



# Microplastics-Induced Eryptosis and Poikilocytosis in Early-Juvenile Nile Tilapia (*Oreochromis niloticus*)

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This study aims to assess the impact of microplastics (MPs) on erythrocytes using eryptosis (apoptosis) and an erythron profile (poikilocytosis and nuclear abnormalities), considered to be novel biomarkers in Nile tilapia (*Oreochromis niloticus*). In this study, four groups of fish were used: The first was the control group. In the second group, 1 mg/L of MPs was introduced to the samples. The third group was exposed to 10 mg/L of MPs. Finally, the fourth group was exposed to 100 mg/L of MPs for 15 days, following 15 days of recovery. The fish treated with MPs experienced an immense rise in the eryptosis percentage, poikilocytosis, and nuclear abnormalities of red blood cells (RBCs) compared with the control group in a concentration-dependent manner. Poikilocytosis of MP-exposed groups included sickle cell shape, schistocyte, elliptocyte, acanthocyte, and other shapes. Nuclear abnormalities of the MPs-exposed groups included micronuclei, binucleated erythrocytes, notched, lobed, blebbed, and hemolyzed nuclei. After the recovery period, a greater percentage of eryptosis, poikilocytotic cells, and nuclear abnormalities in RBCs were still evident in the groups exposed to MPs when crosschecked with the control group. The results show concerning facts regarding the toxicity of MPs in tilapia.

**Keywords:** microplastics, poikilocytosis, apoptosis, tilapia, *Oreochromis niloticus*, erythrocytes

## INTRODUCTION

Plastics consist of small monomers polymerized with supplements of additives, such as stabilizers, plasticizers, and pigments (Xu et al., 2019). Approximately, 300 million tons of manufactured plastics are used in industrial processes and food packing each year (PlasticsEurope., 2015). Most of the plastic wastes are discarded in aquatic ecosystems, especially in developing countries (Karbalaeei et al., 2018).

There is a relatively high presence of microplastics (MPs) in freshwater bodies adjacent to the highly populated urban areas (Eriksen et al., 2013; Yonkos et al., 2014; Zhao et al., 2014; Lasee et al., 2017). Eriksen et al. (2013) found an approximate average value of MPs (mesh size 333 μm)

at 43,000 particles/km<sup>2</sup> with the highest abundance areas containing over 466,000 particles/km<sup>2</sup> adjacent to the major cities. Lasee et al. (2017) found that the average MP concentrations in samples obtained from lakes ranged from non-detectable (ND) to  $5.51 \pm 9.09$  mg/L for varying size fractions. Even though it seems that the MP concentrations are directly proportional to the urban population, the research showed relatively high concentrations of MPs in remote freshwater environments with extremely low population density and industrialization (Free et al., 2014; Yonkos et al., 2014; Fischer et al., 2016). It has been hypothesized that the urban water systems are immensely underrepresented in the MP life cycle despite their potential to significantly alter the microbial distributions by availing downstream transport and increased levels of plastic-consuming microbes in these water systems (McCormick et al., 2014).

The toxic effects of MPs on the aquatic organisms are debatable and differ across the species according to the size and type of MPs as well as the presence of heavy metals and pesticides (Rainieri et al., 2018; Wu et al., 2019). In fish, several adverse effects, include diminished predatory prowess, endocrine inhibition, oxidative stress, hepatic stress, and intestinal alterations (Rochman et al., 2013; Ferreira et al., 2016; Pedà et al., 2016; Hamed et al., 2019a, 2020). Although there are many pieces of research on the MPs effects on fish, still, there is not enough information on the effect of MPs on the alterations of blood cells and molecular damage especially on the Nile tilapia (*Oreochromis niloticus*) (Hamed et al., 2019a, 2020). Hematological, biochemical, and antioxidant alterations recorded in the fish after exposure to MPs are attributed to their toxic effects; however, little is known about the damage mechanism inside the cells and tissues (Hamed et al., 2019a, 2020). Understanding the mechanism of cellular damage associated with MP exposure would help develop protective treatments at both the cellular and tissue levels. Blood and its constituents represent the first point of entry for MPs in fish due to the direct contact between the gills and contaminated water (Barboza et al., 2020). Accordingly, the erythrocytes in target organs offer a unique tool for assessing the cytotoxicity in fish (Fazio, 2019; Soliman and Sayed, 2020). Eryptosis and erythron profiles (poikilocytosis and nuclear abnormalities) of fish erythrocytes are valuable biomarkers of the toxicity of different chemicals, pharmaceutical residues, and ultraviolet radiation (Mekkawy et al., 2011; Morina et al., 2012; Sayed et al., 2013, 2015, 2016, 2017; Sayed, 2016; Sula et al., 2020). The internal health conditions of the fish can be classified according to the indices iterating serum biochemical, hematological, and immunological aspects (Edsall, 1999; Luskova et al., 2002). This allows using these alterations to study the mechanisms underlying the hazardous effects caused by pollutants (Luskova et al., 2002). The hematological properties, such as red blood cell (RBC), hematocrit (Ht), and hemoglobin (Hb) are vital indicators for the evaluation of the health status of fish after being exposed to different environmental stresses, bacterial infections, and chemical toxicity (Kim et al., 2019, 2020). MP invades the circulatory system of fish, causing lethal reactions and metabolic disorders, e.g.,

endocrine disorders, oxidative stress, gene expression, and immune responses (Ma et al., 2020). Fish-absorbed MPs diffuse into the blood system through the cell membrane, causing accumulation in the body. Hence, the hematological aspects can be sensitive indicators of MP exposure in fish (Scanes et al., 2019). Hamed et al. (2019a) stated that the blood properties, e.g., hemoglobin, RBCs, and hematocrit, of *O. niloticus* were altered after being exposed to MP. This phenomenon caused the blood hemodilution after tissue damage and MP toxicity-induced hemolysis.

The Nile tilapia is a prominent fish in aquaculture today. It acts as a valuable experimental model in toxicological studies since it can be easily manipulated and observed under experimental conditions (Soliman, 2017). To the best of our knowledge, this is the first study that aims to examine the cytotoxicity of MPs in tilapia using novel biomarkers, such as eryptosis, poikilocytosis, and nuclear abnormalities of RBCs.

## MATERIALS AND METHODS

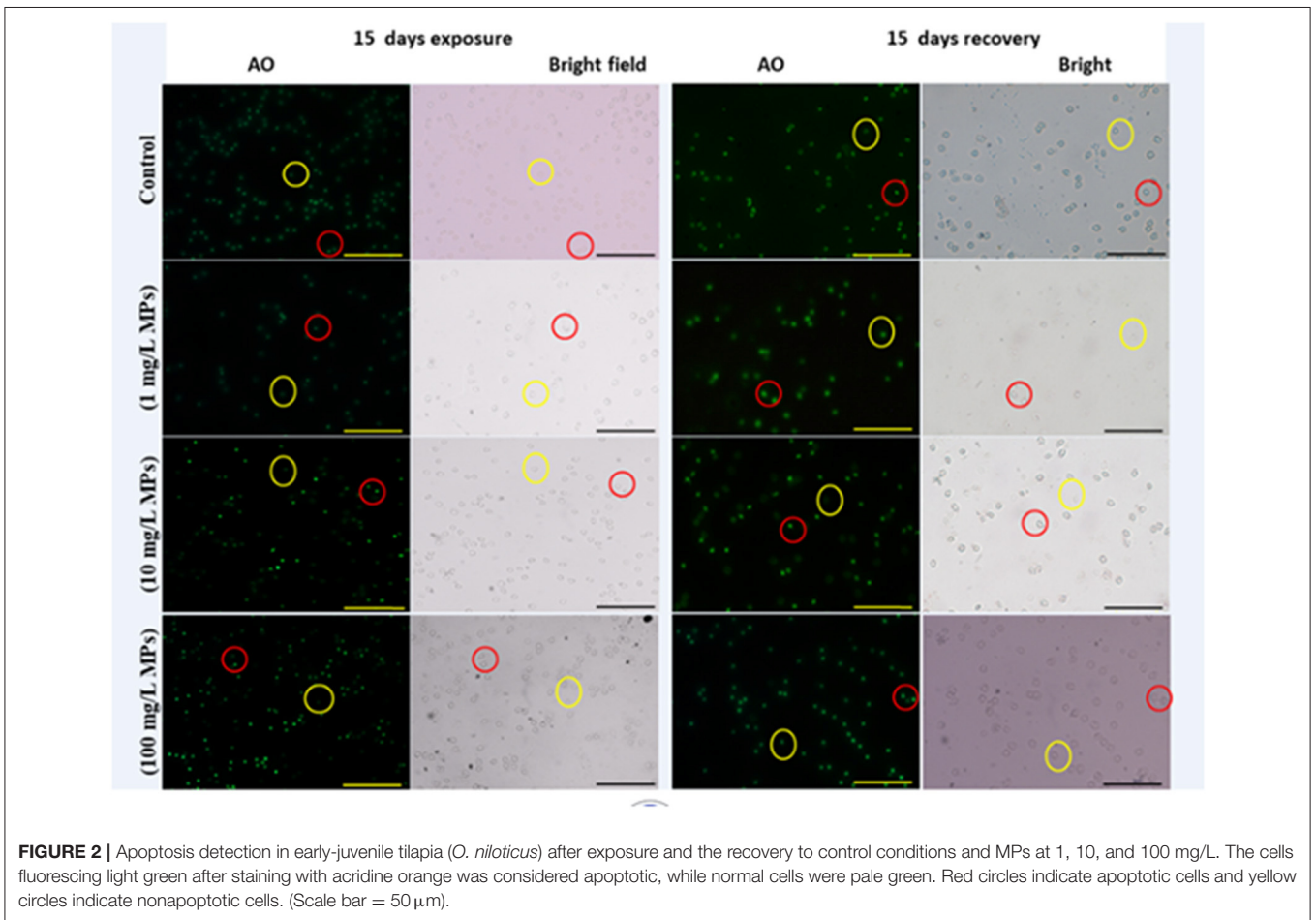
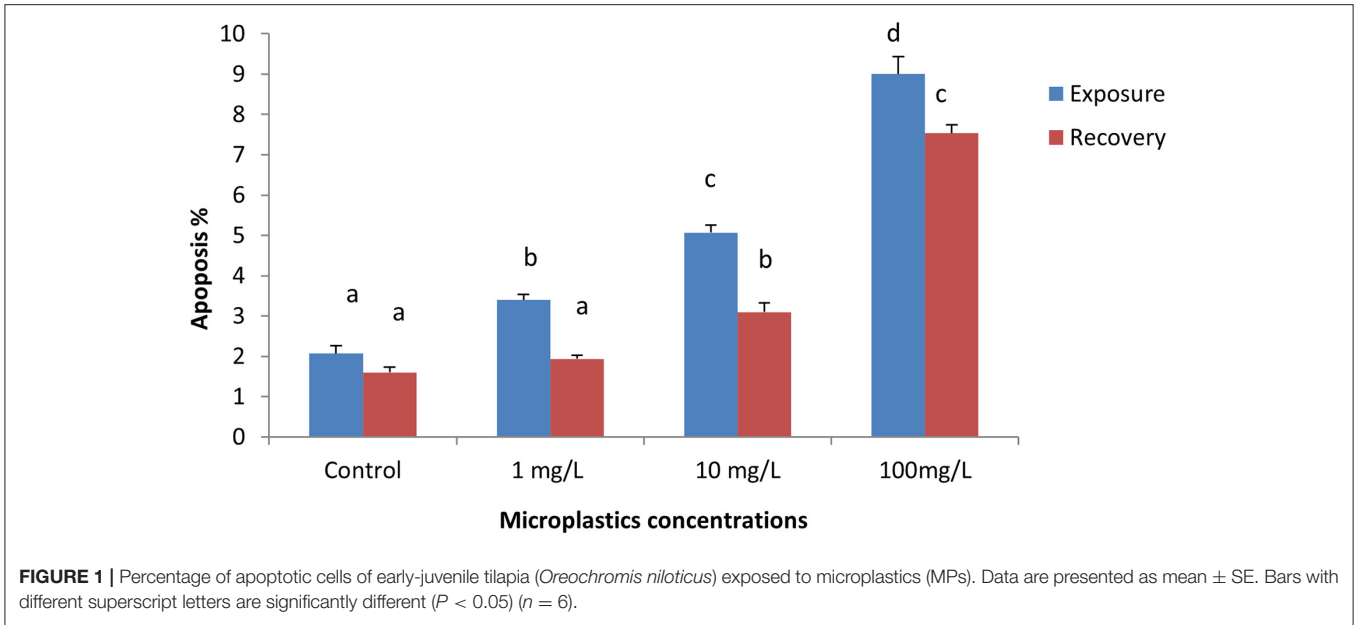
### MPs and Stock Preparation

Microplastics were virgin powder with irregular-shaped particles (more than 90% of MPs and 100 nm in size) purchased from Toxemerge Pty Ltd (Australia). MPs characterization occurred per the procedures described in Hamed et al. (2019a). The stock solution was prepared as per the manufacture protocol, stored at a temperature of 4°C in the dark. The stock solution (2.5 g MP/L) was mixed with purified water (Milli-Q). Then, it was sonicated before each use. A further diluted solution for the test concentrations was prepared from this stock right before the beginning of each experiment.

### Fish Exposure

Early juvenile male tilapia ( $n = 120$  fish, weight  $4.35 \pm 0.067$  g, total length  $3.28 \pm 0.12$  cm) was transported into aired tanks from the Aquaponic Unit at Al-Azhar University to the Laboratory of Fish Biology (Faculty of Science, Assiut University, Egypt). The fish (fed a commercial pellet diet with 3% of body weight per day) were acclimatized for 2 weeks in good-quality water within terms of parameters, such as conductivity ( $260.8 \mu\text{M/cm}$ ), pH (7.4), dissolved oxygen content (6.9 mg/L), temperature (28.5°C),  $\text{NH}_4$  (<1 mg/L),  $\text{NO}^{-3}$  (<1 mg/L), and  $\text{NO}^{-2}$  (<1 mg/L), and photoperiod (12:12 h light:dark). The fish were divided into four groups of 30 individuals each (10 per each replicate): a control group and three groups treated with MPs at concentrations of 1, 10, and 100 mg/L for 15 days followed by 15 days of recovery, according to a method described by Katzenberger and Thorpe (2015). During the exposure period, the water was renewed daily (40%) to reduce the impurities from metabolic wastes and MPs were re-dosed daily, while during the recovery period, the water was renewed daily to diminish the impurities from metabolic wastes without the dosing of MPs.

At the end of the experiment, six fish from each group were randomly handpicked and anesthetized using ice to lessen stress (Hamed et al., 2019a). The blood samples taken from the caudal vein for blood smears were used for eryptosis and erythron



**TABLE 1 |** Percentage of poikilocytosis cells (altered erythrocytes) of juvenile Nile tilapia (*O. niloticus*) after exposure for 15 days to microplastics (MPs) and recovery for 15 days.

	Exposure period				Recovery period			
	Control	1 mg/L of MPs	10 mg/L of MPs	100 mg/L of MPs	Control	1 mg/L of MPs	10 mg/L of MPs	100 mg/L of MPs
Hemolyzed cell	0.7 ± 0.2 <sup>a</sup>	1.7 ± 0.3 <sup>a</sup>	6.7 ± 0.3 <sup>b</sup>	21.7 ± 0.6 <sup>c</sup>	0.3 ± 0.2 <sup>a</sup>	2.2 ± 0.3 <sup>b</sup>	5 ± 0.4 <sup>c</sup>	12.8 ± 0.5 <sup>d</sup>
Sickle cell	1.2 ± 0.2 <sup>a</sup>	3.8 ± 0.6 <sup>b</sup>	6.8 ± 0.5 <sup>c</sup>	15.5 ± 0.4 <sup>d</sup>	0 ± 0 <sup>a</sup>	3 ± 0.4 <sup>b</sup>	3.8 ± 0.4 <sup>b</sup>	10.2 ± 0.5 <sup>c</sup>
Odd shapes	0 ± 0 <sup>a</sup>	4.3 ± 0.5 <sup>b</sup>	9.3 ± 0.7 <sup>c</sup>	23.2 ± 0.7 <sup>d</sup>	0 ± 0 <sup>a</sup>	2.8 ± 0.3 <sup>b</sup>	8.4 ± 0.4 <sup>c</sup>	14.7 ± 0.3 <sup>d</sup>
Schistocyte	0.7 ± 0.2 <sup>a</sup>	4.5 ± 0.4 <sup>b</sup>	6.5 ± 0.4 <sup>c</sup>	20.2 ± 0.7 <sup>d</sup>	0.7 ± 0.2 <sup>a</sup>	4 ± 0.6 <sup>b</sup>	6.6 ± 0.5 <sup>c</sup>	13.3 ± 0.5 <sup>d</sup>
Acanthocyte	1.2 ± 0.2 <sup>a</sup>	5.4 ± 1.0 <sup>b</sup>	10.8 ± 0.5 <sup>c</sup>	19.2 ± 0.7 <sup>d</sup>	0.7 ± 0.2 <sup>a</sup>	5 ± 0.3 <sup>b</sup>	8.8 ± 0.4 <sup>c</sup>	16.7 ± 0.6 <sup>d</sup>
Rouleaux	0.2 ± 0.17 <sup>a</sup>	2 ± 0.5 <sup>b</sup>	1.7 ± 0.3 <sup>b</sup>	6.3 ± 0.3 <sup>c</sup>	0 ± 0 <sup>a</sup>	1 ± 0.3 <sup>b</sup>	1.6 ± 0.2 <sup>b</sup>	4.1 ± 0.3 <sup>c</sup>
Teardrop shape	1.2 ± 0.2 <sup>a</sup>	5.3 ± 0.95 <sup>b</sup>	8.8 ± 0.5 <sup>c</sup>	19.2 ± 0.5 <sup>d</sup>	0.8 ± 0.3 <sup>a</sup>	4.3 ± 0.4 <sup>b</sup>	6.4 ± 0.5 <sup>c</sup>	14.2 ± 0.3 <sup>d</sup>
Filamented	0.5 ± 0.3 <sup>a</sup>	4.3 ± 0.6 <sup>b</sup>	4.3 ± 0.5 <sup>b</sup>	10.3 ± 0.7 <sup>c</sup>	0.5 ± 0.2 <sup>a</sup>	3.3 ± 0.5 <sup>b</sup>	6 ± 0.3 <sup>c</sup>	7.5 ± 0.4 <sup>d</sup>
Heinz bodies	0.2 ± 0.2 <sup>a</sup>	3.3 ± 0.5 <sup>b</sup>	6.2 ± 0.5 <sup>c</sup>	10.7 ± 0.4 <sup>d</sup>	0.5 ± 0.2 <sup>a</sup>	1.8 ± 0.3 <sup>b</sup>	3.8 ± 0.4 <sup>c</sup>	5.3 ± 0.4 <sup>d</sup>
Elliptocyte	1.3 ± 0.2 <sup>a</sup>	3.3 ± 0.3 <sup>b</sup>	6.2 ± 0.4 <sup>c</sup>	19.8 ± 0.8 <sup>d</sup>	0.7 ± 0.3 <sup>a</sup>	3.5 ± 0.4 <sup>b</sup>	7 ± 0.3 <sup>c</sup>	9.5 ± 0.4 <sup>d</sup>
Heart shaped	0 ± 0 <sup>a</sup>	0.5 ± 0.2 <sup>a</sup>	4.8 ± 0.9 <sup>b</sup>	13.5 ± 0.4 <sup>c</sup>	0.5 ± 0.2 <sup>a</sup>	0.8 ± 0.2 <sup>a</sup>	2 ± 0.3 <sup>b</sup>	8.3 ± 0.6 <sup>c</sup>
Eccentric nuclei	1.2 ± 0.2 <sup>a</sup>	5.2 ± 0.4 <sup>b</sup>	9 ± 0.6 <sup>c</sup>	12.7 ± 0.3 <sup>d</sup>	0.7 ± 0.3 <sup>a</sup>	2 ± 0.3 <sup>b</sup>	6.2 ± 0.4 <sup>c</sup>	10 ± 0.6 <sup>d</sup>
Crenated cell	1 ± 0.4 <sup>a</sup>	5 ± 0.4 <sup>b</sup>	16.2 ± 0.9 <sup>c</sup>	21 ± 0.6 <sup>d</sup>	0.8 ± 0.2 <sup>a</sup>	3.7 ± 0.3 <sup>b</sup>	9.4 ± 0.5 <sup>c</sup>	17 ± 0.4 <sup>d</sup>
Kidney shaped	0.7 ± 0.2 <sup>a</sup>	2.8 ± 0.6 <sup>b</sup>	4 ± 0.5 <sup>b</sup>	13 ± 0.5 <sup>c</sup>	0.2 ± 0.2 <sup>a</sup>	1.3 ± 0.2 <sup>b</sup>	3.8 ± 0.4 <sup>c</sup>	10 ± 0.6 <sup>d</sup>
Red cell agglutinate	0.3 ± 0.2 <sup>a</sup>	0.3 ± 0.2 <sup>a</sup>	1.7 ± 0.3 <sup>b</sup>	9.3 ± 0.4 <sup>c</sup>	0 ± 0 <sup>a</sup>	0.2 ± 0.2 <sup>a</sup>	2.2 ± 0.4 <sup>b</sup>	2.8 ± 0.3 <sup>b</sup>
Microcytic	1.3 ± 0.2 <sup>a</sup>	4.8 ± 0.5 <sup>b</sup>	10.3 ± 0.5 <sup>c</sup>	18.3 ± 0.5 <sup>d</sup>	0.3 ± 0.2 <sup>a</sup>	1.7 ± 0.2 <sup>b</sup>	7.4 ± 0.5 <sup>c</sup>	13.2 ± 0.5 <sup>d</sup>
Macrocytic	0.3 ± 0.2 <sup>a</sup>	0.8 ± 0.3 <sup>a</sup>	3.2 ± 0.4 <sup>b</sup>	12.3 ± 0.4 <sup>c</sup>	0.2 ± 0.2 <sup>a</sup>	0.3 ± 0.2 <sup>a</sup>	1.8 ± 0.3 <sup>b</sup>	6.5 ± 0.4 <sup>c</sup>

Data are represented as means ± SE. The values with different superscripted letters during the exposure period in the same row for each parameter are significantly different ( $P < 0.05$ ) ( $n = 6$ ).

The values with different superscripted letters during the recovery period in the same row for each parameter are significantly different ( $P < 0.05$ ) ( $n = 6$ ).

**TABLE 2 |** Percentage of nuclear abnormalities of red blood cells (RBCs) of juvenile Nile tilapia (*O. niloticus*) after exposure for 15 days to MPs and recovery for 15 days.

	Exposure period				Recovery period			
	Control	1 mg/L of MPs	10 mg/L of MPs	100 mg/L of MPs	Control	1 mg/L of MPs	10 mg/L of MPs	100 mg/L of MPs
Micronuclei	1.3 ± 0.2 <sup>a</sup>	5.8 ± 0.3 <sup>b</sup>	9.7 ± 0.4 <sup>c</sup>	15.2 ± 0.4 <sup>d</sup>	1.7 ± 0.2 <sup>a</sup>	4.7 ± 0.3 <sup>b</sup>	6.8 ± 0.3 <sup>c</sup>	9.7 ± 0.3 <sup>d</sup>
Binucleated	1 ± 0.3 <sup>a</sup>	2.3 ± 0.3 <sup>b</sup>	4.8 ± 0.5 <sup>c</sup>	12.3 ± 0.5 <sup>d</sup>	0.7 ± 0.2 <sup>a</sup>	2 ± 0.4 <sup>b</sup>	3.8 ± 0.3 <sup>c</sup>	7.2 ± 0.3 <sup>d</sup>
Blebbled nuclei	0.3 ± 0.2 <sup>a</sup>	1.8 ± 0.3 <sup>b</sup>	4.2 ± 0.3 <sup>c</sup>	8 ± 0.4 <sup>d</sup>	0.3 ± 0.2 <sup>a</sup>	0.8 ± 0.3 <sup>a</sup>	2.3 ± 0.3 <sup>b</sup>	5.2 ± 0.3 <sup>c</sup>
Notched nuclei	0.3 ± 0.2 <sup>a</sup>	2.2 ± 0.5 <sup>b</sup>	6.2 ± 0.5 <sup>c</sup>	13.2 ± 0.3 <sup>d</sup>	0.3 ± 0.2 <sup>a</sup>	1.5 ± 0.2 <sup>b</sup>	2.3 ± 0.2 <sup>c</sup>	6 ± 0.5 <sup>d</sup>
Lobed nuclei	0.3 ± 0.2 <sup>a</sup>	1.8 ± 0.3 <sup>b</sup>	3.8 ± 0.3 <sup>c</sup>	8 ± 0.4 <sup>d</sup>	0.3 ± 0.2 <sup>a</sup>	1.3 ± 0.3 <sup>b</sup>	2.5 ± 0.2 <sup>c</sup>	4.2 ± 0.3 <sup>d</sup>
Hemolyzed nuclei	1.2 ± 0.2 <sup>a</sup>	2. ± 0.3 <sup>a</sup>	5.3 ± 0.4 <sup>b</sup>	13.2 ± 0.6 <sup>c</sup>	0.7 ± 0.2 <sup>a</sup>	1.7 ± 0.2 <sup>b</sup>	3.3 ± 0.3 <sup>c</sup>	8 ± 0.4 <sup>d</sup>

Data are represented as means ± SE. The values with different superscripted letters during the exposure period in the same row for each parameter are significantly different ( $P < 0.05$ ) ( $n = 6$ ).

The values with different superscripted letters during the recovery period in the same row for each parameter are significantly different ( $P < 0.05$ ) ( $n = 6$ ).

profiles (to measure poikilocytosis and nuclear abnormalities). The experimental setup and fish handling were approved by the Research Ethics Committee of the Faculty of Science, Assiut University, Assiut, Egypt.

### Erythrocyte Programmed Cell Death (Eryptosis)

Eryptosis was detected following the procedures described by Sayed et al. (2016). In brief, the blood smears were prepared and stained with acridine orange, and RBCs were then observed under a fluorescence microscope (Zeiss Axioplan2) equipped with a digital 3 CCD color video camera (Sony, AVT-Horn, Japan).

### Erythron Profiles (Poikilocytosis and Nuclear Abnormalities of RBCs)

The blood smears were assorted, stained with hematoxylin and eosin, handpicked, classified with codes, randomized, and scored blindly for erythrocyte alterations and nuclear abnormalities. For every group, 10,000 cells (1,000 per slide minimum) were analyzed at 100× magnification for MN and altered RBCs in the varying groups. The criteria of observation for MN counts were previously detailed by Al-Sabti and Metcalfe (1995) and Schmid (1975).

### Statistical Analysis

All data were assessed in SPSS (SPSS 1998, SPSS Inc., Chicago, IL, USA) at  $p < 0.05$ . Data were subjected to a Shapiro–Wilk



test for normality and Levene's test for homogeneity of variances utilizing the one-way ANOVA test. Should there be variance equality, Fisher's least significant difference *post-hoc* test was employed to crosscheck the treated groups against the control group. Should there be variance inequality, Dunnett's *post-hoc* test instead employed to crosscheck the treated groups against the control group.

## RESULTS

### Erythrocyte Programmed Cell Death (Eryptosis)

The percentage of apoptotic cells increased significantly in the MPs-exposed groups compared with the control group in a concentration-dependent manner (**Figure 1**). RBCs of the control group appeared normal with pale-green nuclei (**Figure 2**), while those of the MPs-exposed groups appeared apoptotic with luminous light-green nuclei (**Figure 2**). After the recovery period, a higher percentage of apoptotic cells was observed in the MPs-treated groups compared with the control group in a concentration-dependent manner (**Figures 1, 2**).

### Erythron Profiles (Poikilocytosis and Nuclear Abnormalities of RBCs)

Exposure to 1, 10, and 100 mg/L of MPs was associated with a significant rise in the percentage of poikilocytosis and nuclear abnormalities of RBCs compared with the control group in a concentration-dependent manner (**Tables 1, 2**). The blood smears of the control group contained normal erythrocytes ellipsoidal in shape with a centrally located ellipsoidal nucleus (**Figure 3A**). The blood smears of the fish exposed to MPs at 1 mg/L displayed poikilocytosis of the erythrocytes (**Figure 3B**) in the form of teardrop-like cells (Tr) with pointed apices, sickle cells (Sk), elongated crescent-shaped RBCs, and cells with eccentric nuclei (deviating or departing from the center of the cell). The alterations of RBCs in fish exposed to MPs at 10 mg/L (**Figures 3C,D**) included Tr, sickle cells, eccentric nuclei, filamented cells, acanthocytes (Ac), spiculated red cells with a few projections of varying sizes and surface distributions, kidney-shaped cells (Kn), elliptocytes (slightly oval to cigar-shaped with blunt ends), and crenated cells (Cr), spherical red cells covered with short and pointed projections. In addition, the alterations of RBCs in the fish exposed to MPs at 100 mg/L (**Figures 3E,F**) included eccentric nuclei, filamented cells, sickle cells, Cr, Kn, elliptocytes, Ac, heart-shaped cells, and schistocytes (fragmented RBCs).

The major nuclear abnormalities of RBCs in fish exposed to MPs at 1 mg/L (**Figures 4B,C**) showed micronuclei (Mn), a round cytoplasmic inclusion with a diameter one-tenth to one-third that of the primary nucleus, and binuclei (Bin). The alterations of the nuclei of RBCs in fish exposed to MPs at 10 mg/L (**Figures 4D–G**) included Mn, Bin, Kn, notched Nuclei (Non), and the cells with clear slits that extend well into the nuclear envelope. Moreover, the alterations of nuclei of RBCs in the fish exposed to MPs at 100 mg/L (**Figures 4H–L**) included Non, irregular nuclei (Irn), elongated nuclei (En), Kn, Bin, lobed

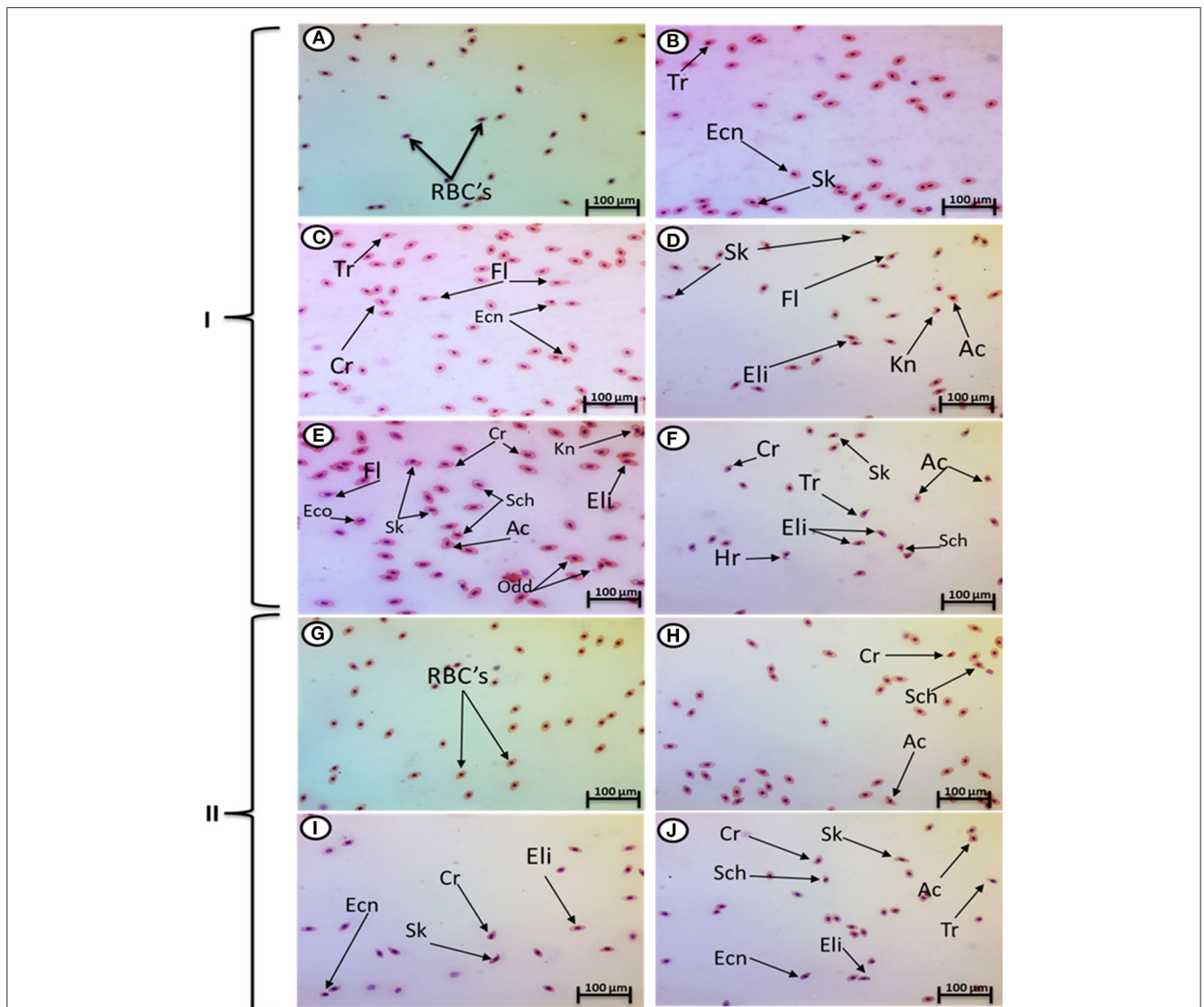
nuclei (Lon), and large evaginations of the nuclear envelope with no clear shape or definition. After the recovery period, a higher percentage of poikilocytosis and nuclear abnormalities of RBCs were found in the MP-treated groups compared with the control group in a concentration-dependent manner (**Tables 1, 2** and **Figures 3G–J, 4A–H**).

## DISCUSSION

Microplastics influenced a significant surge in apoptotic cells percentages, which is in accordance with a report by Gökçe et al. (2018), who observed that polyvinyl chloride (PVC) MPs induced apoptosis in zebrafish (*Danio rerio*) embryos in a dose-dependent manner. Moreover, Xu et al. (2019) reported that the two types of polystyrene nanoparticles (PS-NP25 and PS-NP70) caused an immense increase in apoptosis in human lung epithelial cells. Qu et al. (2019) found that the amino-modified nano polystyrene enhanced the stimulation of apoptosis compared with pristine nano polystyrene in germlines of the nematode *Caenorhabditis elegans*. Sökmen et al. (2020) reported that 20 nm polystyrene plastics can bioaccumulate brain tissue and cause apoptosis in zebrafish embryos.

Several studies reported the existence of mechanisms for apoptosis induction in cells, but details of these mechanisms are not clear. For example, Yirong et al. (2020) found that di (2-ethylhexyl) phthalate can induce apoptosis by stimulating an oxidative burst of neutrophils in carp (*Cyprinus carpio*). This apoptosis was explained by Walpitagama et al. (2019), who supported the assumptions of reactive oxygen species (ROS)-prompted apoptosis demonstrated in our previous study (Hamed et al., 2020). In addition, polystyrene upregulated the expression levels of pro-apoptotic proteins, which are strictly linked to the commencement of cell apoptosis (Xu et al., 2019; Yirong et al., 2020). In the present study, we observed that most of the characteristics of apoptosis reported in the previous studies, such as RBC shrinkage, nuclear and DNA damage, and cell-membrane alteration in erythrocytes (Lang et al., 1998; Bortner and Cidrowski, 1999, 2004; Maeno et al., 2000; Okada et al., 2001; Yu et al., 2001; Javadov and Karmazyn, 2007). These results indicate that the mechanism responsible for apoptosis in erythrocytes may involve alterations of cell components, especially ion channels (Föller et al., 2008). On the other hand, no significant differences were evident between PVC and Mater-Bi micro-debris-exposed groups and control groups concerning apoptosis in *Dreissena polymorpha* (Magni et al., 2020).

In this study, MPs influenced an immense surge in the percentage and nuclear abnormalities of poikilocytosis cells in RBCs compared with the control group. To the best of our knowledge, this is the first study to illustrate the effect of MPs on the poikilocytosis of RBCs. We concluded that MPs or their additives interact with erythrocytes and limit the dehydrogenase of delta-aminolevulinic acid and induce the interruption of plasma membranes, producing abnormally shaped blood cells (poikilocytosis) (Hamed et al., 2019b). The higher production of ROS in erythrocytes could offer a reasonable explanation for this phenomenon; it could be caused by the direct interaction



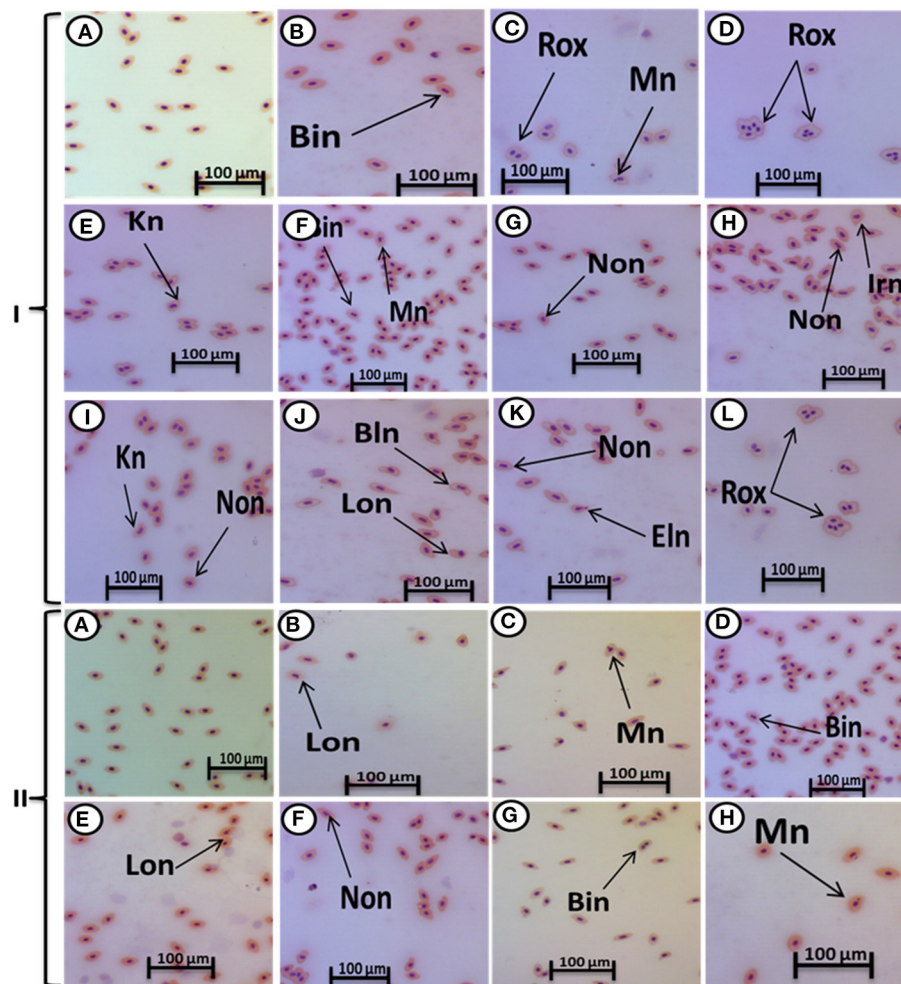
**FIGURE 3** | Hematoxylin and eosin-stained blood smears from juvenile *O. niloticus* showing I; MP exposure: normal erythrocytes (A), deformed ones after exposure to MPs at 1 mg/L for 15 days (B), deformed ones after exposure to MPs at 10 mg/L for 15 days (C,D), and deformed ones after exposure to MPs at 100 mg/L for 15 days (E,F) and II; recovery period: normal erythrocytes (G), deformed ones after recovery from exposure to MPs at 1 mg/L for 15 days (H), deformed ones after recovery from exposure to MPs at 10 mg/L for 15 days (I), and deformed ones after recovery from exposure to MPs at 100 mg/L for 15 days (J). Tr, tear-drop cell; Cr, crenated cell; Ac, acanthocyte; Ecn, eccentric nucleus; Kn, kidney shape; Hr, heart shape; Eli, elliptocytes; odd, odd shape; FL, filamented; Sch, schistocytic and Sk, sickle cell.

between MPs and erythrocyte plasma membranes (da Costa Araújo et al., 2020a).

In our results, the nuclear abnormalities of RBCs included Mn, binucleated erythrocytes, notched, lobed, blebbed, and hemolyzed nuclei. A higher percentage of nuclear abnormalities (and Mn in particular) were observed in *Mytilus galloprovincialis* mussels after exposure to the polyethylene and polystyrene MPs (Avio et al., 2015) and benzo(a)pyrene and/or benzo(a)pyrene-contaminated low-density polyethylene (Pittura et al., 2018). Polyethylene MPs induced nuclear anomalies, e.g., Mn, NoN, blebbed, binucleated or multilobed nuclei, kidney-shaped nuclei, eccentric nuclei, and nuclear vacuoles in erythrocytes of *Danio*

*rerio* after feeding on fry of *Poecilia reticulata* as well as in erythrocytes of *Physalaemus cuvieri* tadpoles (da Costa Araújo et al., 2020b).

The nuclear anomalies may be a product of aneuploidy or disruption of cytokinesis (da Costa Araújo et al., 2020a) as well as the elimination of amplified genes (Crott et al., 2001). Strunjak-Perovic et al. (2009) reported that the nuclear anomalies may result from gene mutations in the nuclear lamina of the nucleus. Moreover, da Costa Araújo et al. (2020a) found that the genotoxicity of polyethylene MPs on erythrocytes was indirectly induced by the formation of free radicals that interfere with DNA integrity, and our previous study linking oxidative stress in



**FIGURE 4 |** Hematoxylin and eosin-stained blood smears from juvenile (*O. niloticus*) showing nuclear abnormalities, I; MP exposure: normal erythrocytes (A), deformed ones after exposure to MPs at 1 mg/L for 15 days (B,C), deformed ones after exposure to MPs at 10 mg/L for 15 days (D–G), and deformed ones after exposure to MPs at 100 mg/L for 15 days (H–L) and II; recovery period: normal erythrocytes after recovery from exposure to MPs for 15 days (A), the deformed ones after recovery from exposure to MPs at 1 mg/L for 15 days (B), deformed ones after recovery from exposure to MPs at 10 mg/L for 15 days (C–E), and deformed ones after recovery from exposure to MPs at 100 mg/L for 15 days (F–H). Bn, blebbed nucleus; En, elongated nucleus; Non, notched nucleus; Mn, micronucleus; Bin, bionuclei; Kn, kidney-shaped nucleus; Irn, irregular nucleus; Lon, lobed nucleus; Rox, rouleaux shape; Eln, elongated nucleus.

tilapia with exposure to MPs (Hamed et al., 2020) reinforces this conclusion. On the contrary, no significant differences between PVC and Mater-Bi micro-debris-exposed groups and the control group were found in *Dreissena polymorpha* Mn (Magni et al., 2020).

After the recovery period, a higher percentage of apoptotic cells and poikilocytosis and nuclear abnormalities of RBCs were still observed in the MP-treated groups in a dose-dependent manner. This could be attributed to the fact that apoptosis is generally believed to be irreversible (Tang et al., 2012).

In addition, Gagnaire et al. (2013) showed that the depleted uranium surged ROS production in adults and larvae even at the low concentrations and the depuration period for adult *D. rerio*. In addition, Stankevičiute et al. (2016) found that Mn incidences were found to immensely surge after a 4-day recovery in all

the tissues. Blood erythrocytes maintained noticeable recovery of all analyzed geno-cytotoxicity endpoints after 8 and 12 days in the 0.25 group; however, the recovery was observed after 12 days of depuration in the liver and kidney erythrocytes. Martins et al. (2018) relayed that the inflammation and formation of hyperplastic foci in fish epithelia showed slower recovery and unexpected increased expression of a *ras* family oncogene homolog following depuration. Removing the chemical intruders could enhance tissue recovery, though it does not completely clear the molecular and histopathological endpoints commonly related to neoplasia. Détrée and Gallardo-Escárate (2018) stated that during the period of recovery, a contrasting response was found with the activation of apoptotic processes and the upregulation of immune receptors and stress-related proteins (glutathione peroxidase, hsp70) in mussels previously exposed



to MPs. This states that the physiological stress and physical damages induced by MPs persist after an event of exposure. A 30-day depuration period induces a tendency to recover initial values of biomarkers determinations in the fish liver of sea bream after long-term exposure to virgin and seawater affected by a MPs enriched-diet. Yet, it seems to suffice for its complete normalization (Capó et al., 2021).

In contrast, Solomando et al. (2021) surmised that the liver and blood biomarkers of *Sparus aurata* managed to recover during the period of depuration, with the majority reaching levels close to those of the control group after long-term exposure to MPs. In addition, Iheanacho and Odo (2020) observed that *Clarias gariepinus* fed a diet containing PVC particles (0, 0.5, 1.5, and 3.0%) for 30 days, presented a reduction in erythrocyte mean cell volume, neutrophil counts, cell hemoglobin, and induced oxidative stress in brain and gills. For 45 days, a depuration period allowed the recovery of most of the analyzed parameters.

## CONCLUSION

The fish RBCs can be sensitive and reliable biomarkers for the measurement of the cytotoxic and genotoxic effects of MPs. Exposure of tilapia to MPs induces apoptosis, poikilocytosis, and nuclear abnormalities of RBCs. Following the period of recovery, almost all the detected variances were still evident in

the MP-treated groups in a dose-dependent manner. This could be attributed to the fact that these changes are generally believed to be irreversible or the necessity of a longer recovery period for the fish.

## DATA AVAILABILITY STATEMENT

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

## ETHICS STATEMENT

The animal study was reviewed and approved by Assiut University. Experimental setup and fish handling were approved within the research by the Ethical Committee of the Faculty of Science, Al-Azhar University, Assiut Branch, Egypt. This study does not contain human or plant subjects, and all applicable international, national, and/or institutional guidelines for the care and use of animals were adhered to.

## AUTHOR CONTRIBUTIONS

MH, AO, HS, and AE-DS: experimental design. MH, HS, and AE-DS: data interpretation. MH: Biochemical analysis. MH, AO, AB, HS, and AE-DS: writing and revision. All authors have read and approved the final manuscript.

## REFERENCES

- Al-Sabti, K., and Metcalfe, C. D. (1995). Fish micronuclei for assessing genotoxicity in water. *Mutat. Res.* 343, 121–135. doi: 10.1016/0165-1218(95)90078-0
- Avio, C. G., Gorbi, S., Milan, M., Benedetti, M., Fattorini, D., d'Errico, G., et al. (2015). Pollutants bioavailability and toxicological risk from microplastics to marine mussels. *Environ. Pollut.* 198, 211–222. doi: 10.1016/j.envpol.2014.12.021
- Barboza, L. G. A., Lopes, C., Oliveira, P., Bessa, F., Otero, V., Henriques, B., et al. (2020). Microplastics in wild fish from North East Atlantic Ocean and its potential for causing neurotoxic effects, lipid oxidative damage, and human health risks associated with ingestion exposure. *Sci. Total Environ.* 717:134625. doi: 10.1016/j.scitotenv.2019.134625
- Bortner, C. D., and Cidlowski, J. A. (1999). Caspase independent/dependent regulation of K(+), cell shrinkage, and mitochondrial membrane potential during lymphocyte apoptosis. *J. Biol. Chem.* 274, 21953–21962. doi: 10.1074/jbc.274.31.21953
- Bortner, C. D., and Cidlowski, J. A. (2004). The role of apoptotic volume decrease and ionic homeostasis in the activation and repression of apoptosis. *Pflugers Arch.* 448, 313–318. doi: 10.1007/s00424-004-1266-5
- Capó, X., Company, J. J., Alomar, C., Compa, M., Sureda, A., Grau, A., et al. (2021). Long-term exposure to virgin and seawater exposed microplastic enriched-diet causes liver oxidative stress and inflammation in gilthead seabream *Sparus aurata*, Linnaeus 1758. *Sci. Total Environ.* 767:144976. doi: 10.1016/j.scitotenv.2021.144976
- Crott, J. W., Mashiyama, S. T., Ames, B. N., and Fenech, M. (2001). The effect of folic acid deficiency and MTHFR C677T polymorphism on chromosome damage in human lymphocytes *in vitro*. *Cancer Epidemiol. Biomarkers Prev.* 10, 1089–1096.
- da Costa Araújo, A. P., de Andrade Vieira, J. E., and Malafaia, G. (2020a). Toxicity and trophic transfer of polyethylene microplastics from *Poecilia reticulata* to *Danio rerio*. *Sci. Total Environ.* 742:140217. doi: 10.1016/j.scitotenv.2020.140217
- da Costa Araújo, A. P., de Melo, N. F. S., de Oliveira Junior, A. G., Rodrigues, F. P., Fernandes, T., de Andrade Vieira, J. E., et al. (2020b). How much are microplastics harmful to the health of amphibians? A study with pristine polyethylene microplastics and *Physalaema cuvieri*. *J. Hazard Mater.* 382, 121066. doi: 10.1016/j.jhazmat.2019.121066
- Détrée, C., and Gallardo-Escárate, C. (2018). Single and repetitive microplastics exposures induce immune system modulation and homeostasis alteration in the edible mussel *Mytilus galloprovincialis*. *Fish Shellfish Immunol.* 83, 52–60. doi: 10.1016/j.fsi.2018.09.018
- Edsall, C. C. (1999). A blood chemistry profile for lake trout. *J. Aquat. Anim. Health.* 11, 81–86.
- Eriksen, M., Mason, S., Wilson, S., Box, C., Zellers, A., Edwards, W., et al. (2013). Microplastic pollution in the surface waters of the Laurentian great lakes. *Mar. Pollut. Bull.* 77, 177–182. doi: 10.1016/j.marpolbul.2013.10.007
- Fazio, F. (2019). Fish hematology analysis as an important tool of aquaculture: a review. *Aquaculture* 500, 237–242. doi: 10.1016/j.aquaculture.2018.10.030
- Ferreira, P., Fonte, E., Soares, M. E., Carvalho, F., and Guilhermino, L. (2016). Effects of multi-stressors on juveniles of the marine fish *Pomatoschistus microps*: gold nanoparticles, microplastics and temperature. *Aquat. Toxicol.* 170, 89–103. doi: 10.1016/j.aquatox.2015.11.011
- Fischer, E. K., Paglialonga, L., Czech, E., and Tamminga, M. (2016). Microplastic pollution in lakes and lake shoreline sediments-A case study on lake Bolsena and lake Chiusi (central Italy). *Environ. Pollut.* 213, 648–657. doi: 10.1016/j.envpol.2016.03.012
- Föllner, M., Huber, S. M., and Lang, F. (2008). Erythrocyte programmed cell death. *IUBMB Life.* 60, 661–668. doi: 10.1002/iub.106
- Free, C. M., Jensen, O. P., Mason, S. A., Eriksen, M., Williamson, N. J., and Boldgiv, B. (2014). High-levels of microplastic pollution in a large, remote, mountain lake. *Mar. Pollut. Bull.* 85, 156–163. doi: 10.1016/j.marpolbul.2014.06.001
- Gagnaire, B., Cavalie, I., Camilleri, V., and Adam-Guillermin, C. (2013). Effects of depleted uranium on oxidative stress, detoxification, and defence parameters of zebrafish *Danio rerio*. *Arch. Environ. Contam. Toxicol.* 64, 140–150. doi: 10.1007/s00244-012-9814-z



- Gökçe, B., Önen, Ö., Kiran, M. S., and Ucuncu, S. I. (2018). The effects of polyvinyl chloride microparticles on the apoptotic processes in zebrafish (*Danio rerio*) embryos. *Int. J. Sci. Technol. Res.* 4, 156–161.
- Hamed, M., Soliman, H. A. M., Osman, A. G. M., and Sayed, A. E. H. (2019a). Assessment of the effect of exposure to microplastics in Nile Tilapia (*Oreochromis niloticus*) early juvenile: I. blood biomarkers. *Chemosphere* 228, 345–350. doi: 10.1016/j.chemosphere.2019.04.153
- Hamed, M., Soliman, H. A. M., Osman, A. G. M., and Sayed, A. E. H. (2020). Antioxidants and molecular damage in Nile Tilapia (*Oreochromis niloticus*) after exposure to microplastics. *Environ. Sci. Pollut. Res. Int.* 27, 14581–14588. doi: 10.1007/s11356-020-07898-y
- Hamed, M., Soliman, H. A. M., and Sayed, A. E. H. (2019b). Ameliorative effect of *Spirulina platensis* against lead nitrate-induced cytotoxicity and genotoxicity in catfish *Clarias gariepinus*. *Environ. Sci. Pollut. Res. Int.* 26, 20610–20618. doi: 10.1007/s11356-019-05319-3
- Iheanacho, S. C., and Odo, G. E. (2020). Neurotoxicity, oxidative stress biomarkers and haematological responses in African catfish (*Clarias gariepinus*) exposed to polyvinyl chloride microparticles. *Comp. Biochem. Physiol. Part C Toxicol. Pharmacol.* 232:108741. doi: 10.1016/j.cbpc.2020.108741
- Javadov, S., and Karmazyn, M. (2007). Mitochondrial permeability transition pore opening as an endpoint to initiate cell death and as a putative target for cardioprotection. *Cell. Physiol. Biochem.* 20, 1–22. doi: 10.1159/000103747
- Karbalaeei, S., Hanachi, P., Walker, T. R., and Cole, M. (2018). Occurrence, sources, human health impacts and mitigation of microplastic pollution. *Environ. Sci. Pollut. Res. Int.* 25, 36046–36063. doi: 10.1007/s11356-018-3508-7
- Katzenberger, T. D., and Thorpe, K. L. (2015). *Assessing the Impact of Exposure to Microplastics in Fish*. Evidence Report - SC120056. Bristol: Environ Agency.
- Kim, J. H., Kim, S. K., and Hur, Y. B. (2019). Hematological parameters and antioxidant responses in olive flounder *Paralichthys olivaceus* in biofloc depend on water temperature. *J. Therm. Biol.* 82, 206–212. doi: 10.1016/j.jtherbio.2019.04.013
- Kim, J. H., Sohn, S., Kim, S. K., and Hur, Y. B. (2020). Effects on hematological parameters, antioxidant and immune responses, AChE, and stress indicators of olive flounders, *Paralichthys olivaceus*, raised in bio-floc and seawater challenged by *Edwardsiella tarda*. *Fish Shellfish Immunol.* 97, 194–203. doi: 10.1016/j.fsi.2019.12.011
- Lang, F., Busch, G. L., Ritter, M., Völkl, H., Waldegger, S., Gulbins, E., et al. (1998). Functional significance of cell volume regulatory mechanisms. *Physiol. Rev.* 78, 247–306. doi: 10.1152/physrev.1998.78.1.247
- Lasee, S., Mauricio, J., Thompson, W. A., Karnjanapiboonwong, A., Kasumba, J., Subbiah, S., et al. (2017). Microplastics in a freshwater environment receiving treated wastewater effluent. *Integr. Environ. Assess. Manage.* 13, 528–532. doi: 10.1002/ieam.1915
- Luskova, V., Svoboda, M., and Kolářová, J. (2002). Effect of diazinon on blood plasma biochemistry in carp (*Cyprinus carpio* L.). *Acta Veterinaria Brno.* 71, 117–123.
- Ma, H., Pu, S., Liu, S., Bai, Y., Mandal, S., and Xing, B. (2020). Microplastics in aquatic environments: toxicity to trigger ecological consequences. *Environ. Pollut.* 261:114089. doi: 10.1016/j.envpol.2020.114089
- Maeno, E., Ishizaki, Y., Kanaseki, T., Hazama, A., and Okada, Y. (2000). Normotonic cell shrinkage because of disordered volume regulation is an early prerequisite to apoptosis. *Proc. Natl. Acad. Sci. U.S.A.* 97, 9487–9492. doi: 10.1073/pnas.140216197
- Magni, S., Bonasoro, F., Della Torre, C., Parenti, C. C., Maggioni, D., and Binelli, A. (2020). Plastics and biodegradable plastics: ecotoxicity comparison between polyvinylchloride and Mater-Bi® micro-debris in a freshwater biological model. *Sci. Total Environ.* 720:137602. doi: 10.1016/j.scitotenv.2020.137602
- Martins, M., Silva, A., Costa, M. H., Miguel, C., and Costa, P. M. (2018). Co-exposure to environmental carcinogens *in vivo* induces neoplasia-related hallmarks in low-genotoxicity events, even after removal of insult. *Sci. Rep.* 8:3649. doi: 10.1038/s41598-018-21975-w
- McCormick, A., Hoellein, T. J., Mason, S. A., Schluep, J., and Kelly, J. J. (2014). Microplastic is an abundant and distinct microbial habitat in an urban river. *Environ. Sci. Technol.* 48, 11863–11871. doi: 10.1021/es503610r
- Mekawy, I. A., Mahmoud, U. M., and Sayed, A. E. H. (2011). Effects of 4-nonylphenol on blood cells of the African catfish *Clarias gariepinus* (Burchell, 1822). *Tissue Cell* 43, 223–229. doi: 10.1016/j.tice.2011.03.006
- Morina, V., Aliko, V., Gavrazaj, F., and Kastrati, D. (2012). Use of blood parameters as biomarkers of contaminant exposure in fish specimens from Sitnica river, Kosovo. *J. Int. Environ. Appl. Sci.* 7, 971–977.
- Okada, Y., Maeno, E., Shimizu, T., Dezaki, K., Wang, J., and Morishima, S. (2001). Receptor-mediated control of regulatory volume decrease (RVD) and apoptotic volume decrease (AVD). *J. Physiol.* 532, 3–16. doi: 10.1111/j.1469-7793.2001.0003g.x
- Pedà, C., Caccamo, L., Fossi, M., Gai, F., Andaloro, F., Genovese, L., et al. (2016). Intestinal alterations in European sea bass *Dicentrarchus labrax* (Linnaeus, 1758) exposed to microplastics: preliminary results. *Environ. Pollut.* 212, 251–225. doi: 10.1016/j.envpol.2016.01.083
- Pittura, L., Avio, C. G., Giuliani, M. E., d'Errico, G., Keiter, S. H., Cormier, B., et al. (2018). Microplastics as vehicles of environmental PAHs to marine organisms: combined chemical and physical hazards to the mediterranean mussels, *Mytilus galloprovincialis*. *Front. Mar. Sci.* 5:103. doi: 10.3389/fmars.2018.00103
- PlasticsEurope. (2015). Plastics - the facts 2014/2015: An analysis of European plastics production, demand and waste data. *PlasticsEurope.* 1–34. doi: 10.1016/j.marpolbul.2013.01.015
- Qu, M., Qiu, Y., Kong, Y., and Wang, D. (2019). Amino modification enhances reproductive toxicity of nanopolystyrene on gonad development and reproductive capacity in nematode *Caenorhabditis elegans*. *Environ. Pollut.* 254:112978. doi: 10.1016/j.envpol.2019.112978
- Rainieri, S., Conledo, N., Larsen, B. K., Granby, K., and Barranco, A. (2018). Combined effects of microplastics and chemical contaminants on the organ toxicity of zebrafish (*Danio rerio*). *Environ. Res.* 162, 135–143. doi: 10.1016/j.envres.2017.12.019
- Rochman, C., Hoh, E., Kurobe, T., and Teh, S. (2013). Ingested plastic transfers hazardous chemicals to fish and induces hepatic stress. *Sci. Rep.* 3:3263. doi: 10.1038/srep03263
- Sayed, A. E. D. H., Elbaghdady, H. A. M., and Zahran, E. (2015). Arsenic-induced genotoxicity in Nile tilapia (*Oreochromis niloticus*); the role of *Spirulina platensis* extract. *Environ. Monit. Assess.* 187:751. doi: 10.1007/s10661-015-4983-7
- Sayed, A. E. D. H., El-Sayed, Y. S., and Ali, H. (2017). Hepatoprotective efficacy of *Spirulina platensis* against lead-induced oxidative stress and genotoxicity in catfish; *Clarias gariepinus*. *Ecotoxicol. Environ. Saf.* 143, 344–350. doi: 10.1016/j.ecoenv.2017.05.013
- Sayed, A. H. (2016). Genotoxicity detection following exposure to silver nanoparticles in African catfish (*Clarias gariepinus*). *Int. J. Nanoparticles* 9, 1–8. doi: 10.1504/IJNP.2016.078531
- Sayed, A. H., Abdel-Tawab, H. S., Abdel Hakeem, S. S., and Mekawy, I. A. (2013). The protective role of quince leaf extract against the adverse impacts of ultraviolet-A radiation on some tissues of *Clarias gariepinus* (Burchell, 1822). *J. Photochem. Photobiol. B Biol.* 119, 9–14. doi: 10.1016/j.jphotobiol.2012.11.006
- Sayed, A. H., Watanabe-Asaka, T., Oda, S., and Mitani, H. (2016). Apoptosis and morphological alterations after UVA irradiation in red blood cells of p53 deficient Japanese medaka (*Oryzias latipes*). *J. Photochem. Photobiol. B Biol.* 161, 1–8. doi: 10.1016/j.jphotobiol.2016.05.004
- Scanes, E., Wood, H., and Ross, P. (2019). Microplastics detected in haemolymph of the Sydney rock oyster *Saccostrea glomerata*. *Mar. Pollut. Bull.* 149:110537. doi: 10.1016/j.marpolbul.2019.110537
- Schmid, W. (1975). The micronucleus test. *Mutat. Res.* 31, 9–15. doi: 10.1016/0165-1161(75)90058-8
- Sökmen, T., Sulukan, E., Türkoglu, M., Baran, A., and Özkaraca, M., Ceyhun, S.B. (2020). Polystyrene nanoplastics (20 nm) are able to bioaccumulate and cause oxidative DNA damages in the brain tissue of zebrafish embryo (*Danio rerio*). *Neurotoxicology* 77, 51–59. doi: 10.1016/j.neuro.2019.12.010
- Soliman, H. A. M. (2017). Protein analysis and heavy metals accumulation in muscles of wild and farmed Nile Tilapia (*Oreochromis niloticus*). *Environ. Res.* 20, 1–10.
- Soliman, H. A. M., and Sayed, A. E.-D. H. (2020). Poikilocytosis and tissue damage as negative impacts of tramadol on juvenile of Tilapia (*Oreochromis niloticus*). *Environ. Toxicol. Pharmacol.* 78:103383. doi: 10.1016/j.etap.2020.103383
- Solomando, A., Capó, X., Alomar, C., Compa, M., Valencia, J. M., Sureda, A., et al. (2021). Assessment of the effect of long-term exposure to microplastics and depuration period in *Sparus aurata* Linnaeus, 1758: liver and blood biomarkers. *Sci. Total Environ.* 786:147479. doi: 10.1016/j.scitotenv.2021.147479

- Stankevičiute, M., Butrimavičiene, L., Valskiene, R., Greiciunaite, J., Baršienė, J., Vosyliene, M. Z., et al. (2016). Analysis of nuclear abnormalities in erythrocytes of rainbow trout (*Oncorhynchus mykiss*) treated with Cu and Zn and after 4-, 8-, and 12-day depuration (post-treatment recovery). *Mut. Res. Gen. Toxicol. Environ. Mutagen.* 797, 26–35. doi: 10.1016/j.mrgentox.2016.01.003
- Strunjak-Perovic, I., Topic Popovic, N., Coz-Rakovac, R., and Jadan, M. (2009). Nuclear abnormalities of marine fish erythrocytes. *J. Fish Biol.* 74, 2239–2249. doi: 10.1111/j.1095-8649.2009.02232.x
- Sula, E., Aliko, V., Pagano, M., and Faggio, C. (2020). Digital light microscopy as a tool in toxicological evaluation of fish erythrocyte morphological abnormalities. *Microsc. Res. Techn.* 83, 362–369. doi: 10.1002/jemt.23422
- Tang, H. L., Tang, H. M., Mak, K. H., Hu, S., Wang, S. S., Wong, K. M., et al. (2012). Cell survival, DNA damage, and oncogenic transformation after a transient and reversible apoptotic response. *Mol. Biol. Cell* 23, 2240–2252. doi: 10.1091/mbc.e11-11-0926
- Walpitagama, M., Carve, M., Douek, A. M., Trestrail, C., Bai, Y., Kaslin, J., et al. (2019). Additives migrating from 3D-printed plastic induce developmental toxicity and neuro-behavioural alterations in early life zebrafish (*Danio rerio*). *Aquat. Toxicol.* 213:105227. doi: 10.1016/j.aquatox.2019.105227
- Wu, P., Huang, J., Zheng, Y., Yang, Y., Zhang, Y., He, F., et al. (2019). Environmental occurrences, fate, and impacts of microplastics. *Ecotoxicol. Environ. Saf.* 184:109612. doi: 10.1016/j.ecoenv.2019.109612
- Xu, M., Halimu, G., Zhang, Q., Song, Y., Fu, X., Li, Y., et al. (2019). Internalization and toxicity: a preliminary study of effects of nanoplastic particles on human lung epithelial cell. *Sci. Total Environ.* 694:133794. doi: 10.1016/j.scitotenv.2019.133794
- Yirong, C., Shengchen, W., Jiaxin, S., Shuting, W., and Ziwei, Z. (2020). DEHP induces neutrophil extracellular traps formation and apoptosis in carp isolated from carp blood via promotion of ROS burst and autophagy. *Environ. Pollut.* 262:114295. doi: 10.1016/j.envpol.2020.114295
- Yonkos, L. T., Friedel, E. A., Perez-Reyes, A. C., Ghosal, S., and Arthur, C. D. (2014). Microplastics in four estuarine rivers in the Chesapeake Bay, USA. *Environ. Sci. Technol.* 48, 14195–14202. doi: 10.1021/es5036317
- Yu, S. P., Canzoniero, L. M., and Choi, D. W. (2001). Ion homeostasis and apoptosis. *Curr. Opin. Cell Biol.* 13, 405–411. doi: 10.1016/S0955-0674(00)00228-3
- Zhao, S., Zhu, L., Wang, T., and Li, D. (2014). Suspended microplastics in the surface water of the yangtze estuary system, China: first observations on occurrence, distribution. *Mar. Pollut. Bull.* 86, 562–568. doi: 10.1016/j.marpolbul.2014.06.032

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