# SCIENTIFIC REPORTS

Received: 04 April 2016 Accepted: 09 August 2016 Published: 30 August 2016

# **OPEN** Synergistic effects of dissolved organic carbon and inorganic nitrogen on methane uptake in forest soils without and with freezing treatment

Haohao Wu<sup>1,2</sup>, Xingkai Xu<sup>1,2</sup>, Cuntao Duan<sup>1</sup>, Tuansheng Li<sup>3</sup> & Weiguo Cheng<sup>4</sup>

There is limited knowledge about how the interaction of dissolved organic carbon (DOC) and inorganic nitrogen (N) released into the soil just after freezing can affect methane ( $CH_4$ ) uptake in forest soils. Here, we present how freezing treatment and glucose, as a DOC source, can affect the roles of NH<sub>4</sub><sup>+</sup>-N and NO<sub>3</sub><sup>--</sup>-N in inhibiting soil CH<sub>4</sub> uptake, by using soil-core incubation experiments. A long-term freezing at low temperature reduced cumulative CH<sub>4</sub> uptake in the soils sampled from two temperate forest stands without carbon (C) and N addition. The inhibition effects of N addition as NH<sub>4</sub>Cl and KNO<sub>3</sub> on the soil CH<sub>4</sub> uptake were much larger than C addition. Freezing treatment eliminated the inhibition effect of NH<sub>4</sub>Cl and KNO<sub>3</sub> addition on CH<sub>4</sub> uptake, and this response was affected by glucose addition and forest types. The addition of glucose eliminated the inhibition effect of  $NO_3^{-}$ -N on CH<sub>4</sub> uptake in the forest soils without and with freezing treatment, while the addition of NH4+-N and glucose inhibited synergistically the soil CH<sub>4</sub> uptake. The results highlight the importance of synergistic effects of DOC and N inputs on the soil CH<sub>4</sub> uptake under forest stands during soil wetting and thawing periods.

Upland soils are regarded as the only biological sink of atmospheric methane (CH<sub>4</sub>) and are responsible for 6% of the global  $CH_4$  consumption<sup>1</sup>. Increasing atmospheric N deposition tends to inhibit the consumption of  $CH_4$ in upland soils, which may partly lead to the rising atmospheric abundance of  $CH_4^{2,3}$ . Ammonium  $(NH_4^+-N)$ and nitrate (NO<sub>3</sub><sup>-</sup>-N) are regarded as main contributors of atmospheric wet N deposition, and the addition of  $NH_4^+$ -N to forest soils can suppress  $CH_4$  uptake<sup>2,4,5</sup>, while the effects of  $NO_3^-$ -N are contradictory with inhibition, no effect or stimulating effect on  $CH_4$  uptake<sup>6-9</sup>. Many researchers have reported that  $NO_3^{-1}$  has a strong inhibitory effect on  $CH_4$  oxidation in forest soils probably due to the toxicity of  $NO_2^-$  on  $CH_4$ -oxidizing bacteria or increased  $NH_4^+$  concentration following  $NO_3^-$  addition<sup>10-15</sup>. The addition of glucose as a labile C source can increase the inhibition by  $NO_3^-$ -N of the soil  $CH_4$  oxidation<sup>14</sup>. The increase in atmospheric carbon dioxide concentration can promote the growth of vegetation, probably resulting in an increased C input into the soil via litter decomposition and turnover of roots. Furthermore, almost half of dissolved organic carbon (DOC) in water extracts of organic layers under temperate forest stands exists in the form of glucose-C (Table S1), which can be normally used by soil microbes. The labile C supply and the variations of soil moisture can stimulate soil microbial activity and the turnover of  $NO_3^{-}-N$  and  $NH_4^{+}-N$  in soil, and the increased carbon dioxide concentration can have different effects on atmospheric CH<sub>4</sub> uptake in unsaturated upland soils<sup>7,16</sup>. The effect of N input on CH<sub>4</sub> uptake appears to depend on the types of added N, but the effects of N addition in combination with C sources on the soil CH<sub>4</sub> uptake are partly understood.

<sup>1</sup>State Key Laboratory of Atmospheric Boundary Layer Physics and Atmospheric Chemistry, Institute of Atmospheric Physics, Chinese Academy of Sciences, Beijing 100029, China. <sup>2</sup>Department of Atmospheric Chemistry and Environmental Science, College of Earth Science, University of Chinese Academy of Sciences, Beijing 100049, China. <sup>3</sup>College of Earth Science and Resources, Chang'an University, Xi'an 710054, China. <sup>4</sup>Faculty of Agriculture, Yamagata University, Tsuruoka 997-8555, Japan. Correspondence and requests for materials should be addressed to X.K.X. (email: xingkai\_xu@mail.iap.ac.cn)

The relationship between  $CH_4$  uptake and DOC concentration in soil is generally elusive<sup>17</sup>. On one hand, the addition of labile C sources can stimulate heterotrophic microbial processes, which cause low oxygen concentration in soil and thus inhibits  $CH_4$  uptake. *Fender et al.*<sup>14</sup> reported that the addition of glucose at a rate of 941.9 g C m<sup>-2</sup> reduced  $CH_4$  uptake by 83% and, more intriguingly, stimulated the inhibiting effect of KNO<sub>3</sub> fertilization on  $CH_4$  uptake in a temperate forest soil. On the other hand, a rapid decomposition of added glucose in the soil probably increases microbial N immobilization by consuming  $NO_3^{-}$ -N and  $NH_4^+$ , which can affect the soil  $CH_4$  uptake. *Xu & Inubushi*<sup>13</sup> reported that the 15-day oxic pre-incubation following addition of glucose at a rate of 10 mg C g<sup>-1</sup> dry soil, stimulated  $CH_4$  uptake rates in temperate volcanic forest soils and this stimulation was larger than 1-day oxic pre-incubation. Furthermore, the stimulating effect of glucose addition on the soil  $CH_4$  uptake varied with the types of forest vegetation, with the largest effect in the Pinus forest soil. Probably, this varying effect of glucose on  $CH_4$  uptake partly depends on the status of inorganic N in the soil following addition of glucose. However, to our knowledge, the roles of  $NH_4^+$  and  $NO_3^-$  in inhibiting  $CH_4$  uptake in forest soils in the presence of soil labile C pool are not known.

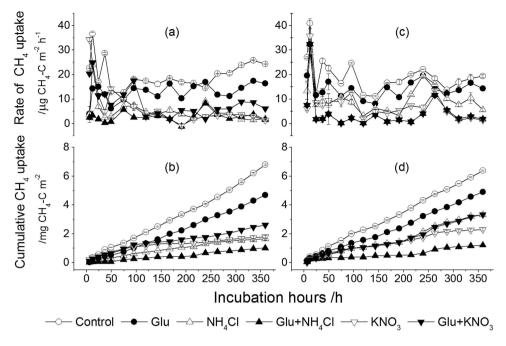
Future climatic change is likely to alter the frequency and intensity of soil drying-wetting and freezing-thawing events<sup>1,18</sup>. Wetting of dry soil and freezing-thawing processes can release labile C and N into the soil<sup>19,20</sup>, but their impacts on the soil CH<sub>4</sub> flux are unknown because they involve methanogenesis and methanotrophy<sup>21,22</sup>. Wetting dry soil can stimulate atmospheric CH<sub>4</sub> oxidation in unsaturated upland soils mainly by alleviating osmotic stress on soil methanotrophs<sup>23,24</sup>, but the effect of the availability of soil C and N upon wetting on soil CH<sub>4</sub> oxidation is not known. *Wu et al.*<sup>25</sup> reported that there was a significant increase in the CH<sub>4</sub> uptake activity following thawing and this increase generally decreased by increasing soil moisture from 32 to 55% WFPS. Methane-oxidizing bacteria are actually facultative and can utilize organic C sources, such as DOC released by wetting and freezing, other than CH<sub>4</sub><sup>17,26</sup>. *Zhu et al.*<sup>27</sup> reported that the change in soil carbon availability during thawing period could affect the dynamic of CH<sub>4</sub> flux from Antarctica soils under laboratory conditions. Furthermore, the release of labile N pools (e.g. NH<sub>4</sub><sup>+</sup>-N) into the soil at thaw<sup>20</sup> can partly limit the capacity of CH<sub>4</sub> oxidation. Due to the many variables involved, understanding the mechanisms involving soil CH<sub>4</sub> flux during wetting and thawing periods.

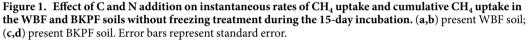
Broadleaf and Korean pine mixed forest (BKPF) is the major component of forest ecosystems in Changbai Mountains, northeastern China. In such district, the mature mixed forest lies in climax community of forest succession, with a greater soil organic matter content and lower bulk density than an adjacent secondary white birch forest (WBF)<sup>20</sup>. Due to relatively lower vegetation coverage and phototaxis property, soil available nutrients, microbial properties and hydrothermal conditions under the white birch forest stand are different from those under the mature mixed forest. The water extracts of organic layer samples collected from the WBF stand contained relatively higher DOC and microbial degradable C pools (e.g. glucose-C) than those from the BKPF stand (Table S1 and Figure S1). The differences in properties of organic layers and mineral soils under the two forest stands may influence the responses of soil CH<sub>4</sub> uptake to the addition of glucose and nitrogen as NH<sub>4</sub>Cl or KNO<sub>3</sub>. Furthermore, whether the increase in DOC input from autumn freshly fallen leaves and in combination with increased N deposition can affect the CH<sub>4</sub> uptake by forest soils during soil wetting and thawing periods has been unknown so far<sup>22</sup>. We hypothesized that freezing treatment and glucose, as a DOC source, can affect the roles of  $NH_4^+$  and  $NO_3^-$  in inhibiting soil CH<sub>4</sub> uptake. For this purpose, a series of laboratory incubation experiments were done to study (1) the single and interactive effect of C and N addition on CH4 uptake in WBF and KBPF soils without and with freezing treatment; (2) the main driving mechanisms of  $CH_4$  uptake during soil wetting and thawing periods by considering the variations of soil properties such as labile C and N pools. The results improve our understanding of how DOC input from forest organic layers and in combination with N deposition can affect the soil CH<sub>4</sub> uptake under forest stands during soil wetting and thawing periods.

## Results

**Changes in soil properties.** The addition of glucose alone reduced  $NH_4^+$ -N and  $NO_3^-$ -N concentrations in the WBF and BKPF soils without and with freezing treatment (P < 0.0001) (Tables S2–S4). The decrease in  $NH_4^+$ -N concentration induced by glucose addition after freezing was higher compared to the unfrozen soils (P < 0.05) (Tables S2 and S3). The addition of glucose alone increased MBC (P < 0.0001) and MBC:MBN ratios (P < 0.001) of the two forest soils without and with freezing treatment (Tables S2–S4).  $NH_4^+$ -N concentration in the two forest soils treated with KNO<sub>3</sub> alone significantly increased without and with freezing treatment, especially in the WBF soil (P < 0.05) (Tables S2–S4). However, this increase did not occur in the KNO<sub>3</sub>+glucose treatment (Tables S2 and S3). The addition of  $NH_4Cl$  and  $KNO_3$  alone significantly decreased soil pH without and with freezing treatment (P < 0.0001), and this decrease varied with vegetation types (P < 0.01) (Tables S2–S4). Freezing treatment caused a release of DOC and DON (P < 0.0001) and a significant decrease in MBC and the MBC:MBN ratios of the two forest soils in all treatments (P < 0.001) (Tables S2–S4). The increased release of DOC induced by freezing varied with vegetation types (P < 0.001) (Tables S2–S4).

**Soil CH**<sub>4</sub> **uptake without and with freezing.** Without freezing treatment, an increased uptake of CH<sub>4</sub> was observed immediately after wetting the WBF and BKPF soils in the control (Fig. 1a,c). The rate of CH<sub>4</sub> uptake in the control was almost the highest among all treatments during the 15-day incubation, thus resulting in significantly higher cumulative CH<sub>4</sub> uptake than that from soils treated with Glu, NH<sub>4</sub>Cl or KNO<sub>3</sub> alone at the end of incubation (P < 0.0001) (Fig. 1b,d and Table 1). Compared to the unfrozen soils (Fig. 1a,c), the peak of CH<sub>4</sub> uptake was delayed and occurred generally within 95 h to 143 h after the beginning of thaw (Fig. 2a,c). The cumulative CH<sub>4</sub> uptake in the control after freezing was significantly higher than that from soils treated with Glu, NH<sub>4</sub>Cl or KNO<sub>3</sub> alone during the 15-day incubation (P < 0.0001) (Fig. 2b,d and Table 1). Without and with freezing treatment, the cumulative CH<sub>4</sub> uptake in the two forest soils treated with glucose alone and in combination





.....

with N sources could be ranked as  $Glu < Glu+KNO_3 < Glu+NH_4Cl$  (P < 0.05), except no difference between Glu and  $Glu+KNO_3$  in the WBF soil after freezing (Figs 1 and 2).

ANOVA analysis showed that the cumulative  $CH_4$  uptake without freezing was affected by the addition of NH<sub>4</sub>Cl or KNO<sub>3</sub> and glucose singularly and interactively (P < 0.0001), and the interaction effect varied with forest types and incubation time (P < 0.001) (Table 1). However, only the addition of KNO<sub>3</sub> and glucose interactively affected the cumulative CH<sub>4</sub> uptake with freezing (P < 0.001) (Table 1).

The average rates of CH<sub>4</sub> uptake in the control without and with freezing were higher than those in the soils treated with C and N alone and in combination during the 15-day incubation, and the lowest rate occurred in the Glu+NH<sub>4</sub>Cl treatment (P < 0.05) (Fig. 3 and Table 2). The addition of NH<sub>4</sub>Cl, KNO<sub>3</sub>, and KNO<sub>3</sub> plus glucose without freezing resulted in a smaller average rate of CH<sub>4</sub> uptake in the WBF soil than in the BKPF soil during the 15-day incubation (P < 0.05) (Fig. 3). However, only the Glu+NH<sub>4</sub>Cl treatment had a lower rate of CH<sub>4</sub> uptake in the WBF soil than that in the BKPF soil with freezing (P < 0.05) (Fig. 3). ANOVA analysis showed that the inhibition of NH<sub>4</sub>Cl or KNO<sub>3</sub> on the average rates of CH<sub>4</sub> uptake varied with vegetation types (P < 0.05) and freezing treatment (P < 0.05) (Table 2).

**The inhibition by N addition of CH**<sub>4</sub> **uptake in forest soils.** The effects of glucose addition and freezing on the inhibition by N addition of CH<sub>4</sub> uptake in forest soils were shown in Table 3. In the absence of freezing and glucose supply, the relative inhibition of CH<sub>4</sub> uptake induced by NH<sub>4</sub>Cl and KNO<sub>3</sub> was both significantly greater in the WBF soil than that in the BKPF soil (P < 0.05). Without glucose, the freezing treatment significantly decreased the absolute and relative inhibition of N addition as NH<sub>4</sub>Cl and KNO<sub>3</sub> on CH<sub>4</sub> uptake in the WBF soil, while it significantly increased the absolute and relative inhibition by NH<sub>4</sub>Cl addition of the CH<sub>4</sub> uptake in the BKPF soil (P < 0.05).

Glucose addition significantly increased the relative inhibition by NH<sub>4</sub>Cl of the CH<sub>4</sub> uptake in the WBF soil with freezing and in the BKPF soil without freezing (P < 0.05) (Table 3). However, glucose addition significantly reduced the absolute and relative inhibition of KNO<sub>3</sub> addition on CH<sub>4</sub> uptake in the two forest soils with and without freezing (P < 0.001) (Tables 3 and 4). ANOVA analysis showed that freezing treatment and in combination with glucose and vegetation type could significantly affect the inhibition of NH<sub>4</sub>Cl addition on the CH<sub>4</sub> uptake (P < 0.05), and that the inhibition of KNO<sub>3</sub> addition on the CH<sub>4</sub> uptake was influenced by freezing and glucose supply (P < 0.001) (Table 4).

**Relationships between CH<sub>4</sub> uptake and soil properties.** The average rates of CH<sub>4</sub> uptake in forest soils without and with freezing treatment were both positively correlated with soil pH (P<0.001) and negatively correlated with the concentrations of soil NH<sub>4</sub><sup>+</sup>-N and DON (P<0.001) (Table 5). According to the results of stepwise regression analysis, 68% of the variability in the soil CH<sub>4</sub> uptake without freezing could be attributed to the soil pH and NH<sub>4</sub><sup>+</sup>-N, with the predominant influence of soil pH (Table 6). Meanwhile, 59% of the variability in the soil CH<sub>4</sub> uptake with freezing treatment could be explained by the soil NH<sub>4</sub><sup>+</sup>N and NO<sub>3</sub><sup>-</sup>-N, and affected by the NH<sub>4</sub><sup>+</sup>-N mostly (Table 6). Together with soil samples without and with freezing, 66% of the variability in the soil CH<sub>4</sub> uptake could be explained by the soil pH, inorganic N and DOC, and affected by the NH<sub>4</sub><sup>+</sup>-N mostly (Table 6).

		Without f	reezing		With freezing treatment				
	Instant rate of	CH4 uptake	Cumulative	CH4 uptake	Instant rate of	f CH <sub>4</sub> uptake	Cumulative	CH <sub>4</sub> uptake	
Source of variation	F	P	F	P	F	P	F	P	
			N addition a	s NH₄Cl					
Between subjects									
Vegetation (VT)	25.6057	<0.0001	131.9421	< 0.0001	1.1760	0.2791	22.6105	< 0.0001	
N addition (N)	854.5645	<0.0001	2593.9278	< 0.0001	1135.8867	<0.0001	3259.7753	< 0.0001	
Glucose (Glu)	203.3456	<0.0001	794.7950	<0.0001	85.5486	<0.0001	407.3302	<0.0001	
$VT \times N$	11.9317	0.0006	37.7731	<0.0001	9.1389	0.0027	51.7877	<0.0001	
VT × Glu	0.3894	0.5331	1.7737	0.1840	13.1318	0.0003	0.8708	0.3515	
N × Glu	0.0014	0.9700	16.5619	0.0001	0.0266	0.8705	1.4048	0.2369	
$VT \times N \times Glu$	14.1612	0.0002	13.4113	0.0003	4.7737	0.0297	0.9401	0.3331	
Within subjects									
Time	19.5037	<0.0001	409.9718	<0.0001	42.9859	<0.0001	980.3995	<0.0001	
VT×Time	6.0406	<0.0001	2.1556	0.0055	4.0185	<0.0001	1.4840	0.0992	
N×Time	3.8538	<0.0001	16.8531	<0.0001	12.3712	<0.0001	53.7808	<0.0001	
Glu×Time	1.7539	0.0336	2.3868	0.0018	1.6525	0.0512	5.0300	<0.0001	
$VT \times N \times Time$	2.3838	0.0019	0.9377	0.5298	3.6115	<0.0001	1.8241	0.0249	
VT  imes Glu  imes Time	3.6389	<0.0001	2.1149	0.0067	2.1210	0.0065	1.1560	0.3006	
N  imes Glu  imes Time	2.0626	0.0086	2.5856	0.0007	4.1774	<0.0001	3.5101	<0.0001	
$VT \times N \times Glu \times Time$	1.3216	0.1775	0.9395	0.5278	5.7621	<0.0001	1.7457	0.0348	
			N addition a	as KNO <sub>3</sub>					
Between subjects									
Vegetation (VT)	9.5220	0.0022	13.3708	0.0003	42.3268	<0.0001	50.1128	<0.0001	
N addition (N)	925.2851	<0.0001	2078.8460	<0.0001	593.6729	<0.0001	1643.8224	<0.0001	
Glucose (Glu)	19.0963	<0.0001	193.7710	<0.0001	7.9758	0.0051	43.4155	<0.0001	
$VT \times N$	1.0196	0.3135	20.6737	<0.0001	14.8393	0.0001	22.7903	<0.0001	
VT×Glu	4.9481	0.0269	7.8644	0.0054	1.1113	0.2927	9.8260	0.0019	
N  imes Glu	207.7277	<0.0001	502.5143	<0.0001	70.4478	<0.0001	195.2618	<0.0001	
$VT \times N \times Glu$	3.5933	0.0590	16.4131	0.0001	0.0179	0.8937	2.4216	0.1208	
Within subjects									
Time	46.5006	<0.0001	987.0992	<0.0001	71.7421	<0.0001	913.3450	<0.0001	
VT×Time	9.9547	<0.0001	2.8894	0.0001	11.4625	<0.0001	3.0983	<0.0001	
N × Time	15.3293	<0.0001	52.4922	<0.0001	5.1846	<0.0001	24.4530	<0.0001	
Glu × Time	6.0925	<0.0001	1.4020	0.1340	4.1249	<0.0001	1.4420	0.1159	
$VT \times N \times Time$	15.0997	<0.0001	9.8376	<0.0001	5.0622	<0.0001	0.6866	0.8158	
VT  imes Glu  imes Time	3.2209	<0.0001	1.1845	0.2761	2.3494	0.0022	0.7016	0.8009	
N  imes Glu  imes Time	4.8770	<0.0001	7.6688	<0.0001	2.6330	0.0005	6.4203	<0.0001	
$VT \times N \times Glu \times Time$	1.7445	0.0350	0.2953	0.9975	2.7847	0.0002	0.7562	0.7427	

Table 1. Summary of ANOVA with repeated measures for the instant rates of  $CH_4$  uptake and cumulative  $CH_4$  uptake in forest soils without and with freezing treatment.

## Discussion

**Effect of freezing on CH<sub>4</sub> uptake in forest soils.** Thawing of frozen soils in the absence of C and N addition significantly increased the release of  $NH_4^+$ -N and DOC into the soils, compared to the unfrozen soils (Tables S2 and S3), and this may reduce CH<sub>4</sub> uptake in forest soils during thawing period<sup>5,26</sup>. The release of labile C and N into the soil after thaw resulted in a significant decrease in cumulative CH<sub>4</sub> uptake in the control throughout the experimental period (Figs 1 and 2). Previous laboratory studies involving freeze-thaw effects have seldomly compared with non-freezing experiment, and in field measurement, the effect of freeze-thaw cycle on soil CH<sub>4</sub> uptake in a desert grassland<sup>28</sup> and a northern hardwood forest<sup>29</sup>. However, *Borken et al.*<sup>30</sup> observed that soil freezing by snow removal increased the rates of soil CH<sub>4</sub> uptake in a temperate forest. And no changes in CH<sub>4</sub> uptake due to freeze-thaw cycle also occurred in a grassland<sup>31</sup>. *Gao et al.*<sup>32</sup> reported that serious freezing at -15 °C had a similar rate compared to the non-freezing treatment. Probably, freezing conditions and the resultant changes in soil gas diffusion and the release of labile C and N pools lead to different changes in CH<sub>4</sub> uptake during thawing period.

Soil moisture is the primary control on the soil  $CH_4$  oxidation<sup>33</sup>, either by affecting gas diffusion or because low soil moisture can cause osmotic stress on soil methanotrophs<sup>23,34</sup>. Soil cores were incubated at 55% WFPS, and thus gas diffusion was not a possible limitation factor of the  $CH_4$  uptake because gas diffusion through soil is restricted over 65% WFPS<sup>35</sup>. Probably, soil  $CH_4$  uptake after freezing under the experimental conditions was

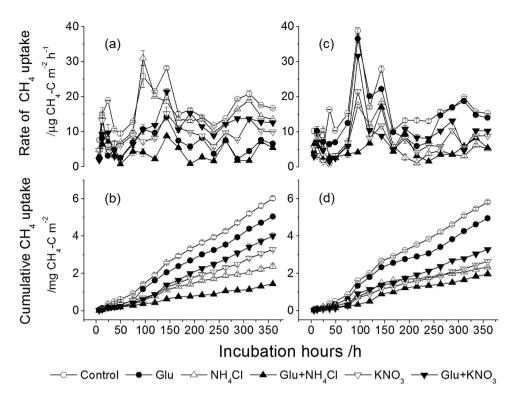


Figure 2. Effect of C and N addition on instantaneous rates of  $CH_4$  uptake and cumulative  $CH_4$  uptake in the WBF and BKPF soils with freezing treatment during the 15-day incubation. (a,b) present WBF soil; (c,d) present BKPF soil. Error bars represent standard error.

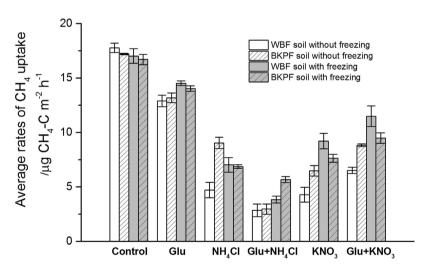


Figure 3. Effect of C and N addition on the average rates of  $CH_4$  uptake in the WBF and BKPF soils without and with freezing treatment during the 15-day incubation. Error bars represent standard error.

attributed to the changes in soil inorganic N and DOC pools and the properties of methanotrophs. At the onset of thaw, the soil CH<sub>4</sub> uptake across all the treatments was smaller but it increased significantly four to six days after thaw (Fig. 2a,c), and the pulse CH<sub>4</sub> uptake after frost was similar to that within 12 h after wetting (Fig. 1a,c). This indicated that there was a delay of several days for a complete reactivation of inactive methanotrophs in the soils frozen at -18 °C for 50 days. This delay probably resulted from the initial pulse release of dissolved organic C (e.g. organic acids) shortly after thaw to be used by facultative methanotrophs<sup>26</sup>. This can limit the recovery of methanotrophs which are capable of oxidizing atmospheric CH<sub>4</sub>. Together with the release of DOC and inorganic N into the soil after severe freezing (Tables S2 and S3) and the changes in soil microbial community as indicated by the microbial biomass C-to-N ratio (Table S4), further research should characterize quality and quantity of DOC and labile N released into the soil at thaw and their relationships to the soil CH<sub>4</sub> uptake as well as the functions of soil methanotrophs<sup>22</sup>.

	N additior	as NH <sub>4</sub> Cl	N additio	n as KNO <sub>3</sub>
Source of variation	F	Р	F	Р
Vegetation (VT)	0.1635	0.6887	0.0310	0.8614
N addition (N)	308.3975	<0.0001	351.7982	<0.0001
Glucose (Glu)	93.8161	<0.0001	5.4137	0.0265
Freezing (F)	4.9874	0.0327	22.6748	0.0000
$VT \times N$	8.8145	0.0056	6.7330	0.0142
VT  imes Glu	0.7289	0.3996	1.3515	0.2536
$VT \times F$	0.1035	0.7498	9.6477	0.0040
N  imes Glu	0.3905	0.5365	81.3734	<0.0001
N  imes F	4.2068	0.0485	18.9566	0.0001
$\mathrm{Glu}\times\mathrm{F}$	10.6002	0.0027	0.7855	0.3821
$VT \times N \times Glu$	1.5187	0.2268	1.1876	0.2840
$VT \times N \times F$	0.1430	0.7078	10.0363	0.0034
$VT \times Glu \times F$	0.7038	0.4077	0.3218	0.5745
$N \times Glu \times F$	0.0151	0.9030	7.3364	0.0108
$VT \times N \times Glu \times F$	5.5761	0.0245	1.1581	0.2899

Table 2. Summary of ANOVA with repeated measures for the average rates of  $CH_4$  uptake in forest soils during the 15-day incubation without and with freezing treatment.

**Effect of glucose addition alone on CH**<sub>4</sub> uptake in forest soils. Glucose is a labile C source that can be easily used by methanogenic bacteria to produce CH<sub>4</sub> and by other microorganisms producing CO<sub>2</sub> and consuming  $O_2$ , probably resulting in the inhibition of  $CH_4$  uptake. Alternatively, the addition of glucose can increase N immobilization in forest soils and reduce NO3--N concentration following preincubation, and this can stimulate  $CH_4$  oxidation<sup>13</sup>. Methanotrphs can utilize multi-C compounds (e.g. glucose) as sole sources of C and energy in the absence of methane<sup>17,36</sup>. The changes induced by the labile C can thus affect the soil  $CH_4$  uptake. Without preincubation, the addition of glucose alone at a rate of  $6.4 \,\mathrm{g\,C}\,\mathrm{m^{-2}}$  significantly inhibited the cumulative atmospheric CH<sub>4</sub> uptake by 23% to 31% in the two forest soils during wetting period (Fig. 1c,d), and this agrees with what reported by Fender at al.<sup>14</sup>. The reason of the inhibition effect of glucose may be the increased microbial respiration<sup>37</sup> and microbial biomass C caused by the added glucose (Table S2) with the decrease in O<sub>2</sub> concentration of soil thus limiting CH<sub>4</sub> oxidation<sup>14</sup>. Additionally, the resultant anaerobic environment in soil micro-sites can stimulate denitrification process with consumption of  $NO_3^-$ ; the intermediate product  $NO_2^-$  of denitrification is toxic to CH<sub>4</sub>-oxidizing bacteria<sup>10,38</sup>. During thawing period, the inhibition of glucose on the cumulative soil CH<sub>4</sub> uptake was reduced by approximately 15% (Fig. 2c,d), which was significantly smaller than that during wetting period (Fig. 1c,d). Probably, the more release of DOC into the soil after frost (Tables S2 and S3) reduced the inhibition of CH<sub>4</sub> uptake in the two forest soils by glucose addition, particularly in the WBF soil (Fig. 3). This is probably related to the relatively high labile C pools (e.g. glucose-C and protein-like substance) in water extracts of forest organic layers under WBF stand than under BKPF stand (Table S1 and Figure S1). According to Wieczorek et al.<sup>26</sup>, the facultative methanotrophs can change their substrate utilization in the presence of different organic C sources and this may reduce  $CH_4$  oxidation. The addition of glucose alone caused a significant increase of MBC:MBN ratios in the forest soils with and without freezing (Tables S2-S4), indicating a shift in microbial community towards more fungal (average C:N ratio, 5-15) than bacteria (average C:N ratio, 3-6)<sup>39,40</sup>. This shift can support the inhibition of glucose on soil CH<sub>4</sub> uptake, because a negative correlation between methane-oxidizing bacteria and fungal biomass was observed in forest soils<sup>41</sup>.

Contrary to the inhibition of CH<sub>4</sub> uptake by glucose addition, Xu & Inubushi<sup>13</sup> reported that adding glucose at a rate of  $10 \text{ mg g}^{-1}$  dry soil significantly stimulated CH<sub>4</sub> uptake in the volcanic forest soils, especially in the *Pinus* forest soil, with a relatively low efficiency in utilizing organic C, and the stimulation depended on forest types and preincubation conditions. Sullivan et al.<sup>17</sup> reported a positive correlation (r = 0.76, P < 0.01) between soil DOC concentration and CH<sub>4</sub> oxidation rate and indicated DOC as an important regulator of CH<sub>4</sub> oxidation in arid soils. Incubation experiment conducted by *Hilger et al.*<sup>42</sup> showed that glucose concentration was positively correlated to  $CH_4$  uptake in a landfill cover soil (r = 0.94, P < 0.05). In our study, there was a significant negative correlation between the average CH<sub>4</sub> uptake without freezing treatment and the soil DOC concentration but not after frost (Table 5). The results by Burke et al.<sup>41</sup> indicated that methane-oxidizing bacteria are more likely found in areas with low C and nutrient cycling rates. According to the results of ANOVA analysis, the inhibition of glucose addition on the soil CH<sub>4</sub> uptake was affected by forest vegetation, freezing, and types of added N (Tables 1 and 2). From the results of this study and previous studies, it can be thus reasonably concluded that the addition of external C such as glucose has variable impacts on  $CH_4$  uptake in unsaturated soils, which depends on microbial C utilization, soil N availability, and hydrothermal conditions. More interesting, the content of glucose in water extracts of organic layers under the WBF and BKPF stands ranged from 2.4 to 4.3 g C m<sup>-2</sup>, accounting for 40% to 55% of DOC pool, and forest organic layers under the WBF stand contained the relatively higher DOC and microbial degradable C pools (Table S4 and Figure S1). Due to the fact that the responses of  $CH_4$  uptake to the addition of glucose varied with forest vegetation and freezing treatment (Tables 1 and 2), the experimental results

		Avera	age rates of $CH_4$	uptake in the WB	F soil	Average rates of CH4 uptake in the BKPF soil					
		NH <sub>4</sub> Cl-induced		KNO <sub>3</sub> -induced		NH <sub>4</sub> Cl-i	nduced	KNO <sub>3</sub> -induced			
Freezing treatment	Glucose addition	Absolute inhibition /µg CH <sub>4</sub> -C m <sup>-2</sup> h <sup>-1</sup>	Relative inhibition /%	Absolute inhibition /μg CH <sub>4</sub> -C m <sup>-2</sup> h <sup>-1</sup>	Relative inhibition/%	Absolute inhibition/µg CH <sub>4</sub> -C m <sup>-2</sup> h <sup>-1</sup>	Relative inhibition/%	Absolute inhibition/µg CH <sub>4</sub> -C m <sup>-2</sup> h <sup>-1</sup>	Relative inhibition/%		
Without freezing	No glucose	$12.31 \pm 0.71$	$72.33 \pm 4.17$	$12.74 \pm 0.68$	$74.86\pm3.97$	$7.98\pm0.51$	$46.85\pm2.99$	$10.54 \pm 0.48$	$61.94 \pm 2.83$		
without freezing	With glucose	$11.68\pm0.57$	$80.42\pm3.94$	$8.01\pm0.30$	$55.12 \pm 2.04$	$11.57\pm0.47$	$79.61\pm3.22$	$5.69\pm0.12$	$39.14 \pm 0.85$		
With freezing	No glucose	$9.99 \pm 0.66$	$58.70\pm3.89$	$7.82\pm0.71$	$45.91 \pm 4.17$	$10.15 \pm 0.17$	$59.65 \pm 1.00$	$9.39\pm0.36$	$55.14 \pm 2.14$		
with neezing	With glucose	$10.69\pm0.32$	$73.55 \pm 2.22$	$3.03 \pm 0.96$	$20.88 \pm 6.58$	$8.85\pm0.29$	$60.94\pm1.97$	$5.04\pm0.48$	$34.69 \pm 3.30$		

Table 3. Effect of freezing treatment and glucose addition on the inhibition of  $CH_4$  uptake by N sources in forest soils.

	Absolute inhibition of	CH <sub>4</sub> uptake	Relative inhibition of C	CH <sub>4</sub> uptake						
Source of variation	F	Р	F	Р						
Inhibition by NH <sub>4</sub> Cl										
Vegetation (VT)	19.1043	0.0005	19.8412	0.0004						
Freezing (F)	7.5539	0.0143	9.7710	0.0065						
Glucose (Glu)	2.8238	0.1123	45.1664	0.0000						
VT  imes F	3.9472	0.0643	3.2305	0.0912						
VT  imes Glu	2.5206	0.1319	1.7649	0.2027						
$F \times \operatorname{Glu}$	6.4774	0.0216	8.8183	0.0090						
$VT \times F \times Glu$	19.6606	0.0004	19.8412	0.0004						
Inhibition by KNO <sub>3</sub>										
Vegetation (VT)	0.3459	0.5647	0.3836	0.5444						
Freezing (F)	53.2792	<0.0001	52.8470	< 0.0001						
Glucose (Glu)	136.2188	<0.0001	73.5058	< 0.0001						
$VT \times F$	25.4935	0.0001	26.1945	0.0001						
VT  imes Glu	0.0385	0.8469	0.0266	0.8726						
F  imes Glu	0.0831	0.7768	0.1286	0.7246						
$VT \times F \times Glu$	0.1217	0.7318	0.6642	0.4271						

Table 4. Summary of ANOVA with repeated measures for the inhibition by N sources on the average rates of  $CH_4$  uptake in forest soils without and with freezing treatment.

		WPFS	pН	NO <sub>3</sub> <sup>-</sup> -N	NH4 <sup>+</sup> -N	DON	DOC	MBN	мвс	MBC:MBN ratio
CH <sub>4</sub> uptake without freezing	r	-0.165	0.801	-0.306	-0.559	-0.662	-0.401	-0.070	-0.125	-0.098
	P value	0.337	< 0.001	0.069	< 0.001	< 0.001	0.015	0.678	0.466	0.532
CH <sub>4</sub> uptake with freezing	r	0.013	0.603	-0.119	-0.679	-0.606	0.043	0.102	0.025	0.359
	P value	0.940	< 0.001	0.491	< 0.001	< 0.001	0.802	0.516	0.882	0.036

.....

Table 5. Pearson correlation (r) coefficients between soil properties and the average rates of CH<sub>4</sub> uptake in forest soils without and with freezing treatment. P value, the significant levels of Pearson correlation coefficients.

indicated that the quality and quantity of DOC released from forest organic layers into the soil can partly affect the soil CH<sub>4</sub> uptake under the two forest stands, especially during spring thawing after winter sincere freezing.

**Effect of N sources and in combination with glucose on CH<sub>4</sub> uptake in forest soils.** There was a significant negative relationship between the average rates of CH<sub>4</sub> uptake and NH<sub>4</sub><sup>+</sup>-N concentration of forest soils in all treatments without and with freezing (Table 5), which showed the inhibition effect of NH<sub>4</sub><sup>+</sup>-N on the CH<sub>4</sub> uptake in unsaturated forest soils<sup>22,43</sup>. The competition of NH<sub>4</sub><sup>+</sup> with CH<sub>4</sub> on the enzyme responsible for both oxidations is considered the reason for the inhibition effect of NH<sub>4</sub><sup>+</sup> on the soil CH<sub>4</sub> uptake<sup>4,5,12</sup>; in addition the concomitant conversion of NH<sub>4</sub><sup>+</sup> to NH<sub>2</sub>OH and NO<sub>2</sub><sup>-</sup> would be toxic to CH<sub>4</sub>-oxidizing bacteria<sup>10,38</sup>. This hypothesis was confirmed by Xu & Invbushi<sup>12</sup>, who reported that the use of nitrification inhibitor eliminated the inhibition of CH<sub>4</sub> uptake by NH<sub>4</sub><sup>+</sup>. Hence, the turnover rather than concentration of NH<sub>4</sub><sup>+</sup> can influence the soil CH<sub>4</sub> uptake under the experimental conditions<sup>3,5</sup>. The smallest average rates of CH<sub>4</sub> uptake occurred in the NH<sub>4</sub>Cl+Glu treatment, with significant differences in the BKPF soil without and with freezing and in the WBF soil without freezing compared to the NH<sub>4</sub>Cl treatment and Glu addition alone (Fig. 3). The results indicated that

Explanatory variable	Coefficient	Standard error	Relative contribution	$R^2$ value	F	P value	Standard errors of estimates			
	Average rates	of CH <sub>4</sub> uptake in t	he soils without freezing	$(n=36)/\mu g$	CH <sub>4</sub> -C n	$h^{-2} h^{-1}$				
Constant	-77.367	14.353		0.68	35.324	< 0.0001	2.9964			
pН	15.811	2.556	75.22%							
$NH_4^+$ -N /g N m <sup>-2</sup>	-0.723	0.354	24.78%							
	Average rates of CH <sub>4</sub> uptake in the soils with freezing ( $n = 36$ ) /µg CH <sub>4</sub> -C m <sup>-2</sup> h <sup>-1</sup>									
Constant	16.486	1.110		0.59	23.326	< 0.0001	2.8817			
$NH_4^+$ -N /g N m <sup>-2</sup>	-2.596	0.385	67.08%							
$NO_3^{-}-N/gNm^{-2}$	-1.097	0.331	32.92%							
Average rates of CH4 up	take in the soil	s with and without	freezing ( $n = 72$ ) /µg CH	$_4$ -C m <sup>-2</sup> h <sup>-</sup>	1					
Constant	-40.639	14.777		0.66	32.036	< 0.0001	2.8876			
pН	9.169	2.460	29.88%							
$NH_4^+$ -N /g N m <sup>-2</sup>	-1.624	0.341	34.98%							
NO <sub>3</sub> <sup>-</sup> -N /g N m <sup>-2</sup>	-0.671	0.305	15.98%							
$DOC/gCm^{-2}$	0.604	0.175	19.16%							

Table 6. Summary of the stepwise regression analysis for the relationships between the average rates of CH<sub>4</sub> uptake against soil properties without and with freezing treatment.  $R^2$  value, determination coefficient of regression.

there was a positive synergistic inhibition effect of  $NH_4^+$ -N and glucose on the soil  $CH_4$  uptake under the experimental conditions and this synergistic effect varied with forest types and freezing treatment.

Besides NH<sub>4</sub><sup>+</sup>-N effect on the soil CH<sub>4</sub> uptake, the significant decrease in soil pH upon N addition without and with freezing (Tables S2–S4) also decreased the soil CH<sub>4</sub> uptake (Table 5)<sup>12,44,45</sup>. Negative effects of soil acidification on soil physical and chemical properties (e.g. Al<sup>3+</sup>) and microbial activities have the potential to reduce CH<sub>4</sub> uptake in forest soils<sup>44,45</sup>.

Nitrate-N addition has a strong inhibition of CH<sub>4</sub> uptake in unsaturated forest soil<sup>8,10-13,15</sup>, and this inhibition varied with vegetation types and freezing treatment (Table 2). An increase in  $NH_4^+$ -N concentration without and with freezing treatment occurred in the two forest soils treated with KNO<sub>3</sub> alone, particularly in the WBF soil (Tables S2-S4), and there was a significant negative correlation between the average rate of CH<sub>4</sub> uptake and the  $NH_4^+$ -N concentration in the soils not treated with  $NH_4Cl$  (r = -0.533, P < 0.01). The inhibition of  $CH_4$  uptake by KNO<sub>3</sub> under the experimental conditions can thus depend on the accumulation of NH<sub>4</sub><sup>+</sup> upon KNO<sub>3</sub> addition, which was also proposed by Fender et al.14. However, Wang & Ineson<sup>10</sup> did not show a significant change in  $NH_4^+$ -N concentration after KNO<sub>3</sub> addition and indicated that  $NO_3^-$ -N rather than  $NH_4^+$ -N nor K<sup>+</sup> was the major responsible inhibitory component for the soil  $CH_4$  uptake. Based on significant differences in concentrations of NH<sub>4</sub><sup>+</sup>-N and MBN between the Glu treatment and Glu+KNO<sub>3</sub> treatment after freezing (Table S3), it was assumed that NO3<sup>-</sup>-N was partly converted into NH4<sup>+</sup>-N in forest soil at thaw and that microbial N immobilization preferable for  $NH_4^+$ -N was increased by adding C. Due to the rapid immobilization of  $NO_3^-$  in forest soils<sup>46</sup>, DON concentration of the two forest soils with and without freezing was significantly increased by adding NO<sub>3</sub><sup>-</sup>-N, and the response varied with vegetation type and freezing treatment (Tables S2–S4). Probably, the conversion of  $NO_3^-$  to  $NH_4^+$  under the experimental conditions resulted from the mineralization of increased DON in the soils. Further research should characterize the accumulation of NH<sub>4</sub>+-N upon NO<sub>3</sub><sup>-</sup>-N addition in forest soils with varying microbial C availability using label <sup>15</sup>N technology and its relationship to the soil CH<sub>4</sub> uptake.

Under the experimental conditions, the addition of glucose significantly weakened the inhibition effect of  $NO_3^{-}-N$  on the CH<sub>4</sub> uptake in the two forest soils without and with freezing (Tables 3 and 4). Simultaneously, compared to the KNO<sub>3</sub> treatment,  $NH_4^+$ -N concentration in the KNO<sub>3</sub>+Glu treatment significantly decreased in the two forest soils with freezing, particularly in the WBF soil (Table S3), and this N concentration was also decreased in the WBF soil without freezing (Table S2). It further supports that  $NH_4^+$ -N rather than  $NO_3^-$ -N itself in the presence of labile C has an inhibition effect on the soil CH<sub>4</sub> uptake. However, Fender et al.<sup>14</sup> demonstrated that glucose addition aggravated the inhibition effect of KNO<sub>3</sub> on the CH<sub>4</sub> uptake in forest soil from 86% to 99.4%. In spite of that, they found an increase in  $NH_4^+$ -N content compared to the soil treated with  $NO_3^-$ -N alone, which is still coincided with the suppose: indirect inhibition effect of NO<sub>3</sub><sup>-</sup>-N on the soil CH<sub>4</sub> uptake by transforming to  $NH_4^+$ -N. Besides the  $NH_4^+$ -N, the decline of soil pH upon  $NO_3^-$  addition can be considered a cause of  $NO_3^{-}$ -N inhibition<sup>44</sup>. However, *Mochizuki et al.*<sup>15</sup> reported that the decrease in soil pH accompanied by the addition of nitrate was not responsible for the strong inhibition by nitrate of CH<sub>4</sub> oxidation. Recently, in situ atmospheric CH<sub>4</sub> oxidation rates in temperate forests from South Korea were reported to be positively correlated with soil nitrate concentration, and the short-term experimental addition of NO<sub>3</sub><sup>-</sup>-N significantly stimulated the atmospheric  $CH_4$  oxidation but inhibited oxidation under high  $CH_4$  concentration<sup>8</sup>. Hence, the mechanisms involving the inhibition by NO<sub>3</sub><sup>-</sup>-N of CH<sub>4</sub> uptake need to be further studied.

Toxicity of  $NO_3^-$  and  $NO_2^-$  produced via  $NO_3^-$  reduction in anaerobic microsites to  $CH_4$ -oxidizing bacteria has been reported to explain the inhibitory effect of  $NO_3^-$  addition on the soil  $CH_4$  uptake<sup>10,38</sup>. However, it is unreal under experimental conditions because the relatively low soil moisture (55% WFPS) and headspace aeration within PVC cylinder at each gas sampling could ensure enough oxygen concentration available in the soil. Certainly, it may create anaerobic environment temporarily after the addition of glucose, because high microbial respiration stimulated by glucose<sup>37</sup> consumes large amounts of oxygen. But, according to simultaneous measurements of nitrous oxide emission from the same experiments, the  $KNO_3$ +Glu treatment had no more cumulative nitrous oxide emission from the two forest soils than the Glu treatment during wetting and thawing periods (data not shown), which indicated the presence of glucose cannot stimulate the denitrification of nitrate-N in the soils at 55% WFPS.

In this study, freezing significantly influenced the inhibition effect of  $NO_3^{-}-N$  and  $NH_4^{+}-N$  on the soil  $CH_4$  uptake (Table 2). This was linked to the increase in soil DOC concentration and a decrease in the microbial biomass C-to-N ratios caused by freezing (Tables S2–S4). The decrease in the ratios shows that freezing can cause a shift in microbial composition towards more bacteria with lower microbial C-to-N ratios than fungi<sup>47</sup>. Together with the changes in soil labile C and N pools and pH, it can be concluded that changes in  $CH_4$  uptake in unsaturated forest soils without and with freezing treatment depended on soil pH, labile C, turnover of N, and microbial community structure.

The quantity and quality of DOC in water extracts of forest organic layers varies with forest types (Table S1 and Figure S1), and the input of the labile C into the soil may affect atmospheric  $CH_4$  uptake under forest stands. Our current studies showed that the addition of forest leaf litters at a dose of  $0.0125 \text{ g} \text{ g}^{-1}$  oven-dried soil resulted in an increase of glucose-C concentration in the soil from 36.3 to 66.6 µg glucose-C g<sup>-1</sup> oven-dried soil during freezing-thawing periods (data not shown). Together with the changes in inorganic N and labile C pools released into the soil after freezing as well as glucose-C as one important form of DOC in water-extracts of forest organic layers (Table S1), the varying synergistic effect of inorganic N and glucose-C addition on soil  $CH_4$  uptake suggested that DOC input from forest organic layers can change the inhibition effect of N deposition on the soil atmospheric  $CH_4$  uptake, which depends on the types of deposited N.

## Methods

Site description and collection of forest soil and organic layer samples. The studying area locates near the National Research Station of Changbai Mountain Forestry Ecosystem (42°24' N, 128°6' E) in Jilin Province, northeastern China with a typical continental temperature climate. The average elevation of the area is 738 m with a flat topography. Based on regular meteorological measurements of the station during the period from 1982 to 2012, daily average air temperature and surface soil temperature in the field from late November to next early March generally rang from -5 °C to -30 °C, and snow depth in winter is normally within the range from 5 cm to 35 cm. For this reason, soil profile in winter can be frozen down to 1.0–1.5 m depth and a complete disappearance of such frozen soil layer normally occurs in the middle of May each year<sup>20</sup>. A mature broadleaf and Korean pine mixed forest and an adjacent white birch forest were selected for soil sampling. The former is of multi-layer structure with canopy density of 0.8 and the average age of dominating trees is about 200 years old; the later as a secondary forest has a more simple structure with canopy density of 0.6 and the average age of dominating trees is about 70 years. Due to relatively lower vegetation coverage and phototaxis property, soil moisture under the white birch forest stand is smaller than that under the mature mixed forest over the year, and the former is characterized by the relatively higher frequency of freezing and thawing cycles during non-growth season period. To collect composite forest soil and organic layer samples, eighteen 1 m imes 1 m plots were selected in each forest stand in October 2012. Mineral soil samples (0-10 cm) in each plot were collected using an 8-cm diameter auger after removing the ground surface mulch, and organic layer samples including fresh and semi-decomposed litter were collected. All samples were kept separately in air-tight plastic bags and rapidly transported to the laboratory within 24 h. The soil samples from each forest stand were mixed thoroughly, sieved (<2 mm) to remove small stones and debris, and then stored in the dark at 4°C prior to incubation and analysis of soil properties. The organic layer samples from each forest stand were dried at 60 °C for 48 h and milled for measurements of plant sample properties.

**Measurements of properties of forest organic layers and mineral soils.** Triplicate soils were dried at 105 °C for 24 h to determine moisture content. Fresh soil pH (soil/water, 1/2.5, w/w) and pH values in water extracts of forest organic layers (sample/water, 1/10, w/w) were respectively measured with a portable pH meter (PB-10, Sartorius, Germany). Total C and N concentrations in forest organic layers and soil samples were measured using an elemental analyzer (vario Macro cube, Elementar, Germany). Concentrations of soil microbial biomass C (MBC) and N (MBN) were measured by the chloroform fumigation and extraction method<sup>48,49</sup>. Fresh forest mineral soils (5.0 g) were extracted by shaking with 25 mL of 0.5 mol L<sup>-1</sup> K<sub>2</sub>SO<sub>4</sub> solution for 30 min and dried organic layer samples (5.0 g) by shaking with 50 mL of deionized water for 24 h on an end-over-end shaker. The suspensions were centrifuged at 4500 g for 5 min and then filtered into 50-mL plastic bottles via cellulose-acetate membrane filters (0.45  $\mu$ m pore size). Concentrations of NH<sub>4</sub><sup>+</sup>-N, NO<sub>3</sub><sup>-</sup>-N, total N (TN), and DOC in the soil extracts and DOC in the organic layer extracts were measured using a continuous flow analyzer (SAN<sup>++</sup>, SKALAR, the Netherlands). Concentrations of soil dissolved organic N (DON) were calculated according to the differences between TN and mineral N ( $NH_4^+$ -N and  $NO_3^-$ -N) concentrations in soil extracts. The soil MBC and MBN were calculated by the differences of K<sub>2</sub>SO<sub>4</sub>-extractable DOC and TN pools between fumigated and non-fumigated soils and divided by 0.4548-50, assuming that fumigation causes a release of microbial N in the same proportion as for microbial C. The glucose concentrations in water extracts of forest organic layers were determined by anthrone-sulfuric acid colorimetric method<sup>51</sup>. UV absorbance at 254 nm (UV<sub>254</sub>) of the water extracts was measured using a spectrophotometer (UV-2800A, Unico, USA). Water extracts of forest organic layers were diluted 40 times for measuring excitation-emission matrix (EEM) fluorescence spectra using a fluorometer (Fluoromax-4, Horiba, USA). For the EEMs, instrumental bias corrections were conducted with S/R model and inner filter corrections<sup>52</sup> were carried out using absorbance spectrum measured with a spectrophotometer (U-2000, Hitachi, Japan). Then, after subtracting the EEM of Milli-water, EEMs of the organic layer water extracts were calibrated to the water Raman signal<sup>53</sup> and expressed in Raman units (RU, nm<sup>-1</sup>) (Figure S1). Humification

Vegetation type	Moisture (%, w/w)	I	Total C (mg C g <sup>-1</sup> )	Total N (mg N g <sup>-1</sup> )	Bulk density (g cm <sup>-3</sup> )	$NO_3^{-}-N$ (µg N g <sup>-1</sup> )	${ m NH_4^+-N}\ (\mu gNg^{-1})$	$\frac{DON}{(\mu gNg^{-1})}$	$\frac{DOC}{(\mu gCg^{-1})}$	MBN (mg N kg <sup>-1</sup> )	MBC (mg C kg <sup>-1</sup> )	MBC:MBN ratio
WBF	31.7	5.67	9.43	0.75	0.73	17.7	20.7	26.9	176.6	312	1985	6.3
BKPF	50.7	5.87	11.78	0.92	0.64	39.4	4.8	28.2	167.7	226	1716	7.7

Table 7. Main soil properties under the two study temperate forests.

.....

index (HIX) was calculated by the ratio of two integrated regions of emission scan (sum of 436 to 480 nm divided by the sum of 300 to 344 nm) with excitation at 255 nm, indicating the relative humification of forest organic layer extracts<sup>54</sup>. The three components which the three fluorescence peaks represent are cited from *Chen et al.*<sup>55</sup>. Properties of water extracts of forest organic layers sampled under WBF and BKPF stands were shown in Table S1.

**Setup of incubation experiments.** Wetting (non-freezing) and freezing-thawing experiments were conducted during November 2012 to January 2013. Packed soil cores were made according to bulk densities of BKPF and WBF soils in the field (Table 7). A factorial design with two forest types (BKPF and WBF) and the addition of nutrients (glucose, namely Glu  $6.4 \text{ g C m}^{-2}$ , NH<sub>4</sub>Cl,  $4.5 \text{ g N m}^{-2}$ , KNO<sub>3</sub>,  $4.5 \text{ g N m}^{-2}$ , Glu+NH<sub>4</sub>Cl, Glu+KNO<sub>3</sub>) was established for the two incubation experiments; no nutrient addition was considered as control. The amounts of added N and C were approximately fourfold annual wet N input and twice glucose concentration in water extracts of organic layers (Table S1) under the two study forests, respectively. Experiments were replicated three times, giving a total of 72 packed soil cores.

Homogenized fresh soils (85 g) were transferred into 100-mL stainless steel cylinders (diameter in 50.5 mm) as a soil core. For each core, appropriate nutrients were precisely sprayed with deionized water as solutions onto the homogenized soil before packing to reach a water-filled pore space (WFPS) level of 55%; this operation were accomplished within 1 hour. In the freezing-thawing experiment 36 soil cores were frozen at -18 °C for 50 days, and the remaining 36 cores were sealed in gastight PVC cylinders (760 mL) with a gas sampling port equipped with 3-way stopcock separately to initiate the non-freezing experiment. Three PVC cylinders without soil served as blank. A long duration of freezing at -18 °C was simulated according to a severe winter frost from late December to next February near the study area<sup>20</sup>. In accordance with air temperature of the study area in late spring and autumn when soil freezing-thawing cycles intensively occur in the field, the soil cores were incubated at 10°C in two incubators (LRH250, Yiheng Instruments, China) for 15 days. Soil moisture at 55% WFPS was simulated according to field moisture at thaw under the two study forests. Deionized water was duly added for each soil core by weighting during the 15-day incubation to avoid evaporation. Gas sampling was performed from each soil core and the blank at 6, 12, 24, 37, 49, 75, 95, 119, 143, 167, 191, 215, 239, 263, 287, 311, 335, and 359 h after the incubation initiated, according to preliminary studies showing a linear decrease in headspace CH<sub>4</sub> concentration within 24 h after closure. Headspace gas samples of 30 ml were collected using 50-ml polypropylene syringes equipped with 3-way stopcock. Each time when gas sampling finished, all PVC cylinders were immediately taken outdoor to be well ventilated for 20 min and then sealed to continue the incubation till the next sampling time. The concentrations of  $CH_4$  in headspace gas samples were quantified by a gas chromatograph (Agilent 7890A, Franklin, USA) equipped with a flame ionization detector. The detector responses were calibrated using a certified gas standard, which contains  $2.11\,\mu L\,L^{-1}\,CH_4$  in air. Main properties of soil cores including moisture, bulk density, pH, NH<sub>4</sub><sup>+</sup>-N, NO<sub>3</sub><sup>-</sup>-N, DON, DOC, MBC, and MBN were measured immediately when the last gas sampling finished (359h), as mentioned above. Soil WFPS inside each core was calculated by soil bulk density and moisture<sup>56</sup>. After freezing at -18 °C for 50 days, all the soil cores from freezing experiment were placed separately inside PVC cylinders and immediately incubated at 10 °C to simulate the soil thawing process. Headspace gases sampling and measurements of headspace CH<sub>4</sub> concentration and soil properties were conducted at the same times of the non-frozen experiment.

**Calculation and statistical analysis.** Instantaneous rates of soil CH<sub>4</sub> uptake were calculated from the differences of headspace CH<sub>4</sub> concentration between the blank and each treatment divided by the period of time from sealing to gas sampling, and were expressed in  $\mu$ g CH<sub>4</sub>-C m<sup>-2</sup> h<sup>-1</sup>. The cumulative uptakes of CH<sub>4</sub> during the 15-day incubation were calculated as the sum of CH<sub>4</sub> uptake for each sampling and were expressed in mg CH<sub>4</sub>-C m<sup>-2</sup>. The average rates of CH<sub>4</sub> uptake during the 15-day incubation were calculated as the sum of CH<sub>4</sub> uptake for each sampling and were expressed in mg CH<sub>4</sub>-C m<sup>-2</sup>. The average rates of CH<sub>4</sub> uptake during the 15-day incubation were calculated by the slopes of linear regressions of cumulative CH<sub>4</sub> uptakes against the incubation time (determination coefficient of regression,  $R^2 > 0.95$ ), and were expressed in  $\mu$ g CH<sub>4</sub>-C m<sup>-2</sup> h<sup>-1</sup>. Means and standard errors for three replicates were calculated. Partial distributed data were normalized prior to statistical analysis. The absolute inhibition by N addition of CH<sub>4</sub> uptake was calculated by the differences of average rates of CH<sub>4</sub> uptake in the presence and absence of N sources (NH<sub>4</sub>Cl or KNO<sub>3</sub>), and its relative inhibition was calculated by the absolute inhibition divided by the average rate of CH<sub>4</sub> uptake in the absence of N addition.

All measured variables were examined for normality (Shapiro-Wilk test) and homogeneity (Levene's test) of variance and transformed where necessary. We used four-factor repeated analysis of variance (ANOVA) with vegetation type, N (NH<sub>4</sub>Cl or KNO<sub>3</sub>) and Glu addition, and freezing as fixed factors to assess their influences on the average rate of CH<sub>4</sub> uptake and soil properties. The another four-factor repeated ANOVA was used with vegetation type, N (NH<sub>4</sub>Cl or KNO<sub>3</sub>) and Glu addition as independent variables between subjects and with sampling time as independent variable within subjects, to assess their influences on the instant rate of CH<sub>4</sub> uptake and cumulative CH<sub>4</sub> uptake during the 15-day incubation. The three-factor repeated ANOVA with vegetation type, freezing and Glu addition as fixed factors was used to assess their impacts on the inhibition by NH<sub>4</sub>Cl and KNO<sub>3</sub>

of the soil CH<sub>4</sub> uptake. Pearson correlation between soil properties and the average rate of CH<sub>4</sub> uptake in forest soils without and with freezing was performed. Stepwise regression analysis was performed to assess the main soil properties which can affect the average rates of CH<sub>4</sub> uptake without and with freezing. Significant effects between treatments in soil properties and CH<sub>4</sub> uptake were determined at the P < 0.05 level using student *T*-test. All statistical analyses were conducted with the software SPSS for Windows (version 19.0, IBM Corp., USA).

### References

- 1. Intergovernmental Panel on Climate Change (IPCC). Fifth Assessment Report, Working Group I: Thephysical Science Basis. Cambridge University Press, Cambridge, pp. 507–507 (2013).
- Steudler, P. A., Bowden, R. D., Melillo, J. M. & Aber, J. D. Influence of nitrogen fertilization on methane uptake in temperate forest soils. *Nature* 341, 314–316 (1989).
- 3. Mosier, A., Schimel, D., Valentine, D., Bronson, K. & Parton, W. Methane and nitrous oxide fluxes in native, fertilized and cultivated grasslands. *Nature* **350**, 330–332 (1991).
- 4. Bodelier, P. L. E. & Laanbroek, H. J. Nitrogen as a regulatory factor of methane oxidation in soils and sediments. *FEMS Microbiol. Ecol.* 47, 265–277 (2004).
- Saari, A., Rinnan, R. & Martikainen, P. J. Methane oxidation in boreal forest soils: kinetics and sensitivity to pH and ammonium. Soil Biol. Biochem. 36, 1037–1046 (2004).
- Hütsch, B. W., Webster, C. P. & Powlson, D. S. Methane oxidation in soil as affected by land use, soil pH and N fertilization. Soil Biol. Biochem. 26, 1613–1622 (1994).
- Phillips, R. L., Whalen, S. C. & Schlesinger, W. H. Response of soil methanotrophic activity to carbon dioxide enrichment in a North Carolina coniferous forest. Soil Biol. Biochem. 33, 793–800 (2001).
- 8. Jang, I., Lee, S., Zoh, K. D. & Kang, H. Methane concentrations and methanotrophic community structure influence the response of soil methane oxidation to nitrogen content in a temperate forest. *Soil Biol. Biochem.* **43**, 620–627 (2011).
- 9. Xu, X. K., Han, L., Luo, X. B. & Han, S. J. Synergistic effects of nitrogen amendments and ethylene on atmospheric methane uptake under a temperate old-growth forest. *Adv. Atmos. Sci.* 28, 843–854 (2011).
- 10. Wang, Z. P. & Ineson, P. Methane oxidation in a temperate coniferous forest soil: effects of inorganic N. Soil Biol. Biochem. 35, 427-433 (2003).
- Reay, D. S. & Nedwell, D. B. Methane oxidation in temperate soils: effects of inorganic N. Soil Biol. Biochem. 36, 2059–2065 (2004).
   Xu, X. K. & Inubushi, K. Effects of N sources and methane concentrations on methane uptake potential of a typical coniferous forest and its adjacent orchard soil. Biol. Fertil. Soils 40, 215–221 (2004).
- Xu, X. K. & Inubushi, K. Effects of nitrogen sources and glucose on the consumption of ethylene and methane by temperate volcanic forest surface soils. *Chin. Sci. Bull.* 52, 3281–3291 (2007).
- 14. Fender, A. C. *et al.* The inhibiting effect of nitrate fertilisation on methane uptake of a temperate forest soil is influenced by labile carbon. *Biol. Fertil. Soils* **48**, 621–631 (2012).
- 15. Mochizuki, Y., Koba, K. & Yoh, M. Strong inhibitory effect of nitrate on atmospheric methane oxidation in forest soils. *Soil Biol. Biochem.* **50**, 164–166 (2012).
- Dijkstra, F. A., Morgan, J. A., von Fischer, J. C. & Follett, R. F. Elevated CO<sub>2</sub> and warming effects on CH<sub>4</sub> uptake in a semiarid grassland below optimum soil moisture. J. Geophy. Res. 116, G01007, doi: 10.1029/2010JG001288 (2011).
- Sullivan, B. W., Selmants, P. C. & Hart, S. C. Does dissolved organic carbon regulate biological methane oxidation in semiarid soils? Glob. Chang. Biol. 19, 2149–2157 (2013).
- Henry, H. A. L. Climate change and soil freezing dynamics: historical trends and projected changes. Clim. Chang. 87, 421–434 (2008).
- 19. Xu, X. K. & Luo, X. B. Effect of wetting intensity on soil GHG fluxes and microbial biomass under a temperate forest floor during dry season. *Geoderma* 170, 118–126 (2012).
- Xu, X. K., Duan, C. T., Wu, H. H., Li, T. S. & Cheng, W. G. Effect of intensity and duration of freezing on soil microbial biomass, extractable C and N pools, and N<sub>2</sub>O and CO<sub>2</sub> emissions from forest soils in cold temperate region. *Sci. China Earth Sci.* 59, 156–169 (2016).
- 21. Ding, W. X. & Cai, Z. C. Methane emission from natural wetlands in China: Summary of years 1995–2004 studies. *Pedosphere* 17, 475–486 (2007).
- Kim, D. G., Vargas, R., Bond-Lamberty, B. & Turetsky, M. R. Effects of soil rewetting and thawing on soil gas fluxes: a review of current literature and suggestions for future research. *Biogeosciences* 9, 2459–2483 (2012).
- West, A. E. & Schmidt, S. K. Wetting stimulates atmospheric CH<sub>4</sub> oxidation by alpine soil. *FEMS Microbiol. Ecol.* 25, 349–353 (1998).
   Blankinship, J. C., Brown, J. R., Dijkstra, P., Allwright, M. C. & Hungate, B. A. Response of terrestrial CH<sub>4</sub> uptake to interactive changes in precipitation and temperature along a climatic gradient. *Ecosystems* 13, 1157–1170 (2010).
- Wu, X., Brueggemann, N., Butterbach-Bahl, K., Fu, B. & Liu, G. Snow cover and soil moisture controls of freeze-thaw-related soil gas fluxes from a typical semi-arid grassland soil: a laboratory experiment. *Biol. Fertil. Soils* **50**, 295–306 (2014).
- Wieczorek, A. S., Drake, H. L. & Kolb, S. Organic acids and ethanol inhibit the oxidation of methane by mire methanotrophs. FEMS Microb. Ecol. 77, 28–39 (2011).
- Zhu, R. B. et al. Greenhouse gas emissions from penguin guanos and ornithogenic soils in coastal Antarctica: Effects of freezing-thawing cycles. Atmos. Environ. 43, 2336–2347 (2009).
- Wang, C. J. et al. Effect of stocking rate on soil-atmosphere CH<sub>4</sub> flux during spring freeze-thaw cycles in a northern desert steppe, China. PLoS One 7, doi: 10.1371/journal.pone.0036794 (2012).
- Groffman, P. M., Hardy, J. P., Driscoll, C. T. & Fahey, T. J. Snow depth, soil freezing, and fluxes of carbon dioxide, nitrous oxide and methane in a northern hardwood forest. *Glob. Chang. Biol.* 12, 1748–1760 (2006).
- Borken, W., Davidson, E. A., Savage, K., Sundquist, E. T. & Steudler, P. Effect of summer throughfall exclusion, summer drought, and winter snow cover on methane fluxes in a temperate forest soil. Soil Biol. Biochem. 38, 1388–1395 (2006).
- Holst, J. et al. Fluxes of nitrous oxide, methane and carbon dioxide during freezing-thawing cycles in an Inner Mongolian steppe. Plant Soil 308, 105–117 (2008).
- Gao, Y. H., Zeng, X. Y., Xie, Q. Y. & Ma, X. X. Release of carbon and nitrogen from alpine soils during thawing periods in the Eastern Qinghai-Tibet Plateau. Water Air Soil Pollut. 226, 209–218 (2015).
- 33. Mancinelli, R. L. The regulation of methane oxidation in soil. Ann. Rev. Microbiol. 49, 581-605 (1995).
- 34. Singh, J. S. *et al.* Effect of soil nitrogen, carbon and moisture on methane uptake by dry tropical forest soils. *Plant Soil* **196,** 115–121 (1997).
- 35. Dunfield, P. F., Topp, E., Archambault, C. & Knowles, R. Effect of nitrogen fertilizers and moisture content on CH<sub>4</sub> and N<sub>2</sub>O fluxes in a humisol: measurements in the field and intact soil cores. *Biogeochemistry* **29**, 199–222 (1995).
- Zhao, S. J. & Hanson, R. S. Variants of the obligate methanotroph isolate 761M capable of growth on glucose in the absence of methane. *Appl. Environ. Microbiol.* 48, 807–812 (1984).
- Wu, H. H., Xu, X. K., Duan, C. T., Li, T. S. & Cheng, W. G. Effect of vegetation type, wetting intensity, and nitrogen supply on external carbon stimulated heterotrophic respiration and microbial biomass carbon in forest soils. *Sci. China Earth Sci.* 58, 1446–1456 (2015).

- Bancroft, K., Grant, I. F. & Alexander, M. Toxicity of NO<sub>2</sub><sup>-</sup> effect of nitrite on microbial activity in an acid soil. Appl. Environ. Microbiol. 38, 940–944 (1979).
- 39. Harris, D., Voroney, R. P. & Paul, E. A. Measurement of microbial biomass N:C by chloroform fumigation-incubation. *Can. J. Soil Sci.* 77, 507–514 (1997).
- 40. Wallander, H., Nilsson, L. O., Hagerberg, D. & Rosengren, U. Direct estimates of C:N ratios of ectomycorrhizal mycelia collected from Norway spruce forest soils. *Soil Biol. Biochem.* **35**, 997–999 (2003).
- Burke, D. J., Smemo, K. A., López-Gutiérrez, J. C. & DeForest, J. L. Soil fungi influence the distribution of microbial functional groups that mediate forest greenhouse gas emission. Soil Biol. Biochem. 53, 112–119 (2012).
- Hilger, H. A., Cranford, D. F. & Barlaz, M. A. Methane oxidation and microbial exopolymer production in landfill cover soil. Soil Biol. Biochem. 32, 457–467 (2000).
- Morishita, T. et al. Effect of nitrogen deposition on CH<sub>4</sub> uptake in forest soils in Hokkaido, Japan. Soil Sci. Plant Nutr. 50, 1187–1194 (2004).
- 44. Benstead, J. & King, G. M. The effect of soil acidification on atmospheric methane uptake by a Maine forest soil. *FEMS Microbiol. Ecol.* **34**, 207–212 (2001).
- 45. Xu, X. K. & Inubushi, K. Responses of ethylene and methane consumption to temperature and pH in temperate volcanic forest soils. *Eur. J. Soil Sci.* **60**, 489–498 (2009).
- Dail, D. B., Davidson, E. A. & Chorover, J. Rapid abiotic transformation of nitrate in an acid forest soil. *Biogeochnistry* 54, 131–146 (2001).
- Larsen, K. S., Jonasson, S. & Michelsen, A. Repeated freeze-thaw cycles and their effects on biological processes in two arctic ecosystem types. *Appl. Soil Ecol.* 21, 187–195 (2002).
- Jenkinson, D. S. The determination of microbial biomass carbon and nitrogen in soil. In: Wilson, J. R. (Ed), Advances in Nitrogen Cycling in Agricultural Ecosystems. CAB International, Wallingford, pp. 368–386 (1988).
- Wu, J., Joergensen, R. G., Pommerening, B., Chaussod, R. & Brookes, P. C. Measurement of soil microbial biomass C by fumigation extraction - an automated procedure. Soil Biol. Biochem. 22, 1167–1169 (1990).
- 50. Jenkinson, D. S., Brooks, P. C. & Powlson, D. S. Measuring soil microbial biomass. Soil Biol. Biochem. 36, 5-7 (2004).
- 51. Deng, S. P. & Tabatabai, M. A. Colorimetric determination of reducing sugars in soils. Soil Biol. Biochem. 26, 473-477 (1994).
- McKnight, D. M. et al. Spectrofluorometric characterization of dissolved organic matter for indication of precursor organic material and aromaticity. *Limnol. Oceanogr.* 46, 38–48 (2001).
- 53. Lawaetz, A. J. & Stedmon, C. A. Fluorescence intensity calibration using the Raman scatter peak of water. Appl. Spectrosc. 63, 936-940 (2009).
- Zsolnay, A., Baigar, E., Jimenez, M., Steinweg, B. & Saccomandi, F. Differentiating with fluorescence spectroscopy the sources of dissolved organic matter in soils subjected to drying. *Chemosphere* 38, 45–50 (1999).
- Chen, W., Westerhoff, P., Leenheer, J. A. & Booksh, K. Fluorescence excitation emission matrix regional integration to quantify spectra for dissolved organic matter. *Environ. Sci. Technol.* 37, 5701–5710 (2003).
- Franzluebbers, A. J. Microbial activity in response to water-filled pore space of variably eroded southern Piedmont soils. *Appl. Soil Ecol.* 11, 91–101 (1999).

# Acknowledgements

This work was supported by the National Natural Science Foundation of China (21228701, 41275166, 41321064, and 41575154).

## **Author Contributions**

X.K.X. and W.G.C. designed the research. H.H.W. and C.T.D. performed the experiments. Writing and data intergration were done by H.H.W. and X.K.X. with significant contributions from W.G.C. and T.S.L. All authors edited and commented on the manuscript. X.K.X. and W.G.C. contributed equally to this work.

## **Additional Information**

Supplementary information accompanies this paper at http://www.nature.com/srep

Competing financial interests: The authors declare no competing financial interests.

**How to cite this article:** Wu, H.H. *et al.* Synergistic effects of dissolved organic carbon and inorganic nitrogen on methane uptake in forest soils without and with freezing treatment. *Sci. Rep.* **6**, 32555; doi: 10.1038/ srep32555 (2016).

This work is licensed under a Creative Commons Attribution 4.0 International License. The images or other third party material in this article are included in the article's Creative Commons license, unless indicated otherwise in the credit line; if the material is not included under the Creative Commons license, users will need to obtain permission from the license holder to reproduce the material. To view a copy of this license, visit http://creativecommons.org/licenses/by/4.0/

© The Author(s) 2016