case for 5 mg L⁻¹ ZnO NPs and capitals for 10 mg L⁻¹ ZnO NPs) are not statistically different (p < 0.05). An asterisk represents a significantly different mean of the same time treatment between 5 and 10 mg L⁻¹ ZnO NPs (p < 0.05). Bars in columns represent standard deviation.

The quantum yield of regulated non-photochemical energy loss (Φ_{NPQ}) increased significantly at 5 mg L⁻¹ ZnO NPs compared to control, while at 10 mg, L⁻¹ increased after 24 h of exposure, but decreased at further exposure time (48 h and 72 h) (Figure 5b). Φ_{NPQ} after 4 h, 48 h and 72 h exposure to the low concentration was significantly higher than in the high concentration (Figure 5b).



Figure 6. The quantum yield of non-regulated energy dissipated in PSII (non-regulated heat dissipation, a loss process due to PSII inactivity) (Φ_{NO}) in control leaf blades of *C. nodosa* (*Cymodocea nodosa*) and in leaf blades exposed to 5 mg L⁻¹ and 10 mg L⁻¹ ZnO NPs for 4 h, 24 h, 48 h and 72 h. Symbol explanations as in Figure 5.

The non-regulated energy loss (Φ_{NO}) did not differ compared to control values after exposure to 5 mg L⁻¹ ZnO NPs (Figure 6), but decreased compared to the control values after 24 h of exposure to the high NPs concentration, and increased at further exposure time (48 h and 72 h) (Figure 6). At 48 h and 72 h treatment Φ_{NO} of *C. nodosa* leaf blades exposed to the high concentration was significantly higher than in the low concentration (Figure 6).

3.4. Electron Transport Rate and Non-Photochemical Quenching in Leaf Blades of Cymodocea nodosa Exposed to ZnO NPs

The relative electron transport rate (ETR) of PSII decreased significantly at both ZnO NPs concentrations, with the exception of the 24 h exposure, where the ETR increased at the high concentration and did not differ compared to control values at the low concentration (Figure 7a). With the 24 h treatment, the ETR of *C. nodosa* leaf blades exposed to the high concentration was significantly higher than in the low concentration (Figure 7a).

The non-photochemical quenching (NPQ) increased at the low concentration compared to control, while at the high concentration it increased after 24 h of exposure, but decreased significantly at further exposure time (48 h and 72 h) (Figure 7b). After 4 h, 48 h and 72 h treatment, the NPQ of *C. nodosa* leaf blades exposed to 5 mg L⁻¹ was significantly higher than 10 mg L⁻¹ (Figure 7b).



Figure 7. The relative electron transport rate of PSII (ETR) (**a**); and the non-photochemical quenching (NPQ) (**b**), in control leaf blades of *C. nodosa* and in leaf blades exposed to 5 mg L^{-1} and 10 mg L^{-1} ZnO NPs for 4 h, 24 h, 48 h and 72 h. Symbol explanations as in Figure 5.

3.5. The Redox State of Plastoquinone Pool of Cymodocea nodosa Leaf Blades Exposed to ZnO NPs

The redox state of plastoquinone (PQ) pool (a measure of PSII open reaction centers) (q_P), did not differ compared to control values after 4 h, 24 h, and 48 h exposure to 5 mg L⁻¹ ZnO NPs but reduced after 72 h (Figure 8). PQ pool reduced compared to control values after 4 h, 48 h and 72 h exposure to

10 mg L⁻¹ ZnO, but increased after a 24 h exposure (Figure 8). After a 72 h exposure of *C. nodosa* to ZnO NPs, the PQ pool was more oxidized at 5 mg L⁻¹ than at 10 mg L⁻¹, but less than controls (Figure 8).



Figure 8. The photochemical quenching (q_P) in control leaf blades of *C. nodosa* and in leaf blades exposed to 5 mg L⁻¹ and 10 mg L⁻¹ ZnO NPs for 4 h, 24 h, 48 h and 72 h. Symbol explanations as in Figure 5.



Figure 9. Representative chlorophyll fluorescence images at 200 µmol photons $m^{-2} s^{-1}$ actinic light of Φ_{PSII} , Φ_{NPQ} , Φ_{NO} , and $q_{\rm P}$; of *C. nodosa* control leaf blades and leaf blades exposed to 5 mg L⁻¹ and 10 mg L⁻¹ ZnO NPs for 4 h, 24 h, 48 h and 72 h. The colour code depicted at the bottom of the images ranges from black (pixel values 0.0) to purple (1.0). The six areas of interest are shown. Average values are presented for each photosynthetic parameter.

3.6. Chlorophyll Fluorescence Images of Cymodocea nodosa Leaf Blades Exposed to ZnO NPs

We did not detect any spatial heterogeneity of Φ_{PSII} , Φ_{NPQ} , Φ_{NO} , and q_P images in the control leaf blades of *C. nodosa* measured at 200 µmol photons m⁻² s⁻¹ actinic light (Figure 9). In addition,

exposure to both NPs concentrations for 4 h, 24 h, 48 h and 72 h did not significantly alter these patterns (Figure 9). A temporal heterogeneity was observed at the images of q_P and Φ_{NPQ} (Figure 9). The lowest Φ_{PSII} with simultaneous low q_P values was observed after a 12 h exposure at the high concentration (Figure 10a). At the same time, the highest levels of H₂O₂ were detected (Figure 10b). Immediately after this, we noticed that the 24 h exposure to 10 mg L⁻¹ ZnO NPs resulted in the highest Φ_{PSII} values with the highest q_P values, both of them being higher than the control values (Figure 9). A parallel decreased Φ_{NO} was detected (Figure 9).



Figure 10. Representative chlorophyll fluorescence images after 5 min illumination at 200 µmol photons $m^{-2} s^{-1}$ actinic light of Φ_{PSII} , Φ_{NPQ} , Φ_{NO} , and q_P of *C. nodosa* leaf blades exposed to 10 mg L⁻¹ ZnO NPs for 12 h. The colour code depicted at the bottom of the images ranges from 0.0 to 1.0. The average values are presented for each photosynthetic parameter (**a**). Below is the representative pattern of H₂O₂ production in a *C. nodosa* leaf blade exposed for 12 h to 10 mg L⁻¹ ZnO NPs. Scale bare: 200 µm. Increased H₂O₂ content is indicated by light green colour (**b**).

3.7. Imaging of Hydrogen Peroxide Production After Exposure of Cymodocea nodosa Leaf Blades to ZnO NPs

No noteworthy quantities of H_2O_2 could be noticed in the control leaf blades of *C. nodosa* (Figure 11a). Exposure at the low concentration for 4 h did not result in any change to H_2O_2 production (Figure 11b), while the same exposure time at the high concentration, resulted in increased production of H_2O_2 (Figure 11c). After a 24-h exposure to 5 mg L⁻¹ ZnO NPs, H_2O_2 levels were the same as the control ones (Figure 11d), while they also dropped at 10 mg L⁻¹ and could not be detected at all (Figure 11e). Later, after a 48-h exposure to ZnO NPs, there was an increase in the accumulation of H_2O_2 being higher at 5 mg L⁻¹ (Figure 11f), than at 10 mg L⁻¹ ZnO NPs (Figure 11g). However, after a 72-h exposure to ZnO NPs the accumulation of H_2O_2 decreased at 5 mg L⁻¹ (Figure 11h), but increased at 10 mg L⁻¹ ZnO NPs (Figure 11i). The highest H_2O_2 accumulation was detected after a 12-h exposure to 10 mg L⁻¹ ZnO NPs (Figure 10b) and the lowest after a 24-h exposure to 10 mg L⁻¹ (Figure 11e).

4. Discussion

Previously, extensive literature survey has demonstrated both the positive and detrimental impacts of NPs on terrestrial and aquatic plants, which are due to size and type of NPs (especially their specific surface area) and the plant species [5,7,8,12,22,25]. Toxicity of ZnO NPs is determined to be due to the dissolution, release and uptake of free Zn ions, but specific nanoparticulate effects may be hard to unravel from effects due to free zinc ions [46,47]. Thus, ZnO NPs effects in *C. nodosa* were correlated to both applied ZnO NPs concentration and to Zn uptake.

In *C. nodosa* cellular, physiological and biochemical measurable responses (biomarkers) to metallic elements (e.g., Cd, Cr, Cu, Ni) have been proposed as early warning signals of alterations in seawater quality [32,33,38]. However, there has been little consideration of the seagrasses, especially *C. nodosa*, as the test material in evaluating the effects of metal oxide nanoparticles [5], despite the fact that coastal ecosystems are expected to be the destination of the majority of the nanoparticles, mainly ZnO and

TiO₂ NPs, discharged by industry [46]. NPs are released into aquatic environment either by direct uses or wastewater plant effluents [47–49].



Figure 11. Representative patterns of H_2O_2 production in *C. nodosa* control leaf blades and after exposure to 5 mg L⁻¹ and 10 mg L⁻¹ ZnO NPs, as indicated by the fluorescence of H_2DCF -DA. The H_2O_2 real-time generation in control leaf blade (**a**); after a 4-h exposure to 5 mg L⁻¹ (**b**); after a 4-h exposure to 10 mg L⁻¹ (**c**); after a 24-h exposure to 5 mg L⁻¹ (**d**); after a 24-h exposure to 10 mg L⁻¹ (**c**); after a 48-h exposure to 10 mg L⁻¹ (**g**); after a 72-h exposure to 5 mg L⁻¹ (**h**); and after a 72-h exposure to 10 mg L⁻¹ (**i**). Scale bare: 200 µm. A higher H_2O_2 content is indicated by light green colour.

Photosynthetic organisms via the process of photosynthesis, transform the light energy into chemical energy with the collaboration of PSII and PSI, while the most susceptible constituent of the photosynthetic apparatus to environmental stresses is believed to be PSII [15–18,50]. PSII functionality estimated by chlorophyll fluorescence imaging has been considered as the most suitable method to identify NPs toxicity effects on plants [5,21,22]. Exposure of plants to NPs can have positive or negative effects on the light reactions of photosynthesis [51].

At the beginning of the experiment, Zn uptake kinetics at 5 mg L⁻¹ ZnO NPs with a more than twice initial rate and higher mean rate (V_c) than 10 mg L⁻¹ (Figure 4, Table 2), resulted in a significant lower quantum efficiency of PSII photochemistry (Φ_{PSII}) compared to 10 mg L⁻¹, after a 24 h exposure (Figure 5a). However, at 5 mg L⁻¹ exposure concentration, Zn uptake reached the equilibrium concentration earlier (T_{eq} = 28 h) than at 10 mg L⁻¹ (T_{eq} = 42 h), resulting in a significantly higher regulated non-photochemical energy loss as heat (Φ_{NPQ}) after 48 h and 72 h exposure, compared to 10 mg L⁻¹ (Figure 5b). This lower Φ_{NPQ} at 10 mg L⁻¹ ZnO NPs, resulted in significantly higher non-regulated energy loss (Φ_{NO}) of *C. nodosa* leaf blades exposed to 10 mg L⁻¹ for 48 h and 72 h (Figure 6), since there was no difference in Φ_{PSII} (Figure 5a). The significantly increased Φ_{NO} (Figure 6) implies higher singlet oxygen (¹O₂) production. Φ_{NO} consists of chlorophyll fluorescence internal conversions and intersystem crossing, indicative of ¹O₂ formation via the triplet state of chlorophyll (³chl*) [42,52–54]. The increased ¹O₂ formation, mostly after 72 h exposure to 10mg L⁻¹ ZnO NPs, was due to an increased Zn uptake compared to 5 mg L⁻¹ (Figure 4).

A reduced photosynthetic efficiency, measured as the maximum quantum efficiency of PSII (F_v/F_m), and the redox state of plastoquinone (PQ) pool (q_P), was also observed previously due to increased Zn accumulation [7]. Furthermore, ZnO NPs treatments enhanced generation of H₂O₂ [7]. A relationship between closed reaction centers (q_P) and increased H₂O₂ generation was noticed (Figures 9–11). The PQ pool is considered as the component integrated in plant antioxidant defense [55], which at a 12 h of exposure mediated the generation of chloroplastic H_2O_2 [22], acting as a fast acclimation-signaling molecule [55,56].

After 12 h exposure of *C. nodosa* to 10 mg L⁻¹ ZnO NPs, leaf Zn uptake was higher than in 5 mg L⁻¹ (Figure 4), resulting in the lower Φ_{PSII} and the lowest q_P values (Figure 10a). At the same time, an increased accumulation of H₂O₂ was detected in the leaves of *C. nodosa* (Figure 10b). Closed reaction centers (q_P) indicate excess photon supply and associated ROS production [57–59]. This photosynthesis derived H₂O₂ is moving throughout leaf veins, serving as a signaling molecule and triggering a stress-defense response [22,35,44,55,56,60,61]. Thus, this stress-defense response triggered a significant increase of the fraction of open PSII reaction centers (q_P) (Figure 8) and a significant increase of Φ_{PSII} (Figure 5a) and ETR (Figure 7a) observed after 24 h exposure of *C. nodosa* to 10 mg L⁻¹ ZnO NPs; while at the same time, H₂O₂ generation decreased to control levels (Figure 11e). However, we have to emphasize that antibacterial activity of ZnO and ZnO NPs is due to ROS generation [62,63], which also orchestrates a regulatory action at various plant developmental stages [64].

Hormesis has been extensively documented in plants, revealing that biphasic dose-responses occur commonly [65,66]. Thus, at low-level stress, plants are activating responses at the cellular and molecular level that enhance adaptation and plant tolerance [65,66]. The photoprotective mechanism of non-photochemical quenching (NPQ) which is closely related to ROS, follows a biphasic dose-response pattern typical of hormesis [67]. NPQ in *C. nodosa* leaf blades exposed to 10 mg L⁻¹ ZnO NPs depicts a hormetic response. A hormetic response suggests that a basal stress level is needed for adaptive responses [68,69]. This basal stress level was 10 mg L⁻¹ ZnO NPs and the time required for the induction of this mechanism was 24 h exposure.

Author Contributions: Conceptualization, P.M. and M.M.; methodology, P.M., I.S. and M.M.; software, P.M.; validation, P.M., K.C. and I.S.; formal analysis, P.M. and K.C.; investigation, P.M., K.C., I.S. and Z.M.; resources, P.M. and M.M.; data curation, P.M. and K.C.; writing—original draft preparation, P.M. and M.M; writing—review and editing, M.M; visualization, P.M., K.C. and M.M.; supervision, P.M. and M.M.; project administration, P.M. and M.M.; funding acquisition, P.M. and M.M.

Funding: This research received no external funding.

Conflicts of Interest: The authors declare no conflict of interest.

References

- Maynard, A.; Aitken, R.J.; Butz, T.; Colvin, V.; Donaldson, K.; Oberdörster, G.; Philbert, M.A.; Ryan, J.; Seaton, A.; Stone, V.; et al. Safe handling of nanotechnology. *Nature* 2006, 444, 267–269. [CrossRef] [PubMed]
- 2. Chang, Y.N.; Zhang, M.; Xia, L.; Zhang, J.; Xing, G. The toxic effects and mechanisms of CuO and ZnO nanoparticles. *Materials* **2012**, *5*, 2850–2871. [CrossRef]
- Kołodziejczak-Radzimska, A.; Jesionowski, T. Zinc oxide—From synthesis to application: A review. *Materials* 2014, 7, 2833–2881. [CrossRef]
- Hong, J.; Rico, C.; Zhao, L.; Adeleye, A.S.; Keller, A.A.; Peralta-Videa, J.R.; Gardea-Torresdey, J.L. Toxic effects of copper-based nanoparticles or compounds to lettuce (*Lactuca sativa*) and alfalfa (*Medicago sativa*). *Environ. Sci. Proc. Impacts* 2015, 17, 177–185. [CrossRef] [PubMed]
- Moustakas, M.; Malea, P.; Haritonidou, K.; Sperdouli, I. Copper bioaccumulation, photosystem II functioning and oxidative stress in the seagrass *Cymodocea nodosa* exposed to copper oxide nanoparticles. *Environ. Sci. Pollut. Res.* 2017, 24, 16007–16018. [CrossRef]
- Ghosh, M.; Sinha, S.; Jothiramajayam, M.; Jana, A.; Nag, A.; Mukherjee, A. Cytogenotoxicity and oxidative stress induced by zinc oxide nanoparticle in human lymphocyte cells in-vitro and Swiss albino male mice in-vivo. *Food Chem. Toxicol.* 2016, 97, 286–296. [CrossRef]
- Tripathi, D.K.; Mishra, R.K.; Singh, S.; Singh, S.; Vishwakarma, K.; Sharma, S.; Singh, V.P.; Singh, P.K.; Prasad, S.M.; Dubey, N.K.; et al. Nitric oxide ameliorates zinc oxide nanoparticles phytotoxicity in wheat seedlings: Implication of the ascorbate-glutathione cycle. *Front. Plant Sci.* 2017, *8*, 1. [CrossRef]

- Pullagurala, V.L.R.; Adisa, I.O.; Rawat, S.; Kim, B.; Barrios, A.C.; Medina-Velo, I.A.; Hernandez-Viezcas, J.A.; Peralta-Videa, J.A.; Gardea-Torresdey, J.L. Finding the conditions for the beneficial use of ZnO nanoparticles towards plants—A review. *Environ. Pollut.* 2018, 241, 1175–1181. [CrossRef]
- Segets, D.; Gradl, J.; Taylor, R.K.; Vassilev, V.; Peukert, W. Analysis of optical absorbance spectra for the determination of ZnO nanoparticle size distribution, solubility, and surface energy. ACS Nano 2009, 3, 1703–1710. [CrossRef]
- 10. Song, W.; Zhang, J.; Guo, J.; Zhang, J.; Ding, F.; Li, L.; Sun, Z. Role of the dissolved zinc ion and reactive oxygen species in cytotoxicity of ZnO nanoparticles. *Toxicol. Lett.* **2010**, *199*, 389–397. [CrossRef]
- 11. Ma, X.; Geiser-Lee, J.; Deng, Y.; Kolmakov, A. Interactions between engineered nanoparticles (ENPs) and plants: Phytotoxicity, uptake and accumulation. *Sci. Total Environ.* **2010**, *408*, 3053–3061. [CrossRef] [PubMed]
- 12. Rajput, V.D.; Minkina, T.M.; Behal, A.; Sushkova, S.N.; Mandzhieva, S.; Singh, R.; Gorovtsov, A.; Tsitsuashvili, V.S.; Purvis, W.O.; Ghazaryan, K.A.; et al. Effects of zinc-oxide nanoparticles on soil, plants, animals and soil organisms: A review. *Environ. Nanotechnol. Monit. Manag.* **2018**, *9*, 76–84. [CrossRef]
- 13. Anderson, J.M. Changing concepts about the distribution of photosystems I and II between grana-appressed and stroma-exposed thylakoid membranes. *Photosynth. Res.* **2002**, *73*, 157–164. [CrossRef]
- 14. Apostolova, E.L.; Dobrikova, A.G.; Ivanova, P.I.; Petkanchin, I.B.; Taneva, S.G. Relationship between the organization of the PSII supercomplex and the functions of the photosynthetic apparatus. *J. Photochem. Photobiol. B* **2006**, *83*, 114–122. [CrossRef]
- 15. Moustaka, J.; Ouzounidou, G.; Sperdouli, I.; Moustakas, M. Photosystem II is more sensitive than photosystem I to Al³⁺ induced phytotoxicity. *Materials* **2018**, *11*, 1772. [CrossRef] [PubMed]
- 16. Ekmekçi, Y.; Tanyolaç, D.; Ayhan, B. Effects of cadmium on antioxidant enzyme and photosynthetic activities in leaves of two maize cultivars. *J. Plant Physiol.* **2008**, *165*, 600–611. [CrossRef] [PubMed]
- 17. Bayçu, G.; Moustaka, J.; Gevrek-Kürüm, N.; Moustakas, M. Chlorophyll fluorescence imaging analysis for elucidating the mechanism of photosystem II acclimation to cadmium exposure in the hyperaccumulating plant *Noccaea caerulescens*. *Materials* **2018**, *11*, 2580. [CrossRef]
- Moustakas, M.; Bayçu, G.; Gevrek-Kürüm, N.; Moustaka, J.; Csatári, I.; Rognes, S.E. Spatiotemporal heterogeneity of photosystem II function during acclimation to zinc exposure and mineral nutrition changes in the hyperaccumulator. *Noccaea caerulescens. Environ. Sci. Pollut. Res.* 2019, 26, 6613–6624. [CrossRef]
- 19. Ding, Z.S.; Zhou, B.Y.; Sun, X.F.; Zhao, M. High light tolerance is enhanced by overexpressed PEPC in rice under drought stress. *Acta Agron. Sin.* **2012**, *38*, 285–292. [CrossRef]
- 20. Zeng, L.D.; Li, M.; Chow, W.S.; Peng, C.L. Susceptibility of an ascorbate-deficient mutant of Arabidopsis to high-light stress. *Photosynthetica* **2018**, *56*, 427–432. [CrossRef]
- Falco, W.F.; Queiroz, A.M.; Fernandes, J.; Botero, E.R.; Falcão, E.A.; Guimarães, F.E.G.; M'Peko, J.C.; Oliveira, S.L.; Colbeck, I.; Caires, A.R.L. Interaction between chlorophyll and silver nanoparticles: A close analysis of chlorophyll fluorescence quenching. *J. Photochem. Photobiol. A* 2015, 299, 203–209. [CrossRef]
- Antonoglou, O.; Moustaka, J.; Adamakis, I.D.; Sperdouli, I.; Pantazaki, A.; Moustakas, M.; Dendrinou-Samara, C. Nanobrass CuZn nanoparticles as foliar spray non phytotoxic fungicides. ACS Appl. Mater. Interfaces 2018, 10, 4450–4461. [CrossRef] [PubMed]
- 23. Rico, C.M.; Hong, J.; Morales, M.I.; Zhao, L.; Barrios, A.C.; Zhang, J.Y.; Peralta-Videa, J.R.; Gardea-Torresdey, J.L. Effect of cerium oxide nanoparticles on rice: A study involving the antioxidant defense system and in vivo fluorescence imaging. *Environ. Sci. Technol.* **2013**, *47*, 5635–5642. [CrossRef] [PubMed]
- 24. Ma, C.X.; White, J.C.; Dhankher, O.P.; Xing, B. Metal-based nanotoxicity and detoxification pathways in higher plants. *Environ. Sci. Technol.* **2015**, *49*, 7109–7122. [CrossRef] [PubMed]
- 25. Tripathi, D.K.; Shweta; Singh, S.; Singh, S.; Pandey, R.; Singh, V.P.; Sharma, N.C.; Prasad, S.M.; Dubey, N.K.; Chauhan, D.K. An overview on manufactured nanoparticles in plants: Uptake, translocation, accumulation and phytotoxicity. *Plant Physiol. Biochem.* **2017**, *110*, 2–12. [CrossRef] [PubMed]
- 26. Moustaka, J.; Moustakas, M. Photoprotective mechanism of the non-target organism Arabidopsis thaliana to paraquat exposure. *Pest. Biochem. Physiol.* **2014**, *111*, 1–6. [CrossRef] [PubMed]
- 27. Moustaka, J.; Ouzounidou, G.; Bayçu, G.; Moustakas, M. Aluminum resistance in wheat involves maintenance of leaf Ca²⁺ and Mg²⁺ content, decreased lipid peroxidation and Al accumulation, and low photosystem II excitation pressure. *BioMetals* **2016**, *29*, 611–623. [CrossRef] [PubMed]

- 28. Foyer, C.H. Reactive oxygen species, oxidative signaling and the regulation of photosynthesis. *Environ. Exp. Bot.* **2018**, 154, 134–142. [CrossRef]
- 29. Moustaka, J.; Panteris, E.; Adamakis, I.D.S.; Tanou, G.; Giannakoula, A.; Eleftheriou, E.P.; Moustakas, M. High anthocyanin accumulation in poinsettia leaves is accompanied by thylakoid membrane unstacking, acting as a photoprotective mechanism, to prevent ROS formation. *Environ. Exp. Bot.* **2018**, *154*, 44–55. [CrossRef]
- 30. Malea, P.; Zikidou, C. Temporal variation in biomass partitioning of the seagrass *Cymodocea nodosa* at the Gulf of Thessaloniki, Greece. *J. Biol. Res.-Thessalon.* **2011**, *5*, 75–90.
- Malea, P.; Kevrekidis, T. Trace element (Al, As, B, Ba, Cr, Mo, Ni, Se, Sr, Tl, U and V) distribution and seasonality in compartments of the seagrass Cymodocea nodosa. *Sci. Total Environ.* 2013, 463–464, 611–623. [CrossRef] [PubMed]
- 32. Malea, P.; Adamakis, I.-D.; Kevrekidis, T. Microtubule integrity and cell viability under metal (Cu, Ni and Cr) stress in the seagrass *Cymodocea nodosa*. *Chemosphere* **2013**, *93*, 1035–1042. [CrossRef] [PubMed]
- Malea, P.; Adamakis, I.-D.; Kevrekidis, T. Effects of lead uptake on microtubule cytoskeleton organization and cell viability in the seagrass *Cymodocea nodoca*. *Ecotoxicol. Environ. Saf.* 2014, 104, 175–181. [CrossRef] [PubMed]
- 34. Moustakas, M.; Malea, P.; Zafeirakoglou, A.; Sperdouli, I. Photochemical changes and oxidative damage in the aquatic macrophyte *Cymodocea nodosa* exposed to paraquat-induced oxidative stress. *Pest. Biochem. Physiol.* **2016**, *126*, 28–34. [CrossRef] [PubMed]
- 35. Batley, G.E.; Kirby, J.K.; McLaughlin, M.J. Fate and risks of nanomaterials in aquatic and terrestrial environments. *Acc. Chem. Res.* **2013**, *46*, 854–862. [CrossRef]
- Perreault, F.; Oukarroum, A.; Melegari, S.P.; Matias, W.G.; Popovic, R. Polymer coating of copper oxide nanoparticles increases nanoparticles uptake and toxicity in the green alga *Chlamydomonas reinhardtii*. *Chemosphere* 2012, *87*, 1388–1394. [CrossRef]
- 37. Dallas, P.; Georgakilas, V.; Niarchos, D.; Komninou, P.; Kehagias, T.; Petridis, D. Synthesis, characterization and thermal properties of polymer/magnetite nanocomposites. *Nanotechnology* **2006**, *17*, 2046–2053. [CrossRef]
- Malea, P.; Adamakis, I.-D.; Kevrekidis, T. Kinetics of cadmium accumulation and its effects on microtubule integrity and cell viability in the seagrass *Cymodocea nodoca*. Aquat. Toxicol. 2013, 144–145, 257–264. [CrossRef]
- 39. Malea, P.; Kevrekidis, T.; Chatzipanagiotou, K.-R.; Mogias, A. Cadmium uptake kinetics in parts of the seagrass *Cymodocea nodosa* at high exposure concentrations. *J. Biol. Res.-Thessalon.* **2018**, 25, 5. [CrossRef]
- 40. Fernández, J.A.; Vázquez, M.D.; López, J.; Carvalleira, A. Modeling the extra and intracellular uptake and discharge of heavy metals in *Fontinalis antipyretica* transplanted along a heavy metal and pH contamination gradient. *Environ. Pollut.* **2006**, *139*, 21–31. [CrossRef]
- 41. Diaz, S.; Villares, R.; Carballeira, A. Uptake kinetics of As, Hg, Sb, and Se in the aquatic moss *Fontinalis antipyretica* Hedw. *Water Air Soil Pollut*. **2012**, 223, 3409–3423. [CrossRef]
- Bayçu, G.; Gevrek-Kürüm, N.; Moustaka, J.; Csatári, I.; Rognes, S.E.; Moustakas, M. Cadmium-zinc accumulation and photosystem II responses of *Noccaea caerulescens* to Cd and Zn exposure. *Environ. Sci. Pollut. Res.* 2017, 24, 2840–2850. [CrossRef] [PubMed]
- 43. Sandalio, L.M.; Rodríguez-Serrano, M.; Romero-Puertas, M.C.; Del Río, L.A. Imaging of reactive oxygen species and nitric oxide in vivo in plant tissues. *Method. Enzymol.* **2008**, *440*, 397–409.
- Moustaka, J.; Tanou, G.; Adamakis, I.D.; Eleftheriou, E.P.; Moustakas, M. Leaf age dependent photoprotective and antioxidative mechanisms to paraquat-induced oxidative stress in *Arabidopsis thaliana*. *Int. J. Mol. Sci.* 2015, *16*, 13989–14006. [CrossRef] [PubMed]
- 45. Villares, R.; Puente, X.; Carballeira, A. Seasonal variarion and background levels of heavy metals in two green seaweeds. *Environ. Pollut.* **2002**, *119*, 79–90. [CrossRef]
- 46. Miller, R.J.; Lenihan, H.S.; Muller, E.B.; Tseng, N.; Hanna, S.K.; Keller, A.A. Impacts of metal oxide nanoparticles on marine phytoplankton. *Environ. Sci. Technol.* **2010**, *44*, 7329–7334. [CrossRef] [PubMed]
- 47. Matranga, V.; Corsi, I. Toxic effects of engineered nanoparticles in the marine environment: Model organisms and molecular approaches. *Mar. Environ. Res.* **2012**, *76*, 32–40. [CrossRef] [PubMed]
- 48. Wong, S.W.Y.; Leung, P.T.; Djurišić, A.B.; Leung, K.M.Y. Toxicities of nano zinc oxide to five marine organisms: Influences of aggregate site and ion solubility. *Anal. Bioanal. Chem.* **2010**, *396*, 609–618. [CrossRef]

- Klaine, S.J.; Alvarez, P.J.J.; Batley, G.E.; Fernandes, T.F.; Handy, R.D.; Lyon, D.Y.; Mahendra, S.; Mclaughlin, M.J.; Lead, J.R. Nanomaterials in the environment: Behavior, fate, bioavailability, and effects: A critical review. *Environ. Toxicol. Chem.* 2008, 27, 1825–1851. [CrossRef]
- 50. Zlobin, I.E.; Ivanov, Y.V.; Kartashov, A.V.; Sarvin, B.A.; Stavrianidi, A.N.; Kreslavski, V.D.; Kuznetsov, V.V. Impact of weak water deficit on growth, photosynthetic primary processes and storage processes in pine and spruce seedlings. *Photosynth. Res.* **2019**, *139*, 307–323. [CrossRef]
- 51. Hatami, M.; Kariman, K.; Ghorbanpour, M. Engineered nanomaterial-mediated changes in the metabolism of terrestrial plants. *Sci. Total Environ.* **2016**, *571*, 275–291. [CrossRef]
- 52. Apel, K.; Hirt, H. Reactive oxygen species: Metabolism, oxidative stress and signal transduction. *Annu. Rev. Plant Biol.* **2004**, *55*, 373–399. [CrossRef] [PubMed]
- Gawroński, P.; Witoń, D.; Vashutina, K.; Bederska, M.; Betliński, B.; Rusaczonek, A.; Karpiński, S. Mitogen-activated protein kinase 4 is a salicylic acid-independent regulator of growth but not of photosynthesis in *Arabidopsis. Mol. Plant* 2014, 7, 1151–1166. [CrossRef] [PubMed]
- 54. Sperdouli, I.; Moustakas, M. Differential blockage of photosynthetic electron flow in young and mature leaves of *Arabidopsis thaliana* by exogenous proline. *Photosynthetica* **2015**, *53*, 471–477. [CrossRef]
- 55. Borisova-Mubarakshina, M.M.; Vetoshkina, D.V.; Ivanov, B.N. Antioxidant and signaling functions of the plastoquinone pool in higher plants. *Physiol. Plant.* **2019**, *166*, 181–198. [CrossRef] [PubMed]
- 56. Wilson, K.E.; Ivanov, A.G.; Öquist, G.; Grodzinski, B.; Sarhan, F.; Huner, N.P.A. Energy balance, organellar redox status, and acclimation to environmental stress. *Can. J. Bot.* **2006**, *84*, 1355–1370. [CrossRef]
- 57. Sperdouli, I.; Moustakas, M. A better energy allocation of absorbed light in photosystem II and less photooxidative damage contribute to acclimation of *Arabidopsis thaliana* young leaves to water deficit. *J. Plant Physiol.* **2014**, *171*, 587–593. [CrossRef]
- 58. Agathokleous, E.; Mouzaki-Paxinou, A.C.; Saitanis, C.J.; Paoletti, E.; Manning, W.J. The first toxicological study of the antiozonant and research tool ethylene diurea (EDU) using a *Lemna minor* L. bioassay: Hints to its mode of action. *Environ. Pollut.* **2016**, *213*, 996–1006. [CrossRef]
- 59. Sperdouli, I.; Moustakas, M. Leaf developmental stage modulates metabolite accumulation and photosynthesis contributing to acclimation of *Arabidopsis thaliana* to water deficit. *J. Plant Res.* **2014**, 127, 481–489. [CrossRef]
- 60. Mittler, R.; Vanderauwera, S.; Suzuki, N.; Miller, G.; Tognetti, V.B.; Vandepoele, K.; Gollery, M.; Shulaev, V.; Van Breusegem, F. ROS signaling: The new wave? *Trends Plant Sci.* **2011**, *16*, 300–309. [CrossRef]
- 61. Foyer, C.H.; Shigeoka, S. Understanding oxidative stress and antioxidant functions to enhance photosynthesis. *Plant Physiol.* **2011**, *155*, 93–100. [CrossRef] [PubMed]
- 62. Fiedot-Tobola, M.; Ciesielska, M.; Maliszewska, I.; Rac-Rumijowska, O.; Suchorska-Wozniak, P.; Teterycz, H.; Bryjak, M. Deposition of zinc oxide on different polymer textiles and their antibacterial properties. *Materials* **2018**, *11*, 707. [CrossRef] [PubMed]
- Gavrilenko, E.A.; Goncharova, D.A.; Lapin, I.N.; Nemoykina, A.L; Svetlichnyi, V.A.; Aljulaih, A.A.; Mintcheva, N.; Kulinich, S.A. Comparative study of physicochemical and antibacterial properties of ZnO nanoparticles prepared by laser ablation of Zn target in water and air. *Materials* 2019, 12, 186. [CrossRef] [PubMed]
- 64. Singh, R.; Singh, S.; Parihar, P.; Mishra, R.K.; Tripathi, D.K.; Singh, V.P.; Chauhan, D.K.; Prasad, S.M. Reactive oxygen species (ROS): Beneficial companions of plants' developmental processes. *Front. Plant Sci.* **2016**, *7*, 1299. [CrossRef] [PubMed]
- 65. Agathokleous, E. Environmental hormesis, a fundamental non-monotonic biological phenomenon with implications in ecotoxicology and environmental safety. *Ecotoxicol. Environ. Saf.* **2018**, *148*, 1042–1053. [CrossRef]
- 66. Agathokleous, E.; Kitao, M.; Calabrese, E.J. Hormesis: A compelling platform for sophisticated plant science. *Trends Plant Sci.* **2019**, *24*, 318–326. [CrossRef] [PubMed]
- 67. Agathokleous, E.; Kitao, M.; Harayama, H. On the nonmonotonic, hormetic photoprotective response of plants to stress. *Dose-Response* **2019**, *17*, 1559325819838420. [CrossRef] [PubMed]

- 68. Kitao, M.; Tobita, H.; Kitaoka, S.; Harayama, H.; Yazaki, K.; Komatsu, M.; Agathokleous, E.; Koike, T. Light energy partitioning under various environmental stresses combined with elevated CO₂ in three deciduous broadleaf tree species in Japan. *Climate* **2019**, *7*, 79. [CrossRef]
- 69. Agathokleous, E.; Kitao, M.; Calabrese, E.J. Environmental hormesis and its fundamental biological basis: Rewriting the history of toxicology. *Environ. Res.* **2018**, *165*, 274–278. [CrossRef] [PubMed]



© 2019 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (http://creativecommons.org/licenses/by/4.0/).