



Dichlorodiphenyltrichloroethane Impairs Amyloid Beta Clearance by Decreasing Liver X Receptor α Expression

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Abnormal amyloid beta (A β) clearance is a distinctive pathological mechanism for Alzheimer's disease (AD). ATP-binding cassette transporter A1 (ABCA1), which mediates the lipidation of apolipoprotein E, plays a critical role in AB clearance. As an environmental factor for AD, dichlorodiphenyltrichloroethane (DDT) can decrease ATP-binding cassette transporter A1 (ABCA1) expression and disrupt A β clearance. Liver X receptor α $(LXR\alpha)$ is an autoregulatory transcription factor for ABCA1 and a target of some environmental pollutants, such as organophosphate pesticides. In this study, we aimed to investigate whether DDT could affect A β clearance by targeting LXR α . The DDT-pretreated H4 human neuroglioma cells and immortalized astrocytes were incubated with exogenous AB to evaluate AB consumption. Meanwhile, cytotoxicity and $LXR\alpha$ expression were determined in the DDT-treated cells. Subsequently, the antagonism of DDT on LXRa agonist T0901317 was determined in vitro. The interaction between DDT and LXRa was predicted by molecular docking and molecular dynamics simulation technology. We observed that DDT could inhibit A β clearance and decrease the levels of LXRa mRNA and LXRa protein. Moreover, DDT is supposed to strongly bind to LXR α and exert antagonistic effects on LXR α . In conclusion, this study firstly presented that DDT could inhibit LXRα expression, which would contribute to Aβ clearance decline in vitro. It provides an experimental basis to search for potential therapeutic targets of AD.

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INTRODUCTION

Alzheimer's disease (AD) is a neurodegenerative disease characterized by progressive loss of memory and other cognitive domains (Erkkinen et al., 2018). As the most common cause of dementia, AD affects about 30–46 million people worldwide (Caputo et al., 2020). With a global aging population, 75 million people are expected to suffer from AD in 2030 (Realdon et al., 2016). Owing to the lack of effective treatments for dementia, the current goal of AD is early detection and treatment to prevent the exacerbation of cognitive impairment. Except for early-onset familial AD, the majority of late-onset AD is caused by the combination of genetic and environmental factors (Alzheimer's Association, 2013). As one of the most persistent and widely used organochlorine pesticides,

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dichlorodiphenyltrichloroethane (DDT) seems to be a critical environmental factor for AD (van Wendel De Joode et al., 2001; Singh et al., 2013; Richardson et al., 2014). Kim et al. (2015) have reported that the elders with high serum DDT concentration had about three times higher risks of low cognition. Based on a small study, DDT was found more often in AD brains (n = 7) than the control (n = 14) (Fleming et al., 1994). To screen out the potential targets of DDT-induced AD could be beneficial for the early detection and treatment of AD. Nevertheless, it remains unclear how DDT leads to AD.

"Amyloid hypothesis" is the most widely accepted pathological mechanism for AD. According to this hypothesis, amyloid beta (A β) pathology in the brain is followed by the development of neurofibrillary tau pathology and commences several years before obvious memory loss (Jack et al., 2009). As an initial step for AD, the increased A β level is attributed to an imbalance between A β production and A β clearance (Bates et al., 2009). More notably, the vast majority of A β accumulation is the result from deficient A β clearance in microglia and astrocytes (Corona et al., 2016). ATP-binding cassette transporter A1 (ABCA1), which mediates the lipidation of apolipoprotein E (ApoE), is critical for A β clearance (Koldamova et al., 2014). In AB protein precursor transgenic mice, ABCA1 deletion aggravates AB deposition and ABCA1 overexpression decreases AB deposition (Wahrle et al., 2005, 2008). In our previous study (Li et al., 2015), DDT significantly decreased ABCA1 protein expression. In addition, aberrant A^β clearance could be ameliorated when ABCA1 protein levels were rescued. Consequently, we proposed that ABCA1 could play an important role in deficient AB clearance caused by DDT. However, few studies have investigated how DDT affects ABCA1 expression.

As an autoregulatory transcription factor, liver X receptor α (LXRα) can activate ABCA1 expression by forming heterodimers with retinoic X receptor (Laffitte et al., 2001; Terwel et al., 2011; Wang et al., 2014). In transgenic mouse models of AD, the LXR agonist can increase ABCA1 expression, decrease Aβ deposition, and improve cognitive performance (Jiang et al., 2008). LXRα-null mice have lower brain ABCA1 levels at the age of 12 weeks old, when no amyloid plaques can be detected. It suggested that LXRa-regulated ABCA1 expression could be an early factor for AD progression (Zelcer et al., 2007). Recently, some environmental pollutants such as organophosphate pesticides can target the oxysterol-binding domain of LXRa to inhibit LXRa expression directly (Mozzicafreddo et al., 2015). Similar to organophosphate pesticides, DDT has obvious neurotoxicity and can regulate the expression of several nuclear receptors (Kelce et al., 1995; Medina-Diaz et al., 2007; Kazantseva et al., 2013). Therefore, we suspected that DDT could regulate LXRa expression directly. In this study, we aimed to explore the role of LXRa in DDT-induced AB deposition. H4 human neuroglioma cells and immortalized astrocytes were selected to evaluate A β clearance and LXR α expression after DDT treatment. The interaction between DDT and LXRa was predicted by molecular docking simulation technology. This study will provide evidence for elucidating the underlying mechanism of DDT-induced A β deposition, and for the potential the rapeutic target of AD.

MATERIALS AND METHODS

Reagents

The following antibodies were used in this study: Actin (AC-15) from Sigma-Aldrich (St Louis, MO, USA), Aβ (82E1) from IBL International (Hamburg, Germany), and LXRa (ab41902) from Abcam Inc. (Cambridge, MA, USA). T0901317 (ab142808) was from Abcam Inc. (Cambridge, MA, USA). $A\beta_{40}$ monomers were purchased from American Peptide Company (Sunnyvale, CA, USA). 44-DDT (386340) from Sigma-Aldrich (St. Louis, MO, USA) was dissolved into 10 mM stock solution in dimethyl sulfoxide (DMSO). Alamarblue[®] cell viability reagent, Dulbecco's modified Eagle's medium (DMEM), and Trizol reagent were purchased from Invitrogen Company (Carlsbad, CA, USA). High capacity cDNA transcription kit was supplied by Applied Biosystems (Foster City, CA, USA). Tris-Tricine gel, Tris-Glycine extended (TGXTM) gel, and 0.2µm pore size nitrocellulose membrane were supplied by Bio-Rad Laboratories Incorporation (Hercules, CA, USA). RIPA buffer was purchased from Millipore Corporation (Bedford, MA, USA). Fresh serum-free DMEM supplemented N2 was obtained from BRL (Grand Island, NY, USA). EDTA-free protease inhibitor cocktail was provided by Roche (Mannheim, Germany). BCA protein assay kit was obtained from Thermo Fisher Scientific (Waltham, MA, USA). Lumigen TMA-100 ECL detection kit was from Lumigen Inc. (Southfield, MI, USA). Real-time reversetranscription PCR (RT-PCR) kit (including Premix Ex Taq and SYBR Green I) was purchased from Takara Biotechnology Company Limited, Ltd. (Dalian, China). The other chemical reagents were of analytical grade.

Cell Culture

H4 human neuroglioma cells were cultured in DMEM with 10% fetal bovine serum and 1% penicillin/streptomycin. Immortalized astrocytes, which were derived from ApoE3targeted replacement mice (Zhao et al., 2014), were maintained in DMEM/F-12 with 20% fetal bovine serum, 2 mM L-glutamine, glutamine, 1 mM sodium pyruvate, $1 \times$ nonessential amino acids, and 1% penicillin/streptomycin. All the cells were incubated at 37°C in a humidified 5% CO₂ incubator. For cytotoxicity assay, cells were seeded and cultured overnight in 96-well plates. For protein and mRNA detection, cells were seeded on the day before treatments in 12-well plates. When the cells were at a confluence of 70%, the incubated cells were treated with 0.1% DMSO vehicle (as the negative control), 1 μ M DDT, and 10 μ M DDT, respectively. Based on our previous study (Li et al., 2015), 1 μ M DDT slightly increased A β levels and 10 μ M DDT had the most dramatic effect on Aß levels.

Cytotoxicity Assay

The incubated cells were pretreated with DDT or DMSO vehicle for 48 h. Based on the manufacturer's guide, the *in vitro* cytotoxic effects of DDT were measured by the Alamarblue[®] cell viability reagent.

A_β Clearance Assay

The incubated cells were pretreated with DDT or DMSO vehicle for 24 h. After the removal of media, cells were treated with the fresh serum-free DMEM supplemented N2 containing DDT or DMSO and incubated with 200 nM AB40 monomers dissolved in DMSO for 24 h. Subsequently, the media were collected and separated on a 16.5% Tris-Tricine gel. The cellular lysates were separated on a 4–20% TGXTM gel. The levels of A β were determined via Western blotting analysis. All the experiments were repeated independently three times. The cells were divided into three groups and were respectively treated with T0901317, T0901317 with DDT, and DMSO vehicle for 24 h. After the removal of media, cells were treated with the fresh serum-free DMEM supplemented N2 containing T0901317, T0901317 with DDT, or DMSO vehicle and then incubated with 200 nM A β_{40} monomers dissolved in DMSO for another 24 h. The cellular lysates were separated on a 4–20% TGXTM gel. The levels of A β in the remaining medium were determined via Western blotting analysis.

RNA Isolation and Real-Time RT-PCR

The incubated cells were pretreated with DDT or DMSO vehicle for 48 h. Total RNAs were extracted using the Trizol reagent and reverse transcribed with the high-capacity cDNA transcription kit following the manufacturer's instruction. Real-time RT-PCR was performed in triplicate in 20-µl volumes in 96-well plates with StepOnePlus system from Applied Biosystems. Glyceraldehyde phosphate dehydrogenase (GAPDH) was selected as a normalization control. Relative mRNA levels were calculated by comparative Ct method. The primers were designed by Methyl Primer Express 1.0 from Applied Biosystems. The primers of human GAPDH were as follows: forward 5'-CAAGGGCATCCTGGGCTAC-3'; reverse 5'-TTGAAGTCAGAGGAGACCACCTG-3'. The primers of mouse GAPDH were as follows: forward 5'-AGGTCGGTGT GAACGGATTTG-3'; reverse 5'-TGTAGACCATGTAGTTGA GGTCA-3'. The primers of human LXRa were as follows: forward 5'-TTGCCTTGCTCATTGCT-3'; reverse 5'-CATCCG TGGGAACATCA-3'. The primers of mouse LXRa were as follows: forward 5'-GCTCATTGCCATCAGCAT-3'; reverse 5'-AGCATCCGTGGGAACATCA-3'.

Western Blotting Analysis

The incubated cells were pretreated with T0901317, T0901317 with DDT, or DMSO vehicle for 48 h. Subsequently, the incubated cells were lysed in RIPA buffer with EDTA-free protease inhibitor cocktail at 4°C for 30 min. Subsequently, the incubated cells were centrifuged at $17,000 \times g$ for 10 min and the supernatants were collected. Protein concentration was quantified using the BCA protein assay kit. Equal amounts of total protein for each lysate were separated on a 4–20% TGXTM gel and transferred to 0.2-µm pore size nitrocellulose membranes. Membranes were treated with specific primary antibodies and peroxidase-conjugated secondary antibodies, and developed using the Lumigen TMA-100ECL detection kit. LAS-4000 was employed for image collection. All the experiments were repeated independently at least three times.

Molecular Docking and Molecular Dynamics Simulation

Molecular docking study was performed to investigate the binding mode between the DDT and the LXR α using AutoDock Vina 1.1.2 (Trott and Olson, 2010). The three-dimensional (3D) structure of the LXR α (PDB ID: 3IPQ) was downloaded from the RCSB database¹. The 3D structure of the DDT was drawn by ChemBioDraw Ultra 14.0 and ChemBio3D Ultra 14.0 software. The AutoDockTools 1.5.6 package (Sanner, 1999; Morris et al., 2009) was employed to generate the docking input files. The binding site of the LXR α was identified as center_x: 43.079, center_y: 16.295, and center_z: -6.463 with dimensions size_x: 15, size_y: 15, and size_z: 15. In order to increase the docking accuracy, the value of exhaustiveness was set to 20. For Vina docking, the default parameters were used if it was not mentioned. Then, the molecular dynamics study was performed to revise the docking result.

The Amber 14 (Götz et al., 2012; Pierce et al., 2012; Salomon-Ferrer et al., 2013) and AmberTools 15 programs were used for molecular dynamics simulation of the selected docked pose. Cryptotanshinone was first prepared by ACPYPE (Sousa Da Silva and Vranken, 2012), a tool based on ANTECHAMBER (Wang et al., 2004, 2006) for generating automatic topologies and parameters in different formats for different molecular mechanics programs, including calculation of partial charges. Then, the force field "leaprc.gaff" (generalized amber forcefield) was used to prepare the ligand, while "leaprc.ff14SB" was used for the receptor. The system was placed in a rectangular box (with a 10.0-Å boundary) of TIP3P water using the "SolvateOct" command with the minimum distance between any solute atoms. Equilibration of the solvated complex was done by carrying out a short minimization (2,000 steps of each steepest descent and conjugate gradient method), 1,000 ps of heating, and 500 ps of density equilibration with weak restraints using the GPU accelerated PMEMD (Particle Mesh Ewald Molecular Dynamics) module. At last, 40 ns of molecular dynamics simulations was carried out. All the molecular dynamics analysis was performed on Dell Precision T5500 workstation.

Binding Free Energy and Energy Decomposition Per Residue Calculation

The binding free energies (ΔG_{bind} in kcal/mol) were calculated using the Molecular Mechanics/Generalized Born Surface Area (MM/GBSA) method, implemented in AmberTools 15. Moreover, to identify the key protein residues responsible for the ligand binding process, the binding free energy was decomposed on a per-residue basis. For each complex, the binding free energy of MM/GBSA was estimated as follows: $\Delta G_{\text{bind}} = G_{\text{complex}} - G_{\text{protein}} - G_{\text{ligand}}$. ΔG_{bind} was the binding free energy and G_{complex} , G_{protein} , and G_{ligand} were the free energies of complex, protein, and ligand, respectively.

Statistical Analysis

SPSS 15.0 from SPSS Science Incorporation (Chicago, IL, USA) and GraphPad Prism 5.0 from GraphPad Software (La Jolla, CA,

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USA) were employed for data analysis. All the data were shown as mean \pm SEM. First, the equal variance test (Levene Median test) and normality test (Kolmogorov–Smirnov test) were used to determine the statistical significance. The two-tailed Student's *t*-test was used when the data meet the assumptions of parametric testing. Statistical significance was set at P < 0.05.

RESULTS

Cytotoxicity Under DDT Exposure

The acute high dose of DDT exposure may result in nonspecific cytotoxicity, which is responsible for the poor cell viability and performance. Thus, Alamarblue[®] assay was used to investigate whether the DDT doses in this study could trigger nonspecific cytotoxicity in H4 human neuroglioma cells (**Figure 1A**) and immortalized astrocytes (**Figure 1B**). Compared to the negative control DMSO group, no obvious cell toxicity was shown in cells treated with either 1 μ M DDT or 10 μ M DDT.

A_β Clearance Under DDT Exposure

Because H4 human neuroglioma cells and immortalized astrocytes cannot secrete detectable AB endogenously, we employed these cells to assess the ability of $A\beta$ clearance. At first, the incubated cells were pretreated with DDT or DMSO vehicle for 24 h. After the removal of media with serum, cells were treated with the fresh serum-free media containing DDT or DMSO and incubated with 200 nM AB40 for another 24 h. The levels of the remaining $A\beta$ in the media were then examined by Western blotting analysis. As shown in Figures 2A,B, the DDT treatment group exhibited much more A^β remaining in the media than the control group, especially the 10 μ M DDT group. According to the increasing tendency from the semiquantitative assay (**Figures 2C,D**, P < 0.01), the higher dose of DDT exposure could cause more $A\beta$ remaining in the media. These results suggested that DDT could impair Aβ clearance *in vitro*. Exposure to 10 µM DDT could result in more obvious AB deposition without nonspecific cytotoxicity, which provides a suitable DDT dose for our further in vitro experiments about evaluating the effects of DDT on LXRa.

LXR_{\alpha} Expression Under DDT Exposure

Our previous study demonstrated that ABCA1 expression is downregulated by DDT treatment. Because ABCA1 is a critical target of LXR α in brain, we proposed that LXR α may account for the DDT-ABCA1 molecular pathway. According to the results in **Figure 3**, H4 human neuroglioma cells and immortalized astrocytes were treated with 10 μ M DDT to evaluate the effects of DDT on LXR α expression. Compared to the DMSO control group, the DDT group exhibited significantly lower concentration of LXR α mRNA (**Figures 3A,B**, *P* < 0.001). Likewise, the DDT group demonstrated a lower content of LXR α







protein than the control (**Figures 3C,D**). The semiquantitative assay of LXR α protein also showed a statistically significant difference between these two groups (**Figures 3E,F**, *P* < 0.05). It suggested that DDT could decrease LXR α expression *in vitro*.

Autoregulation of LXR α and Antagonistic Effects of DDT on LXR α

Autoregulation of the LXRa gene is an important component of lipid-inducible efflux pathway in human macrophages (Laffitte et al., 2001). It is unknown whether this autoregulation exists in H4 human neuroglioma cells. According to the results in Figure 4A, H4 human neuroglioma cells were treated with 10 μ M LXR α agonist T0901317. Compared to the DMSO control group, the T0901317 group significantly increased ABCA1 and LXRa protein levels, as well as enhanced AB clearance. The T0901317+DDT group exhibited lower ABCA1 protein levels, lower LXRa protein levels and poorer Aß clearance than the T0901317 group. These results implied that DDT exerted an antagonistic effect on this autoregulation. The semiquantitative assay of ABCA1, LXR α , and A β protein also showed a statistically significant difference between these groups (Figure 4B, P < 0.01). These data suggested that LXR α agonist T0901317 could activate LXRa expression and LXRa may bind to its own promoter to enhance its transformation. DDT may compete for the binding site of LXRa with LXRa agonist to exert antagonistic effects.

Molecular Dynamics Results

To determine the competition of DDT with LXR α agonist T0901317 and explore the potential binding mode between the DDT and the LXR α , molecular docking and molecular dynamics simulation were performed using the AutoDock Vina 1.1.2 and Amber14 software package. The binding mechanism of LXR α with DDT was determined by 40-ns molecular dynamics simulation based on the docking results. To explore the dynamic stability of the complex and to ensure the rationality of the sampling strategy, the root-mean-square deviation (RMSD) values of the protein backbone based on the starting structure along the simulation time were calculated and plotted in **Figure 5A**. As shown in **Figure 5A**, the protein structures of the two systems were stabilized during the 40-ns simulation.

The root-mean-square fluctuations (RMSF) of the residues of the whole protein in the LXR α -DDT complex and in the free LXR α were calculated to reveal the flexibility of the residues. The RMSF of these residues are shown in **Figure 5B**, clearly depicting different flexibilities in the binding site of LXR α between the presence and absence of the DDT. The majority of the residues in the LXR α binding site that bind with DDT showed a small degree of flexibility with a RMSF of less than 2 Å when compared with the free LXR α , indicating that these residues seem to be more rigid as a result of binding to DDT.

To gain more information about the residues surrounding the binding site and their contribution to the system, the electrostatic, Van der Waals, solvation, and total contribution of the residues to the binding free energy were calculated with the MMGBSA method. The summations of the per-residue interaction free energies were separated into Van der Waals (ΔE_{vdw}) , solvation (ΔE_{sol}) , electrostatic (ΔE_{ele}) , and total contribution (ΔE_{total}). In the LXR α -DDT complex, the residues Phe-315 and Phe-326 with the ΔE_{vdw} of <-1.8 kcal/mol (Figure 5C) showed an appreciable Van der Waals interactions with the DDT because of the close proximity between the residues and the DDT, forming $CH-\pi$ interaction and Cl- π interaction, respectively (Figure 5D). Except for the residues Phe-315 and Phe-326, the majority of the decomposed energy interaction originated from Van der Waals interactions, apparently through hydrophobic interactions (i.e., Phe-257, Leu-260, Ala-261, Met-298, Leu-299, Leu-331, Ile-336, Ile-339, Leu-428, and Trp-443). In addition, the total binding free energy for the LXRa-DDT complex calculated according to the MMGBSA approach, and the estimated ΔG_{bind} of -36.5 kcal/mol was found for DDT, suggesting that DDT could strongly bind to LXRa and interact with the binding site of LXRa. In summary, the above molecular simulation gave a rational explanation of the interactions between DDT and LXRa, which provided valuable information for the LXRα antagonist.

DISCUSSION

In the past few decades, AD pathogenesis has been focused on the genetic mechanisms. Recently, environmental factors contributing to AD are drawing more and more attention (Hayden et al., 2010; Iraola-Guzman et al., 2011; Singh et al.,







complex. (D) The predicted binding mode of DDT in LXR α binding pocket obtained from molecular dynamics simulation.

2013; Billioti De Gage et al., 2014; Richardson et al., 2014). As an organochlorine pesticide strongly associated with AD, DDT has been extensively used as a broad-spectrum insecticide in agriculture since World War II (Davies et al., 2007; Greenwood, 2014). The common use of DDT in agriculture was banned in the 1970s–1980s, but DDT is still used for controlling disease vectors responsible for malaria and the Zika virus in developing countries (Wnuk et al., 2020). DDT is known to bioaccumulate in human and is negatively associated with cognitive scores (Kiyosawa et al., 2008; Kim et al., 2015). DDT will be metabolized into a major and more stable derivative, dichlorodiphenyldichloroethylene (DDE), which can accumulate in organs for a long time that are rich in adipose tissue, such as the brain (Di et al., 2017). Recently, the concept of an AD exposome was reviewed by Finch and Kulminski (2019). They addressed major gaps in understanding environmental contributions to the AD. AD is characterized by the impaired clearance of brain Aβ (Wildsmith et al., 2013). Our previous study has shown that DDT could dramatically decrease A β clearance in H4 human neuroglioma cells, which provides a potential mechanism for



DDT-induced AD (Li et al., 2015). Astrocytes play an essential role in AB clearance and are widely used for AB study (Zhao et al., 2014). These two types of cells cannot secrete detectable A β endogenously, in which the clearance of exogenous A β can be measured accurately without the confounding effects of endogenous AB. Thus, the abovementioned cells were selected to investigate how DDT impairs AB clearance in this study. It is not practical for a long-term chronic DDT exposure in cell culture. As to acute high-dose exposure, it is essential to select a suitable dose in cell experiments. In line with our previous study, 10 μ M DDT can inhibit A β clearance much heavier than 1 μ M DDT without obvious cell toxicity. Approximately 10.2 µM DDE has been found in human serum samples from elderly individuals (Kim et al., 2015); most of the subjects were exposed to DDT for the whole lifetime, the exposure levels to DDT in older persons were presumably much higher when they were younger, and they have likely experienced a higher body burden during their lifetime than what we can estimate from the current serum concentrations (Kim et al., 2015). In addition, we used 10 μ M DDE to perform the same assay and had the much similar results (data not shown). In our opinion, 10 µM DDT exposure should be reasonable in the current study.

Due to the similar rate of A β production and clearance, even minor deficits in A β clearance will lead to A β accumulation (Corona et al., 2016). Cumulatively, ABCA1-mediated ApoE lipidation is thought to be necessary for the efficient clearance of A β and the improvement of cognitive impairment (Wahrle et al., 2008; Corona et al., 2016; Fitz et al., 2017). Interestingly, dysfunctional ABCA1 contributes to the development of type 2 diabetes through increased cholesterol levels in pancreatic β -cells (Koldamova et al., 2014). In addition, ABCA1 can interact with other genetic risk factors (such as ApoE4) to worsen the AD phenotype (Fitz et al., 2017). Previously, we have identified for the first time that ABCA1 is the downstream target gene adversely affected by DDT (Li et al., 2015). However, it remains unclear how DDT regulates ABCA1 expression. In AD mouse models, the agonists of nuclear receptors have been proven to reduce A β pathology and improve cognition (Skerrett et al., 2014). To our knowledge, ABCA1 is transcriptionally regulated by LXR α , which is a nuclear receptor. LXR activation has been regarded as a top biological pathway to ameliorate AD-related cognitive impairment and A β accumulation (Niculescu et al., 2020). Among the LXR subtypes, LXR α mainly mediates the metabolism of cholesterol and A β in the brain (Terwel et al., 2011). In our present study, we observed significantly lower concentration of LXR α mRNA and protein in H4 human neuroglioma cells and immortalized astrocytes treated with 10 μ M DDT. Combined with our previous results, we speculated that DDT could inhibit LXR α expression to disrupt ABCA1 expression and A β clearance.

Furthermore, we are wondering how DDT affects LXRa expression. To our knowledge, LXRa is a ligand-activated transcription factor with the ability of positive autoregulation (Laffitte et al., 2001). Some environmental pollutants, such as organophosphates, have been reported to act like ligands or modulators to affect LXRa activity via binding to the oxysterolbinding domain of LXRa directly (Mozzicafreddo et al., 2015). DDT and organophosphates have structural similarities in their chemical features. Our present study also implied that DDT could have the antagonistic effects on LXRa. It has been reported that DDT could regulate the activity of estrogen receptors by binding to estrogen receptors directly (Frigo et al., 2005; Matsushima, 2018). Additionally, our present molecular docking simulation analysis suggested that DDT could bind to LXRa directly and become a stable bimolecular complex. Moreover, the binding site (i.e., Phe-257, Leu-260, Ala-261, Met-298, Leu-299, Leu-331, Ile-336, Ile-339, Leu-428, and Trp-443) is in the LXRa C-terminal, which is the ligand binding domain. It also implied that DDT could compete for the ligand binding site of LXR α with LXR α agonist. Therefore, we speculated that DDT may impair ABCA1 expression by targeting LXRa directly. In summary, we identified for the first time that DDT could inhibit LXR α expression and interfere with A β clearance. Furthermore, our results implied that DDT may have direct antagonizing effects on LXR α expression by binding to it (**Figure 6**). Our present study provides evidence for identifying the new targets of DDT-induced molecular pathway in AD. It also provides a valuable experimental basis to explain how environmental factors, especially pesticides, affect AD. In the future, more definitive studies on mice and human are warranted to assess the effects of environmental factors on neurodegenerative disorders.

DATA AVAILABILITY STATEMENT

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

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AUTHOR CONTRIBUTIONS

All the authors listed have made a substantial, direct, and intellectual contribution to this work, and approved the submitted version for publication.

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Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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