# Bacterial Community Affects Toxin Production by *Gymnodinium catenatum*



## Maria E. Albinsson<sup>1,2</sup>\*, Andrew P. Negri<sup>3</sup>, Susan I. Blackburn<sup>2</sup>, Christopher J. S. Bolch<sup>1</sup>

1 National Centre for Marine Conservation and Resource Sustainability, Australian Maritime College, University of Tasmania, Launceston, Tasmania, Australia, 2 Commonwealth Scientific and Industrial Research Organisation, Marine and Atmospheric Research, Hobart, Tasmania, Australia, 3 Australian Institute of Marine Science, Townsville, Queensland, Australia

## Abstract

The paralytic shellfish toxin (PST)-producing dinoflagellate Gymnodinium catenatum grows in association with a complex marine bacterial community that is both essential for growth and can alter culture growth dynamics. Using a bacterial community replacement approach, we examined the intracellular PST content, production rate, and profile of G. catenatum cultures grown with bacterial communities of differing complexity and composition. Clonal offspring were established from surface-sterilized resting cysts (produced by sexual crosses of strain GCDE06 and strain GCLV01) and grown with: 1) complex bacterial communities derived from each of the two parent cultures; 2) simplified bacterial communities composed of the G. catenatum-associated bacteria Marinobacter sp. strain DG879 or Alcanivorax sp. strain DG881; 3) a complex bacterial community associated with an untreated, unsterilized sexual cross of the parents. Toxin content (STX-equivalent per cell) of clonal offspring (134–197 fmol STX cell<sup>-1</sup>) was similar to the parent cultures (169–206 fmol STX cell<sup>-1</sup>), however cultures grown with single bacterial types contained less toxin (134–146 fmol STX cell<sup>-1</sup>) than offspring or parent cultures grown with more complex mixed bacterial communities (152–176 fmol STX cell<sup>-1</sup>). Specific toxin production rate (fmol STX day<sup>-1</sup>) was strongly correlated with culture growth rate. Net toxin production rate (fmol STX cell<sup>-1</sup> day<sup>-1</sup>) did not differ among treatments, however, mean net toxin production rate of offspring was 8-fold lower than the parent cultures, suggesting that completion of the sexual lifecycle in laboratory cultures leads to reduced toxin production. The PST profiles of offspring cultures were most similar to parent GCDE06 with the exception of cultures grown with Marinobacter sp. DG879 which produced higher proportions of dcGTX2+3 and GC1+2, and lower proportions of C1+2 and C3+4. Our data demonstrate that the bacterial community can alter intracellular STX production of dinoflagellates. In G. catenatum the mechanism appears likely to be due to bacterial effects on dinoflagellate physiology rather than bacterial biotransformation of PST toxins.

Citation: Albinsson ME, Negri AP, Blackburn SI, Bolch CJS (2014) Bacterial Community Affects Toxin Production by Gymnodinium catenatum. PLoS ONE 9(8): e104623. doi:10.1371/journal.pone.0104623

Editor: Brett Neilan, University of New South Wales, Australia

Received May 31, 2013; Accepted July 15, 2014; Published August 12, 2014

**Copyright:** © 2014 Albinsson et al. This is an open-access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

**Funding:** This research was sponsored by the Aquafin CRC Project 4.2(2). A whole-of-ecosystem assessment of environmental issues for salmonid aquaculture. M. E. Albinsson was supported by an Australian Postgraduate Scholarship – The Thomas Crawford Memorial Scholarship while doing this research. The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing Interests: The authors have declared that no competing interests exist.

\* Email: Maria.Albinsson@csiro.au

## Introduction

Paralytic shellfish toxins (PST) are neurotoxic alkaloids produced by several dinoflagellates, including a number of species of the genus *Alexandrium*, *Pyrodinium bahamense*, and the unarmoured species *Gymnodinium catenatum* [1], [2], [3], [4], [5]. The toxins consist of the parent compound saxitoxin (STX) and at least 22 derivatives of varying oral toxicity [6].

Putative biosynthetic genes and pathways for PST production are now known from several cyanobacteria [7], [8], [9], and homologous genes have been detected from PST-producing dinoflagellates including *G. catenatum* [10]. There is currently no convincing structural or molecular evidence that heterotrophic bacteria produce PST autonomously yet the bacterial community is known to indirectly influence dinoflagellate PST toxicity [11], [12], [13] either by biotransformation from one PST derivative to another [14], [15], [16] or potentially via their effect on dinoflagellate growth and physiology [17], [18]. For example, axenic *Alexandrium* and *Protogonyaulax* cultures have been shown to contain a higher PST content than their non-axenic counterparts, suggesting that the associated bacterial community may reduce dinoflagellate toxin production under some circumstances [18], [19].

Gymnodinium catenatum is capable of producing at least 20 PSTs: the N-sulfocarbamoyl gonyautoxins (GTXs); the N-sulfocarbamoyl-11-hydroxysulfate C-toxins; the hydroxyl-benzoate toxins (GC-toxins); and a number of non-sulfated saxitoxin analogues [5], [20], [21], [22]. The range of PSTs produced varies considerably within and between populations, and culture-induced variation is also evident [13], [21]. The reasons for this high level of variation are poorly understood, but may be due to gene variation in the saxitoxin biosynthesis pathway [23], other genetic factors controlling toxin biosynthesis, or the direct or indirect effects of the microbial community [11], [21], [24].

Laboratory-grown *G. catenatum* cultures are associated with a suite of associated bacteria [11], [24], in which the two - proteobacteria, *Marinobacter* sp. and *Alcanivorax* sp., are constant components [11], [24]. Using strains originally isolated from the Tasmanian *G. catenatum* strain GCDE08 [11], Bolch et al., [25] found that *G. catenatum* has an obligate requirement for bacterial

associates and that both *Marinobacter* sp. DG879 and *Alcanivorax* sp. DG881 are capable of supporting survival and growth of the vegetative *G. catenatum* cells. Given this high reliance on bacterial associates, we anticipated that the bacterial community could alter toxin content and/or production of the dinoflagellate cell.

Our earlier studies have found no consistent link between community membership and cellular toxicity [11]; however, these studies did not examine PST production rates and were potentially confounded by isolation, geographic and culture-related effects. Here we use controlled uni-bacterial community replacement [25] to show that changes to the associated bacterial community modify both PST content and production by the dinoflagellate *G. catenatum*.

## **Materials and Methods**

## Production of resting cysts

Two sexually compatible, PST-producing, Tasmanian strains, GCDE06 (CS-301/06) and GCLV01 (CS-800), were provided by the Australian National Algae Culture Collection (ANACC; http://www.csiro.au/ANACC). Sexual crosses were carried out in sterile 55 mm diameter plastic Petri dishes containing 10 mL of nitrate and phosphate deficient GSe medium [26] and incubated at  $21^{\circ}C \pm 2^{\circ}C$ , at 80 µmol PAR m<sup>-2</sup> sec<sup>-1</sup> with a 18:6 h light:dark cycle for 3 weeks to promote cyst formation. Resting cysts were harvested by manual micropipetting, washed several times in fresh GSe medium, surface-sterilized in 0.5% H<sub>2</sub>O<sub>2</sub>, and washed 3 times to remove residual H<sub>2</sub>O<sub>2</sub> [25]. Successful surface-sterilization of cysts was checked by spread-plating 10 µL of the sterilized sample onto ZM1 agar incubated at 24°C for 4 days. If

bacterial growth was observed after incubation, the cysts were discarded.

## Unibacterial and community replacement cultures

A community replacement approach [25] was used to establish a series of clonal offspring cultures with the following modified bacterial communities: 1) complex bacterial communities derived from each of the two parent cultures GCDE06 and GCLV01; 2) simplified uni-bacterial communities of the *G. catenatum*-associated bacteria *Marinobacter* sp. DG879 or *Alcanivorax* sp. DG881; or 3) a complex bacterial community derived from unsterilized cysts from a sexual cross of both parent cultures.

Table 1 summarises all cultures established in this study. Bacterial strains Marinobacter sp. DG879 and Alcanivorax sp. DG881 were originally isolated from the G. catenatum strain GCDE08 [24] and maintained on modified ZoBell's Marine agar (ZM1) [25], [27] at 20°C in total darkness. Groups of ten to fifteen sterilized G. catenatum cysts were each placed into 55 mm diameter sterile polystyrene Petri dishes containing 10 mL GSe medium using a micropipette. A specific bacterial community was then added to each of the Petri dishes, with triplicate Petri dishes established for each of the four different bacterial communities as follows: Community 1) 1 mL of 10<sup>5</sup> cells mL<sup>-1</sup> of Marinobacter sp. DG879; Community 2) 1 mL of 10<sup>5</sup> cells mL<sup>-1</sup> of Alcanivorax sp. DG881; Community 3) 1 mL of  $10^5$  cells mL<sup>-1</sup> of a 5  $\mu$ m filtrate from the non-axenic parent strain GCDE06; and Community 4) 1 mL of  $10^5$  cells mL<sup>-1</sup> of a 5  $\mu$ m filtrate from the non-axenic parent strain GCLV01. Two sets of experimental control cultures were also established as follows:

Table 1. Summary of established G. catenatum cultures used in this study.

			Bacterial
Culture Treatment name	Original <i>G. catenatum</i> cultures	Bacterial community in culture	concentration at time of culture establishment
	Parent cultures		
DE06	GCDE06	Mixed culture-associated community	$\sim 10^5 \text{ cells mL}^{-1}$
LV01	GCLV01	Mixed culture-associated community	$\sim 10^5 \text{ cells mL}^{-1}$
	Cyst-derived cultures		
Gc/Mar*	GCDE06×GCLV01	Marinobacter sp. DG879	$10^5$ cells mL <sup>-1</sup>
Gc/Alc*	GCDE06×GCLV01	Alcanivorax sp. DG881	$10^5$ cells mL <sup>-1</sup>
DE06 filtrate	GCDE06×GCLV01	Filtrate (5 μm) from the non-axenic parent strain GCDE06	$10^5$ cells mL <sup>-1</sup>
LV01 filtrate	GCDE06×GCLV01	Filtrate (5 μm) from the non-axenic parent strain GCLV01	$10^5$ cells mL <sup>-1</sup>
Positive control	GCDE06×GCLV01	Mixed culture-associated community from both parent cultures	$\sim 10^5 \text{ cells mL}^{-1}$
Sterility control**	GCDE06×GCLV01	None- sterilized cysts placed in sterile seawater	N/A

The seven final cultures included two parent cultures and five cyst-derived cultures with altered bacterial communities.

\*Gc in the culture/treatment name is short for *Gymnodinium catenatum*.

\*\*The sterility control is non-viable as G.catenatum cannot grow axenically. Used as a control to assess contamination by bacteria from intracellular or other sources. doi:10.1371/journal.pone.0104623.t001 Table 2. Specific growth rate (µ) of Gymnodinium catenatum (Day 0 to Day 25) when grown with different microbial communities.

Culture	Specific growth rate ( $\mu$ ; $\pm$ SE)	
DE06	0.114 (±0.005)	
LV01	0.073 (±0.006)	
Gc/Mar	0.097 (±0.004)	
Gc/Alc	0.127 (±0.008)	
Gc/DE06 filtrate	0.084 (±0.004)	
Gc/LV01 filtrate	0.104 (±0.005)	
Positive control	0.077 (±0.006)	

doi:10.1371/journal.pone.0104623.t002

- Positive controls: Offspring cultures established from unsterilized cysts allowed to excyst in sterile GSe medium. These controls are equivalent of offspring from a typical crossing experiment. Survival and growth indicates the resting cyst viability.
- 2) Sterility controls: Surface-sterilized cysts germinated in sterile (0.2 μm) filtered seawater. Survival after germination is not expected as *G. catenatum* cannot grow in the absence of bacteria [25]. Survival and growth indicates bacterial contamination either from failed surface-sterilization, intracellular bacteria released at germination, or from other sources such as contaminated growth medium.

Petri dishes were incubated for 3 weeks at  $21^{\circ}C \pm 2^{\circ}C$ , 80 µmol PAR m<sup>-2</sup> sec<sup>-1</sup> with a 18:6 h light:dark cycle to allow excystment of the resting cysts. Clonal isolates were established from the mixed offspring cultures by dilution in sterile GSe medium (to 1 cell mL<sup>-1</sup>) and aseptic distribution of 1 mL aliquots into wells of sterile 24-multiwell plates containing 2 mL sterile GSe medium. Wells were examined using a Leitz Labovert FS inverted microscope (200× magnification), those containing a single cell/chain were labelled for later isolation, and the plate incubated for a further week. Clonal cultures were established from labelled wells by aseptic transfer into 50 mL Erlenmeyer flasks containing 25 mL

GSe medium, and 2 weeks later to 50 mL flasks containing 40 mL GSe medium.

A total of 5 clonal isolates were established for each of the 5 bacterial treatments, together with 5 clonal cultures of each of the two parent strains (GCDE06 and GCLV01). The clonal cultures were maintained on three week aseptic transfer intervals in 50 mL Erlenmeyer flasks at  $18^{\circ}$ C, 80 µmol PAR m<sup>-2</sup> sec<sup>-1</sup> with an 18:6 h light:dark cycle (Phillips cool-white fluorescent). Transfers to fresh medium were made every 2 weeks.

## Toxin production of cultures

Cultures for toxin content and production rate estimates were transferred to 175 mL Erlenmeyer flasks and maintained for 50 days under the same light and temperature conditions described earlier to establish the period of exponential growth phase of cultures prior to the toxin production experiment. At Day 0, 10 mL of late exponential phase cultures was inoculated into 150 mL of sterile GSe medium and the remaining inoculum culture volumes retained for triplicate cell counts (Leitz Labovert FS microscope, 200 × magnifications) and toxin analysis. Subsequent cultures were grown at 18°C and 80  $\mu$ mol PAR m<sup>-2</sup> sec<sup>-1</sup> light with a 18:6 h light:dark cycle until Day 25 (late-exponential phase) and harvested for toxin analysis. Cell concentration was



Figure 1. Toxin content (fmol cell<sup>-1</sup>) between Day 0 and Day 25 of *G. catenatum* cultures grown with different bacterial communities. Day 0 and Day 25 toxin content data for the offspring cultures were pooled as there was no significant difference. Toxicity of parent cultures (Day 25) included for comparison only. Superscripts indicate significant differences (p<0.05), means labelled with the same letter are not significantly different. Error bars ± SE, n = 5. doi:10.1371/journal.pone.0104623.g001



Figure 2. Relationship between specific growth rate ( $\mu$ ) and specific toxin production rate ( $\mu_{toxr}$ , fmol day<sup>-1</sup>). The relationship was established using 35 cultures of *Gymnodinium catenatum* measured between Day 0 and Day 25. doi:10.1371/journal.pone.0104623.g002

estimated every 5 days from triplicate cell counts using a Sedgwick-Rafter counting chamber (Graticules Ltd, UK). Culture samples (100 mL) were filtered through precombusted (400°C, 4 h) 47 mm GF/C (Whatman) filters, placed in 15 mL screw-cap polycarbonate centrifuge tubes containing 5 mL of 0.05 M acetic acid. Samples were sonicated for 30 s on ice several times using an ultrasonic cell disruptor (Braunsonic, 150 W, 5 mm probe), centrifuged at 5000×g for 5 min, the supernatants filtered through 0.45 µm filters, and then frozen at  $-20^{\circ}$ C until analysis.

#### HPLC analysis of PSTs

PSTs including C-toxins, gonyautoxins (GTX) and the hydroxybenzoate (GC) toxins were analysed by HPLC using the methods of Negri and Jones [28] and Negri et al. [20]. Briefly, toxins were separated using a Waters 600 HPLC, with postcolumn reactor (Pickering PCX 5100) using a 5  $\mu$ m, 250 mm×4.6 mm Alltima ODS column (Alltech, IL, USA) with a flow rate of 0.8 mL min<sup>-1</sup>. Post-column oxidation was performed according to the method of Oshima et al. [4]. Derivative PST fluorescence was detected with a Linear LC305 spectrofluorometric detector (excitation at 330 nm and emission at 390 nm). The retention times and fluorescent intensity of the PSTs were compared with PST standards (NRC, Canada) and identity of PST compounds confirmed by sample spiking experiments and removing post-column oxidation and observing the disappearance of peaks. The quantification of toxins was achieved by comparing peak areas with those of authentic standards and combining this data with cell counts from the time of harvest then converting to total PST expressed as finol cell<sup>-1</sup>. As the proportion of  $\alpha$ -epimers is often very small and some interconversion is expected, concentrations of epimer pairs were combined comparing toxin profiles.

## Calculations and statistical analyses

Specific growth rates  $(\mu)$  of the cultures from Day 0 to Day 25 were calculated over the exponential growth phase (Day 0 to Day 25) using the equation:







**Figure 4. Toxin profiles (mol%) of each toxin group identified.** The profiles of the identified toxin groups (C1+2, C3+4, GC1+2, dcGTX2+3, GTX2+3, and GTX1+4) from HPLC analysis on cultures at Day 25 (harvest), Error bars  $\pm$  SE, n = 5. Superscripts indicate significant differences (p<0.05), means labelled with the same letter are not significantly different. doi:10.1371/journal.pone.0104623.g004

$$\mu = \ln (N2/N1)/(t2-t1)$$

where NI and N2 = cell concentration at time 1 (t1) and time 2 (t2) respectively [29]. Specific toxin production rates ( $\mu_{tox}$ ) and net toxin production rates between Day 0 and 25 were calculated using the equations and methods described by Anderson et al. [30]. Briefly, the STX content of cells (fmol cell<sup>-1</sup>) at Day 0 and at Day 25 was multiplied by  $N_t$  (cell concentration at time t) to yield  $T_t$ , the total toxin concentration (fmol STX mL<sup>-1</sup> culture) at time t. Toxin concentrations (fmol STX mL<sup>-1</sup>) at Day 0 were calculated from the STX content (fmol cell<sup>-1</sup>) of the inocula and included a correction for dilution at inoculation. Values of T were then used to calculate specific toxin rate  $\mu_{tox}$  over each time interval:

$$\mu_{tox} = ln (T_2/T_1)/(t_2-t_1)$$

The net toxin production rate  $R_{tox}$  (fmol toxin cell<sup>-1</sup> d<sup>-1</sup>) was determined using:

$$R_{tox} = (T_2 - T_1)/((N')(\Delta t))$$

where N' is the ln average of the cell concentration (below):

$$N' = (N_2 - N_1)/(\ln N_2) - (\ln N_1)$$

and  $\Delta t \ (=t_2-t_1)$  the interval between Day 0 and Day 25. The ln average concentrations are used to account for exponential growth over the time period  $t_1-t_2$  [30].

The correlation between the specific toxin production rate and algal growth rate was done using linear regression. Significant differences in toxin production among bacterial culture treatments were compared using one-way ANOVA using the statistical software package R (Version 2.9.0), with treatment and time (Day 0 and 25) as factors, followed by Tukey's post hoc tests of significance.



**Figure 5. HPLC chromatogram comparison.** Comparison of (a) *Gymnodinium catenatum* positive control (DE06 x LV01 cross with natural bacterial assemblage) and (b) with the addition of *Marinobacter* sp. DG879 (Gc/Mar). doi:10.1371/journal.pone.0104623.g005

Differences in culture toxin profiles were examined using PERMANOVA+ [31], a multivariate analysis of variance with significance testing by random permutation, as implemented in the PRIMER-6 software package (www.primer-e.com). The toxin profile dataset (mol% STX, Day 25) was standardised prior to analysis to account for unequal variances among treatments. Canonical analysis of principal coordinates (CAP) [32] was performed using principal coordinates (PCO) calculated from a Euclidean distance matrix, and the first two canonical axes plotted. Over-parameterisation of the CAP was limited by restricting the number of PCO axes to that which maximised the leave-one-out allocation to groups [33]. All PERMANOVA and CAP tests used 9999 unrestricted random permutations of the raw data.

## Results

## Culture growth dynamics

Both the positive and the sterility control cultures behaved as expected. The sterility control did not survive, indicating that growth media was not contaminated. The positive controls germinated (in the presence of G. catenatum representative bacterial community), survived and were used to establish five

replicate cultures. The experimental cultures of *G. catenatum* exhibited average exponential growth rates (K') between Day 0 and 25 of  $0.095\pm0.005$  to  $0.14\pm0.008$  Div. day<sup>-1</sup> with no significant variation among cultures grown with different bacterial communities (Table 2) (F = 1.96, df = 6, p = 0.092).

## Toxin content and production

There was no significant difference in the toxin content of the offspring cultures between Day 0 and 25, (F = 3.416, df = 4, p = 0.072) therefore the data from both time periods was pooled. Cultures established with mixed bacterial communities contained a significantly higher toxin content than cultures grown with single bacterial types (F = 5.89, df = 6, p = 0.007) (Fig. 1). The specific toxin production rate ( $\mu_{tox}$ ; finol day<sup>-1</sup>) varied from 0.07 to 0.13 finol day<sup>-1</sup> (F = 6.35, df = 34, p = 0.0003) (Fig. 1) and was strongly correlated with the exponential growth rate of the cultures (df = 34, p = <0.001, R<sup>2</sup> = 0.82) (Fig. 2). Net toxin production rate ( $R_{tox}$ ; amol cell<sup>-1</sup> day<sup>-1</sup>) varied considerably among replicates and consequently no significant difference was detected among the different treatments (F = 2.44, df = 6, p = 0.064). However, the net toxin production rate of offspring cultures was



Figure 6. Canonical discriminant analysis of principal coordinates of toxin profiles of *G. catenatum* parent and offspring cultures. doi:10.1371/journal.pone.0104623.g006

almost 8-fold lower than the parent cultures (t = 2.83, df = 33, p = 0.008) (Fig. 3).

#### Toxin profiles

Toxin profiles (Day 25) were generally very consistent among replicates (Fig. 4). Parent strain GCDE06 produced higher proportions of C1+2 (F = 63.21, df = 9, p = <0.001) and GC1+2 (F = 42.42, df = 9, p = <0.001) and lower proportions of C3+4 (F = 53.96, df = 9, p = <0.001), GTX2+3 (F = 6.59, df = 9, df = 9)p = 0.033), and GTX1+4 (F = 5.41, df = 9, p = 0.048) than strain GCLV01 (Fig. 4). The toxin profile of the majority of offspring cultures were most similar to that of the GCDE06 parent cultures, with the exception of those grown with Marinobacter sp. DG879 which exhibited lower proportions of C1+2 (F = 435.35, df = 9, p = <0.001) and C3+4 (F = 38.7, df = 9, p = <0.001) and greater proportions of GC1+2 (F = 46.19, df = 9, p = <0.001) and dcGTX2+3 (F = 297.40, df = 9, p = <0.001) (Fig. 4). None of the cultures contained detectable concentrations of dcSTX, STX or GC3. Example HPLC chromatograms from cultures grown with a mixed bacterial community and with Marinobacter sp. DG879 are shown in Figure 5. Offspring positive control cultures were dominated by C2 with dcGTX3 as a minor component, whereas dcGTX3 was a major component and C2 a minor component in cultures grown with Marinobacter sp. DG879.

The CAP analysis of toxin profile data for all 35 cultures resolved three distinct toxin profile groupings. Parent cultures GCDE06 and GCLV01 were clearly separated along axis 2 (y-axis), primarily due to the different ratios of C1+2 versus C3+4 (Fig. 6). Cultures grown with *Marinobacter* sp. DG879 clustered separately along axis 1 (x-axis), and all other offspring cultures

clustered with or near parent GCDE06, or between the GCDE06 and GCLV01 parent clusters (Fig. 6).

#### Discussion

## Bacterial effects on toxicity of G. catenatum

A number of previous studies have also shown that bacteria change or induce cellular toxicity of axenic and non-toxic algal strains [12], [18], [34], [35], [36] however our study is the first to show that altering the culture-associated bacterial community results in changes to both PST content and PST production rate. The toxin content of G. catenatum vegetative cells varies considerably [4], [21], [37] and is influenced by a range of culture-related and environmental factors such as isolation method, growth medium, nutrient type, concentration, salinity and temperature [21], [38], [39], [40]. All of these factors were controlled across our treatments, therefore our data shows that change in both cellular toxicity, net PST production and PST profile of the dinoflagellate is due to the modification of the bacterial community. Culture-based microbial diversity of the unibacterial cultures routinely recovered only the single bacterial type added in each case (Alcanivorax or Marinobacter) and we are confident that the cultivable communities were overwhelmingly dominated by the added bacterial strain.

Previous work shows that *G. catenatum* culture bacterial communities are typically composed of 15–24 distinct bacterial genotypes and a consistent community structure dominated by  $\alpha$ - and  $\gamma$ -Proteobacteria and Bacteroidetes [11], [24]. While we do not have detailed microbial community data for the mixed community cultures in the present experiment, it is reasonable to assume similarly complex communities associated with parent and

the mixed community offspring cultures. The observed changes in toxicity could then be related to changes in bacterial associate composition (i.e. the specific bacteria present), the reduction in community complexity, or both. Which factor is most important is difficult to determine from our experimental design and data, however it is interesting to note that the two uni-bacterial cultures showed the lowest cellular toxicity suggesting that reduced bacterial diversity may be partially responsible for reduced toxin content of *G. catenatum*. Axenic and antibiotic treated *A. catenella* and *A. tamarense* cultures with substantially reduced bacterial diversity also show reduced PST content [18], [41] indicating that this response is not confined to *G. catenatum*.

Earlier studies of G. catenatum have shown up to 40-fold reduction in the cellular toxicity of G. catenatum cultures established from wild resting cysts [21]. We did not observe the same scale of reduction in cyst germinated offspring, however, we did detect an 8-fold reduction in mean net toxin production rate  $(\text{amol cell}^{-1} \text{ day}^{-1})$  compared to parent cultures. Studies of PST profiles of resting cysts indicate that PST production is reduced or stops during gamete and resting cyst formation [42], [43]. Sonication and washing of cysts during isolation may remove specific microbial associates (and their metabolites) required to fully induce dinoflagellate PST synthesis pathways after germination [21], [24]. Alternatively, reduced PST production may be associated with down-regulation of many secondary metabolite synthesis pathways during cyst formation and dormancy, or a form of reversible gene-silencing as described in some mycotoxinproducing fungi [44], [45].

The strong linear relationship between specific toxin production rate  $(\mu_{tox})$  and specific growth rate  $(\mu)$ , and the poor correlation between net toxin production rate  $(R_{tox})$  and  $\mu$  found for G. catenatum in this study, is similar to that of Alexandrium fundyense [30]. Anderson et al. [30] suggested this pattern resulted from nutritional deficiencies in batch cultures, causing an uncoupling of toxin synthesis from cell division, leading to highly variable rates of toxin accumulation. We sought to minimise nutritional limitation by harvesting cultures while in logarithmic growth and using saturating light intensity for G. catenatum at 21°C (80-90 µmol PAR  $m^{-2} \sec^{-1}$  [46], however, absolute growth rates observed in our experimental cultures are relatively low compared to other culture studies at similar temperatures and some of our cultures may have been light-limited leading to nutrient-limited growth in some cases. Alternatively, PST synthesis has been shown to occur only during the G1 phase of cell division [47], therefore the poor correlation may be due to differences in division synchrony among cultured treatments. Secondary metabolites (such as toxins) are also often subject to induction effects (up regulation) in response to unfavourable conditions [48], [49], further reducing the likelihood of a simple correlation with growth rate. Current knowledge of saxitoxin gene expression in dinoflagellates is limited, but recent studies suggest transcriptional or pre-translational regulations are major components. Taroncher-Oldenburg and Anderson [49] identified more than 20 transcriptionally regulated genes that were differentially expressed throughout the cell cycle of Alexandrium fundyense which were either up- or down regulated during toxigenesis. Similarly, Zhuang et al. [50] found that genes related to toxin production were expressed at different levels at different time points of the diel cycle.

## Bacterial effects on toxin profile

The dominance in our cultures of sulfocarbamoyl toxins C1+2 and C3+4 and the more recently characterised hydroxy-benzoate toxins (GC1+2) is consistent with the range of toxins evident from earlier studies of Australian *G. catenatum* [4], [21]. The

substantially different profile produced by cultures grown with *Marinobacter* sp. DG879 (Gc/Mar; lower C1+2 and C3+4; greater dcGTX2+3 and GC1+2) in comparison to the other cultures, could be due to conversion of C1 and C2 to their decarbamoyl derivatives dcGTX2 and dcGTX3. The GC-toxins, GC1 and GC2 are p-hydroxybenzoate analogues of the carbamate toxins GTX2 and GTX3 (respectively) but the biosynthetic pathway of GC-toxins is not yet fully understood [9], [10], [23].

Whether the observed PST congener transformations are carried out by *G. catenatum* only in the presence of *Marinobacter* sp. DG879 or by the bacterium itself is not clear. Bacterial PST production has been speculated or suggested by many studies [51], [52], [53] but there is currently no convincing structural evidence to confirm PST production by heterotrophic bacteria. Our previous studies have also shown that *Marinobacter* sp. DG879 does not produce PST-like toxins or activity [24] therefore the PST profile changes observed in GC/Mar cultures is not explained by production of particular PST congeners by *Marinobacter*. The recent discovery of dinoflagellate homologues of saxitoxin synthesis genes (e.g. SxtA1, SxtA4, SxtG) in *Alexandrium, Pyrodinium bahamense* and *Gymnodinium catenatum* [9], [10], [54] also appears to confirm that saxitoxin synthesis in our cultures is primarily or exclusively of dinoflagellate origin.

The changes may be mediated by bacterial enzymatic transformation of PSTs [14], [15], [16]. For example an uncharacterised carbamoylase activity has been demonstrated in bivalve-associated gut enzymes transform GC1, GC2 and GC3 to the more toxic dcGCX2, dcGTX3 and dcSTX [16]. This pathway could explain transformation within G. catenatum of C1+2 to dcGTX2+3, but does not explain the high proportions of GC1+2. On the other hand, proportional increase in GC1 and GC2 is balanced by a similar reduction in C1 and C2, suggesting separate biosynthetic pathways or more complex enzymatic transformations. To date all bacterial PST transformation activities in earlier studies have, however, been observed in supernatants of bacterial cultures or shellfish extracts. This indicates that activity is mediated by extracellular enzymes, presumably only acting on PST in the surrounding medium or cell boundary layer. We examined only intracellular PST in this work therefore any bacterial mediated PST transformation would need to be mediated by enzymes produced by intra-cellular bacteria or by surface bacteria releasing enzymes that are actively transported into the dinoflagellate cell. Alternatively particular bacterial associates may produce metabolites that have an indirect influence of Marinobacter sp. DG879 via their action on toxin synthesis pathways or physiology of the dinoflagellate cell.

Multivariate comparison of parent PST profiles with the 25 offspring from our experiments contradict previous parent-offspring studies of *G. catenatum* [5] and *Alexandrium* species [55], [56] which all report segregation of parental toxin profiles in a 2:2 Mendelian pattern. If this was the case for our study, the CAP ordination should show offspring clustering with or near the parent profiles in equal approximately numbers. Excluding the GC/Mar offspring that cluster separately on the axis 1 of the CAP, the majority of other offspring profiles cluster almost exclusively with parent GCDE06 (see Fig. 6) and indicate a uni-parental inheritance pattern in our experiment.

Resting cysts were not produced in self-crosses of either parent culture so it is highly unlikely that a majority of the offspring were products of self-crosses of GCDE06. A 2:2 segregation pattern also assumes the major determinants of toxin profile are inherited as single locus. Given the complexity of the core saxitoxin biosynthetic pathway (>20 genes), the additional predicted tailoring enzymes, and also a probable presence of multiple

expressed copies [10], it would in retrospect be surprising to find that PST profiles were inherited as a single non-recombining unit in dinoflagellates.

Our controlled bacterial community manipulation and germination study demonstrates that interactions with bacteria result in distinct and reproducible changes in cellular toxicity and toxin production rate, and cause significant changes to the intracellular PST profile of *G. catenatum*. How the observed changes are mediated remains unclear, however the balance of evidence

#### References

- Oshima Y, Hasegawa M, Yasumoto T, Hallegraeff G, Blackburn S (1987) Dinoflagellate *Gymnodinium catenatum* as the source of paralytic shellfish toxins in Tasmanian shellfish. Toxicon 25: 1105–1111.
- Cembella AD (1988) Ecophysiology and metabolism of paralytic shellfish toxins in marine microalgae. In: Anderson DM, Cembella AD, Hallegraeff GM, editors. Physiological Ecology of Harmful Algal Blooms. NATO ASI Series Vol. G4, Springer, Berlin. 381–403.
- Maclean JL (1989) An overview of *Pyrodinium* red tides in the western Pacific. In: Hallegraeff GM and Maclean JL, editors. Biology, Epidemiology and Management of *Pyrodinium* Red Tides. ICLARM Conference Proceedings 21. Fisheries Department, Ministry of Development, Brunei Darussalam, and International Center for Living Aquatic Resources Management, Manila, Philippines, 1–8.
- Oshima Y, Blackburn SI, Hallegraeff GM (1993a) Comparative study on paralytic shellfish toxin profiles of the dinoflagellate *Gymnodinium catenatum* from three different countries. Mar. Biol. 116: 471–476.
- Oshima Y, Itakura H, Lee KC, Yasumoto T, Blackburn S, et al. (1993b) Toxin production by the dinoflagellate *Gymnodinium catenatum*. In: Smayda TJ, Shimizu Y, editors. Toxic phytoplankton blooms in the sea. Proceedings of the fifth international conference on toxic marine phytoplankton. Elsevier, Amsterdam, 907–912.
- 6. Shimizu Y (1993) Microbial metabolites. Chem. Rev. 93: 1685-1689.
- Kellmann R, Mihali TK, Jeon YJ, Pickford R, Pomati F, et al. (2008) Biosynthetic intermediate analysis and functional homology reveal a saxitoxin gene cluster in cyanobacteria. Appl. Environ. Microbiol. 74: 4044–4053.
- Mihali TK, Kellmann R, Neilan BA (2009) Characterisation of the paralytic shellfish toxin biosynthesis gene clusters in *Anabaena circinalis* AWQC131C and *Aphanizomenon* sp. NH-5. BMC Biochem. 10: 8.
- Hackett JD, Wisccaver JH, Brosnahan ML, Kulis DM, Anderson DM, et al. (2012) Evolution of saxitoxin synthesis in cyanobacteria and dinoflagellates. Mol. Biol. Evol. 30(1): 70–78.
- Stüken A, Orr RJS, Kellmann R, Murray SA, Neilan BA, et al. (2011) Discovery of nuclear-encoded genes for the neurotoxin saxitoxin in dinoflagellates. PLoS ONE 6(5): e20096.
- Green DH, Hart MC, Blackburn SI, Bolch CJS (2010) Bacterial diversity of *Gymnodinium catenatum* and its relationship to dinoflagellate toxicity. Aquat. Microb. Ecol. 61: 73–87.
- Kodama M (2010) Paralytic shellfish poisoning toxins: Biochemistry and origin. Aqua-BioSci. Monogr. 3(1): 1–38.
- Hallegraeff GM, Blackburn SI, Doblin M, Bolch CJS (2012) Global toxicology, ecophysiology and population relationships of the chainforming PST dinoflagellate *Gymnodinium catenatum*. Harmful Algae 14: 130–143.
- Kotaki Y, Oshima Y, Yasumoto T (1985a) Bacterial transformation of paralytic shellfish toxins in coral reef crabs and a marine snail. B. Jpn. Soc. Sci. Fish. 51: 1009–1013.
- Kotaki Y, Oshima Y, Yasumoto T (1985b) Bacterial transformation of paralytic shellfish toxins. In: Anderson DM, White AW, Baden DG, editors. Toxic Dinoflagellates, Elsevier Science Publishing Co. Amsterdam. 287–292.
- Vale P (2008) Fate of benzoate paralytic shellfish poisoning toxins from *Gymnodinium catenatum* in shellfish and fish detected by pre-column oxidation and liquid chromatography with fluorescence detection. J. of Chromatogr A 1190: 191–197.
- Doucette GJ, Kodama M, Franca S, Gallacher S (1998) Bacterial interactions with harmful algal bloom species: bloom ecology, toxigenesis, and cytology. In: Anderson DM, Cembella AD, Hallegraeff GM, editors. Physiological Ecology of Harmful Algal Blooms. Springer, Berlin NATO ASI Series. 619–647.
- Uribe P, Espejo RT (2003) Effects of associated bacteria on the growth and toxicity of Alexandrium catenella. Appl. Environ. Microbiol. 69(1): 659–662.
- Singh HT, Oshima Y, Yasumoto T (1982) Growth and toxicity of *Protogonyaulax tamarensis* in axenic cultures. B. Jpn. Soc. Sci. Fish. 48: 1341–134.
- Negri A, Stirling D, Quilliam M, Blackburn S, Bolch C, et al. (2003) Three novel hydroxybenzoate saxitoxin analogues isolated from the dinoflagellate *Gymnodinium catenatum*. Chem. Res. Toxicol. 16(8): 1029–1033.
- Negri AP, Bolch CJS, Geier S, Green DH, Park T-G, et al. (2007) Widespread presence of hydrophobic paralytic shellfish toxins in *Gymnodinium catenatum*. Harmful Algae 6: 774–780.
- 22. Vale P (2008) Complex profiles of hydrophobic paralytic shellfish poisoning compounds in *Gymnodinium catenatum* identified by liquid chromatography

suggests bacterial influences on dinoflagellate cell physiology and growth are the most likely mechanism.

## **Author Contributions**

Conceived and designed the experiments: MEA APN SIB CJSB. Performed the experiments: MEA APN. Analyzed the data: MEA APN SIB CJSB. Contributed reagents/materials/analysis tools: MEA APN SIB CJSB. Wrote the paper: MEA CJSB APN SIB.

with fluorescence detection and mass spectrometry. J. Chromatogr. A 1195(1-2): 85–93.

- Orr RJS, Stüken A, Murray SA, Jakobsen KS (2013) Evolution and distribution of saxitoxin biosynthesis in dinoflagellates. Mar. Drugs 11: 2814–2828.
- Green DH, Llewellyn LE, Negri AP, Blackburn SI, Bolch CJS (2004) Phylogenetic and functional diversity of the cultivable bacterial community associated with the paralytic shellfish poisoning dinoflagellate *Gymnodinium catenatum*. FEMS Microbiol. Ecol. 47: 345–357.
- Bolch CJS, Subramanian TA (2011) The toxic dinoflagellate Gymnodinium catenatum (Dinophyceae) requires marine bacteria for growth. J. Phycol. 47: 1009–1022.
- Blackburn SI, Bolch CJS, Haskard KA, Hallegraeff GM (2001) Reproductive compatibility among four global populations of the toxic dinoflagellate *Gymnodinium catenatum* (Dinophyceae). Phycologia 40(1): 78–87.
- Zobell DE (1941) Studies on marine bacteria I. The cultural requirements of heterotrophic aerobes. J. Mar. Res. 5: 42–75.
- Negri AP, Jones GJ (1995) Bioaccumulation of paralytic shellfish poisoning (PSP) toxins from the cyanobacterium *Anabaena circinalis* by the freshwater mussel *Alathyria condola*. Toxicon 33: 667–678.
- Levasseur M, Thompson PA, Harrison PJ (1993) Physiological acclimation of marine phytoplankton to different nitrogen sources. J. Phycol. 29: 587–595.
- Anderson DM, Kulis DM, Sullivan JJ, Hall S, Lee C (1990) Dynamics and physiology of saxitoxin production by the dinoflagellate *Alexandrium* spp. Mar. Biol. 104: 511–524.
- Anderson MJ, Gorley RN, Clarke KR (2005) PERMANOVA+ for PRIMER: Guide to Software and Statistical Methods. PRIMER-E, Plymouth, UK, 214.
- Anderson MJ, Willis TJ (2003) Canonical analysis of principal coordinates: a useful method of constrained ordination for ecology. Ecology 84: 511–525.
- Anderson MJ, Robinson J (2003) Generalized discriminant analysis based on distances. Aust. NZ. J. Stat. 45: 301–318.
- Carrasquero-Verde JR (1999) Role of associated bacteria in *Heterosigma* carterae toxicity to salmonids. Aquat. Toxicol. 45: 19–34.
- Pérez-Guzmán L, Pérez-Matos AE, Rosado W, Tosteson TR, Govind NS (2008) Bacteria associated with toxic clonal cultures of the dinoflagellate Ostreopsis lenticularis. Mar. Biotechnol. 10: 492–496.
- Jasti S, Sieracki E, Poulton NJ, Giewat MW, Rooney-Varga JN (2005) Phylogenetic diversity and specificity of bacteria closely associated with *Alexandrium* spp. and other phytoplankton. Appl. Environ. Microbiol. 71(7): 3483–3494.
- 37. Negri AP, Bolch CJS, Blackburn SI, Dickman M, Llewellyn LE, et al. (2001) Paralytic shellfish toxins in *Gymnodinium catenatum* strains from six countries. In: Hallegraeff GM, Blackburn SI, Bolch CJ, Lewis RJ, editors. Harmful Algal Blooms 2000. Intergovernmental Oceanographic Commission of UNESCO 2001. United Nations Educational, Scientific, and Cultural Organization, 210– 213.
- Band-Schmidt CJ, Bustillos-Guzmán J, Morquecho L, Gárate-Lizárraga I, Alonso-Rodriguez R, et al. (2006) Variations of PSP toxin profiles during different growth phases in *Gymnodinium catenatum* (Dinophyceae) strains isolated from three locations in the Gulf of California, Mexico. J. Phycol. 42: 757–768.
- Ellegaard M, Christensen NF, Moestrup Ø (1993) Temperature and salinity effects on growth of a non-chain-forming strain of *Gymnodinium catenatum* (Dinophyceae) established from recent sediments in the sound (Øresund), Denmark. J. Phycol. 29: 418–426.
- Flynn KJ, Flynn K, John EH, Reguera B, Reyero I, et al. (1996) Changes in toxins, intracellular and dissolved free amino acids of the toxic dinoflagellate *Gymnodinium catenatum* in response to changes in inorganic nutrients and salinity. J. Plankton Res. 18(11): 2093–2111.
- Hold GL, Smith EA, Birkbeck TH, Gallacher S (2001) Comparison of paralytic shellfish toxin (PST) production by the dinoflagellate *Alexandrium lusitanicum* NEPCC 253 and *Alexandrium tamarense* NEPCC 407 in the presence and absence of bacteria. FEMS Microbiol. Ecol. 36: 223–234.
- Oshima Y, Bolch CJ, Hallegraeff GM (1992) Toxin composition of resting cysts of *Alexandrium tamarense* (Dinophyceae). Toxicon 30(12): 1539–1544.
  Bravo I, Franco JM, Reyero MI (1998) PSP toxin composition of three life cycle
- 43. Bravo I, Franco JM, Reyero MI (1998) PSP toxin composition of three life cycle stages of *Gymnodinium catenatum*. In: Reguera B, Blanco J, Fernández ML, Wyatt T, editors. Harmful Alga. Xunta de Galicia and IOC of UNESCO, Grafisant, Santiago de Compostela, 356–358.

- Bok JW, Noordermeer D, Kale SP, Keller NP (2006) Secondary metabolic gene cluster silencing in Aspergillus nidulans. Mol. Microbiol. 6: 1636–1645.
- 45. Smith CA, Woloshuk CP, Robertson D, Payne GA (2007) Silencing of the aflatoxin gene cluster in a diploid strain of *Aspergillus flavus* is suppressed by ectopic aflR expression. Genetics 176: 2077–2086.
- Yamamoto T, Oh SJ, Kataoka Y (2002) Effects of temperature, salinity and irradiance on the growth of the toxic dinoflagellate *Gymnodinium catenatum* (Dinophyceae) isolated from Hiroshima Bay, Japan. Fisheries Sci. 68: 356–363.
- Taroncher-Oldenburg G, Kulis DM, Anderson DM (1997) Toxin variability during the cell cycle of the dinoflagellate *Alexandrium fundyense*. Limnol. Oceanogr. 42(5, part 2): 1178–1188.
- Yang I, John U, Beszteri S, Gloeckner G, Krock B (2010) Comparative gene expression in toxic versus non-toxic strains of the marine dinoflagellate *Alexandrium minutum*. BMC Genomics 11: 248.
- Taroncher-Oldenburg G, Anderson DM (2000) Identification and characterization of three differentially expressed genes, encoding S-Adenosylhomocysteine Hydrolase, Methionine Aminopeptidase, and a histone-like protein, in the toxic dinoflagellate Alexandrium fundyense. Appl. Envir. Microbiol. 66(5): 2105– 2112.
- Zhuang Y, Zhang H, Hannick L, Lin S (2012) Cell cycle, saxitoxin, and proton pump rhodopsin: insight from gene expression profiling for an *Alexandrium fundyense* culture and a natural bloom. J. of. Phyc. 48(1): S27.

- Gallacher S, Flynn KJ, Leftley J, Lewis J, Munro PD, et al. (1996) Bacterial production of sodium channel blocking toxins. In: Yasumoto T, Oshima Y, Fukuyo Y, editors. Harmful and Toxic Algal Blooms. Intergovernmental Oceanographic Commission of UNESCO. 355–358.
- Gallacher S, Flynn KJ, Franco JM, Brueggemann EE, Hines HB (1997) Evidence for production of paralytic shellfish toxins by bacteria associated with *Alexandrium* spp. (Dinophyta) in culture. Appl. Environ. Microb. 63(3): 239– 245.
- Töbe K, Ferguson C, Kelly M, Gallacher S, Medlin L (2001) Seasonal occurrence at a Scottish PSP monitoring site of purportedly toxic bacteria originally isolated from the toxic dinoflagellate genus *Alexandrium*. Eur. J. Phycol. 36: 243–256.
- Orr RJS, Stüken A, Murray S, Jakobsen KS (2013) Evolutionary acquisition and loss of saxitoxin biosynthesis in Dinoflagellates: lessons from the second "core" gene - sxtG. Appl. Environ. Microbiol. 79: 2128–2136.
- Sako Y, Kim C-H, Ishida Y (1992) Mendelian inheritance of paralytic shellfish poisoning toxin in the marine dinoflagellate *Alexandrium catenella*. Biosci. Biotech. Bioch. 56: 692–694.
- Ishida Y, Kim CH, Sako Y, Hirooka N, Uchida A (1993) PSP toxin production is chromosome dependent in Alexandrium spp, In: Smayda TJ and Shimizu Y, editors. Toxic phytoplankton blooms in the sea. Proceedings of the 5th International Conference on Toxic Marine Phytoplankton, Elsevier, Amsterdam. 881–87.