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RESEARCH ARTICLE

Removal of dyes (BG, MG, and SA) from aqueous solution using a novel adsorbent macrocyclic compound

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Abstract

The use of macrocyclic compounds to remove organic dyes is fascinating because they have a wide surface area range and can be used for different things. new (14E, 34E)-7-Hydroxy-7, 8, 22, 23, 24, 25, 26, 27-Octahydro-6H, 16H, 33H Tetrabenzo[f,k,u,z][1,5,13,20] Tetraoxacycloheptacosine-16,33-Dione (HOTTD) was obtained by a simple high-dilution method, and characterized by FTIR, ¹H-NMR, FESEM, EDX, and XRD. It worked well in removing organic dyes from aqueous solutions. Contact time, pH, dosage, initial concentration and temperature were studied. The optimum conditions were achieved by using 20 mg/ L dye concentration, 50 mg dose of adsorbent and pH 9.0 at room temperature. The adsorption process was remarkably fast and reached equilibrium within 10 min for both Brilliant Green and Malachite Green while 70 min for Safranin. The batch adsorption experiments followed a pseudo 2nd order and Langmuir model with maximum adsorption capacity 19.26 mg/g, 18.28 mg/g, and 14.35 mg/g for Brilliant Green, Malachite green and Safranin respectively. The process was endothermic and spontaneous in nature. Adsorbent regeneration test provides an excellent value 5 times.

Introduction

Nowadays, a severe problem is the colored waters, especially the wastewaters from dyeing industries. Water pollution is still a significant global issue that distresses the worldwide human population [1]. Dyeing effluent is produced by numerous sectors such as leather, plastics, cosmetics, and textiles due to fast economic and industrial growth. Most dyes and their metabolites are hazardous to aquatic living species, posing a severe risk to the natural environment and human safety. As a result, dyes must be removed from the wastewater solution to an acceptable level [2,3]. Various techniques for removing these contaminants have been established, namely chemical precipitation, membrane separation, solvent extraction, and physical adsorption. Some of these methods have significant drawbacks, such as limited removal efficiency, high removal cost, and extended removal time [4].

Adsorption is a surface phenomenon in a multi-component fluid such as gas or liquid where solutes are attached to the surface of a solid substance by chemical and physical bonds [5]. Adsorbents can be usually classified as natural adsorbents, synthetic adsorbents, and semisynthetic adsorbents [6,7]. Synthetic adsorbent such as polymeric resins, zeolites, or aluminosilicates; supramolecular macrocyclic compounds, magnetic nanoparticles (Fe₃O₄), and modified magnetic nanoparticles is the most popular adsorbent used by the researcher because of the high adsorption capacities [8,9]. Adsorption has the benefits of being simple to operate, efficient, and inexpensive compared to other approaches. Several efficient adsorbents have been produced in recent years because the adsorbent's characteristics significantly impact its performance [4]. The adsorbent chosen in the adsorption process is essential; however, specific adsorbents have a high cost, poor recyclability, and limited capacity, limiting their application.

Macrocyclic compounds are defined as cyclic compounds with nine or more members (including all hetero atoms) and with three or more donor atoms [10]. Macrocyclic compounds including crown ethers, cyclodextrin, cucurbituril, calixarene, and columnarene. have always been the main body of supramolecular chemistry research [11]. Supramolecular macrocyclic compounds with macrocyclic cavities as host molecules have a strong adsorption effect on many guest molecules, and they have received extremely wide attention as dye adsorbents. Among many dye treatment technologies, adsorption shows unique advantages in the treatment of difficult-to-degrade dyes [11,12].

The cavity of the macrocyclic compound has a specific adsorption effect on many dyes, and the surface groups of the macrocyclic compound will also form strong non-covalent interactions with the dye, such as electrostatic interaction, $\pi - \pi$ interaction, and hydrogen bonding. Their macrocyclic cavities can provide host-guest chemical effects, making them useful in separation and enrichment, molecular recognition, and supramolecular self-recognition. It has a high application value in assembly [12]. Because of the advantages of easy modification and inclusion effect, macrocyclic compounds have great application prospects in the adsorption and removal of dyes [13]. Macrocyclic compounds used as dye adsorbents can successfully achieve efficient and rapid dye adsorption and effectively promote the recycling of industrial wastewater [11]. When macrocyclic compounds are applied as dye adsorbents, they may accomplish efficient and quick dye adsorption while enhancing industrial wastewater recycling [14]. Many researchers have reported the use of supramolecular compounds for removal of organic dyes, Ting X. et.al. [2], And that reported in Hong-LiuJianget.al. And Rong G. et.al. [15,16].

The purpose of the paper is to prepare a new macrocycle compound [(14E, 34E)-7-hydroxy-7, 8, 22, 23, 24, 25, 26, 27-octahydro-6H, 16H, 33H-tetrabenzo[f,k,u,z][1,5,13,20] tetraoxacycloheptacosine-16,33-dione] using high-dilution method which is low cost, high product yield with less toxicity and highest chemical stability. The new crown compound have electron on the heteroatoms of the ring, which leads to increase the ability of the molecule with the ability to form complexes with the cations. The high-dilution principle, according to which low concentrations of the starting precursor favor cyclization over chain formation, was first developed by Paul Ruggli and Karl Ziegler for the cyclization of small organic molecules [17]. The prepared compounds were characterized by FTIR, ¹H-NMR, FESEM, EDX, and XRD. The novel prepared crown ether is used as the adsorbent for the removal of dyes such as Brilliant Green, Malachite green, and Safranin. The adsorption kinetics, isotherms, thermodynamics, and regeneration of the adsorbent to remove dyes were studied in detail. Additionally, contact time, starting dye concentrations, pH, temperature, and adsorbent dosage were utilized for this study. The uptake of the adsorbent was applied in synthetic dyes sample.

Material and methods

Reagents

All reagents used were of analytical reagent grade stated. Brilliant Green (BG) $(C_{27}H_{33}N_2HO_4S, M.wt = 482.65 \text{ g/mol})$ and Malachite Green (MG) $(C_{23}H_{25}ClN_2, M.$

wt = 365.0 g/mol) were purchased from Fluka. Safranin (SA) ($C_{20}H_{19}N_4Cl$, M.wt = 350.84 g/ mol) and Sodium carbonate (Na₂CO₃, \geq 99.5%, M.wt = 105.99 g/mol) were purchased from BDH. Sodium hydroxide (NaOH, 97%, M.wt = 40.0 g/mol), potassium hydroxide (KOH, 85–100.5%, M.wt = 56.11 g/mol) were purchased from Scharlau. Hydrochloric acid (HCl, 37%, M. wt = 36.45 g/mol) was purchased from Merck. 1,3-dichloropropane (98%, M.wt = 128.98 g/ mol) was purchased from Riedel-De Haen AG. Absolute Ethanol (C₂H₅OH, M.wt = 46.07 g/ mol), Methanol (CH₃OH, M.wt = 32.04 g/mol), 2-hydroxybenzaldehyde (C₇H₆O₂, 98%, M. wt = 122.123 g/mol), 2-hydroxy acetophenone (C₈H₈O₂, 99%, M.wt = 136.15 g/mol), and 1,6-dibromohexane (C₆H₁₂Br₂, 96%, M.wt = 243.96 g/mol) were purchased from Sigma-Aldrich.

Preparation of new materials

According to the modified method [18], compound (1) [2,2²-((2-hydroxypropane-1,3-diyl) bis (oxy))dibenzaldehyde] was prepared by dissolving (9.77 g, 0.08 mol) of 2- hydroxybenzaldehyde in 90 mL ethanol. The solution was heated in a round bottle flask, and the compound was activated by adding anhydrous Na_2CO_3 (33.92 g, 0.32 mol). The 2-hydroxy1,3-dichloropropane (5.16 g, 0.04 mol) was added to the mixture and diluted to 40 mL of absolute ethanol. The mixture was refluxed by stirring for 8.0 hours. Then, the mixture was poured into an ice bath. The residue was filtrated, dried, and recrystallized using chloroform/methanol solution (1:1) to obtain the solid white product (yield 85%, m.p. 115–117°C).

Compound (2) [1,1-((hexane-1,6-diylbis(oxy))bis(2,1-phenylene))bis(ethan-1-one)] was prepared depending on the modified method [19], in which 2-hydroxy acetophenone (10.89 g, 0.08 mol) was dissolved in 100 mL absolute ethanol, heated and stirred on a magnetic stirrer. (33.92 g, 0.32 mol) of Na₂CO₃ was added slowly for activation of 2-hydroxy acetophenone. Then 1,6-dibromohexane (5.16 g, 0.04 mol) in ethanol (30 mL) was added dropwise. First, the mixture was refluxed for ten-hour at 180–200°C (the effect of the temperature was optimized). The mixture was put in an ice bath. The resulting precipitate was collected by filtration, dried, and recrystallized using chloroform/methanol solution (1:1) to give a white product (yield 82%).

Compound (3) [(14E, 34E)-7-Hydroxy-7, 8, 22, 23, 24, 25, 26, 27-Octahydro-6H, 16H, 33H Tetrabenzo[f,k,u,z][1,5,13,20]Tetraoxacycloheptacosine-16,33-Dione (HOTTD)] as macrocyclic compound (MC) was prepared according to the modified method [20], A mixture of compound (1) (75 mg; 0.25 mmol) and compound (2) (88.6 mg; 0.25 mmol) was dissolved in a KOH solution (10%, 130–160 ml) in MeOH/H₂O (3:1) and the mixture was refluxed for 6.0 h. The reaction mixture was left at room temperature with stirring for five days; then, the solvent was reduced to nearly half volume under reduced pressure. The resulting precipitate was collected by filtration, dried, and recrystallized using chloroform/methanol solution (1:1) to obtain yellow precipitation (yield 82%).

Characterization of prepared materials

The synthesized macrocyclic compound was characterized in a way that covers a large area. Fourier transfer infrared spectra were carried out (Shimadzu IRAffinity-1 Fourier Transform Infrared (FTIR) spectrometer). The morphologies of particles were observed using FESEM coupled with EDX, (TESCAN MIRA3 FEG-SEM, Czech Republic at 15kV under a low vacuum after coating with a gold thin film, with an SE detector for EDX), and The powder XRD patterns were recorded by PANalytical X'Pert Pro diffractometer with Cu-k α radiation (1.5406 Å, 45 kV, 40 mA) and a 20 between 5 to 80°. Finally, 1H-NMR (broker AVANCENEO (400MHz) spectrometer).

Batch adsorption procedure

In this investigation, all experiments of adsorption were examined, including dosage effect using various amounts of sorbent 25–100 mg, pH effect from 2.0–11.0, the impact of initial dyes concentration 5–150 mg/L, temperature effect from 5–50°C and contact time 0–24 h.

Batch adsorption process applied with 50 mg of adsorbent and 10 mL of 20 mg/L of dye solution at room temperature and pH 9.0. Except for the adsorption kinetics experiment, the tube was kept in a water bath oscillator thermostat for 24 h at 200 rpm. The centrifuge was used to separate the sorbent and sorbate when the adsorption was completed at the specified time. After the adsorption process, the concentration of BG, MG, and SA dyes was determined using a Spectrophotometer at 618, 625, and 520 nm, respectively. The quantity of BG, MG, and SA sorbed onto sorbent and removal percentage were computed according to the following formula Eqs 1 and 2:

$$q_e = \frac{(C_i - C_e) \times V}{m} \tag{1}$$

$$R\% = \frac{(C_{\rm i} - C_{\rm e})}{C_{\rm i}} \times 100$$
(2)

Where equilibrium capacity of a sorbent is represented by qe (mg/g), V (L) is a volume of solution, m (g) means sorbent mass, both C_i and C_e (mg/L) were initial, and equilibrium BG, MG, and SA concentration, respectively and R was the removal percent of dyes.

Isotherms are mathematical equations that describe the adsorption behavior of a particular adsorbent-adsorbate combination. Langmuir, Freundlich, and Temkin models were used to agree with the adsorption data to investigate the mechanism of the adsorption process [21,22]. The linear version of the Langmuir equation depicts in Eq 3.

$$\frac{C_{e}}{q_{e}} = \frac{1}{qm \, \text{KL}} + \frac{C_{e}}{qm} \tag{3}$$

 K_L is a Langmuir constant related to adsorption capacity (mg/g), qm. That can be connected with variations in the appropriate area and porosity of the adsorbent, implying that adsorption capacity will be more significant if the surface area and volume are large. The separation factor R_L is a dimensionless constant that expresses in Eq.4 the essential characteristics of the Langmuir isotherm.

$$RL = \frac{1}{1 + KL C_e}$$
(4)

 R_L values indicate the adsorption to be unfavorable when $R_L > 1$, linear when $R_L = 1$, favorable when $0 < R_L > 1$, and irreversible when $R_L = 0$.

While the following Eqs 5 and 6 represent Freundlich and Timken isotherms linear forms:

$$\log q_e = \log KF + \frac{1}{n} \log C_e \tag{5}$$

$$q_e = at + 2.303bt \log C_e \tag{6}$$

1/n indicates adsorption capacity, while the K_F represents adsorption capacity (L/mg). The Temkin constant bt is related to the heat of sorption (J/mol) and the Temkin isotherm constant at (L/g). Thermodynamic parameters such as ΔG° , ΔH° , and ΔS° were calculated using

Eqs <u>7</u>–<u>9</u>.

$$\mathrm{Kd} = \frac{\mathbf{q}_{\mathrm{e}}}{C_{\mathrm{e}}} \tag{7}$$

$$\ln Kd = \frac{\Delta S^{\circ}}{R} - \frac{\Delta H^{\circ}}{RT}$$
(8)

$$\Delta G^{\circ} = -RT \ln Kd \tag{9}$$

Where T is the absolute temperature K; R is the universal gas constant 8.314 J/mol. K; K_d is the adsorption distribution constant derived using qe/Ce and qe mg/g [23].

This research investigated the adsorption kinetics by fitting the experimental data to pseudo 1^{st} order, pseudo 2^{nd} order, and intra-particle diffusion kinetic models developed in the previous literature. Then, using linearized pseudo 1^{st} order, pseudo 2^{nd} order, and intra-particle diffusion, represent them in the following Eqs 10–12, [24].

$$\log(\mathbf{q}_{e} - \mathbf{qt}) = \log \mathbf{q}_{e} - \frac{\mathbf{k}_{1}\mathbf{t}}{2.303} \tag{10}$$

$$\frac{t}{qt} = \frac{1}{q_e} + \frac{1}{k_2 q_e^2}$$
(11)

$$qt = k_3 t^{0.5} + C (12)$$

The k_1 (min⁻¹) and k_2 (g/mg.min) are connected to the pseudo 1st and 2nd order kinetic models. The intercept and the slope of the plot of log(q_e- qt) against t were used to determine the value of qe. k_3 is the intra-particle diffusion rate constant (mg/g.min^{1/2}) which can be obtained from the slope of the straight line and C is an intra-particle diffusion rate constant related to the boundary layer thickness (mg/g) which can be obtained from the intercept of the straight line [25].

The adsorption/desorption study was cared by 50 mg of MC placed in a conical flask containing dyes solution 10 mL, 20 mg/L and the mixture was shaken in a water bath using thermostat shaker for 70 min at room temperature. The mixture was separated by centrifuge and the final dye concentration was found. The adsorbent was recycled by washing with ethanol and water under constant stirring for three times (1.0 h), respectively. The experiment repeated and dipped the adsorbent into another dye solution to begin a new adsorption process. The recyclability of the material was followed up to five consecutive adsorption-desorption cycles.

Results and discussion

Characterization of new adsorbent

A new crown ether was prepared by condensation reaction between the aldehyde and ketone.

FT-IR

The new aldehyde (2, 2-(propane-1,3-diylbis(oxy) dibenzaldehyde) (Compound 1) was prepared in good yield (85%) based on Williamson ether synthesis between 2-hydroxy Benzaldehyde and 1,3 dichloro-2-propanol [18].



The structure of the synthesized compound was determined using FT-IR techniques, FT-IR spectrum of aldehyde compound (1) (**Fig 1**), the appearance of broadband at 3460.3 cm⁻¹ is attributed to the OH group, and the strong band at 1678 cm⁻¹ refers to the carbonyl group while 2877, 2949 cm⁻¹ C-H Aliphatic.

A new ketone 1,1-((hexane-1,6-diylbis(oxy))bis(2,1-phenylene))bis(ethan-1-one) compound (2) was prepared in good yield (82%) based on Williamson ether synthesis between 2-hydroxy acetophenone and 1,6-dibromohexane.



FT-IR characterized the structure of the new synthesized compound (2) (Fig 1), the appearance of a strong band at 1662 cm⁻¹ refers to the carbonyl group of the ketone, while 2870, 2939.5 cm⁻¹ refers to C-H aliphatic.

According to the reported procedure [20]. A new crown ether was synthesized by modifying Classen–Schmidt condensation between 2,2'-((2-hydroxypropane-1,3-diyl) bis(oxy)) dibenzaldehyde (Compound 1) and 1,1-((hexane-1,6-diylbis(oxy))bis(2,1-phenylene))bis (ethan-1-one) in a basic medium was synthesized in good yield (82%).



tetrabenzo[f,k,u,z][1,5,13,20]tetraoxacycloheptacosine-16,33-dione

FT-IR characterized of a new synthesized crown ether compound (3) structure, as shown in **Fig 1**. FT-IR spectrum of compound shows shifting of the carbonyl group to the lower value 1678 and 1662 cm⁻¹ to the 1654 cm⁻¹ corresponds to the carbonyl group of α , β -unsaturated ketone, that is a useful guide to identify conjugated enone as described in the literature.

¹H-NMR. The ¹H-NMR spectrum of compound (1) (S1 Fig), shows a signal at 4.26 ppm refers to the hydroxyl group, while 4.31 ppm belongs to methylene proton, a quintet at 4.55 ppm refers to CH- aliphatic protons, a multiple at 7.0–7.8 ppm corresponds to the eight protons of two aromatic ring, and a single signal at 10.48 ppm belongs to the proton of an aldehyde group.

The ¹H-NMR spectrum of compound (2) (S2 Fig), as illustrated by a signal at 2.65 ppm, refers to (s, 6H, $2 \times$ -COCH3), 4.10 ppm belonging to methylene proton, a multiple at 1.62–1.92 ppm refers to eight CH- aliphatic protons and multiples at 6.95–7.03 ppm corresponds to the eight protons of two aromatic ring.



Fig 1. FT-IR spectrum of Crown ether, Aldehyde, and Ketone.

The ¹H-NMR spectrum of compound (3) (S3 Fig) shows the disappearance of an aldehyde group at 10.48 ppm and the appearance of multiple signals in the range 0.83–1.3 ppm and 2.5 ppm, which refers to eight CH-aliphatic, multiple protons in the range 3.47–4.02 ppm refers to nine protons of (4X-OCH₂-, and X-CH-OH), a signal at 5.48 ppm refers to X–OH, and a multiple at 6.91–7.81 ppm fits 16 protons of four aromatic ring and 4 protons of α , β unsaturated enones.

FESEM. FESEM (Field Emission Scanning Electron Microscopy) has been a primary tool for characterizing the surface morphology and fundamental physical properties of the adsorbent before and after the adsorption of dyes. Fig 2(A) Shows many pores in the adsorbent surface with various grooves favorable for adsorption. A significant change is observed after dyes adsorption Fig 2(B)–2(D) in the adsorbent structure. The adsorbent surface was changed, and dye molecules coated on all the surfaces and pores of the adsorbent decreased after adsorption. This surface property should be considered a factor for dye binding.

EDX. The EDX analysis indicated the presence of certain components in the composition, which were then determined using the spectrum. As illustrated in Fig 3(A), C and O are the two most essential components in producing macrocyclic compounds. In Fig 3(B)-3(D) appearing new atoms such as N, S, and Cl, the dye molecules successfully sorption on the sorbent.

XRD. The XRD analysis was completed, and the findings are shown to examine the crystallographic structure of the adsorbent. XRD patterns were presented in Fig 4. When comparing the XRD pattern of MC before adsorption and after adsorption of BG, MG, and SA dyes. XRD pattern of MC has broad peaks (~22° or ~25°), demonstrating the semicrystalline or amorphous structure of macrocyclic compounds [26]. After the adsorption of dyes, it was discovered that they significantly changed with a decrease in the crystalline structure. Consequently, dye molecules preferentially adsorb via chemisorption and only partly through physisorption. While XRD analysis shows no significant changes in peak locations $2\theta > 30$,



Fig 2. FESEM images of (a) MC before adsorption of dye (magnification: 25 kx, $2 \mu m$), (b) MC after adsorption of BG (magnification: 25 kx, $2 \mu m$), (c) MC after adsorption of MG (magnification: 25 kx, $2 \mu m$), and (d) MC after adsorption of SA (magnification: 25 kx, 500 nm).

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Fig 4. Demonstrates the X-ray diffraction spectra for MC (black line), MC after adsorption of BG (red line), MC after adsorption of MG (blue line), and MC after adsorption of SA (green line).

only an increase in intensity due to dyes adsorption of adsorbent, this indicates that the crystal structure of the adsorbent was not destroyed after adsorption of dyes.

Adsorption time and kinetic study

The impact of contact time on the removal percentage of BG, MG, and SA was achieved over the time interval of 0-24 h under fixed conditions values (pH = 9, adsorbent dosage 50mg/ 10mL and 20 mg/L concentration of dye), the results are shown in Fig 5.

From the chart, we can see in the initial minutes, a removal percent value of about 89.00% for BG, 90.50% for MG, and 24.60% for SA was attained, when the contact time reached 10





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Kinetic models	Parameter	BG	MG	SA
	q _e (exp.) (mg/g)	arameter BG exp.) (mg/g) 4.0 $cal (mg/g)$ 0.1803 $c_{1} (min^{-1})$ 1.714 R ² 0.6622 $cal (mg/g)$ 4.0 (g/mg.min) 3.27 R ² 1.0 C (mg/g) 3.651 mg/g.min ^{1/2}) 0.0609	4.0	3.869
	$q_{cal} (mg/g)$	0.1803	0.1881	1.622
Pseudo-first order	$k_1 (min^{-1})$	1.714	0.1096	0.0553
	R ²	0.6622	0.7443	0.8433
	q _{cal} (mg/g)	4.0	4.001	3.869
Pseudo-second order	k ₂ (g/mg.min)	3.27	2.715	0.1637
	R ²	1.0	0.9996	1.0
	C (mg/g)	3.651	3.572	0.2389
Intra-particle diffusion	$k_3 (mg/g.min^{\frac{1}{2}})$	0.0609	0.0612	0.596
	R ²	0.9907	0.9832	0.9854

Table 1. Kinetic characteristics for the adsorption of BG, MG, and SA.

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min for both BG and MG, and 70 min for SA, the value of removal increase to 99.66% and 99.86%, and 96.19% of BG, MG, and SA respectively. By further increasing the contact time after 1.0 min, the removal percentage was increased slightly reaching the maximum value at 10 min, after this time there was no significant change by increasing the time. The optimized equilibrium time is about 10 min for BG and MG, enough for further experiments.

While for SA, According to the diagram, the removal percent increased rapidly when contact time reached 70 min, then increased slightly, reaching a maximum value of about 96.19%. Thus, the subsequent adsorption experiments were performed at 70 min.

The result showed that the prepared adsorbent has short equilibrium time which shows the adsorbent have enough number of active site for a given concentration. Compared with the literature, this equilibrium time is better than the ones previously reported by Ahmed et al. [27] and Abbas [28] for BG, Palapa et al. [29] and Alorabi [30] for MG and Abukhadra and Mohamed [31] and Mullerova et al. [32] for SA.

The kinetic studies can give information about the nature of the adsorption system (chemical or physical) and the strength of held between the adsorbate and the used adsorbent [31]. Pseudo-first-order, pseudo-second-order, and intra-particle diffusion kinetic models were employed to match the experimental sorption data in this work, **S4 Fig** and **Table 1** provide the fitting parameters for the kinetic models. When Pseudo-first-order is compared to the pseudo-second-order kinetic model, the R² value of the pseudo-second-order kinetic model is more significant for BG, MG, and SA; the value is 1.0, 0.9996 and 1.0 respectively, suggesting that the adsorption of all dyes onto crown ether adsorbent occurred by the chemisorption mechanism. Moreover, the qe, exp value derived by pseudo-second-order was consistent with the qe, cal. value, supporting the pseudo-second-order kinetic model. Similar trends have been reported by other adsorbents [28, 33] for BG, [30,34] for MG and [32] for SA.

The intra-particle diffusion explains the movement of dye molecules from the bulk solution phase to the solid phase. The constants k_3 and C are calculated and listed in Table 1. As shown in S4 Fig the plots of qt vs. t^{1/2} for BG, MG, and SA dye molecules are shown in two stages. The first straight portion illustrates macropore diffusion and the second stage describes micropore diffusion. All the intra-particle diffusion rate constants C in this work are not zero which implies that the adsorption method may not be predominantly controlled by intra-particle diffusion [35].



Fig 6. Effect of pH on removal percentage of BG, MG, and SA onto adsorbent.

Effect of adsorption conditions

Effect of pH on adsorption. A significant factor in the adsorption experiment is the pH solution. The dye becomes more unstable at a particular pH and becomes easier to remove from the aqueous solution via selective sorbent. pH has a significant impact on the ionization and charge of the functional group of the adsorbate [36]. To look at the effects of solution pH on the adsorption of BG, MG, and SA dyes, the pH of the adsorbate solution was changed from 2.0 to 11.0 and initially adapted with either HCl or NaOH 1.0 M and 0.1 M. The experiment was performed at a fixed amount of adsorbent 50 mg/10 mL, an initial dye concentration of 20 mg/L, and under a contact time of 70 min. Fig 6 illustrates the removal efficiency of BG increase 58.9% - 96.46% with an increasing pH range of 2.0–9.0, then the removal efficiency constant until it reaches pH 11.0. At the same time, the removal percent for MG 45.9% - 99.19% was observed in pH 2.0–9.0. After that, adsorption is almost constant until it reaches pH 11.0. The removal percent for SA 30.9% - 95.88% was observed in pH 2.0–10.0; then, it was constant. pH 9 was chosen as the optimum pH for the removal of dyes onto the adsorbent. Similar findings have been reported in other publications for BG [37], MG [38], and SA [39] dyes.

Concentration of initial dyes effect. The impact of initial dye concentrations ranging from 5.0 mg/L to 150 mg/L and under the optimized values of the other experimental conditions (adsorbent dosage of 50 mg/10 mL, initial pH of 9, and contact time of 70 min) on removal percent was studied to assess the adsorption performance of dye molecules on the adsorbent. As presented in Fig 7, raising the starting dye concentration reduced the adsorption percent. As a result, when the initial concentration rises from 10 mg/L to 60 mg/L for BG, MG, and SA, the removal percent between 99.99% - 96.61%, 99.96% - 97.13%, and 100% - 84.22% respectively, but when the concentration rises beyond 60 mg/L, the percentage drops slightly. This behavior may be described in the following way, the degree of active sites was sufficient to



Fig 7. Effect of concentration on adsorption of BG, MG, and SA onto adsorbent.

occupy dye molecules at low concentrations, resulting in high removal efficiency. Conversely, with higher starting dye concentrations and the same adsorbent mass, a scarcity of adsorption sites may develop, resulting in a decrease in removal effectiveness due to the active site of adsorbent saturated by dye molecules at extremely high starting concentrations [40,41].

Adsorbent dose effect. On dye molecule sorption, the action of adsorbent weight was studied. As observed in Fig 8, the optimal weight for adsorbent was 50 mg, which resulted in 99.68%, 100%, and 96.5% removal for BG, MG, and SA, respectively. Initially, the adsorption process increases as the adsorbent mass increases, but after an optimal dose is achieved, it stays constant. Due to a more significant number of active sites, clearance efficiency is anticipated to rise. Any further addition of the adsorbent seemed to have no significant effect on adsorption, which might be due to adsorption site overlapping due to adsorbent particle crowding [42,43]. Generally, 50 mg/10mL was taken as an optimum quantity for this work.

Effect of temperature. The mobility of molecules and ions in a solution is affected by temperature. This can include ion adsorption since ions must be mobile to 'collide/interact' with the adsorbent and enhance adsorption, which is especially important in batch adsorption investigations. Therefore, the effect of temperature on the removal % of dyes was studied, and the results are shown in **Fig 9**. The measurements were carried out at varying temperatures ranging from 5.0 to 45°C. As observed from the figure, the removal % increased by increasing temperature from 5.0 to 25°C for BG and 5.0 to 30°C for both MG and SA. Then removal % remained constant for BG and MG, above 30°C for SA. There is an apparent decrease in the removal percentage. As a result, throughout the rest of this study, the room temperature was given maximum removal, and it was selected as the optimum removal temperature for all dyes [39,44,45].



According to basic thermodynamic principles, energy cannot be acquired or lost in an isolated system, and entropy change is the only driving factor. Therefore, environmental engineers must examine both energy and entropy factors to determine which process will occur naturally. The effect of temperature on the BG, MG, and SA dyes' adsorption onto sorbent was investigated using thermodynamics. Kd, the thermodynamic equilibrium constant, may be used to calculate the thermodynamic parameters Eq (7). The Eqs (8) and (9) are used to compute the standard Gibbs free energy ΔG° kJ/mol, standard enthalpy change ΔH° kJ/mol, and standard entropy change ΔS° J/mol K.

The slope and intercept of the linear plot of ln Kd versus 1/T were used to get the values of Δ H and Δ S, whereas the values of Δ G were determined using the equation above. As may be seen, the adsorption of dyes is more favorable at higher temperatures, implying that the adsorption process is endothermic, as shown by the positive values of Δ H (Table 2). The results suggest that chemisorption might be dominant. Following that, a positive value of Δ S implies a high degree of randomness at the interface between the solid and the solution. The negative values of Δ G indicate that the adsorption of all dyes onto the adsorbent occurred as a result of spontaneous adsorption [46,47].

Adsorption isotherms models study. The equilibrium of sorption is an essential physicochemical feature in explaining adsorption behavior in solid-liquid systems. The object of adsorption isotherms is to establish a relationship between the adsorbate concentration in bulk and the quantity adsorbed at the interface [46]. The results were analyzed using the Langmuir,



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Fig 9. Effect temperature on dye removal percent.
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Freundlich, and Temkin isotherms. The parameters are determined and presented in Table 3. The applicability of the isotherm to describe the adsorption process was judged by the value of correlation coefficient R^2 . According to the results, the adsorption of all dyes is best fitted with

Dyes	T(K)	ΔG (kJ/mol)	ΔH(kJ/mol)	ΔS (J/mol K)
	278	+1.183	+ 153.38	+ 564.89
BG-MC	288	-0.0893		
	298	-13.998		
	303	-15.09		
	308	-15.90		
	318	-18.24		
	278	+3.143	+166.72	+585.56
MG-MC	288	-1.032		
	298	-4.303		
	303	-13.344		
	308	-16.65		
	318	-18.26		
	278	+1.160	+48.45	+170.2
SA-MC	288	+0.0244		
	298	-2.116		
	303	-4.454		
	308	-4.545		
	318	-4.783		

Table 2. Thermodynamic parameters for the adsorption of BG, MG, and SA dye molecules onto the adsorbent.

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Isotherm models	Parameter	BG	MG	SA
	q _m (mg/g)	19.26	18.28	14.35
Langmuir	$K_L (min^{-1})$	3.99	3.055	2.010
	R ²	0.9992	0.9994	0.9991
	R _L	(0.0244-0.0020)	(0.0316-0.0027)	(0.0455-0.0041)
	1/n	0.2282	0.2656	0.3633
Freundlich	K _f	9.166	8.218	4.304
	R ²	0.9807	0.9308	0.9618
	a _t	11.467	10.381	5.426
Temkin	b _t	1.74	2.0570	2.438
	R ²	0.9118	0.9781	0.9949

Table 3.	The parameters of	of the Langmuir and	Freundlich and	Temkin models for	r dye adsorpti	on onto adsorbent
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the Langmuir isotherm. It is concluded that reactive zones on the adsorbent sites are homogenously covered by dyes molecule. R_L , separation factor is favorable for adsorbent toward BG, MG, and SA with ranges (0.0244–0.0020), (0.0316–0.0027), and (0.0455–0.0041), respectively, within the concentration range of 10 to 120 mg/L, pH = 9.0, adsorbent dosage 50 mg/10 mL and room temperature.

Regeneration and reusability. Regeneration and reusability of valuable adsorbates is an important aspect of adsorption technology. Here, desorption of adsorbent was tested by evaluating the dyes adsorption efficiency through treatment with four different eluents (deionized distilled water, ethanol, methanol, and nitric acid); the best eluent is ethanol. After the desorption test, the adsorbent's reusability was tested for up to five adsorption-desorption cycles for all dyes as shown in **Fig 10**. The bar chart show the relationship between the number of recycles and the removal efficiency of the regenerated adsorbent. It can be seen that the removal efficiency was observed during the five adsorption-desorption cycles for BG and MG. These data are encouraging and propose that the adsorbent in this experimental study has good reusability. While for SA dye molecule after the third adsorption-desorption cycle the removal efficiency decreased gradually and stayed at 70% for five cycles.

The chemical interaction- mechanical suggestion. A suggested adsorption mechanism between target dye molecules and functional groups of crown ether as shown in S5 Fig. As can be seen, electrostatic attraction and dipole–dipole interactions are found to be the dominant interactions between target dye and adsorbent surface. At higher solution pH values, the surface of crown ether particles become a negatively charge (-OH group) and the electropositive groups (nitrogen (N) atom) on the cationic dye molecules, As a result, dye ions are thus more favorably adsorbed onto the surfaces of the adsorbent at relative higher pH values. FESEM and EDX give another support for this mechanism (host-guest) by occupied all the pores in the adsorbent surface and present of dye atoms in EDX chart. A similar mechanism has been previously reported by Fayazi M. et al [39].

Comparison with other adsorbents. BG, MG, and SA adsorption capacity were considerable compared to other adsorbents found in the literature, as shown in Table 4.

From Table 4, various adsorbents have been used for uptake of BG, MG and SA. Each bio sorbent has different adsorption capacity depends on the adsorbent nature and functional groups. The new prepared crown ether show high removal capacity during 10 min at room temperature and pH 9.0 for both BG and MG while 70 min for SA.



Adsorbents	Dyes	Adsorption capacity (mg/g)	References
Mn0.5Cu0.5Fe ₂ O ₄ nanospinels	BG	0.89	[48]
Rambutan peels	-	9.64	[49]
Tannin gel (TG) mine modified tannin gel (ATG)		20.41 8.55	[50]
Salix alba leaves (SAL)		15.89	[51]
MC (HOTTD)		19.26	This study
Treated sawdust		21.7	[52]
Rice husk treated with NaOH	MG	12.16	[53]
ZnO-NR-AC		20	[54]
Activated carbon (AC)		28.5	[55]
MC (HOTTD)		18.28	This study
Rice husk (treated with NaOH)		9.77	[56]
Kaolinite (NRK) Clay	SA	16.23	[57]
Iron oxide/sepiolite		18.48	[39]
Bambusa tulda		32.26	[58]
Nano FeO/CA		1.910	[59]
MC (HOTTD)		14.35	This study

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Conclusions

The effectiveness of a novel macrocyclic molecule as an adsorbent for removing BG, MG, and S dye was proven. The effect of adsorbent dose, contact duration, and solution pH on dye absorption was determined. The following parameters were found to be optimal after optimizing the variables indicated above: 50 mg of adsorbent, contact time of 10 minutes for both BG and MG, and 70 minutes for SA, and pH >8.0. A high percentage of dye removal was observed based on this improved process. The equilibrium adsorption data revealed that both models, pseudo-second-order, and Langmuir, very effectively described the dyes adsorption process of the adsorbent. According to the Langmuir model, the maximum adsorption capacity (q_m) for BG, MG, and SA was 19.29, 18.28, and 14.35 mg/g, respectively. Overall, this research suggests that the synthesized adsorbent has a high potential for usage in removing dyes in wastewater treatment.

Supporting information

S1 Fig. ¹H-NMR spectrum of compound (1). (DOCX)

S2 Fig. ¹H-NMR spectrum of compound (2). (DOCX)

S3 Fig. ¹H-NMR spectrum of compound (3). (DOCX)

S4 Fig. (a) pseudo-first-order kinetics model, (b) pseudo-second-order kinetics model, and (c) Intra-particle diffusion kinetic model for adsorption of BG, MG, and SA by MC adsorbent. (DOCX)

S5 Fig. Possible Safranin (SA) adsorption mechanism on the adsorbent structure. (DOCX)

Author Contributions

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References

- Anastopoulos I, Hosseini-Bandegharaei A, Fu J, Mitropoulos AC, Kyzas GZ. Use of nanoparticles for dye adsorption. Journal of Dispersion Science and Technology. 2018; 39(6):836–47.
- Xu T, He Y, Qin Y, Zhao C, Peng C, Hu J, et al. Facile preparation of porous organic copolymer based on triptycene and crown ether for efficient organic dye adsorption. RSC advances. 2018; 8(9):4963–8. https://doi.org/10.1039/c7ra12495c PMID: 35539550
- Li B, Wang Q, Guo J-Z, Huan W-W, Liu L. Sorption of methyl orange from aqueous solution by protonated amine modified hydrochar. Bioresource technology. 2018; 268:454–9. <u>https://doi.org/10.1016/j.</u> biortech.2018.08.023 PMID: 30107359
- Liu Y, Wang S, Zhou H, Li X, Duan L, Chen S, et al. Simultaneous removal of methyl orange and Cr (VI) using polyethyleneimine-modified corncob-derived carbon material. BioResources. 2020; 15(4):7342– 56.
- Sasaki T, Iizuka A, Watanabe M, Hongo T, Yamasaki A. Preparation and performance of arsenate (V) adsorbents derived from concrete wastes. Waste management. 2014; 34(10):1829–35. <u>https://doi.org/ 10.1016/j.wasman.2014.01.001</u> PMID: 24472713
- Cojocariu B, Mocanu AM, Nacu G, Bulgariu L. Possible utilization of PET waste as adsorbent for Orange G dye removal from aqueous media. Desalination and Water Treatment. 2018; 104:338–45.
- 7. Bulgariu L, Escudero LB, Bello OS, Iqbal M, Nisar J, Adegoke KA, et al. The utilization of leaf-based adsorbents for dyes removal: A review. Journal of Molecular Liquids. 2019; 276:728–47.
- Crini G, Lichtfouse E, Wilson LD, Morin-Crini N. Conventional and non-conventional adsorbents for wastewater treatment. Environmental Chemistry Letters. 2019; 17(1):195–213.
- Sazali N, Harun Z, Sazali N. A review on batch and column adsorption of various adsorbent towards the removal of heavy metal. Journal of Advanced Research in Fluid Mechanics and Thermal Sciences. 2020; 67(2):66–88.
- 10. Schettini R. Novel macrocyclic systems in asymmetric phase-transfer catalysis. 2017.
- Li P, Zhang D, Jia Q. Research advances in supramolecular macrocyclic compounds for dye adsorption. Se pu = Chinese Journal of Chromatography. 2020; 38(3):297–306. https://doi.org/10.3724/SP.J. 1123.2019.10018 PMID: 34213209
- Zhang H, Zou R, Zhao Y. Macrocycle-based metal-organic frameworks. Coordination Chemistry Reviews. 2015; 292:74–90.
- Meo PL, Lazzara G, Liotta L, Riela S, Noto R. Cyclodextrin–calixarene co-polymers as a new class of nanosponges. Polymer Chemistry. 2014; 5(15):4499–510.
- Ma J, Zhang Y, Zhao B, Jia Q. Supramolecular adsorbents in extraction and separation techniques-A review. Analytica Chimica Acta. 2020; 1122:97–113. https://doi.org/10.1016/j.aca.2020.04.054 PMID: 32503749
- Jiang H-L, Lin J-C, Hai W, Tan H-W, Luo Y-W, Xie X-L, et al. A novel crosslinked β-cyclodextrin-based polymer for removing methylene blue from water with high efficiency. Colloids and Surfaces A: Physicochemical and Engineering Aspects. 2019; 560:59–68.
- Guo R, Wang R, Yin J, Jiao T, Huang H, Zhao X, et al. Fabrication and highly efficient dye removal characterization of beta-cyclodextrin-based composite polymer fibers by electrospinning. Nanomaterials. 2019; 9(1):127. https://doi.org/10.3390/nano9010127 PMID: 30669533
- Fan Q, Wang T, Dai J, Kuttner J, Hilt G, Gottfried JM, et al. On-surface pseudo-high-dilution synthesis of macrocycles: principle and mechanism. ACS nano. 2017; 11(5):5070–9. <u>https://doi.org/10.1021/</u> acsnano.7b01870 PMID: 28419801
- Ilhan S, Temel H, KIIIc A. Synthesis and spectral studies of macrocyclic Cu (II) complexes by reaction of various diamines, copper (II) perchlorate and 1, 4-bis (2-carboxyaldehyde phenoxy) butane. Journal of Coordination Chemistry. 2008; 61(2):277–84.
- Mondal R, Mandal TK, Mallik AK. Simple synthesis of a new family of 22-to 28-membered macrocycles containing two chalcone moieties. Arkivoc. 2012; 2012(9):95–110.
- Mohamed SK, Akkurt M, Hawaiz FE, Ayoob MM, Hosten E. Crystal structure of (6E, 20E)-3, 24difluoro-13, 14, 28, 29-tetrahydro-5H, 22H-tetrabenzo [e, j, p, u][1, 4, 12, 15] tetraoxacyclodocosine-5, 22-dione. Acta Crystallographica Section E: Crystallographic Communications. 2017; 73(1):13–6.
- 21. Ibrahim BM. Heavy metal ions removal from wastewater using various low-cost agricultural wastes as adsorbents: a survey. Zanco Journal of Pure and Applied Sciences. 2021; 33(2):76–91.

- Ayawei N, Ebelegi AN, Wankasi D. Modelling and interpretation of adsorption isotherms. Journal of chemistry. 2017;2017.
- He Y, Wu P, Xiao W, Li G, Yi J, He Y, et al. Efficient removal of Pb (II) from aqueous solution by a novel ion imprinted magnetic biosorbent: Adsorption kinetics and mechanisms. PloS one. 2019; 14(3): e0213377. https://doi.org/10.1371/journal.pone.0213377 PMID: 30917141
- Al-Musawi TJ, Mengelizadeh N, Al Rawi O, Balarak D. Capacity and modeling of acid blue 113 dye adsorption onto chitosan magnetized by Fe2O3 nanoparticles. Journal of Polymers and the Environment. 2022; 30(1):344–59.
- 25. Yilmaz M, Al-Musawi TJ, Khatibi AD, Baniasadi M, Balarak D. Synthesis of activated carbon from Lemna minor plant and magnetized with iron (III) oxide magnetic nanoparticles and its application in removal of Ciprofloxacin. Biomass Conversion and Biorefinery. 2022:1–14.
- Xu B, Zheng H, Zhou H, Wang Y, Luo K, Zhao C, et al. Adsorptive removal of anionic dyes by chitosanbased magnetic microspheres with pH-responsive properties. Journal of Molecular Liquids. 2018; 256:424–32.
- Ragab A, Ahmed I, Bader D. The removal of brilliant green dye from aqueous solution using nano hydroxyapatite/chitosan composite as a sorbent. Molecules. 2019; 24(5):847. <u>https://doi.org/10.3390/</u> molecules24050847 PMID: 30823358
- Abbas M. Removal of brilliant green (BG) by activated carbon derived from medlar nucleus (ACMN)– Kinetic, isotherms and thermodynamic aspects of adsorption. Adsorption Science & Technology. 2020; 38(9–10):464–82.
- 29. Palapa NR, Mohadi R, Rachmat A. Adsorption study of malachite green removal from aqueous solution using Cu/M³⁺ (M³⁺ = AI, Cr) layered double hydroxide. Mediterranean Journal of Chemistry. 2020; 10 (1):33–45.
- Alorabi AQ. Effective removal of malachite green from aqueous solutions using magnetic nanocomposite: synthesis, characterization, and equilibrium study. Adsorption Science & Technology. 2021;2021.
- Abukhadra MR, Mohamed AS. Adsorption removal of safranin dye contaminants from water using various types of natural zeolite. Silicon. 2019; 11(3):1635–47.
- Mullerova S, Baldikova E, Prochazkova J, Pospiskova K, Safarik I. Magnetically modified macroalgae Cymopolia barbata biomass as an adsorbent for safranin O removal. Materials Chemistry and Physics. 2019; 225:174–80.
- Giri BS, Gun S, Pandey S, Trivedi A, Kapoor RT, Singh RP, et al. Reusability of brilliant green dye contaminated wastewater using corncob biochar and Brevibacillus parabrevis: hybrid treatment and kinetic studies. Bioengineered. 2020; 11(1):743–58. https://doi.org/10.1080/21655979.2020.1788353 PMID: 32631112
- Abate GY, Alene AN, Habte AT, Getahun DM. Adsorptive removal of malachite green dye from aqueous solution onto activated carbon of Catha edulis stem as a low cost bio-adsorbent. Environmental Systems Research. 2020; 9(1):1–13.
- Sheha R, El-Zahhar A. Synthesis of some ferromagnetic composite resins and their metal removal characteristics in aqueous solutions. Journal of Hazardous Materials. 2008; 150(3):795–803. <u>https://doi.org/ 10.1016/j.jhazmat.2007.05.042</u> PMID: 17630189
- Laskar N, Kumar U. Removal of Brilliant Green dye from water by modified Bambusa Tulda: adsorption isotherm, kinetics and thermodynamics study. International Journal of Environmental Science and Technology. 2019; 16(3):1649–62.
- Ishaq M, Sultan S, Ahmad I, Saeed K. Removal of Brilliant Green Dye from Aqueous Medium by Untreated Acid treated and Magnetite Impregnated Bentonite Adsorbents. Journal of the Chemical Society of Pakistan. 2017; 39(5).
- Agarwal S, Tyagi I, Gupta VK, Mashhadi S, Ghasemi M. Kinetics and thermodynamics of Malachite Green dye removal from aqueous phase using iron nanoparticles loaded on ash. Journal of Molecular Liquids. 2016; 223:1340–7.
- Fayazi M, Afzali D, Taher M, Mostafavi A, Gupta V. Removal of Safranin dye from aqueous solution using magnetic mesoporous clay: Optimization study. Journal of Molecular Liquids. 2015; 212:675–85.
- Amar I, Sharif A, Ali M, Alshareef S, Altohami F, Abdulqadir M, et al. Removal of methylene blue from aqueous solutions using nano-magnetic adsorbent based on zinc-doped cobalt ferrite. Chemical Methodologies. 2020; 4(1):1–18.
- 41. Yadav S, Asthana A, Chakraborty R, Jain B, Singh AK, Carabineiro SA, et al. Cationic dye removal using novel magnetic/activated charcoal/β-cyclodextrin/alginate polymer nanocomposite. Nanomaterials. 2020; 10(1):170. https://doi.org/10.3390/nano10010170 PMID: 31963657

- Qiao J, Gao S, Yao J, Zhang L, Li N. A rapid and green method for the removal of anionic dyes from aqueous solution using sulfobetaine-modified magnetic nanoparticles. AIP Advances. 2019; 9 (6):065308.
- **43.** Zad ZR, Moosavi B, Taheri A. Synthesis of monodisperse magnetic hydroxyapatite/Fe₃O₄ nanospheres for removal of Brilliant green (BG) and Coomassie brilliant blue (CBB) in the single and binary systems. Advanced Journal of Chemistry-Section B. 2020; 2(3):159–71.
- 44. Zhang L, Zhang Y, Tang Y, Li X, Zhang X, Li C, et al. Magnetic solid-phase extraction based on Fe3O4/ graphene oxide nanoparticles for the determination of malachite green and crystal violet in environmental water samples by HPLC. International Journal of Environmental Analytical Chemistry. 2018; 98 (3):215–28.
- **45.** Rezayati Zad Z, Moosavi B, Taheri A. Synthesis of monodisperse magnetic hydroxyapatite/Fe₃O₄ nanospheres for removal of Brilliant green (BG) and Coomassie brilliant blue (CBB) in the single and binary systems. Advanced Journal of Chemistry-Section B. 2020; 2(3):159–71.
- Salem MA, Elsharkawy RG, Hablas MF. Adsorption of brilliant green dye by polyaniline/silver nanocomposite: Kinetic, equilibrium, and thermodynamic studies. European polymer journal. 2016; 75:577–90.
- Kataria N, Garg V. Application of EDTA modified Fe₃O₄/sawdust carbon nanocomposites to ameliorate methylene blue and brilliant green dye laden water. Environmental research. 2019; 172:43–54.
- Hashemian S, Dehghanpor A, Moghahed M. Cu0. 5Mn0. 5Fe₂O₄ nano spinels as potential sorbent for adsorption of brilliant green. Journal of Industrial and Engineering Chemistry. 2015; 24:308–14.
- Nor NM, Hadibarata T, Yusop Z, Lazim ZM. Removal of brilliant green and procionred dyes from aqueous solutionby adsorption using selected agricultural wastes. Jurnal Teknologi. 2015; 74(11).
- Akter N, Hossain MA, Hassan MJ, Amin M, Elias M, Rahman MM, et al. Amine modified tannin gel for adsorptive removal of Brilliant Green dye. Journal of environmental chemical engineering. 2016; 4 (1):1231–41.
- Fiaz R, Hafeez M, Mahmood R. Removal of brilliant green (BG) from aqueous solution by using low cost biomass Salix alba leaves (SAL): thermodynamic and kinetic studies. Journal of Water Reuse and Desalination. 2020; 10(1):70–81.
- Garg V, Gupta R, Yadav AB, Kumar R. Dye removal from aqueous solution by adsorption on treated sawdust. Bioresource technology. 2003; 89(2):121–4. <u>https://doi.org/10.1016/s0960-8524(03)00058-0</u> PMID: 12699929
- Chowdhury S, Mishra R, Saha P, Kushwaha P. Adsorption thermodynamics, kinetics and isosteric heat of adsorption of malachite green onto chemically modified rice husk. Desalination. 2011; 265(1–3):159– 68.
- 54. Ghaedi M, Azad FN, Dashtian K, Hajati S, Goudarzi A, Soylak M. Central composite design and genetic algorithm applied for the optimization of ultrasonic-assisted removal of malachite green by ZnO Nanorod-loaded activated carbon. Spectrochimica Acta Part A: Molecular and Biomolecular Spectros-copy. 2016; 167:157–64. https://doi.org/10.1016/j.saa.2016.05.025 PMID: 27318150
- 55. Lam SS, Liew RK, Wong YM, Yek PNY, Ma NL, Lee CL, et al. Microwave-assisted pyrolysis with chemical activation, an innovative method to convert orange peel into activated carbon with improved properties as dye adsorbent. Journal of cleaner production. 2017; 162:1376–87.
- Chowdhury S, Misra R, Kushwaha P, Das P. Optimum sorption isotherm by linear and nonlinear methods for safranin onto alkali-treated rice husk. Bioremediation Journal. 2011; 15(2):77–89.
- Adebowale KO, Olu-Owolabi BI, Chigbundu EC. Removal of safranin-O from aqueous solution by adsorption onto kaolinite clay. Journal of Encapsulation and Adsorption Sciences. 2014; 4(03):89.
- Laskar N, Kumar U. Adsorption of Safranin (Cationic) dye from water by Bambusa tulda: Characterization and ANN modeling. Environmental Engineering Science. 2018; 35(12):1361–75.
- 59. Hussain AFA, Halboos MH, editors. Adsorption of safranin dye from their aqueous solutions by using CA and Nano FeO/CA. Journal of Physics: Conference Series; 2020: IOP Publishing.